

Coordinated Aquatic Monitoring Program

CAMP Twelve Year Data Report (2008-2019) Technical Document 6: Churchill River Diversion Region

Prepared by

Manitoba Hydro

And

North/South Consultants Inc.

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CAMP TWELVE YEAR DATA REPORT (2008-2019)

TECHNICAL DOCUMENT 6: CHURCHILL RIVER DIVERSION REGION

Prepared by

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EXECUTIVE SUMMARY

This report presents the results of monitoring conducted under the Coordinated Aquatic Monitoring Program (CAMP) for years 1 through 12 (i.e., 2008/2009 through 2019/2020) in the Churchill River Diversion Region is composed of the Rat/Burntwood River system from downstream of Notigi Lake and the Notigi Lake Control Structure (CS) to First Rapids on the Burntwood River, approximately 20 km upstream of Split Lake. Waterbodies and sites monitored in this region over this period included three on-system and one off-system waterbodies as follows:

- Threepoint Lake;
- Apussigamasi Lake;
- Footprint lake; and
- Leftrook Lake (off-system).

Monitoring on-system waterbodies addresses the primary objective of CAMP – to monitor aquatic ecosystem health along Manitoba Hydro's hydraulic operating system. The off-system waterbodies were included in CAMP to provide regional information collected in a manner consistent with monitoring of on-system waterbodies that will assist in interpreting any observed environmental changes over time. Such comparisons are intended to help distinguish between hydroelectric-related effects and other external factors (e.g., climate change) in each CAMP region.

Monitoring was conducted annually at some waterbodies and river reaches and on a three-year rotation at other sites. Components monitored under CAMP in the Churchill River Diversion Region presented in this report include water regime, water quality, benthic macroinvertebrates, fish community, and mercury in fish. Results of sedimentation monitoring conducted at the Wuskwatim Generating Station (GS) are also included. Climatological data for the region are also included to provide supporting information to assist with interpretation of CAMP monitoring results.



TABLE OF CONTENTS

| 1.0 | INTE | RODUCT | rion | 1-1 |
|-----|------|---------|-----------------------------|------|
| 2.0 | PHY | SICAL E | NVIRONMENT | 2-1 |
| | 2.1 | Introd | luction | 2-1 |
| | 2.2 | Climat | te | 2-2 |
| | | 2.2.1 | Temperature | 2-2 |
| | | 2.2.2 | Precipitation | 2-5 |
| | 2.3 | Water | Regime | 2-7 |
| | | 2.3.1 | Flow | 2-9 |
| | | 2.3.2 | Water Level and Variability | 2-11 |
| | | 2.3.3 | Water Temperature | 2-24 |
| | 2.4 | Sedim | nentation | 2-26 |
| | | 2.4.1 | Continuous Turbidity | 2-26 |
| | | 2.4.2 | Suspended Sediment Load | 2-29 |
| 3.0 | WAT | TER QUA | ALITY | 3-1 |
| | 3.1 | Introd | luction | 3-1 |
| | 3.2 | Dissol | ved Oxygen | 3-4 |
| | | 3.2.1 | Dissolved Oxygen | 3-4 |
| | 3.3 | Water | · Clarity | 3-27 |
| | | 3.3.1 | Secchi Disk Depth | 3-27 |
| | | 3.3.2 | Turbidity | 3-35 |
| | | 3.3.3 | Total Suspended Solids | 3-40 |
| | 3.4 | Nutrie | ents and Trophic Status | 3-45 |
| | | 3.4.1 | Total Phosphorus | 3-45 |
| | | 3.4.2 | Total Nitrogen | 3-56 |
| | | 3.4.3 | Chlorophyll a | 3-61 |



| 4.0 | BEIN | THIC INVERTEBRATES | 4- | | |
|-----|-----------------|------------------------------------|------|--|--|
| | 4.1 | Introduction | 4-1 | | |
| | 4.2 | Abundance | 4-2 | | |
| | | 4.2.1 Total Invertebrate Abundance | 4-2 | | |
| | 4.3 | Community Composition | 4-8 | | |
| | | 4.3.1 Relative Abundance | 4-8 | | |
| | | 4.3.2 EPT Index | 4-15 | | |
| | | 4.3.3 O+C Index | 4-18 | | |
| | 4.4 | Richness | 4-2 | | |
| | | 4.4.1 Total Taxa Richness | 4-2 | | |
| | | 4.4.2 EPT Taxa Richness | 4-24 | | |
| | 4.5 | Diversity | 4-27 | | |
| | | 4.5.1 Hill's Effective Richness | 4-27 | | |
| 5.0 | FISH COMMUNITY5 | | | | |
| | 5.1 | Introduction | 5-1 | | |
| | 5.2 | Abundance | 5-4 | | |
| | | 5.2.1 CPUE | 5-4 | | |
| | 5.3 | Condition | 5-20 | | |
| | | 5.3.1 Fulton's Condition Factor | 5-20 | | |
| | | 5.3.2 Relative Weight | 5-32 | | |
| | 5.4 | Growth | 5-43 | | |
| | | 5.4.1 Length-at-Age | 5-43 | | |
| | 5.5 | Recruitment | 5-52 | | |
| | | 5.5.1 Relative Year-Class Strength | 5-52 | | |
| | 5.6 | Diversity | 5-58 | | |
| | | 5.6.1 Relative Species Abundance | 5-58 | | |
| | | 5.6.2 Hill's Effective Richness | 5-70 | | |



| 6.0 | MER | CURY IN | N FISH | 6-1 |
|-----|------|---------|--|-------------|
| | 6.1 | Introd | uction | 6-1 |
| | 6.2 | Mercu | ry in Fish | 6-4 |
| | | 6.2.1 | Arithmetic Mean Mercury Concentration | 6-4 |
| | | 6.2.2 | Length-Standardized Mean Concentration | 6-13 |
| 7.0 | LITE | RATURE | E CITED | 7 -1 |



LIST OF TABLES

| Table 1-1. | Churchili River Diversion Region Calvip monitoring summary | 1 -2 |
|---------------|---|------|
| Table 2.1-1. | Physical Environment indicators and metrics | 2-1 |
| Table 2.2-1. | Thompson mean monthly and annual air temperature (in °C) compared to | |
| | 1981-2010 normal | 2-4 |
| Table 2.2-2. | Thompson total monthly and annual precipitation (in mm) compared to | |
| | 1981-2010 normal | 2-6 |
| Table 2.3-1. | Churchill River Diversion at the Notigi CS monthly average flow (cms) | 2-10 |
| Table 2.3-2. | Threepoint Lake monthly average water level (m) | 2-13 |
| Table 2.3-3. | Threepoint Lake monthly water level range (m) | 2-14 |
| Table 2.3-4. | Footprint Lake monthly average water level (m). | 2-15 |
| Table 2.3-5. | Footprint Lake monthly average water level (m). | 2-16 |
| Table 2.3-6. | Apussigamasi Lake monthly water average level (m) | 2-17 |
| Table 2.3-7. | Apussigamasi Lake monthly water level range (m) | 2-18 |
| Table 2.3-8. | Leftrook Lake monthly average water level (m) | 2-19 |
| Table 2.3-9. | Leftrook Lake monthly water level range (m) | 2-20 |
| Table 2.3-10. | 2017-19 Wuskwatim GS water temperature ranges | 2-25 |
| Table 2.4-1. | 2008-2019 sedimentation sampling inventory | 2-26 |
| Table 2.4-2. | Sedimentation indicators and metrics. | 2-26 |
| Table 2.4-3. | 2017-2019 Wuskwatim GS average monthly turbidity | 2-27 |
| Table 2.4-4. | 2017-19 Wuskwatim GS average monthly sediment load | 2-30 |
| Table 3.1-1. | 2008-2019 Water quality sampling inventory | 3-2 |
| Table 3.1-2. | Water quality indicators and metrics | 3-2 |
| Table 3.2-1. | 2008-2019 On-system sites summary of thermal stratification and DO | |
| | concentrations | 3-9 |
| Table 3.2-2. | 2008-2019 On-system sites DO, water depth, and ice thickness summary | |
| | statistics | 3-10 |
| Table 3.2-3. | 2008-2019 Off-system sites summary of thermal stratification and DO | |
| | concentrations | 3-12 |
| Table 3.2-4. | 2008-2019 Off-system sites DO, water depth, and ice thickness summary | |
| | statistics | 3-13 |



| Table 3.3-1. | 2008-2019 On-system sites water clarity summary statistics | 3-29 |
|--------------|--|------|
| Table 3.3-2. | 2008-2019 Off-system sites water clarity metric summary statistics | 3-30 |
| Table 3.4-1. | 2008-2019 On-system sites TP, TN, and chlorophyll $\it a$ summary statistics | 3-48 |
| Table 3.4-2. | 2008-2019 On-system trophic status based on TP, TN, and chlorophyll $\it a$ | |
| | open-water season mean concentrations. | 3-49 |
| Table 3.4-3. | 2008-2019 Off-system sites TP, TN and chlorophyll $\it a$ summary statistics | 3-50 |
| Table 3.4-4. | 2008-2019 Off-system trophic status based on TP, TN, and chlorophyll $\it a$ | |
| | open-water season mean concentrations. | 3-51 |
| Table 4.1-1. | 2010 to 2019 Benthic invertebrate sampling inventory | 4-2 |
| Table 4.1-2. | Benthic invertebrate indicators and metrics. | 4-2 |
| Table 4.3-1. | 2010 to 2019 Threepoint Lake nearshore benthic invertebrate relative | |
| | abundance | 4-11 |
| Table 4.3-2. | 2010 to 2019 Threepoint Lake offshore benthic invertebrate relative | |
| | abundance | 4-11 |
| Table 4.3-3. | 2010 to 2019 Footprint Lake nearshore benthic invertebrate relative | |
| | abundance | 4-12 |
| Table 4.3-4. | 2010 to 2019 Footprint Lake offshore benthic invertebrate relative | |
| | abundance | 4-12 |
| Table 4.3-5. | 2010 to 2019 Apussigamasi Lake nearshore benthic invertebrate relative | |
| | abundance | 4-13 |
| Table 4.3-6. | 2010 to 2019 Apussigamasi Lake offshore benthic invertebrate relative | |
| | abundance | 4-13 |
| Table 4.3-7. | 2010 to 2019 Leftrook Lake nearshore benthic invertebrate relative | |
| | abundance | 4-14 |
| Table 4.3-8. | 2010 to 2019 Leftrook Lake offshore benthic invertebrate relative | |
| | abundance | 4-14 |
| Table 5.1-1. | 2008-2019 Inventory of fish community sampling | 5-2 |
| Table 5.1-2. | Fish community indicators and metrics | 5-2 |
| Table 5.2-1. | 2009-2019 Catch-per-unit-effort | 5-12 |
| Table 5.3-1. | 2009-2019 Fulton's condition factor of target species | 5-26 |
| Table 5.3-2. | 2009-2019 relative weight of target species. | 5-37 |
| Table 5.4-1. | 2009-2019 Fork length-at-age of target species | 5-47 |
| Table 5.6-1. | Inventory of fish species | 5-61 |



| Table 5.6-2. | 2009-2019 Relative species abundance in standard gang index gill nets in | |
|---------------|---|------|
| | Threepoint Lake | 5-62 |
| Table 5.6-3. | 2009-2019 Relative species abundance in small mesh index gill nets in | |
| | Threepoint Lake | 5-63 |
| Table 5.6-4. | 2009-2019 Relative species abundance in standard gang index gill nets in | |
| | Footprint Lake. | 5-64 |
| Table 5.6-5. | 2009-2019 Relative species abundance in small mesh index gill nets in | |
| | Footprint Lake. | 5-65 |
| Table 5.6-6. | 2009-2019 Relative species abundance in standard gang index gill nets in | |
| | Apussigamasi Lake | 5-66 |
| Table 5.6-7. | 2009-2019 Relative species abundance in small mesh index gill nets in | |
| | Apussigamasi Lake | 5-67 |
| Table 5.6-8. | 2009-2019 Relative species abundance in standard gang index gill nets in | |
| | Leftrook Lake | 5-68 |
| Table 5.6-9. | 2009-2019 Relative species abundance in small mesh index gill nets in | |
| | Leftrook Lake | 5-69 |
| Table 5.6-10. | 2009-2019 Hill's effective species richness | 5-72 |
| Table 6.1-1. | 2008-2019 Inventory of fish mercury sampling. | 6-2 |
| Table 6.1-2. | Mercury in fish indicators and metrics | 6-2 |
| Table 6.2-1. | 2010-2019 Fork length, age, and mercury concentrations of Lake Whitefish, | |
| | Northern Pike, and Walleye | 6-6 |
| Table 6.2-2. | 2010-2019 Fork length and mercury concentrations of 1-year-old Yellow | |
| | Perch | 6-8 |



LIST OF FIGURES

| rigure 1-1. | On-system and on-system waterbodies and river reaches sampled under | 4.0 |
|---------------|---|------|
| | CAMP in the Churchill River Diversion Region: 2008-2019 | 1-3 |
| Figure 2.2-1. | Thompson mean monthly air temperature (in °C) compared to 1981-2010 | |
| | normal | 2-4 |
| Figure 2.2-2. | Thompson total monthly precipitation (in mm) compared to 1981-2010 | |
| | normal | 2-6 |
| Figure 2.3-1. | Hydrometric and continuous water quality monitoring stations in the | |
| | Churchill River Diversion Region | 2-8 |
| Figure 2.3-2. | 2008-2020 Churchill River Diversion at the Notigi CS daily mean flow | 2-11 |
| Figure 2.3-3. | 2008-2020 Upper Churchill at the Notigi CS daily mean flow and Threepoint | |
| | Lake daily mean water level. | 2-21 |
| Figure 2.3-4. | 2008-2020 Upper Churchill River at the Notigi CS daily mean flow and | |
| | Footprint Lake daily mean water level. | 2-22 |
| Figure 2.3-5. | 2008-2020 Upper Churchill River at the Notigi CS daily mean flow and | |
| | Apussigamasi Lake daily mean water level | 2-23 |
| Figure 2.3-6. | 2008-2020 Leftrook Lake daily mean water level | 2-24 |
| Figure 2.3-7. | 2017-2019 Wuskwatim GS continuous water temperature | 2-25 |
| Figure 2.4-1. | 2017-2019 Wuskwatim GS monthly turbidity | 2-28 |
| Figure 2.4-2. | 2017-2019 Wuskwatim GS continuous turbidity | 2-29 |
| Figure 2.4-3. | 2017-2019 Wuskwatim GS monthly sediment load | 2-30 |
| Figure 2.4-4. | 2017-2019 Wuskwatim GS daily sediment load | 2-31 |
| Figure 3.1-1. | 2008-2019 Churchill River Diversion Region water quality sites | 3-3 |
| Figure 3.2-1. | 2008-2019 On-system and off-system water temperature depth profiles | 3-14 |
| Figure 3.2-2. | 2008-2019 On-system and off-system dissolved oxygen depth profiles and | |
| | comparison to instantaneous minimum objectives for the protection of | |
| | aquatic life | 3-15 |
| Figure 3.2-3. | 2008-2019 Threepoint Lake surface and bottom dissolved oxygen | |
| - | concentrations with comparison to instantaneous minimum objectives for | |
| | the protection of aquatic life | 3-16 |



| Figure 3.2-4. | 2008-2019 On-system seasonal surface and bottom dissolved oxygen | |
|----------------|--|------|
| | concentrations with comparison to instantaneous minimum objectives for | |
| | the protection of aquatic life. | 3-17 |
| Figure 3.2-5. | 2008-2019 On-system seasonal surface and bottom dissolved oxygen | |
| | saturation. | 3-18 |
| Figure 3.2-6. | 2008-2019 On-system open-water season surface and bottom dissolved | |
| | oxygen saturation. | 3-19 |
| Figure 3.2-7. | 2008-2019 On-system ice-cover season surface and bottom dissolved | |
| | oxygen saturation. | 3-20 |
| Figure 3.2-8. | 2008-2019 Footprint Lake surface and bottom dissolved oxygen | |
| | concentrations with comparison to instantaneous minimum objectives for | |
| | the protection of aquatic life | 3-21 |
| Figure 3.2-9. | 2008-2019 Apussigamasi Lake surface and bottom dissolved oxygen | |
| | concentrations with comparison to instantaneous minimum objectives for | |
| | the protection of aquatic life | 3-22 |
| Figure 3.2-10. | 2008-2019 Leftrook Lake surface and bottom dissolved oxygen | |
| | concentrations with comparison to instantaneous minimum objectives for | |
| | the protection of aquatic life. | 3-23 |
| Figure 3.2-11. | 2008-2019 Off-system seasonal surface and bottom dissolved oxygen | |
| | concentrations with comparison to instantaneous minimum objectives for | |
| | the protection of aquatic life. | 3-24 |
| Figure 3.2-12. | 2008-2019 Off-system seasonal surface and bottom dissolved oxygen | |
| | saturation. | 3-25 |
| Figure 3.2-13. | 2008-2019 Off-system open-water and ice-cover season surface and | |
| | bottom dissolved oxygen saturation | 3-26 |
| Figure 3.3-1. | 2008-2019 On-system open-water season Secchi disk depths | 3-31 |
| Figure 3.3-2. | 2008-2019 On-system seasonal Secchi disk depth, turbidity, and TSS | |
| | concentrations | 3-32 |
| Figure 3.3-3. | 2008-2019 Off-system open-water season Secchi disk depths | 3-33 |
| Figure 3.3-4. | 2008-2019 Off-system seasonal Secchi disk depth, turbidity, and TSS | |
| | concentrations | 3-34 |
| Figure 3.3-5. | 2008-2019 On-system open-water and ice-cover season turbidity levels | 3-38 |
| Figure 3.3-6. | 2008-2019 Off-system open-water and ice-cover season turbidity levels | 3-39 |



| Figure 3.3-7. | 2008-2019 On-system open-water and ice-cover season TS: | |
|---------------|---|------|
| | concentrations | |
| Figure 3.3-8. | 2008-2019 Off-system open-water and ice-cover season TS | |
| | concentrations | 3-44 |
| Figure 3.4-1. | 2008-2019 On-system open-water and ice-cover season TP | |
| | concentrations | 3-52 |
| Figure 3.4-2. | 2008-2019 On-system seasonal total phosphorus, total nitrogen, and | |
| | chlorophyll a concentrations | 3-53 |
| Figure 3.4-3. | 2008-2019 Off-system open-water and ice-cover season TP | |
| | concentrations | 3-54 |
| Figure 3.4-5. | 2008-2019 On-system open-water and ice-cover season TN | ١ |
| | concentrations | 3-59 |
| Figure 3.4-6. | 2008-2019 Off-system open-water and ice-cover season TN | ١ |
| | concentrations | 3-60 |
| Figure 3.4-7. | | |
| | concentrations | 3-64 |
| Figure 3.4-8. | 2008-2019 Off-system open-water and ice-cover season chlorophyll | а |
| | concentrations | 3-65 |
| Figure 4.1-1. | 2010-2019 Benthic invertebrate nearshore (NS) and offshore (OS) sampling | 9 |
| | sites | 4-3 |
| Figure 4.2-1. | 2010 to 2019 Nearshore benthic invertebrate abundance (total no. pe | r |
| | sample) | 4-7 |
| Figure 4.2-2. | 2010 to 2019 Offshore benthic invertebrate abundance (density; total no |). |
| | per m ²) | 4-7 |
| Figure 4.3-1. | 2010 to 2019 Nearshore benthic invertebrate EPT Index | 4-17 |
| Figure 4.3-2. | 2010 to 2019 Offshore benthic invertebrate EPT Index | 4-17 |
| Figure 4.3-3. | 2010 to 2019 Nearshore benthic invertebrate O+C Index | 4-20 |
| Figure 4.3-4. | 2010 to 2019 Offshore benthic invertebrate O+C Index | 4-20 |
| Figure 4.4-1. | 2010 to 2019 Nearshore benthic invertebrate total richness (family level) | 4-23 |
| Figure 4.4-2. | 2010 to 2019 Offshore benthic invertebrate total richness (family-level) | 4-23 |
| Figure 4.4-3. | 2010 to 2019 Nearshore benthic invertebrate EPT richness (family level) | 4-26 |
| Figure 4.4-4. | 2010 to 2019 Offshore benthic invertebrate EPT richness (family level) | 4-26 |
| Figure 4.5-1. | 2010 to 2019 Nearshore benthic invertebrate diversity (family level) | 4-29 |



| Figure 4.5-2. | 2010 to 2019 Offshore benthic invertebrate diversity (family level) | 4-29 |
|----------------|---|------|
| Figure 5.1-1. | 2008-2019 Fish community sampling sites | 5-3 |
| Figure 5.2-1. | 2009-2019 Catch-per-unit-effort (CPUE) of standard gang index gill nets | 5-13 |
| Figure 5.2-2. | 2009-2019 Catch-per-unit-effort (CPUE) of small mesh index gill nets | 5-14 |
| Figure 5.2-3. | 2009-2019 Catch-per-unit-effort (CPUE) of Lake Whitefish | 5-15 |
| Figure 5.2-4. | 2009-2019 Catch-per-unit-effort (CPUE) of Northern Pike | 5-16 |
| Figure 5.2-5. | 2009-2019 Catch-per-unit-effort (CPUE) of Sauger | 5-17 |
| Figure 5.2-6. | 2009-2019 Catch-per-unit-effort (CPUE) of Walleye | 5-18 |
| Figure 5.2-7. | 2009-2019 Catch-per-unit-effort (CPUE) of White Sucker | 5-19 |
| Figure 5.3-1. | 2009-2019 Fulton's condition factor (KF) of Lake Whitefish | 5-27 |
| Figure 5.3-2. | 2009-2019 Fulton's condition factor (KF) of Northern Pike | 5-28 |
| Figure 5.3-3. | 2017-2019 Fulton's condition factor (KF) of Sauger | 5-29 |
| Figure 5.3-4. | 2009-2019 Fulton's condition factor (KF) of Walleye | 5-30 |
| Figure 5.3-5. | 2010-2019 Fulton's condition factor (KF) of White Sucker | 5-31 |
| Figure 5.3-6. | Relative weight (Wr) of Lake Whitefish | 5-38 |
| Figure 5.3-7. | Relative weight (Wr) of Northern Pike | 5-39 |
| Figure 5.3-8. | Relative weight (Wr) of Sauger | 5-40 |
| Figure 5.3-9. | Relative weight (Wr) of Walleye. | 5-41 |
| Figure 5.3-10. | Relative weight (Wr) of White Sucker | 5-42 |
| Figure 5.4-1. | 2009-2019 Fork length-at-age (FLA) 4 of Lake Whitefish | 5-48 |
| Figure 5.4-2. | 2009-2019 Fork length-at-age (FLA) 4 of Northern Pike | 5-49 |
| Figure 5.4-3. | 2017-2019 Fork length-at-age (FLA) 3 of Sauger | 5-50 |
| Figure 5.4-4. | 2009-2019 Fork length-at-age (FLA) 3 of Walleye | 5-51 |
| Figure 5.5-1. | Relative year-class strength (RYCS) of Lake Whitefish | 5-54 |
| Figure 5.5-2. | Relative year-class strength (RYCS) of Northern Pike | 5-55 |
| Figure 5.5-3 | Relative year-class strength (RYCS) of Sauger | 5-56 |
| Figure 5.5-4. | Relative year-class strength (RYCS) of Walleye | 5-57 |
| Figure 5.6-1. | 2009-2019 Hill's effective species richness. | 5-73 |
| Figure 6.1-1. | 2008-2019 Fish mercury sampling sites. | 6-3 |
| Figure 6.2-1. | 2010-2019 Mercury concentration versus fork length of Lake Whitefish | 6-9 |
| Figure 6.2-2. | 2010-2019 Mercury concentration versus fork length of Northern Pike | 6-10 |
| Figure 6.2-3. | 2010-2019 Mercury concentration versus fork length of Walleye | 6-11 |



CAMP 12 YEAR DATA REPORT

2024

| Figure 6.2-4. | 2010-2019 Mercury concentrations of 1-year-old Yellow Perch | . 6-12 |
|---------------|---|--------|
| Figure 6.2-5. | 2010-2019 Length-standardized mean mercury concentrations (±95% | |
| | confidence intervals) of Lake Whitefish | . 6-16 |
| Figure 6.2-6. | 2010-2019 Length-standardized mean mercury concentrations (±95% | |
| | confidence intervals) of Northern Pike | . 6-17 |



LIST OF PHOTOGRAPHS

| Photograph 1. | Threepoint Lake | 1-4 |
|---------------|-------------------|-----|
| Photograph 2. | Apussigamasi Lake | 1-4 |
| Photograph 3. | Footprint Lake | 1-5 |
| Photograph 4. | Leftrook Lake | 1-5 |



LIST OF APPENDICES

| Appendix 2-1. | Seasonal and annual temperature normals derived from ERA5-Land | |
|---------------|---|------|
| | data | 2-32 |
| Appendix 2-2. | Seasonal and precipitation normals derived from RA5-Land data | 2-36 |
| Appendix 3-1. | Water quality sampling sites: 2008-2019 | 3-66 |
| Appendix 4-1. | Benthic invertebrate nearshore and offshore sampling sites: 2008- | |
| | 2019 | 4-30 |
| Appendix 4-2. | Benthic invertebrate nearshore and offshore supporting substrate | |
| | data by year | 4-35 |
| Appendix 5-1. | Gillnetting site information and locations | 5-74 |



ABBREVIATIONS, ACRONYMS, AND UNITS

| ANN | Annual | | | | | | | |
|----------------|--|--|--|--|--|--|--|--|
| CAMP | Coordinated Aquatic Monitoring Program | | | | | | | |
| CCME | Canadian Council of Ministers of the Environment | | | | | | | |
| CL(s) | Confidence limit(s) | | | | | | | |
| cms | Cubic metres per second | | | | | | | |
| CONT | Continuous | | | | | | | |
| CPUE | Catch-per-unit-effort | | | | | | | |
| CRD | Churchill River Diversion | | | | | | | |
| CS | Control structure(s) | | | | | | | |
| DELTs | Deformities, Erosion, Lesions, and Tumours | | | | | | | |
| DL(s) | Detection limit(s) | | | | | | | |
| DO | Dissolved oxygen | | | | | | | |
| ECCC | Environment and Climate Change Canada | | | | | | | |
| EPT | Ephemeroptera (mayflies), Plecoptera (stoneflies), and Trichoptera (caddisflies) | | | | | | | |
| FLA | Fork length-at-age | | | | | | | |
| FNU | Formazin nephelometric unit | | | | | | | |
| GN | Standard gang index gill net | | | | | | | |
| GS(s) | Generating station(s) | | | | | | | |
| h | hour | | | | | | | |
| IC | Ice-cover season | | | | | | | |
| IQR | Interquartile range | | | | | | | |
| KF | Fulton's Condition Factor | | | | | | | |
| m | Metre | | | | | | | |
| m^2 | Metre squared | | | | | | | |
| Max | Maximum | | | | | | | |
| μg/L | Micrograms per litre | | | | | | | |
| mg/L | Milligrams per litre | | | | | | | |
| Min | Minimum | | | | | | | |
| mm | Millimetre | | | | | | | |
| MWQSOGs | Manitoba Water Quality Standards, Objectives, and Guidelines | | | | | | | |
| MWS | Manitoba Water Stewardship | | | | | | | |
| n | Sample size or number of samples | | | | | | | |
| n _F | Number of fish | | | | | | | |
| no. | Number | | | | | | | |
| ND | No data | | | | | | | |



| ns | Number of sites | | | | | | |
|------------------|---|--|--|--|--|--|--|
| NS | Nearshore | | | | | | |
| n _{spp} | Number of species caught in standard and small mesh gill nets | | | | | | |
| NTU | Nephelometric turbidity units | | | | | | |
| O+C | Oligochaeta and Chironomidae | | | | | | |
| OECD | Organization for Economic Cooperation and Development | | | | | | |
| OS | Offshore | | | | | | |
| OW | Open-water season | | | | | | |
| PAL | Protection of aquatic life | | | | | | |
| ppm | Parts per million | | | | | | |
| RCEA | Regional cumulative effects assessment | | | | | | |
| ROT | Rotational | | | | | | |
| RSA | Relative species abundance | | | | | | |
| RYCS | Relative year-class strength | | | | | | |
| SD | Standard deviation | | | | | | |
| SE | Standard error | | | | | | |
| SN | Small mesh index gillnet gang | | | | | | |
| SP | Spring | | | | | | |
| SU | Summer | | | | | | |
| spp. | species | | | | | | |
| T/day | Tonnes per day | | | | | | |
| TN | Total nitrogen | | | | | | |
| TOC | Total organic carbon | | | | | | |
| TP | Total phosphorus | | | | | | |
| TSS | Total suspended solids | | | | | | |
| unid. | unidentified | | | | | | |
| WI | Winter | | | | | | |
| Wr | Relative weight | | | | | | |
| °C | Degree Celsius | | | | | | |



WATERBODY ABBREVIATIONS

| Abbreviation | Waterbody |
|--------------|-------------------|
| 3PT | Threepoint Lake |
| FOOT | Footprint Lake |
| APU | Apussigamasi Lake |
| LEFT | Leftrook Lake |



FISH SPECIES LIST

| Abbreviation | Common Species Name | Species Name |
|--------------|----------------------------|--------------------------|
| BURB | Burbot | Lota lota |
| CISC | Cisco | Coregonus artedi |
| EMSH | Emerald Shiner | Notropis atherinoides |
| GOLD | Goldeye | Hiodon alosoides |
| LKCH | Lake Chub | Couesius plumbeus |
| LKWH | Lake Whitefish | Coregonus clupeaformis |
| LNSC | Longnose Sucker | Catostomus catostomus |
| MOON | Mooneye | Hiodon tergisus |
| NRPK | Northern Pike | Esox lucius |
| SAUG | Sauger | Sander canadensis |
| SHRD | Shorthead Redhorse | Moxostoma macrolepidotum |
| SPSH | Spottail Shiner | Notropis hudsonius |
| TRPR | Trout-perch | Percopsis omiscomaycus |
| WALL | Walleye | Sander vitreus |
| WHSC | White Sucker | Catostomus commersonii |
| YLPR | Yellow Perch | Perca flavescens |



1.0 INTRODUCTION

This report presents the results of monitoring conducted under the Coordinated Aquatic Monitoring Program (CAMP) for years 1 through 12 (i.e., 2008/2009 through 2019/2020) in the Churchill River Diversion Region. The Churchill River Diversion Region is composed of the Rat/Burntwood River system from downstream of Notigi Lake and the Notigi Lake Control Structure (CS) to First Rapids on the Burntwood River, approximately 20 km upstream of Split Lake. Waterbodies and sites monitored in this region over this period included three on-system and one off-system waterbodies as follows:

- Threepoint Lake;
- Apussigamasi Lake;
- Footprint Lake; and
- Leftrook Lake (off-system).

Monitoring on-system waterbodies addresses the primary objective of CAMP – to monitor aquatic ecosystem health along Manitoba Hydro's hydraulic operating system. The off-system waterbodies were included in CAMP to provide regional information collected in a manner consistent with monitoring of on-system waterbodies that will assist in interpreting any observed environmental changes over time. Such comparisons are intended to help distinguish between hydroelectric-related effects and other external factors (e.g., climate change) in each CAMP region.

A summary of monitoring conducted by waterbody or river reach presented in this data report is provided in Table 1-1 and monitoring areas are shown in Figure 1-1. As noted in Table 1-1, monitoring was conducted annually at some waterbodies and river reaches and on a three-year rotation at other sites. Components monitored under CAMP in the Churchill River Diversion Region presented in this report include the physical environment (water regime and sedimentation), water quality, benthic macroinvertebrates, fish community, and mercury in fish. Climatological data for the region are also included to provide supporting information to assist with interpretation of CAMP monitoring results.



Table 1-1. Churchill River Diversion Region CAMP monitoring summary.

| Waterbody/ | | On/Off | -System | Component | | | | | | | | |
|--------------------------------------|--------------|---------------|----------------|-----------------|---------------|------------------|--------------------------|-------------------|-----------------|--|--|--|
| Area | Abbreviation | On- System | Off- System | Water Regime | Sedimentation | Water Quality | Benthic Invertebrates | Fish Community | Fish Mercury | | | |
| Notigi Lake CS | NOT CS | • | | CONT | | | | | | | | |
| Threepoint Lake | 3PT | • | | CONT | | ANN | ANN | ANN | ANN | | | |
| Wuskwatim Generating Station (GS) | WUSK GS | • | | | CONT | | | | | | | |
| Apussigamasi Lake | APU | • | | CONT | | ROT | ROT | ROT | | | | |
| Footprint Lake | FOOT | • | | CONT | | ROT | ROT | ROT | | | | |
| Leftrook Lake | LEFT | | • | CONT | | ANN | ANN | ANN | ANN | | | |

Notes:

1. CONT = site monitored continuously; ANN = site sampled each year; ROT = site sampled every 3 years.



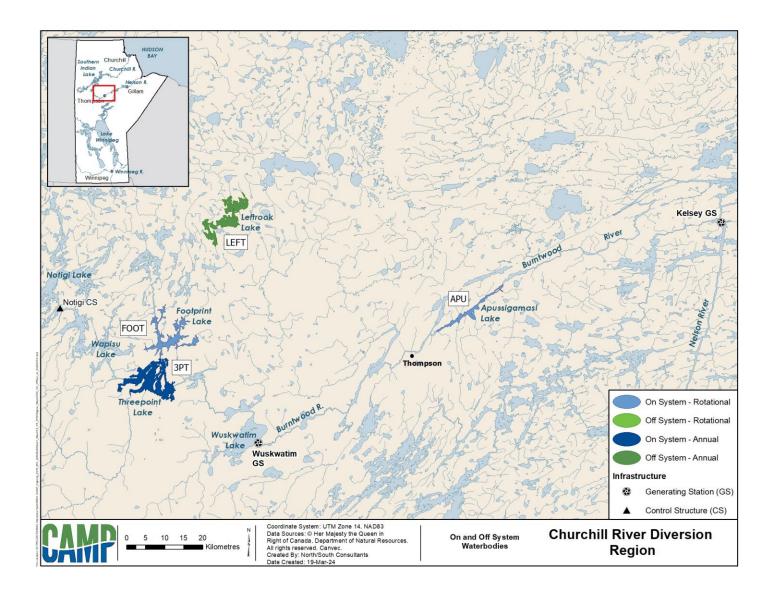


Figure 1-1. On-system and off-system waterbodies and river reaches sampled under CAMP in the Churchill River Diversion Region: 2008-2019.









Photograph 1. Threepoint Lake.





Photograph 2. Apussigamasi Lake.







Photograph 3. Footprint Lake.





Photograph 4. Leftrook Lake.



2.0 PHYSICAL ENVIRONMENT

2.1 INTRODUCTION

The following presents the results of the physical environment monitoring conducted from 2008 to 2019 in the Churchill River Diversion Region. Four waterbodies were monitored in the Churchill River Diversion Region: three on-system sites (Threepoint, Footprint, and Apussigamasi lakes); and one off-system site (Leftrook Lake; Figure 1-1). In addition, a continuous water quality monitoring station is located at the Wuskwatim GS. Though CAMP does not directly monitor climate, data from Environment and Climate Change Canada (ECCC) is included in reporting to contextualize the data collected under each CAMP component. For the Churchill River Diversion Region, meteorological conditions from ECCC's Thompson station are reported.

Three indicators (climate; water regime; and sedimentation) were selected for detailed reporting (Table 2.1-1). Metrics for these indicators include temperature, precipitation, water flow, level and variability, water temperature, continuous turbidity, and suspended sediment load (Table 2.1-1). A detailed description of these indicators in provided in CAMP (2024).

A detailed description of the program design and sampling methods is provided in Technical Document 1, Sections 2.1 and 2.2.

Table 2.1-1. Physical Environment indicators and metrics.

| Indicator | Metric | Units |
|----------------------|-----------------------------------|--|
| Climate ¹ | Temperature | Degrees Celsius (°C) |
| Climate ² | Precipitation | Millimetres (mm) |
| | • Flow | Cubic meters per second (cms) |
| Water Regime | Water Level and Variability | Metres (m) |
| Ü | Water Temperature | Duration of temperature in 5-degree Celsius increments (#days/5 °C) |
| Codinocutation | Continuous Turbidity | Formazin nephelometric unit (FNU) |
| Sedimentation | Suspended Sediment Load | Tonnes/day (T/day) |

Notes:

1. Climate is not monitored through CAMP; data are included for reporting purposes only.



2.2 CLIMATE

In this section, mean monthly air temperatures and total monthly precipitation for each year in the monitoring program (2008-2020) are compared to ECCC climate normals to provide a summary of the Thompson station meteorological conditions. Climate normals are used to summarize the average climatic conditions of a particular location. As recommended by the World Meteorological Organization, ECCC calculates climate normals using a 30-year period (e.g., 1981-2010). The Thompson station is used herein to illustrate climate conditions in the Churchill River Diversion Region.

Historical monthly average air temperature and total monthly precipitation during the monitoring period were calculated based on available daily data from ECCC at multiple stations. It is important to note that the use of multiple stations could introduce inhomogeneities in observations between various stations and the station used for climate normals (Climate ID: 5062922). For instances where datasets were missing more than 10% of the daily data in a month, monthly values were gap-filled using ERA5-Land data (Muñoz Sabater 2019). Seasonal and annual maps derived from ERA5-Land data are also provided in Appendices 2-1 and 2-2 to complement the station data and offer a broader spatial representation of temperature and precipitation conditions across Manitoba. Although the ERA5-Land data correlated reasonably well with the actual observed ECCC data for the Thompson station, it should be noted that ERA5-Land is a gridded reanalysis product, meaning the dataset combines modelled data with observations, and therefore may not provide an entirely accurate representation of observed climate

2.2.1 TEMPERATURE

Figure 2.2-1 illustrates the mean monthly air temperatures (in °C) for each year during the monitoring period compared to the 1981-2010 normal mean temperature. As shown, air temperatures at this location follow a distinct seasonal pattern; warmer in the summer (warmest in July) and cooler in the winter (coldest in January). In general, recorded air temperatures for the monitoring period were consistent with the climate normal pattern. Some deviations can be seen, for example, 2010 recorded considerably warmer temperatures from January to April.

Table 2.2-1 summarizes the mean monthly air temperature data and categorizes each month in the monitoring period as "below normal", "near normal" or "above normal" conditions. It should be noted that the "near normal" category was subjectively defined as +/- 1°C of the ECCC climate normal. Months "below normal" are highlighted in blue, "near normal" are highlighted in grey,



and "above normal" are highlighted in orange. Over the monitoring period, the months of January and September generally experienced warmer than normal conditions (≥ 7 out of 13 months above normal). On an annual basis, no distinct patterns in the data were identified as most years in the monitoring period were near normal conditions; 2010 had the warmest annual average temperature at -0.2°C, while 2014 had the coolest annual average temperature at -4.2°C. The maximum and minimum monthly average air temperatures during the monitoring period were 18.0°C (July 2012 and July 2020) and -29.5°C (December 2013), respectively.



Table 2.2-1. Thompson mean monthly and annual air temperature (in °C) compared to 1981-2010 normal.

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec | Annual |
|---------------------|-------|-------|-------|------|-----|------|------|------|------|------|-------|-------|--------|
| 2008 | -22.4 | -24.9 | -17.2 | -3.0 | 4.9 | 13.0 | 15.7 | 16.5 | 6.3 | 2.8 | -9.6 | -27.5 | -3.8 |
| 2009 | -22.4 | -21.0 | -16.6 | -2.2 | 1.6 | 11.9 | 13.9 | 13.3 | 12.1 | -0.6 | -5.8 | -22.9 | -3.2 |
| 2010 | -18.1 | -16.6 | -4.9 | 3.5 | 5.8 | 13.2 | 17.8 | 14.2 | 7.2 | 2.7 | -8.9 | -18.7 | -0.2 |
| 2011 | -24.8 | -20.3 | -16.0 | -2.1 | 5.8 | 13.7 | 16.5 | 15.5 | 11.3 | 3.1 | -9.7 | -16.6 | -2.0 |
| 2012 | -20.6 | -16.2 | -9.6 | -1.3 | 6.9 | 13.4 | 18.0 | 15.0 | 8.9 | -0.5 | -14.1 | -22.9 | -1.9 |
| 2013 | -25.6 | -20.0 | -14.4 | -6.4 | 7.7 | 15.1 | 16.1 | 14.9 | 11.2 | -0.1 | -14.5 | -29.5 | -3.8 |
| 2014 | -25.4 | -23.4 | -19.5 | -7.5 | 6.9 | 12.6 | 15.8 | 15.0 | 7.0 | 1.6 | -14.9 | -17.9 | -4.2 |
| 2015 | -23.2 | -25.2 | -12.7 | -2.5 | 5.3 | 13.0 | 16.0 | 14.2 | 9.3 | 1.5 | -7.3 | -13.3 | -2.1 |
| 2016 | -19.0 | -21.4 | -10.5 | -4.0 | 8.1 | 13.3 | 16.4 | 14.1 | 10.1 | -0.9 | -3.5 | -21.5 | -1.6 |
| 2017 | -18.1 | -18.4 | -13.0 | -4.3 | 5.9 | 12.2 | 16.7 | 15.1 | 9.9 | 1.0 | -15.5 | -23.2 | -2.6 |
| 2018 | -24.1 | -23.3 | -12.0 | -5.7 | 7.4 | 14.3 | 16.3 | 13.2 | 4.0 | -3.2 | -15.6 | -17.3 | -3.8 |
| 2019 | -25.5 | -24.2 | -9.7 | -2.5 | 4.6 | 12.8 | 16.0 | 13.8 | 8.1 | 0.4 | -12.4 | -20.8 | -3.3 |
| 2020 | -19.4 | -18.9 | -13.5 | -7.1 | 4.5 | 11.8 | 18.0 | 15.1 | 7.2 | -2.5 | -13.0 | -18.8 | -3.0 |
| 1981-2010 Normal | -23.9 | -20.1 | -12.5 | -2.2 | 6.1 | 12.6 | 16.2 | 14.5 | 7.8 | 0.1 | -12.0 | -20.9 | -2.9 |

Below Normal Near Normal Above Normal

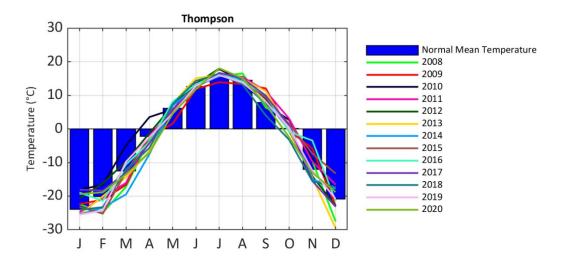


Figure 2.2-1. Thompson mean monthly air temperature (in °C) compared to 1981-2010 normal.



2.2.2 PRECIPITATION

Figure 2.2-2 illustrates the monthly total precipitation (in mm) for each year during the monitoring period compared to the 1981-2010 normal total precipitation. Total precipitation refers to the water equivalent of all types of precipitation. The total precipitation at Thompson follows a noticeable seasonal pattern, where generally the highest amounts of precipitation fall during the summer months (July and August) and the lowest amounts fall during the winter months (January and February). Overall, recorded precipitation for the monitoring period followed similar patterns to the climate normal, although deviations can be seen, such as 2018, where the recorded total precipitation for June and July was much higher than normal and for 2017 (August), which recorded total precipitation well below the normal condition.

Table 2.2-2 summarizes the total monthly precipitation data and categorizes each month in the monitoring period as "below normal", "near normal" or "above normal" conditions. It should be noted that the "near normal" was subjectively defined as +/- 10% of the ECCC climate normal. Months "below normal" are highlighted in light brown, "near normal" are highlighted in grey, and "above normal" are highlighted in green. Over the monitoring period, July and October generally experienced more than normal precipitation (≥ 7 out of 13 months above normal), while February, April, May, June, November, and December generally experienced less than normal precipitation (≥ 7 out of 13 months below normal). On an annual basis, no distinct patterns in the data were identified as most years experienced near normal conditions, however there were more years with above normal precipitation than below normal in the monitoring period; 2020 had the highest annual total precipitation (746.2 mm), while 2013 had the lowest annual total precipitation (427.4 mm). The maximum and minimum monthly total precipitation recorded during the monitoring period were 236.1 mm (August 2010) and 2.8 mm (April 2008), respectively.



Table 2.2-2. Thompson total monthly and annual precipitation (in mm) compared to 1981-2010 normal.

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec | Annual |
|---------------------|------|------|------|------|------|-------|-------|-------|-------|-------|------|------|--------|
| 2008 | 35.2 | 27.2 | 35.2 | 2.8 | 21.4 | 46.5 | 150.7 | 108.7 | 52.6 | 23.2 | 17.0 | 13.6 | 534.1 |
| 2009 | 21.1 | 45.4 | 35.0 | 16.8 | 46.2 | 49.4 | 146.0 | 41.8 | 44.1 | 51.0 | 14.9 | 10.2 | 521.9 |
| 2010 | 32.5 | 10.8 | 14.2 | 13.4 | 39.0 | 9.1 | 91.1 | 236.1 | 59.8 | 60.0 | 26.2 | 31.8 | 624.0 |
| 2011 | 9.4 | 14.0 | 7.4 | 13.2 | 17.0 | 53.0 | 133.5 | 185.4 | 35.0 | 44.3 | 18.6 | 20.5 | 551.3 |
| 2012 | 32.7 | 8.4 | 45.0 | 18.6 | 19.2 | 117.1 | 43.7 | 108.1 | 62.0 | 55.4 | 16.4 | 25.8 | 552.4 |
| 2013 | 10.7 | 9.4 | 9.4 | 9.6 | 18.0 | 33.6 | 77.4 | 59.4 | 63.0 | 63.4 | 65.4 | 8.1 | 427.4 |
| 2014 | 34.4 | 15.6 | 27.0 | 15.4 | 42.6 | 94.6 | 62.8 | 131.8 | 26.6 | 43.8 | 22.7 | 16.2 | 533.5 |
| 2015 | 18.9 | 8.8 | 23.6 | 24.0 | 25.6 | 65.6 | 114.2 | 52.3 | 113.8 | 33.3 | 11.1 | 17.0 | 508.2 |
| 2016 | 12.6 | 13.2 | 20.1 | 3.0 | 47.1 | 44.3 | 112.9 | 38.1 | 98.4 | 114.6 | 19.1 | 23.0 | 546.4 |
| 2017 | 19.0 | 23.6 | 72.4 | 40.4 | 24.2 | 60.4 | 46.5 | 17.2 | 39.8 | 82.2 | 50.9 | 31.5 | 508.1 |
| 2018 | 52.0 | 14.7 | 7.8 | 13.2 | 7.6 | 161.6 | 171.6 | 47.6 | 67.6 | 50.4 | 25.4 | 14.6 | 634.1 |
| 2019 | 28.4 | 12.2 | 9.3 | 38.2 | 47.9 | 59.6 | 102.8 | 72.9 | 31.1 | 39.6 | 37.3 | 19.6 | 498.9 |
| 2020 | 20.1 | 11.6 | 23.1 | 72.1 | 16.2 | 117.8 | 173.6 | 117.7 | 61.2 | 33.7 | 62.5 | 36.6 | 746.2 |
| 1981-2010 Normal | 19.5 | 16.5 | 22.5 | 29.0 | 47.4 | 67.8 | 80.9 | 70.7 | 62.1 | 37.1 | 32.9 | 22.8 | 509.2 |

Below Normal Near Normal Above Normal

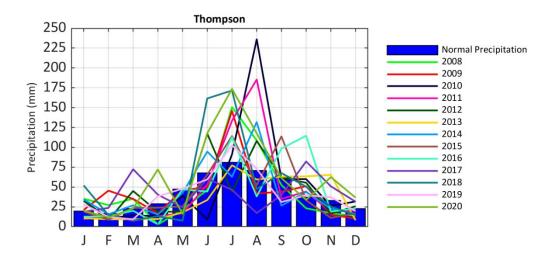


Figure 2.2-2. Thompson total monthly precipitation (in mm) compared to 1981-2010 normal.



2.3 WATER REGIME

The Churchill River Diversion (CRD) improves downstream hydropower generation by transferring the majority of the water flow from the Churchill River to the Nelson River via the Rat River and the Burntwood River. The amount of water diverted to the Nelson River is regulated by the Notigi CS, while Southern Indian Lake is used as a reservoir. Local inflows also contribute to the total water flowing from the Burntwood River into the Nelson River. Additional information on the Churchill River Diversion water regime can be found in the Physical Environment Part IV section of the Regional Cumulative Effects Assessment (RCEA) – Phase II Report (RCEA 2015).

On-System Sites

On-system CAMP monitoring occurred on Threepoint Lake, Footprint Lake, and Apussigamasi Lake (Figure 2.3-1).

Continuous water temperature is measured at the Wuskwatim GS continuous water quality monitoring site. Monitoring started in 2018 and consists of measuring water temperature every 5 minutes and monthly site visits to verify the data. For the water temperature indicator, the continuous water temperature and the duration, in days, that water is below 1 °C and five-degree increments is reported.

Off-System Sites

CAMP monitors Leftrook Lake as the off-system waterbody for this region (Figure 2.3-1).



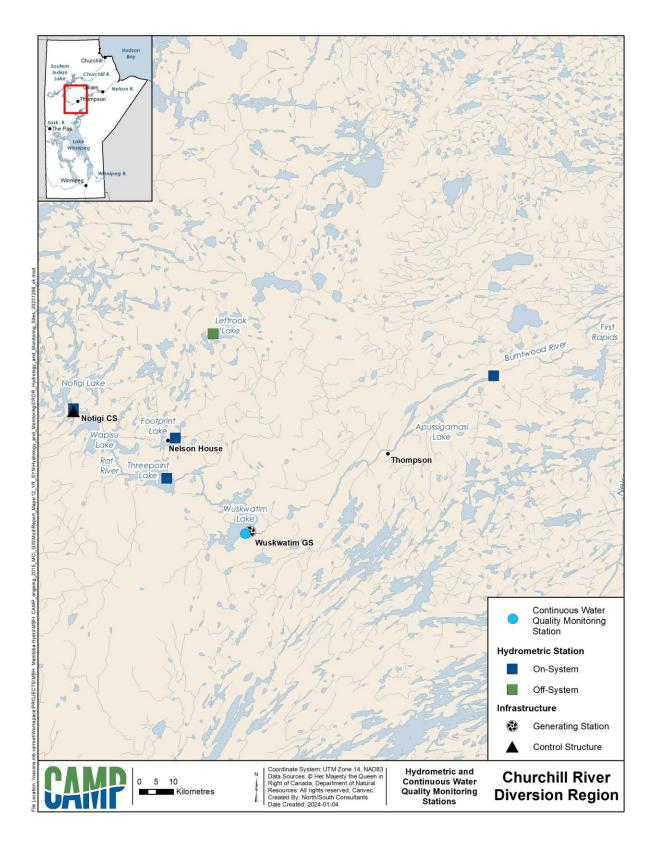


Figure 2.3-1. Hydrometric and continuous water quality monitoring stations in the Churchill River Diversion Region.



2.3.1 FLOW

2.3.1.1 ON-SYSTEM SITES

Notigi Control Structure

From 2008 to 2020, Churchill River Diversion flow conditions ranged from dry to very wet and were more frequently above average than below average compared to the reference period from 1986 to 2015 (Figure 2.3-2 and Table 2.3-1). Monthly mean flow ranged from 523 to 2,392 cms with the overall mean from 2008 to 2020 at 954 cms. Very dry flow conditions, defined as lower than 10th percentile, did not occur in any months during the 2008 to 2020 CAMP monitoring period (Table 2.3-1). Flow conditions were very wet, defined as above the 90th percentile, in parts of eleven years during CAMP, during the following months; July to August 2008, June to December 2009, January 2010, September 2011, June to August and November to December 2012, January, August and October 2013, June to July 2014, May to December 2017 and 2018, January, May to July and November to December 2019, and May to December 2020 (Table 2.3-1).

2.3.1.2 OFF-SYSTEM SITES

There are no off-system flows reported for this region.



Churchill River Diversion at the Notigi CS monthly average flow (cms). Table 2.3-1.

| Year | Annual | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|--------|------|-----|-----|-----|------|------|------|------|------|------|------|------|
| 2008 | 836 | 717 | 719 | 727 | 727 | 836 | 1029 | 1251 | 1100 | 780 | 686 | 708 | 741 |
| 2009 | 1095 | 709 | 699 | 727 | 734 | 889 | 1138 | 1177 | 1465 | 1667 | 1527 | 1291 | 1099 |
| 2010 | 787 | 1010 | 951 | 900 | 879 | 822 | 768 | 653 | 638 | 761 | 738 | 675 | 657 |
| 2011 | 782 | 722 | 735 | 657 | 621 | 732 | 727 | 697 | 951 | 1103 | 926 | 791 | 717 |
| 2012 | 945 | 785 | 770 | 737 | 751 | 970 | 1139 | 1228 | 1037 | 935 | 926 | 1015 | 1040 |
| 2013 | 896 | 977 | 901 | 843 | 789 | 800 | 781 | 891 | 1030 | 967 | 1048 | 903 | 822 |
| 2014 | 831 | 794 | 800 | 793 | 693 | 881 | 983 | 997 | 955 | 840 | 805 | 725 | 702 |
| 2015 | 710 | 717 | 706 | 709 | 732 | 818 | 741 | 763 | 742 | 698 | 662 | 629 | 601 |
| 2016 | 670 | 568 | 543 | 523 | 525 | 685 | 668 | 704 | 851 | 718 | 599 | 764 | 883 |
| 2017 | 1210 | 861 | 836 | 818 | 798 | 1282 | 1654 | 1760 | 1634 | 1473 | 1252 | 1087 | 1030 |
| 2018 | 1239 | 957 | 912 | 859 | 819 | 1155 | 1231 | 1275 | 1568 | 1670 | 1562 | 1258 | 1092 |
| 2019 | 965 | 989 | 923 | 871 | 861 | 1004 | 1046 | 1006 | 943 | 875 | 950 | 1078 | 1025 |
| 2020 | 1490 | 972 | 928 | 887 | 862 | 1046 | 1331 | 1827 | 2392 | 2254 | 2029 | 1817 | 1502 |

| Very Dry Lower than 10th percentile Dry 10th to 30th percentile Average 30th to 70 percentil | 90th |
|--|------|
|--|------|

Notes:

1. Percentiles calculated using 1986-2015 as the reference period.



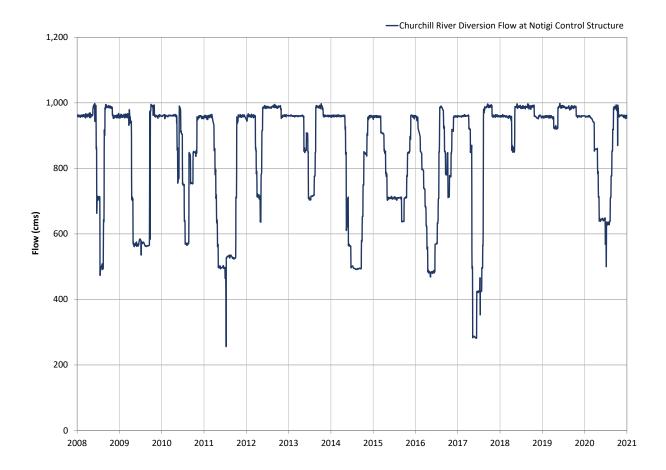


Figure 2.3-2. 2008-2020 Churchill River Diversion at the Notigi CS daily mean flow.

2.3.2 WATER LEVEL AND VARIABILITY

2.3.2.1 ON-SYSTEM SITES

Threepoint Lake

Threepoint Lake is located along the Burntwood River between Notigi and Wuskwatim Lake. Water levels generally follow the pattern set by the CRD flow at Notigi (Figure 2.3-3). During the period from 2008-2020, Threepoint Lake monthly average water levels were more than 0.5 m above the 2008-2020 average in 14 months and were lower than 0.5 m below the 2008-2020 average in 34 months (Table 2.3-2). Threepoint Lake monthly water level variability was lower (below 0.25 m) in 87 months, moderate (between 0.25 and 0.75 m) in 48 months and higher (above 0.75 m) in 20 months (Table 2.3-3).



Footprint Lake

Footprint Lake is located along the Footprint River and is affected by backwater effects of CRD. Water levels generally follow the pattern set by the CRD flow at Notigi (Figure 2.3-4), and closely follow water levels on Threepoint Lake (Figure 2.3-3). During the period from 2008-2020, Footprint Lake monthly average water levels were more than 0.5 m above the 2008-2020 average in 14 months and were lower than 0.5 m below the 2008-2020 average in 34 months (Table 2.3-4). Footprint Lake monthly water level variability was lower (below 0.25 m) in 87 months, moderate (between 0.25 and 0.75 m) in 49 months and higher (above 0.75 m) in 20 months (Table 2.3-5).

Apussigamasi Lake

Apussigamasi Lake is located on the Burntwood River, just downstream from the City of Thompson. The water level gauge on Apussigamasi Lake was established in 2009 as part of CAMP. Water levels generally follow the pattern set by the CRD flow at Notigi (Figure 2.3-5). During the period from 2009-2020, Apussigamasi Lake monthly average water levels were more than 0.5 m above the June 2009-2020 average in 43 months and lower than 0.5 m below the June 2009-2020 average in 30 months (Table 2.3-6). Apussigamasi Lake monthly water level variability was lower (below 0.25 m) in 45 months, moderate (between 0.25 and 0.75 m) in 73 months, and higher (above 0.75 m) in 21 months (Table 2.3-7).

2.3.2.2 OFF-SYSTEM SITES

Leftrook Lake

Water levels on Leftrook Lake vary with precipitation in the local drainage basin (Figure 2.3-6). During the period when data is available from 2010-2020, Leftrook Lake monthly average water levels were more than 0.5 m above the 2010-2020 average in 1 month and not lower than 0.5 m below the 2010-2020 average in any months (Table 2.3-8). Leftrook Lake monthly water level variability was lower (below 0.25 m) in 107 months, moderate (between 0.25 and 0.75 m) in 6 months, and higher (above 0.75 m) in 1 month (Table 2.3-9).



Table 2.3-2. Threepoint Lake monthly average water level (m).

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|
| 2008 | 243.41 | 243.49 | 243.49 | 243.50 | 243.52 | 243.20 | 242.42 | 242.13 | 243.23 | 243.40 | 243.32 | 243.31 |
| 2009 | 243.52 | 243.52 | 243.50 | 243.29 | 242.25 | 242.16 | 242.12 | 241.91 | 241.84 | 243.12 | 243.33 | 243.33 |
| 2010 | 243.47 | 243.41 | 243.34 | 243.31 | 243.16 | 243.09 | 242.49 | 241.84 | 242.84 | 242.96 | 243.26 | 243.42 |
| 2011 | 243.66 | 243.76 | 243.62 | 243.02 | 241.84 | 241.34 | 241.32 | 241.63 | 241.61 | 242.22 | 243.22 | 243.27 |
| 2012 | 243.32 | 243.39 | 243.33 | 242.73 | 242.88 | 243.55 | 243.46 | 243.35 | 243.36 | 243.34 | 243.31 | 243.42 |
| 2013 | 243.68 | 243.76 | 243.48 | 243.38 | 243.28 | 242.99 | 242.43 | 242.38 | 243.21 | 243.52 | 243.40 | 243.32 |
| 2014 | 243.42 | 243.54 | 243.51 | 243.46 | 243.57 | 242.52 | 241.82 | 241.45 | 241.29 | 242.04 | 242.88 | 243.27 |
| 2015 | 243.42 | 243.73 | 243.65 | 243.23 | 242.74 | 242.48 | 242.47 | 242.65 | 242.51 | 242.45 | 242.86 | 243.27 |
| 2016 | 243.55 | 243.52 | 242.98 | 242.22 | 241.59 | 241.59 | 241.93 | 243.23 | 243.21 | 242.82 | 242.93 | 243.36 |
| 2017 | 243.48 | 243.37 | 243.41 | 243.37 | 243.24 | 241.80 | 241.58 | 241.84 | 243.26 | 243.39 | 243.31 | 243.29 |
| 2018 | 243.42 | 243.56 | 243.44 | 243.37 | 243.38 | 243.52 | 243.52 | 243.48 | 243.45 | 243.41 | 243.34 | 243.33 |
| 2019 | 243.62 | 243.90 | 243.54 | 243.34 | 243.47 | 243.56 | 243.46 | 243.37 | 243.42 | 243.40 | 243.29 | 243.25 |
| 2020 | | 243.31 | 243.30 | 243.11 | 242.93 | 242.75 | 242.79 | 242.51 | 243.18 | 243.51 | 243.44 | 243.46 |

| Lower Lower than 0.5 m below average | Average Within 0.5 m below and above average | Higher More than 0.5 m above average |
|--|--|--|
|--|--|--|



Table 2.3-3. Threepoint Lake monthly water level range (m).

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|------|------|------|------|------|------|------|------|------|------|------|------|
| 2008 | 0.09 | 0.05 | 0.02 | 0.04 | 0.12 | 0.95 | 0.67 | 0.93 | 0.58 | 0.04 | 0.09 | 0.13 |
| 2009 | 0.15 | 0.06 | 0.05 | 0.95 | 0.37 | 0.24 | 0.19 | 0.37 | 0.84 | 0.82 | 0.06 | 0.15 |
| 2010 | 0.12 | 0.06 | 0.07 | 0.07 | 0.55 | 0.43 | 1.11 | 0.82 | 0.36 | 0.27 | 0.37 | 0.11 |
| 2011 | 0.43 | 0.29 | 0.14 | 1.08 | 0.89 | 0.17 | 0.37 | 0.28 | 0.30 | 1.61 | 0.17 | 0.07 |
| 2012 | 0.13 | 0.06 | 0.21 | 0.58 | 0.68 | 0.32 | 0.21 | 0.04 | 0.09 | 0.09 | 0.11 | 0.23 |
| 2013 | 0.27 | 0.28 | 0.16 | 0.09 | 0.34 | 0.33 | 0.44 | 0.61 | 0.51 | 0.21 | 0.15 | 0.08 |
| 2014 | 0.14 | 0.07 | 0.09 | 0.04 | 0.85 | 0.97 | 0.46 | 0.25 | 0.09 | 1.27 | 0.44 | 0.19 |
| 2015 | 0.23 | 0.37 | 0.51 | 0.35 | 0.50 | 0.11 | 0.37 | 0.25 | 0.13 | 0.18 | 0.46 | 0.29 |
| 2016 | 0.29 | 0.33 | 0.62 | 0.95 | 0.16 | 0.11 | 0.92 | 0.79 | 0.51 | 0.28 | 0.53 | 0.21 |
| 2017 | 0.10 | 0.05 | 0.11 | 0.24 | 1.26 | 1.15 | 0.44 | 1.62 | 0.40 | 0.07 | 0.13 | 0.04 |
| 2018 | 0.23 | 0.07 | 0.13 | 0.15 | 0.16 | 0.14 | 0.10 | 0.13 | 0.12 | 0.04 | 0.10 | 0.10 |
| 2019 | 0.45 | 0.23 | 0.33 | 0.12 | 0.22 | 0.11 | 0.13 | 0.08 | 0.10 | 0.07 | 0.10 | 0.15 |
| 2020 | | 0.03 | 0.04 | 0.31 | 0.23 | 0.34 | 0.31 | 0.29 | 0.82 | 0.11 | 0.02 | 0.29 |

| Lower Variability | Moderate Variability | Higher Variability |
|-------------------|----------------------|--------------------|
| Below 0.25 m | 0.25 to 0.75 m | Above 0.75 m |



Table 2.3-4. Footprint Lake monthly average water level (m).

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|
| 2008 | 243.39 | 243.47 | 243.48 | 243.50 | 243.52 | 243.23 | 242.43 | 242.08 | 243.21 | 243.38 | 243.30 | 243.30 |
| 2009 | 243.49 | 243.49 | 243.48 | 243.29 | 242.24 | 242.15 | 242.11 | 241.89 | 241.84 | 243.12 | 243.34 | 243.32 |
| 2010 | 243.45 | 243.41 | 243.33 | 243.28 | 243.13 | 243.06 | 242.46 | 241.81 | 242.80 | 242.93 | 243.24 | 243.40 |
| 2011 | 243.64 | 243.74 | 243.60 | 243.02 | 241.83 | 241.35 | 241.35 | 241.61 | 241.60 | 242.20 | 243.20 | 243.27 |
| 2012 | 243.31 | 243.36 | 243.31 | 242.71 | 242.84 | 243.50 | 243.43 | 243.34 | 243.36 | 243.36 | 243.32 | 243.41 |
| 2013 | 243.65 | 243.74 | 243.45 | 243.36 | 243.26 | 242.97 | 242.41 | 242.36 | 243.19 | 243.49 | 243.37 | 243.30 |
| 2014 | 243.40 | 243.52 | 243.48 | 243.43 | 243.54 | 242.48 | 241.77 | 241.40 | 241.26 | 242.01 | 242.86 | 243.24 |
| 2015 | 243.39 | 243.71 | 243.63 | 243.19 | 242.70 | 242.44 | 242.40 | 242.62 | 242.47 | 242.43 | 242.84 | 243.24 |
| 2016 | 243.53 | 243.49 | 242.96 | 242.19 | 241.55 | 241.56 | 241.89 | 243.19 | 243.17 | 242.78 | 242.91 | 243.34 |
| 2017 | 243.45 | 243.33 | 243.38 | 243.36 | 243.23 | 241.78 | 241.56 | 241.82 | 243.24 | 243.37 | 243.30 | 243.27 |
| 2018 | 243.40 | 243.54 | 243.42 | 243.35 | 243.34 | 243.49 | 243.48 | 243.44 | 243.41 | 243.37 | 243.31 | 243.31 |
| 2019 | 243.59 | 243.86 | 243.52 | 243.33 | 243.45 | 243.53 | 243.42 | 243.34 | 243.39 | 243.36 | 243.26 | 243.25 |
| 2020 | 243.35 | 243.31 | 243.28 | 243.09 | 242.91 | 242.72 | 242.75 | 242.48 | 243.14 | 243.48 | 243.41 | 243.43 |

| Lower Lower than 0.5 m below average | Average Within 0.5 m below and above average | Higher More than 0.5 m above average |
|--|--|--|
|--|--|--|



Table 2.3-5. Footprint Lake monthly average water level (m).

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|------|------|------|------|------|------|------|------|------|------|------|------|
| 2008 | 0.09 | 0.05 | 0.02 | 0.04 | 0.11 | 0.95 | 0.66 | 0.92 | 0.56 | 0.03 | 0.08 | 0.13 |
| 2009 | 0.15 | 0.05 | 0.06 | 0.94 | 0.38 | 0.23 | 0.19 | 0.36 | 0.84 | 0.79 | 0.06 | 0.14 |
| 2010 | 0.13 | 0.06 | 0.08 | 0.04 | 0.52 | 0.43 | 1.11 | 0.80 | 0.35 | 0.26 | 0.37 | 0.11 |
| 2011 | 0.43 | 0.29 | 0.14 | 1.08 | 0.90 | 0.16 | 0.26 | 0.27 | 0.27 | 1.61 | 0.18 | 0.07 |
| 2012 | 0.12 | 0.05 | 0.21 | 0.58 | 0.68 | 0.31 | 0.17 | 0.05 | 0.06 | 0.09 | 0.12 | 0.21 |
| 2013 | 0.27 | 0.28 | 0.16 | 0.08 | 0.34 | 0.32 | 0.45 | 0.60 | 0.50 | 0.21 | 0.14 | 0.08 |
| 2014 | 0.14 | 0.05 | 0.10 | 0.04 | 0.85 | 1.00 | 0.46 | 0.22 | 0.11 | 1.29 | 0.44 | 0.19 |
| 2015 | 0.23 | 0.37 | 0.51 | 0.36 | 0.50 | 0.10 | 0.38 | 0.25 | 0.12 | 0.17 | 0.47 | 0.29 |
| 2016 | 0.28 | 0.32 | 0.63 | 0.96 | 0.16 | 0.12 | 0.91 | 0.79 | 0.50 | 0.28 | 0.54 | 0.21 |
| 2017 | 0.11 | 0.06 | 0.12 | 0.24 | 1.25 | 1.16 | 0.42 | 1.59 | 0.40 | 0.06 | 0.13 | 0.04 |
| 2018 | 0.23 | 0.07 | 0.14 | 0.15 | 0.16 | 0.13 | 0.09 | 0.14 | 0.11 | 0.04 | 0.09 | 0.10 |
| 2019 | 0.45 | 0.22 | 0.31 | 0.11 | 0.21 | 0.11 | 0.13 | 0.07 | 0.09 | 0.06 | 0.08 | 0.16 |
| 2020 | 0.06 | 0.09 | 0.03 | 0.30 | 0.22 | 0.35 | 0.31 | 0.29 | 0.80 | 0.11 | 0.02 | 0.29 |

| Lower Variability | Moderate Variability | Higher Variability |
|-------------------|----------------------|--------------------|
| Below 0.25 m | 0.25 to 0.75 m | Above 0.75 m |



Table 2.3-6. Apussigamasi Lake monthly water average level (m).

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|
| 2008 | | | | | | | | | | | | |
| 2009 | | | | | | 186.39 | 186.33 | 185.94 | 185.75 | 186.67 | 187.11 | 187.80 |
| 2010 | 188.33 | 188.09 | 187.72 | 187.28 | 187.06 | 186.81 | 186.53 | 185.89 | 186.68 | 186.78 | 187.11 | 187.91 |
| 2011 | 187.90 | 188.10 | 187.94 | 187.25 | 186.11 | 185.43 | 185.37 | 185.83 | 185.77 | 185.87 | 187.06 | 188.20 |
| 2012 | 188.02 | 187.90 | 187.58 | 187.02 | 186.74 | 187.31 | 187.26 | 187.11 | 187.13 | 187.12 | 187.45 | 188.07 |
| 2013 | 187.98 | 187.85 | 187.55 | 187.44 | 187.27 | 186.82 | 186.36 | 186.17 | 186.83 | 187.32 | 187.38 | 187.95 |
| 2014 | 187.80 | 187.82 | 187.76 | 187.56 | 187.83 | 186.69 | 185.96 | 185.51 | 185.32 | 185.81 | 186.93 | 187.86 |
| 2015 | 187.79 | 187.81 | 187.86 | 187.41 | 186.75 | 186.44 | 186.36 | 186.66 | 186.56 | 186.40 | 186.74 | 187.72 |
| 2016 | 188.12 | 188.11 | 187.43 | 186.59 | 185.67 | 185.70 | 185.82 | 186.89 | 187.04 | 186.82 | 186.74 | 187.61 |
| 2017 | 188.00 | 187.90 | 187.77 | 187.62 | 187.73 | 186.17 | 185.67 | 185.55 | 186.87 | 187.13 | 187.41 | 187.69 |
| 2018 | 187.98 | 188.06 | 187.73 | 187.50 | 187.42 | 187.35 | 187.46 | 187.30 | 187.25 | 187.18 | 187.39 | 187.92 |
| 2019 | 187.93 | 187.95 | 187.82 | 187.41 | 187.36 | 187.37 | 187.27 | 187.15 | 187.23 | 187.16 | 187.34 | 187.78 |
| 2020 | 188.01 | 187.89 | 187.61 | 187.23 | 187.10 | 186.87 | 186.91 | 186.53 | 186.93 | 187.31 | 187.79 | 187.86 |



Table 2.3-7. Apussigamasi Lake monthly water level range (m).

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|------|------|------|------|------|------|------|------|------|------|------|------|
| 2008 | | | | | | | | | | | | |
| 2009 | | | | | | 0.04 | 0.09 | 0.24 | 0.25 | 1.11 | 0.06 | 1.26 |
| 2010 | 0.32 | 0.30 | 0.36 | 0.44 | 0.14 | 0.21 | 0.71 | 0.39 | 0.64 | 0.17 | 0.87 | 0.28 |
| 2011 | 0.18 | 0.38 | 0.44 | 0.81 | 1.16 | 0.33 | 0.29 | 0.57 | 0.47 | 1.13 | 1.16 | 0.53 |
| 2012 | 0.07 | 0.27 | 0.29 | 0.75 | 0.25 | 0.61 | 0.30 | 0.15 | 0.07 | 0.17 | 0.61 | 0.45 |
| 2013 | 0.34 | 0.17 | 0.24 | 0.07 | 0.54 | 0.23 | 0.57 | 0.33 | 0.58 | 0.40 | 0.68 | 0.15 |
| 2014 | 0.05 | 0.07 | 0.17 | 0.12 | 0.97 | 1.04 | 0.49 | 0.25 | 0.19 | 1.13 | 1.11 | 0.47 |
| 2015 | 0.23 | 0.25 | 0.47 | 0.30 | 0.73 | 0.24 | 0.46 | 0.37 | 0.30 | 0.17 | 1.01 | 0.76 |
| 2016 | 0.10 | 0.47 | 0.74 | 1.04 | 0.36 | 0.21 | 0.73 | 0.76 | 0.23 | 0.35 | 0.35 | 1.10 |
| 2017 | 0.61 | 0.48 | 0.29 | 0.28 | 1.11 | 1.55 | 0.39 | 1.10 | 0.61 | 0.15 | 0.72 | 0.27 |
| 2018 | 0.29 | 0.17 | 0.40 | 0.11 | 0.71 | 0.32 | 0.23 | 0.25 | 0.20 | 0.10 | 0.82 | 0.45 |
| 2019 | 0.33 | 0.38 | 0.57 | 0.18 | 0.26 | 0.14 | 0.10 | 0.20 | 0.10 | 0.10 | 0.66 | 0.38 |
| 2020 | 0.14 | 0.21 | 0.35 | 0.35 | 0.49 | 0.43 | 0.41 | 0.53 | 0.71 | 0.22 | 1.00 | 0.62 |

| Lower Variability | Moderate Variability | Higher Variability |
|-------------------|----------------------|--------------------|
| Below 0.25 m | 0.25 to 0.75 m | Above 0.75 m |



Table 2.3-8. Leftrook Lake monthly average water level (m).

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|
| 2008 | | | | | | | | | | | | |
| 2009 | | | | | | | | | | | | |
| 2010 | | | | | | 253.68 | 253.59 | | | | | |
| 2011 | | | 253.77 | 253.75 | 253.88 | 253.83 | 253.81 | 254.01 | 254.13 | 253.98 | 253.92 | 253.91 |
| 2012 | 253.89 | 253.86 | 253.83 | 253.86 | 254.07 | 254.10 | 254.04 | 253.88 | 253.95 | 253.96 | 253.99 | 253.95 |
| 2013 | 253.88 | 253.83 | 253.80 | 253.77 | 253.87 | 253.87 | 253.80 | 253.79 | 253.78 | 253.98 | 254.12 | 254.05 |
| 2014 | 253.96 | 253.90 | 253.86 | 253.82 | 253.88 | | | | | | | 253.93 |
| 2015 | 253.91 | 253.87 | 253.82 | 253.82 | 254.00 | 253.96 | 253.93 | 254.18 | 254.22 | 254.23 | 254.14 | 254.05 |
| 2016 | 253.98 | 253.94 | 253.91 | 253.89 | 254.00 | 254.11 | 254.02 | 253.99 | 253.92 | 254.12 | 254.10 | 254.06 |
| 2017 | 253.99 | 253.95 | 253.97 | 253.96 | 254.39 | 254.67 | 254.22 | 253.96 | 253.83 | 253.80 | 253.86 | 253.87 |
| 2018 | 253.86 | 253.84 | 253.79 | 253.77 | 254.04 | 254.01 | 254.04 | 254.03 | 253.92 | 253.87 | 253.87 | 253.84 |
| 2019 | 253.81 | 253.79 | 253.77 | 253.78 | 253.93 | 253.95 | 253.90 | 253.87 | 253.87 | 253.88 | 253.85 | 253.83 |
| 2020 | 253.82 | 253.80 | 253.78 | 253.81 | 254.03 | 254.27 | 254.42 | 254.37 | 254.21 | 254.13 | 254.06 | 253.99 |



Table 2.3-9. Leftrook Lake monthly water level range (m).

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|------|------|------|------|------|------|------|------|------|------|------|------|
| 2008 | | | | | | | | | | | | |
| 2009 | | | | | | | | | | | | |
| 2010 | | | | | | 0.11 | 0.02 | | | | | |
| 2011 | | | 0.04 | 0.06 | 0.11 | 0.12 | 0.12 | 0.35 | 0.19 | 0.08 | 0.04 | 0.02 |
| 2012 | 0.02 | 0.05 | 0.02 | 0.08 | 0.18 | 0.15 | 0.19 | 0.08 | 0.09 | 0.05 | 0.03 | 0.06 |
| 2013 | 0.06 | 0.04 | 0.03 | 0.03 | 0.19 | 0.11 | 0.06 | 0.04 | 0.06 | 0.31 | 0.03 | 0.10 |
| 2014 | 0.06 | 0.05 | 0.04 | 0.04 | 0.20 | | | | | | | 0.01 |
| 2015 | 0.04 | 0.04 | 0.05 | 0.20 | 0.05 | 0.04 | 0.22 | 0.06 | 0.15 | 0.08 | 0.10 | 0.08 |
| 2016 | 0.05 | 0.03 | 0.03 | 0.04 | 0.15 | 0.10 | 0.08 | 0.10 | 0.01 | 0.00 | 0.00 | 0.07 |
| 2017 | 0.05 | 0.03 | 0.06 | 0.02 | 0.97 | 0.50 | 0.34 | 0.18 | 0.10 | 0.08 | 0.03 | 0.02 |
| 2018 | 0.01 | 0.04 | 0.05 | 0.11 | 0.22 | 0.04 | 0.09 | 0.14 | 0.05 | 0.03 | 0.02 | 0.03 |
| 2019 | 0.02 | 0.02 | 0.02 | 0.09 | 0.12 | 0.04 | 0.07 | 0.03 | 0.09 | 0.05 | 0.02 | 0.03 |
| 2020 | 0.02 | 0.02 | 0.01 | 0.04 | 0.45 | 0.15 | 0.23 | 0.27 | 0.10 | 0.07 | 0.05 | 0.07 |

| Lower Variability | Moderate Variability | Higher Variability |
|-------------------|----------------------|--------------------|
| Below 0.25 m | 0.25 to 0.75 m | Above 0.75 m |



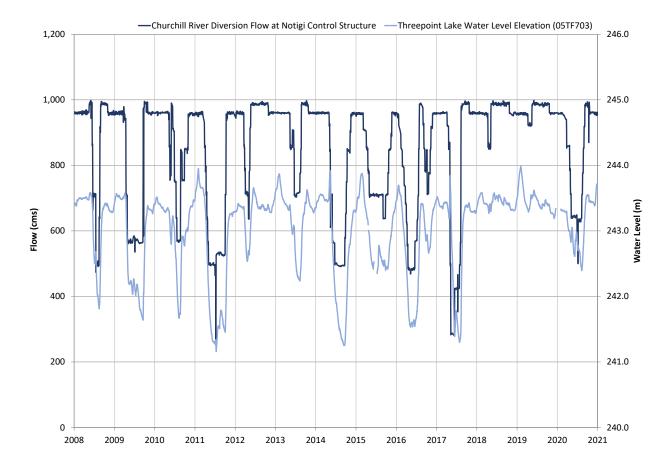


Figure 2.3-3. 2008-2020 Upper Churchill at the Notigi CS daily mean flow and Threepoint Lake daily mean water level.



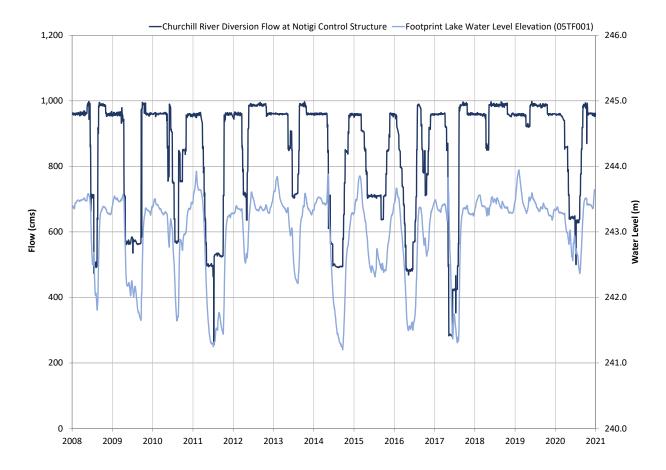


Figure 2.3-4. 2008-2020 Upper Churchill River at the Notigi CS daily mean flow and Footprint Lake daily mean water level.



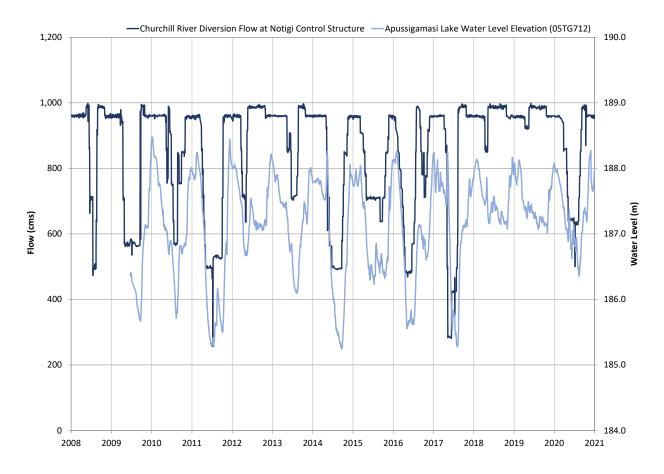


Figure 2.3-5. 2008-2020 Upper Churchill River at the Notigi CS daily mean flow and Apussigamasi Lake daily mean water level.



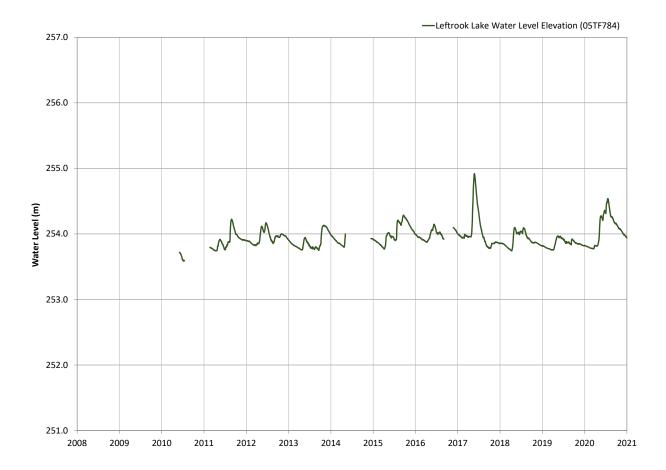


Figure 2.3-6. 2008-2020 Leftrook Lake daily mean water level.

2.3.3 WATER TEMPERATURE

2.3.3.1 ON-SYSTEM SITES

Wuskwatim Generating Station

Water temperature in the Churchill River Diversion Region is monitored at the continuous water quality monitoring station located at the Wuskwatim GS (Figure 2.3-1). Water temperatures drop to near 0 °C during the winter period and begin to increase in April (Figure 2.3-7). Temperatures peaked around 19 °C in August during the two summers since monitoring has started, returning to near 0 °C in early November.

The duration, in days, that water temperature is within different temperature ranges is used as a metric (Table 2.3-10). The number of days that the water temperature was below 1 °C, which is



used as a proxy-metric for the duration of the ice-cover period, ranged from 176 to 193 days. In summer, the there were no days above 20 °C in the two years of monitoring.

2.3.3.2 OFF-SYSTEM SITES

There are no continuous monitoring off-system sites in this region.

Table 2.3-10. 2017-19 Wuskwatim GS water temperature ranges.

| Monitoring | Number of Days in Temperature Range ² | | | | | | | | | |
|-------------------|--|--------|---------|---------|---------|--------|--|--|--|--|
| Year ¹ | <1 °C | 1-5 °C | 5-10 °C | 10-15°C | 15-20°C | >20 °C | | | | |
| 2017 | | | | | | | | | | |
| 2018 | 176 | | | | | | | | | |
| 2019 | 193 | 38 | 42 | 46 | 66 | 0 | | | | |

Notes:

- 1. Period <1°C is for the entire winter period (e.g., 2017 monitoring year is from Nov 2017 to May 2018).
- 2. The duration has been estimated using data from nearby gauging stations to infill missing data when available.

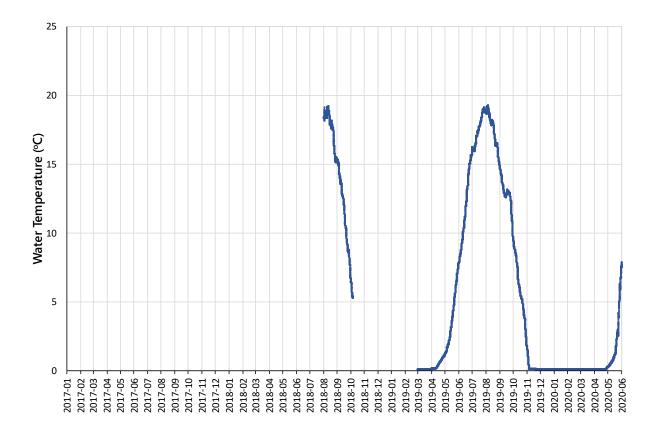


Figure 2.3-7. 2017-2019 Wuskwatim GS continuous water temperature.



2.4 SEDIMENTATION

The following presents the results of sedimentation monitoring conducted in the Churchill River Diversion Region. Monitoring occurred on-system at the continuous water quality monitoring site located at the Wuskwatim GS (Figure 2.3-1). Monitoring started in 2018 (Table 2.4-1) and consists of measuring turbidity every 5 minutes and monthly site visits to verify the data and collect water samples for measuring total suspended solids (TSS) used in calculating the sediment load. For the sedimentation indicator, two metrics (continuous turbidity and suspended sediment load) were selected for detailed reporting (Table 2.4-2).

Table 2.4-1. 2008-2019 sedimentation sampling inventory.

| Waterbody/ | | Sampling Year | | | | | | | | | | |
|--------------|------|---------------|------|------|------|------|------|------|------|------|------|------|
| Area | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
| Wuskwatim GS | | | | | | | | | | | • | • |

Table 2.4-2. Sedimentation indicators and metrics.

| Indicator | Metric | Units |
|---------------|-------------------------|-------|
| Cadinantation | Continuous turbidity | FNU |
| Sedimentation | Suspended sediment load | T/day |

2.4.1 CONTINUOUS TURBIDITY

2.4.1.1 ON-SYSTEM SITES

Wuskwatim Generating Station

Turbidity in the Churchill River Diversion Region is monitored at the continuous water quality monitoring station located at the Wuskwatim GS. The average monthly turbidity ranged from 8.0 to 30.7 FNU (Table 2.4-3, Figure 2.4-1) with the hourly turbidity ranging from 7 to 38 FNU (Figure 2.4-2).

The continuous data set at the Wuskwatim GS is relatively short and will take more data to confirm patterns and trends in the data. The data set shows greater variability and higher turbidity during the open water period and declining trends during the ice-cover months. During the two summers



monitored, peak turbidity was over 30 FNU with the highest occurring in 2019 when it reached over 35 on several occasions. Minimum turbidity reached 7–9 FNU during the two winters monitored.

2.4.1.2 OFF-SYSTEM SITES

There are no continuous monitoring off-system sites in this region.

Table 2.4-3. 2017-2019 Wuskwatim GS average monthly turbidity.

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|------|------|------|-----|------|------|------|------|------|------|------|------|
| 2017 | | | | | | | | | | | | |
| 2018 | | | | | | | | 20.7 | 19.0 | | | 19.5 |
| 2019 | | | 9.8 | 8.0 | 10.5 | 14.0 | 20.4 | 23.4 | 24.2 | 29.0 | 30.7 | 26.5 |
| 2020 | 17.8 | 13.0 | 10.0 | 8.1 | | | | | | | | |

Notes:

1. Monthly data only shown for months with more than 15 days of data.



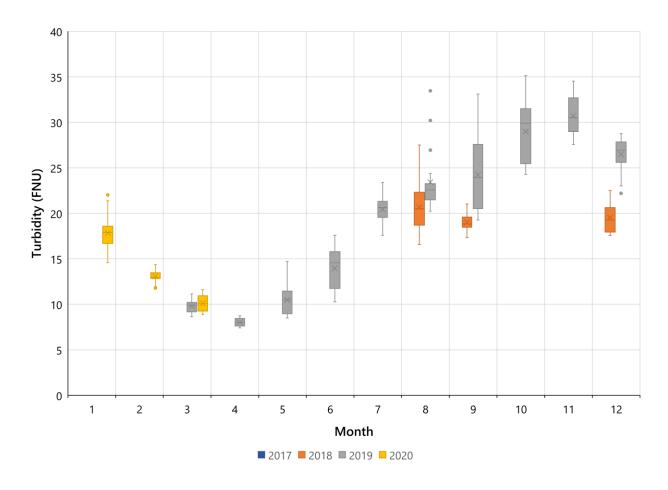


Figure 2.4-1. 2017-2019 Wuskwatim GS monthly turbidity.



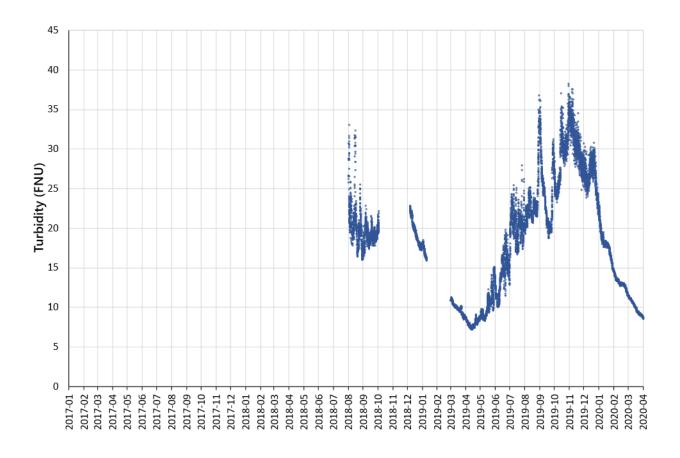


Figure 2.4-2. 2017-2019 Wuskwatim GS continuous turbidity.

2.4.2 SUSPENDED SEDIMENT LOAD

2.4.2.1 ON-SYSTEM SITES

Wuskwatim Generating Station

Sediment load is estimated using the discharge data, continuous turbidity data (Figure 2.4-2) and water samples collected to correlate the turbidity to TSS. The average monthly sediment load ranged from 158 to 836 T/day (Table 2.4-4, Figure 2.4-3). While there is limited data to make observations from, the open-water sediment load was several magnitude times higher than the ice-cover period. There were two peak periods during the 2019 open-water period, occurring in June and November (Figure 2.4-4).



2.4.2.2 OFF-SYSTEM SITES

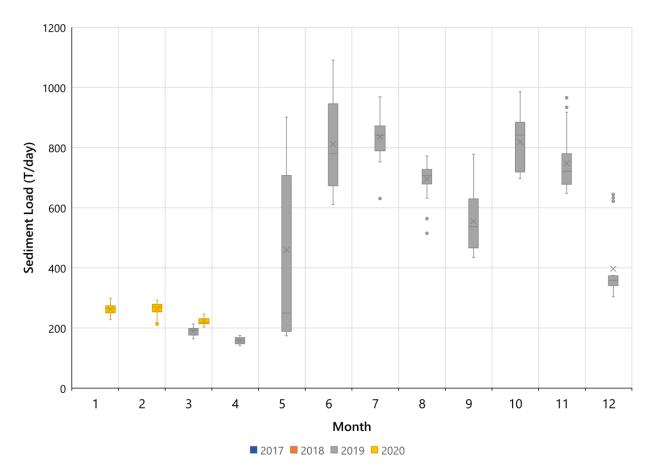
There are no continuous monitoring off-system sites in this region.

Table 2.4-4. 2017-19 Wuskwatim GS average monthly sediment load.

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|-----|-----|-----|-----|-----|-----|-----|-----|-----|-----|-----|-----|
| 2017 | | | | | | | | | | | | |
| 2018 | | | | | | | | | | | | |
| 2019 | | | 189 | 158 | 459 | 811 | 836 | 695 | 555 | 818 | 747 | 397 |
| 2020 | 262 | 263 | 221 | 200 | | | | | | | | |

Notes:

- 1. Monthly average only shown for months with more than 15 days of data.
- 2. Some months are missing TSS measurements to estimate the load.



*Monthly data only shown for months with more than 15 days of data.

Figure 2.4-3. 2017-2019 Wuskwatim GS monthly sediment load.



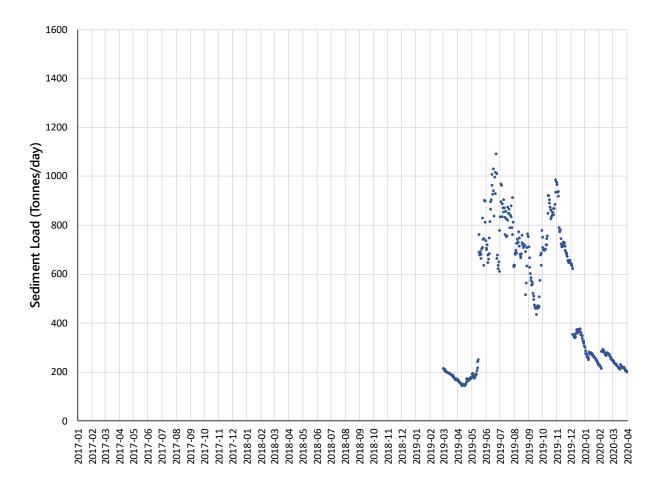
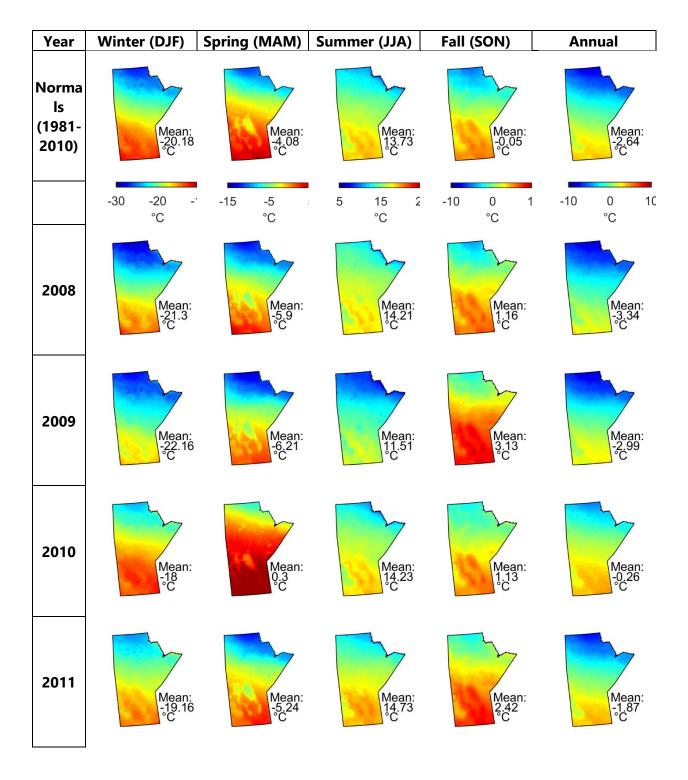


Figure 2.4-4. 2017-2019 Wuskwatim GS daily sediment load.



APPENDIX 2-1. SEASONAL AND ANNUAL TEMPERATURE NORMALS DERIVED FROM ERA5-LAND DATA







| Year | Winter (DJF) | Spring (MAM) | Summer (JJA) | Fall (SON) | Annual |
|--------------------------------|-----------------------|----------------------|----------------|---------------------|----------------------|
| Norma Is (1981- 2010) | Mean: -20.18 | Mean: -4.08 | Mean: 13.73 | Mean: -0.05 | Mean: -2,64 |
| | -30 -20 - °C | · -15 -5 °C | 5 15 2 °C | -10 0 1 | -10 0 10 °C |
| 2012 | Mean: -17.18 °C | Mean: -2.21 | Mean: 14.67 | Mean: -0,56 | Mean: -1.6 °C |
| 2013 | Mean: -21.14 | Mean: -5.44 °C | Mean: 14.77 °C | Mean: 0.18 °C | Mean: -3.4 °C |
| 2014 | Mean: -23.96 | Mean: -6.82 °C | Mean: 14.22 °C | Mean: -1, 1 | Mean: -3,57 °C |
| 2015 | Mean: -20.59 | Mean: -4.27 °C | Mean: 13.93 | Mean: 1.85 | Mean: -2,01 |



| Year | Winter (DJF) | Spring (MAM) | Summer (JJA) | Fall (SON) | Annual |
|--------------------------------|-----------------|---------------------|----------------|----------------------|----------------------|
| Norma Is (1981- 2010) | Mean: -20.18 | Mean: -4.08 | Mean: 13.73 | Mean: -0.05 | Mean: -2.64 |
| | -30 -20 - °C | r -15 -5 °C | 5 15 2 °C | -10 0 1 | -10 0 10 °C |
| 2016 | Mean: -17.27 | Mean: -3.97 | Mean: 14.12 °C | Mean: 3.06 | Mean: -1.52 °C |
| 2017 | Mean: -17.4 | Mean: -4.7 °C | Mean: 14.65 | Mean: -0.13 | Mean: -2.06 |
| 2018 | Mean: -21.53 | Mean: -3,95 | Mean: 14.64 | Mean: -2.88 °C | Mean: |
| 2019 | Mean: -20.97 | Mean: | Mean: 13.83 | Mean: 0.05 C | Mean: -2.86 °C |
| 2020 | Mean: -18.09 | Mean: -5.53 | Mean: 14.49 | Mean: -1.15 °C | Mean: -2:39 °C |



APPENDIX 2-2. SEASONAL AND PRECIPITATION NORMALS DERIVED FROM ERA5-LAND DATA



| Year | Winter (DJF) | Spring (MAM) | Summer (JJA) | Fall (SON) | Annual |
|--------------------------------|---------------------|-----------------------|-----------------------|-----------------------|-----------------------|
| Norma Is (1981- 2010) | Mean: 74.31 mm | Mean: 125.65 mm | Mean: 258.17 mm | Mean: 172.15 mm | Mean: 630.29 mm |
| | 0 75 1 mm | 0 125 24 mm | 0 250 50 mm | 0 175 3: mm | 300 600 900 mm |
| 2008 | Mean: 78.4 mm | Mean: 104.61 mm | Mean: 290.53 mm | Mean: 178.35 mm | Mean: 648.56 mm |
| 2009 | Mean: 65.86 mm | Mean: 132.3 mm | Mean: 267.91 mm | Mean: 133.36 mm | Mean: 602.31 mm |
| 2010 | Mean: 68.2 mm | Mean: 130.98 mm | Mean: 325.23 mm | Mean: 191.98 mm | Mean: 726.41 mm |
| 2011 | Mean: 78.2 mm | Mean: 121.34 mm | Mean: 256.51 mm | Mean: 156.49 mm | Mean: 610.91 mm |



| Year | Winter (DJF) | Spring (MAM) | Summer (JJA) | Fall (SON) | Annual |
|--------------------------------|----------------------|-----------------------|-----------------|-----------------------|-----------------------|
| Norma Is (1981- 2010) | Mean: 74.31 mm | Mean: 125.65 mm | Mean: 258.17 mm | Mean: 172.15 mm | Mean: 630.29 mm |
| | 0 75 1 mm | 0 125 2 mm | 0 250 5 mm | 0 175 39 mm | 300 600 900 mm |
| 2012 | Mean: 71.65 mm | Mean: 150.46 mm | Mean: 257.34 mm | Mean: 187.43 mm | Mean: 677.42 mm |
| 2013 | Mean: 83.27 mm | Mean: 111.29 mm | Mean: 205.49 mm | Mean: 196.31 mm | Mean: 573.79 mm |
| 2014 | Mean: 75.15 mm | Mean: 112.99 mm | Mean: 262.94 mm | Mean: 167.02 mm | Mean: 620.67 mm |
| 2015 | Mean: 64.52 mm | Mean: 122.35 mm | Mean: 277.73 mm | Mean: 191.73 mm | Mean: 662.9 mm |



| Year | Winter (DJF) | Spring (MAM) | Summer (JJA) | Fall (SON) | Annual |
|--------------------------------|----------------------|-----------------------|-----------------------|-----------------------|-----------------------|
| Norma Is (1981- 2010) | Mean: 74.31 mm | Mean: 125.65 mm | Mean: 258.17 mm | Mean: 172.15 mm | Mean: 630.29 mm |
| | 0 75 1 mm | 0 125 29 mm | 0 250 5 mm | 0 175 3: mm | 300 600 900 mm |
| 2016 | Mean: 72.14 mm | Mean: 119.58 mm | Mean: 245.3 mm | Mean: 237.28 mm | Mean: 674.47 mm |
| 2017 | Mean: 89.39 mm | Mean: 158.25 mm | Mean: 187.25 mm | Mean: 194.52 mm | Mean: 629.66 mm |
| 2018 | Mean: 71.69 mm | Mean: 82.03 mm | Mean: 270.02 mm | Mean: 152.2 mm | Mean: 570.82 mm |
| 2019 | Mean: 68.35 mm | Mean: 92.78 mm | Mean: 285.73 mm | Mean: 186.43 mm | Mean: 631 mm |
| 2020 | Mean: 67.53 mm | Mean: 128.77 mm | Mean: 315.82 mm | Mean: 167.59 mm | Mean: 689.97 mm |



3.0 WATER QUALITY

3.1 INTRODUCTION

The following presents the results of water quality monitoring conducted from 2008 to 2019 in the Churchill River Diversion Region. Four waterbodies were monitored in the Churchill River Diversion Region: one on-system annual site (Threepoint Lake); two on-system rotational sites (Footprint and Apussigamasi lakes); and, one off-system annual site (Leftrook Lake; Table 3.1-1 and Figure 3.1-1). Annual sites are sampled each year, whereas rotational sites are sampled once every three years on a rotational basis and are therefore limited to three or four years of data for the 12-year period.

The CAMP water quality program includes four sampling periods (referred to as spring, summer, fall, and winter) per monitoring year (i.e., April-March) typically at a single location within each waterbody or area of a waterbody/river reach. Sampling in the Churchill River Diversion Region was initiated in 2009 and therefore there are only 11 years of monitoring over the 12-year period. During this time, water quality sampling was conducted at each sampling location during each sampling period (i.e., n=44 for annual sites) with two exceptions (Table 3.1-1; Appendix 3-1):

- sampling could not be completed in Threepoint Lake due to slush and thin ice in the winter of 2019 therefore only 10 winter samples were collected over the 12-year period (i.e., n=43); and,
- sampling could not be completed in Apussigamasi Lake due to thin ice in the winter of 2012, therefore only three winter samples were collected over the 12-year period (i.e., n=15).

A detailed description of the program design and sampling methods is provided in Technical Document 1, Section 2.3.

Three indicators (dissolved oxygen [DO]; water clarity; and nutrients/trophic status) were selected for detailed reporting (Table 3.1-2). Metrics for these indicators include DO and its supporting metric temperature/stratification, Secchi disk depth, turbidity, TSS, total phosphorus (TP), total nitrogen (TN), and chlorophyll *a* (Table 3.1-2). A detailed description of these indicators is provided in CAMP (2024).



Table 3.1-1. 2008-2019 Water quality sampling inventory.

| Waterbody/ | Sampling Year ¹ | | | | | | | | | | | |
|------------|----------------------------|------|------|------|-----------------------|------|------|------|------|------|------|------|
| Area | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
| 3PT | | • | • | • | • | • | • | • | • | • | • | ●2 |
| FOOT | | | • | | | • | | | • | | | • |
| APU | | • | | | ● ² | | | • | | | • | |
| LEFT | | • | • | • | • | • | • | • | • | • | • | • |

- 1. Sampling year is from April-March.
- 2. No winter sample collected due to unsafe ice conditions.

Table 3.1-2. Water quality indicators and metrics.

| Indicator | Metric | Units | | | | |
|------------------------------|---|--|--|--|--|--|
| Dissolved Oxygen | Dissolved oxygen (DO) | milligrams per litre (mg/L) and percent (%) saturation | | | | |
| | Temperature/stratification ¹ | °C | | | | |
| | Secchi disk depth | m | | | | |
| Water Clarity | • Turbidity | Nephelometric turbidity units (NTU) | | | | |
| | Total suspended solids (TSS) | mg/L | | | | |
| | Total phosphorus (TP) | mg/L | | | | |
| Nutrients and Trophic Status | Total nitrogen (TN) | mg/L | | | | |
| | • Chlorophyll <i>a</i> | micrograms per litre (μg/L) | | | | |

Notes:

1. Supporting metric.



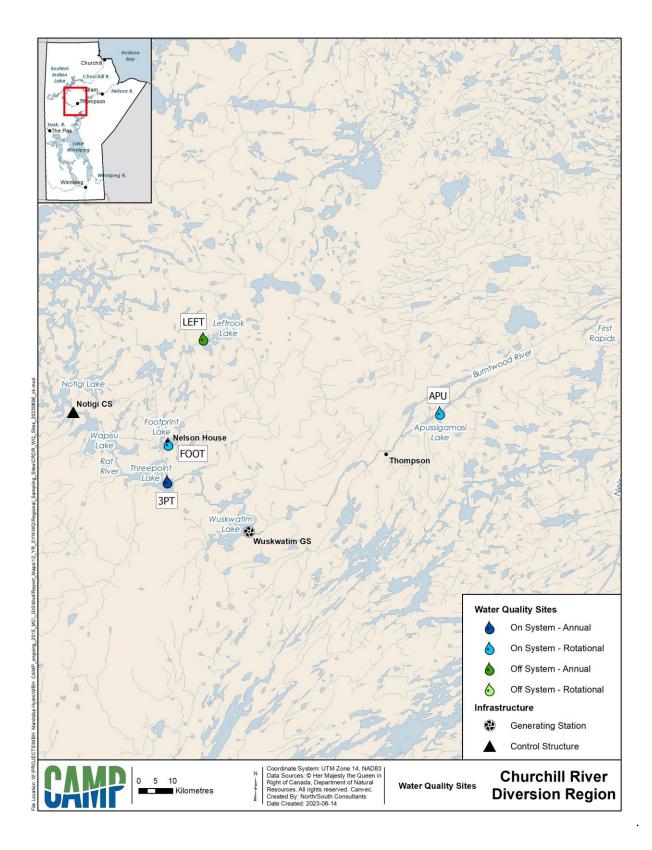


Figure 3.1-1. 2008-2019 Churchill River Diversion Region water quality sites.



3.2 DISSOLVED OXYGEN

3.2.1 DISSOLVED OXYGEN

3.2.1.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Threepoint Lake was well-oxygenated year-round and DO concentrations throughout the water column consistently met the Manitoba Water Quality Standards, Objectives, and Guidelines (MWQSOGs) instantaneous minimum objectives for cool- and cold-water aquatic life during the open-water and ice-cover seasons (Manitoba Water Stewardship [MWS] 2011; Table 3.2-1).

Threepoint Lake was isothermal (i.e., thermal stratification was not observed) and DO concentrations were similar throughout the water column during each sampling period (Table 3.2-1, and Figures 3.2-1 and 3.2-2). During the open-water season, DO concentrations ranged from 8.38 to 12.31 mg/L at the surface and 8.06 to 12.32 mg/L near the bottom (maximum site water depth = 8.0 m). During the ice-cover season, DO concentrations ranged from 15.03 to 16.42 mg/L at the surface and 15.03 to 16.91 mg/L near the bottom (Table 3.2-2 and Figure 3.2-3).

DO concentrations varied between seasons with seasonal mean DO concentrations being higher in winter and spring when the water was cooler, and lower in the summer and fall when the water was warmer (Figure 3.2-4).

DO saturation was near 100% at both the surface and near the bottom during each season sampled (Figure 3.2-5). During the open-water season, surface DO saturation ranged from 91.2 to 115.3% with a mean of 100.5% and a median of 100.6% over the 11 years of monitoring. Mean surface DO saturation levels in the open-water season were similar from year to year ranging from 94.6 to 105.3% and were within or near the interquartile range (IQR) of 95.9 to 102.8%. Bottom DO saturation during the open-water season ranged from 89.5 to 113.0% with a both a mean and median of 99.5% over the 11 years of monitoring. Mean bottom DO saturation levels in the open-water season were similar from year to year ranging from 93.9 to 104.1% and were within or near the IQR of 94.6 to 103.6% (Table 3.2-2 and Figure 3.2-6).



During the ice-cover season, DO saturation at the surface ranged from 105.6 to 121.6% with a mean of 113.2% and a median of 113.7%. The IQR was 107.9 to 117.3%. Bottom DO saturation during the ice-cover season ranged from 105.8 to 119.6% with a mean of 113.6% and a median of 115.5%. The IQR was 108.6 to 118.0% (Table 3.2-2 and Figure 3.2-7).

ROTATIONAL SITES

Footprint Lake

Footprint Lake was well-oxygenated near the surface and DO concentrations near the surface met the MWQSOGs during all sampling periods. DO concentrations decreased with water depth and fell below the MWQSOGs instantaneous minimum objective for cool- and/or cold-water aquatic life in both the open-water and ice-cover seasons in some years.

Footprint Lake was isothermal with the exception of two spring sampling events and one summer sampling event over the four years of monitoring. Specifically, stratification was observed in spring 2010, spring 2013, and summer 2019 (Table 3.2-1 and Figure 3.2-1).

During the open-water season Footprint Lake was well-oxygenated near the surface. Typically DO concentrations were similar throughout the water column; however, DO decreased with water depth during some sampling events (Figure 3.2-2). Specifically, DO concentrations were lower near the bottom than at the surface in summer 2013 and 2019. DO concentrations near bottom met MWQSOGs instantaneous minimum objectives for cool- and cold-water aquatic life (5.0 and 4.0 mg/L, respectively) in summer 2013 but were below both of these objectives in summer 2019 (Table 3.2-1). During the open-water season, DO concentrations ranged from 8.54 to 10.70 mg/L at the surface and 2.93 to 10.48 mg/L near the bottom (maximum site water depth = 14.1 m; Table 3.2-2, and Figure 3.2-8).

During the ice-cover season, Footprint Lake was well-oxygenated near the surface while DO concentrations decreased with water depth and fell below the MWQSOGs instantaneous minimum objectives for cold-water aquatic life near the bottom of the water column in some winters (Figure 3.2-2). Specifically, DO concentrations near the bottom were below the MWQSOGs instantaneous minimum objectives for cold-water aquatic life (8.0 mg/L) in the winters of 2010 and 2019 (Table 3.2-1). In the ice-cover season, DO concentrations ranged from 12.82 to 16.42 mg/L at the surface and 3.60 to 13.33 mg/L near the bottom (Table 3.2-2 and Figure 3.2-8) The decrease in DO concentrations with depth occurred despite the lake being isothermal in winter (Table 3.2-1 and Figure 3.2-1).



During the open-water season, surface DO saturation ranged from 90.2 to 109.7% with a mean of 98.8% and a median of 98.4% over the four years of monitoring. Mean surface DO saturation levels in the open-water season ranged from 96.5 to 101.3% and were within or near the IQR of 95.9 to 100.2%. Bottom DO saturation during the open-water season ranged from 26.7 to 98.4% with a mean of 83.5% and a median of 91.9% over the four years of monitoring. Mean bottom DO saturation levels in the open-water season ranged from 70.2 to 94.4% and were within or near the IQR (88.5 to 94.3%) in two of four years. Mean DO saturation levels near the bottom were below the IQR in 2013 and 2019 (Table 3.2-2 and Figure 3.2-6).

During the ice-cover season, DO saturation at the surface ranged from 89.7 to 122.4% with a mean of 107.4%. Bottom DO saturation during the ice-cover season ranged from 27.0 to 97.2% with a mean of 60.5% (Table 3.2-2 and Figure 3.2-7).

Apussigamasi Lake

Apussigamasi Lake was well-oxygenated year-round and DO concentrations throughout the water column consistently met the MWQSOGs instantaneous minimum objectives for cool- and coldwater aquatic life during the open-water and ice-cover seasons (Table 3.2-1).

Apussigamasi Lake was isothermal and DO concentrations were similar throughout the water column during each sampling period (Table 3.2-1, and Figures 3.2-1 and 3.2-2). During the openwater season, DO concentrations ranged from 9.02 to 11.70 mg/L at the surface and 8.90 to 11.68 mg/L near the bottom (maximum site water depth = 11.6 m). During the ice-cover season, the DO concentration was 15.86 mg/L both near the surface and near the bottom in 2018 (Table 3.2-2 and Figure 3.2-9).

DO saturation in Apussigamasi Lake was near 100% at both the surface and near the bottom of the water column during each season sampled. During the open-water season, surface DO saturation ranged from 99.0 to 112.0% with a mean of 106.5% and a median of 107.5% over the four years of monitoring. Mean surface DO saturation levels in the open-water season were similar from year to year ranging from 101.2 to 109.2% and were within or near the IQR of 102.7 to 109.6%. Bottom DO saturation during the open-water season ranged from 97.5 to 111.7% with a mean of 106.3% and median of 108.6% over the four years of monitoring. Mean bottom DO saturation levels in the open-water season were similar from year to year ranging from 100.2 to 110.0% and were within or near the IQR of 101.8 to 109.6% (Table 3.2-2 and Figure 3.2-6).



During the ice-cover season, DO saturation at the surface was 111.6% both near the surface and near the bottom in 2018 (Table 3.2-2 and Figure 3.2-7).

3.2.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

Leftrook Lake was well-oxygenated near the surface and DO concentrations met the MWQSOGs during all open-water sampling periods. However, DO concentrations decreased with water depth and fell below the MWQSOGs instantaneous minimum objectives for cool- and cold-water aquatic life during some open-water sampling events. DO concentrations below the MWQOSOGs instantaneous minimum objectives for cool- and cold-water aquatic life were observed throughout the water column during the ice-cover season.

Leftrook Lake was thermally stratified during the spring or summer sampling events in eight openwater periods as well as approximately half the ice-cover sampling events over the 11 years of monitoring (Figure 3.2-1). Stratification was observed in two spring sampling events (2010 and 2013), six summer sampling events (2011, 2012, 2014, 2017, 2018, and 2019), and five winter sampling events (2009, 2011, 2012, 2015, and 2016; Table 3.2-3).

During summer of most years, DO concentrations decreased down the water column to levels below one or more of the MWQSOGs instantaneous minimum objectives for cool- and cold-water aquatic life (5.0 and 4.0 mg/L, respectively) at approximately 6-9 m from the surface (Table 3.2-3 and Figure 3.2-2). Specifically, DO concentrations near the bottom were below at least one of these objectives in the summers of 2011, 2012, 2013, 2014, 2016, 2017, 2018 and 2019. Additionally, DO concentrations near the bottom were below both objectives in spring 2010 (Table 3.2-3). During the open-water season, DO concentrations ranged from 7.94 to 11.40 mg/L at the surface and from 0.00 to 10.54 mg/L near the bottom (maximum site water depth = 11.6 m; Table 3.2-4 and Figure 3.2-10).

During the ice-cover season, DO concentrations in Leftrook Lake decreased rapidly and fell below the MWQSOGs instantaneous minimum objectives for cold- and cool-water aquatic life at approximately 0.3-4 m below the ice (Figure 3.2-2). DO concentrations near the surface were below the MWQSOGs instantaneous minimum objective for cold-water aquatic life (8.0 mg/L) in winter of 2017, and DO concentrations near the bottom were below this objective in the winters



of 2009, 2010, 2012, 2016, 2017, 2018, and 2019. Additionally, DO concentrations near the bottom were below the MWQSOGs instantaneous minimum objective for cool-water aquatic life (3.0 mg/L) in the winters of 2009, 2010, 2012, 2017 and 2018 (Table 3.2-3). In the ice-cover season, DO concentrations ranged from 7.80 to 12.82 mg/L at the surface and from 0.00 to 7.54 mg/L near the bottom (Table 3.2-4 and Figure 3.2-10).

Surface DO concentrations varied between seasons with seasonal mean DO concentrations being higher in spring and fall when the water was colder, and lower in the summer when the water was warmer. Despite colder water in winter, DO concentrations in winter were similar to those in spring and fall (Figure 3.2-11).

DO saturation was near 100% at the surface during the open-water season; however, surface DO saturation was lower in winter (mean = 73.6%; Figure 3.2-12). In the open-water season, surface DO saturation ranged from 86.6 to 121.3% with a mean of 98.5% and a median of 97.2% over the 11 years of monitoring. Mean surface DO saturation levels in the open-water season ranged from 89.9 to 103.4% and were within or near the IQR of 93.3 to 101.9%. During the ice-cover season, surface DO saturation ranged from 56.1 to 89.1% with a mean of 73.6% and a median of 74.5%. The IQR for the ice-cover season was 70.0 to 77.9% (Table 3.2-4 and Figure 3.2-13).

Seasonal differences in both DO concentration and percent saturation occurred near the bottom of the water column where, over the 11 years of monitoring, mean DO saturation was lower in summer and winter (30.0 and 23.1%, respectively) than in spring and fall (84.1 and 92.9%, respectively; Figures 3.2-11 and 3.2-12). During the open-water season, bottom DO saturation ranged from 0.0 to 103.0% with a mean of 68.5% and a median of 88.4% over the 11 years of monitoring. Mean bottom DO saturation levels ranged from 60.1 to 88.1% and were within the IQR for the open-water season (32.1 to 93.3%); however, bottom DO saturation in summer tended to be below the IQR for the open-water season. During the ice-cover season, bottom DO saturation ranged from 0.0 to 56.7% with a mean of 23.1% and a median of 19.4%. The IQR for the ice-cover season was 11.5 to 31.5% (Table 3.2-4 and Figure 3.2-13).

ROTATIONAL SITES

There are no off-system rotational sites in this region.



CAMP 12 YEAR DATA REPORT

CHURCHILL RIVER DIVERSION REGION
2024

Table 3.2-1. 2008-2019 On-system sites summary of thermal stratification and DO concentrations.

| | | Surface | 3РТ | | | | FOOT | | | | APU | | | |
|----------------|----------|--------------|-----|------------|-----|-----------|------|------------|-----|-----------|------------|-----|-----|-----------|
| Metric | Sampling | or Bottom | | Open-Water | | Ice-Cover | | Open-Water | | Ice-Cover | Open-Water | | | Ice-Cover |
| | Year | | SP | SU | FA | WI | SP | SU | FA | WI | SP | su | FA | WI |
| | 2008 | | | | | | | | | | | | | |
| | 2009 | | No | No | No | No | | | | | No | No | No | No |
| | 2010 | | No | No | No | No | 2010 | No | No | No | | | | |
| | 2011 | | No | No | No | No | | | | | | | | |
| | 2012 | | No | No | No | No | | | | | No | No | No | ND |
| Thermal | 2013 | | No | No | No | No | 2013 | No | No | No | | | | |
| Stratification | 2014 | | No | No | No | No | | | | | | | | |
| | 2015 | | No | No | No | No | | | | | No | No | No | No |
| | 2016 | | No | No | No | No | No | No | No | No | | | | |
| | 2017 | | No | No | No | No | | | | | | | | |
| | 2018 | | No | No | No | No | | | | | No | No | No | No |
| | 2019 | | No | No | No | ND | No | 2019 | No | No | | | | |
| | | Surface | | | | | | | | | | | | |
| | 2008 | Bottom | | | | | | | | | | | | |
| | | Surface | ND | ND | ND | Yes | | | | | ND | ND | ND | ND |
| | 2009 | Bottom | ND | ND | ND | Yes | | | | | ND | ND | ND | ND |
| | | Surface | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | | | | |
| | 2010 | Bottom | Yes | Yes | Yes | Yes | Yes | Yes | Yes | 2010 | | | | |
| | | Surface | Yes | Yes | Yes | ND | | | | | | | | |
| | 2011 | Bottom | Yes | Yes | Yes | ND | | | | | | | | |
| | | Surface | Yes | Yes | Yes | Yes | | | | | Yes | Yes | Yes | ND |
| | 2012 | Bottom | Yes | Yes | Yes | Yes | | | | | Yes | Yes | Yes | ND |
| | | Surface | Yes | Yes | Yes | ND | Yes | Yes | Yes | ND | | | | |
| DO met MWQSOGs | 2013 | Bottom | Yes | Yes | Yes | ND | ND | Yes | Yes | ND | | | | |
| PAL objectives | | Surface | Yes | Yes | Yes | ND | | | | | | | | |
| | 2014 | Bottom | Yes | Yes | Yes | ND | | | | | | | | |
| | | Surface | Yes | Yes | Yes | ND | | | | | Yes | Yes | Yes | ND |
| | 2015 | Bottom | Yes | Yes | Yes | ND | | | | | Yes | Yes | Yes | ND |
| | | Surface | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | | | | |
| | 2016 | Bottom | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | | | | |
| | | Surface | Yes | Yes | Yes | Yes | | | | | | | | |
| | 2017 | Bottom | Yes | Yes | Yes | Yes | | | | | | | | |
| | | Surface | Yes | Yes | Yes | Yes | | | | | Yes | Yes | Yes | Yes |
| | 2018 | Bottom | Yes | Yes | Yes | Yes | | | | | Yes | Yes | Yes | Yes |
| | 2.5 | Surface | Yes | Yes | Yes | ND | Yes | Yes | Yes | Yes | | | | |
| | 2019 | Bottom | Yes | Yes | Yes | ND | Yes | 2019 | Yes | 2019 | | | | |

- 1. SP = spring; SU = summer; FA = fall; WI = winter.
- 2 ND = No data
- 3. MWQSOGs = Manitoba Water Quality Standards, Objectives, and Guidelines; PAL = Protection of aquatic life.
- 4. DO concentrations were compared to the most stringent MWQSOGs instantaneous minimum PAL objectives for each season; i.e., 5 mg/L for cool-water early life for the open-water season and 8 mg/L for cold-water early life the ice-cover season.
- 5. Cells with a year indicated denote instances of stratification or non-compliance with MWQSOGs instantaneous minimum objectives.
- 6. _____ = Sampling did not occur.



Table 3.2-2. 2008-2019 On-system sites DO, water depth, and ice thickness summary statistics.

| | | | | | Dissolv | ved Oxygen | | | | Wa | iter | Ice |
|------|----------------|------------------------|-------|-------|-----------------------|------------|--------------------|------------------|-------------------|----------------------|------|--------------------------|
| Site | Statistic | DO – Surface (mg/L) | | | DO – Bottom (mg/L) | | ration - ce (%) | DO Satu Botto | ration - m (%) | Depth at Site (m) | | Thickness at Site (m) |
| | | ow | IC | ow | IC | ow | IC | ow | IC | ow | IC | IC |
| | Mean | 9.82 | 15.70 | 9.79 | 15.73 | 100.5 | 113.2 | 99.5 | 113.6 | 5.3 | 5.8 | 0.71 |
| | Median | 9.63 | 15.72 | 9.60 | 15.55 | 100.6 | 113.7 | 99.5 | 115.5 | 5.5 | 5.7 | 0.70 |
| | Minimum | 8.38 | 15.03 | 8.06 | 15.03 | 91.2 | 105.6 | 89.5 | 105.8 | 3.2 | 4.9 | 0.58 |
| | Maximum | 12.31 | 16.42 | 12.32 | 16.91 | 115.3 | 121.6 | 113.0 | 119.6 | 8.0 | 6.5 | 0.90 |
| 207 | SD | 0.972 | 0.576 | 1.03 | 0.731 | 6.00 | 6.43 | 5.94 | 5.95 | 0.96 | 0.56 | 0.095 |
| 3PT | SE | 0.177 | 0.235 | 0.188 | 0.298 | 1.10 | 2.63 | 1.08 | 2.43 | 0.17 | 0.18 | 0.030 |
| | Lower Quartile | 9.17 | 15.21 | 9.10 | 15.18 | 95.9 | 107.9 | 94.6 | 108.6 | 4.5 | 5.4 | 0.66 |
| | Upper Quartile | 10.29 | 16.12 | 10.47 | 16.10 | 102.8 | 117.3 | 103.6 | 118.0 | 5.9 | 6.2 | 0.77 |
| | n | 30 | 6 | 30 | 6 | 30 | 6 | 30 | 6 | 33 | 10 | 10 |
| | % Detections | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | - | - | - |
| | Mean | 9.67 | 14.95 | 8.46 | 8.06 | 98.8 | 107.4 | 83.5 | 60.5 | 11.0 | 11.0 | 0.80 |
| | Median | 9.93 | - | 9.04 | - | 98.4 | - | 91.9 | - | 10.9 | - | - |
| | Minimum | 8.54 | 12.82 | 2.93 | 3.60 | 90.2 | 89.7 | 26.7 | 27.0 | 7.9 | 3.6 | 0.68 |
| | Maximum | 10.70 | 16.42 | 10.48 | 13.33 | 109.7 | 122.4 | 98.4 | 97.2 | 13.5 | 14.1 | 1.02 |
| FOOT | SD | 0.713 | 1.89 | 2.28 | 4.92 | 4.92 | 16.5 | 22.4 | 35.2 | 2.04 | 4.94 | 0.156 |
| FOOT | SE | 0.206 | 1.09 | 0.689 | 2.84 | 1.42 | 9.54 | 6.75 | 20.3 | 0.59 | 2.47 | 0.078 |
| | Lower Quartile | 9.09 | - | 8.78 | - | 95.9 | - | 88.5 | - | 9.0 | - | - |
| | Upper Quartile | 10.16 | - | 9.68 | - | 100.2 | - | 94.3 | - | 13.1 | - | - |
| | n | 12 | 3 | 11 | 3 | 12 | 3 | 11 | 3 | 12 | 4 | 4 |
| | % Detections | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | - | - | - |



Table 3.2-2. continued.

| | | | Dissolved Oxygen | | | | | | | | | Ice | |
|------|----------------|-------|------------------------|-------|-----------------------|-------|--------------------------------|-------|-------------------------------|------|--------------|--------------------------|--|
| Site | Statistic | | DO – Surface (mg/L) | | DO – Bottom (mg/L) | | DO Saturation - Surface (%) | | DO Saturation - Bottom (%) | | pth e (m) | Thickness at Site (m) | |
| | | ow | IC | ow | IC | ow | IC | ow | IC | ow | IC | IC | |
| | Mean | 10.51 | 15.86 | 10.51 | 15.86 | 106.5 | 111.6 | 106.3 | 111.6 | 8.7 | 10.9 | 0.70 | |
| | Median | 10.63 | - | 10.77 | - | 107.5 | - | 108.6 | - | 8.7 | - | - | |
| | Minimum | 9.02 | 15.86 | 8.90 | 15.86 | 99.0 | 111.6 | 97.5 | 111.6 | 6.8 | 10.4 | 0.65 | |
| | Maximum | 11.70 | 15.86 | 11.68 | 15.86 | 112.0 | 111.6 | 111.7 | 111.6 | 10.1 | 11.6 | 0.73 | |
| ADII | SD | 0.787 | - | 0.835 | - | 4.33 | - | 4.94 | - | 0.94 | 0.61 | 0.042 | |
| APU | SE | 0.262 | - | 0.278 | - | 1.44 | - | 1.65 | - | 0.27 | 0.35 | 0.024 | |
| | Lower Quartile | 10.11 | - | 10.03 | - | 102.7 | - | 101.8 | - | 8.2 | - | - | |
| | Upper Quartile | 11.02 | - | 11.00 | - | 109.6 | - | 109.6 | - | 9.5 | - | - | |
| | n | 9 | 1 | 9 | 1 | 9 | 1 | 9 | 1 | 12 | 3 | 3 | |
| | % Detections | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | - | - | - | |



^{1.} OW = Open-water season; IC = Ice-cover season.

^{2.} SD = standard deviation; SE = standard error; n = number of samples.

Table 3.2-3. 2008-2019 Off-system sites summary of thermal stratification and DO concentrations.

| | | Surface | LEFT | | | | | | |
|--------------------|------------------|---------|------|------------|-----|-----------|--|--|--|
| Metric | Sampling Year | or | | Open-Water | | Ice-Cover | | | |
| | Teal | Bottom | SP | SU | FA | WI | | | |
| | 2008 | | | | | | | | |
| | 2009 | | No | No | No | 2009 | | | |
| | 2010 | | 2010 | No | No | No | | | |
| | 2011 | | No | 2011 | No | 2011 | | | |
| | 2012 | | No | 2012 | No | 2012 | | | |
| Thermal | 2013 | | 2013 | No | No | No | | | |
| Stratification | 2014 | | No | 2014 | No | No | | | |
| | 2015 | | No | No | No | 2015 | | | |
| | 2016 | | No | No | No | 2016 | | | |
| | 2017 | | No | 2017 | No | No | | | |
| | 2018 | | No | 2018 | No | No | | | |
| | 2019 | | No | 2019 | No | No | | | |
| | 2000 | Surface | | | | | | | |
| | 2008 | Bottom | | | | | | | |
| | 2009 | Surface | ND | ND | ND | Yes | | | |
| | | Bottom | ND | ND | ND | 2009 | | | |
| | 2010 | Surface | Yes | Yes | Yes | Yes | | | |
| | 2010 | Bottom | 2010 | Yes | Yes | 2010 | | | |
| | 2011 | Surface | Yes | Yes | Yes | ND | | | |
| | 2011 | Bottom | Yes | 2011 | Yes | ND | | | |
| | 2012 | Surface | Yes | Yes | Yes | Yes | | | |
| | 2012 | Bottom | Yes | 2012 | Yes | 2012 | | | |
| | 2013 | Surface | Yes | Yes | Yes | ND | | | |
| DO met MWQSOGs PAL | 2013 | Bottom | ND | 2013 | Yes | ND | | | |
| objectives | 2014 | Surface | Yes | Yes | Yes | ND | | | |
| | 2014 | Bottom | Yes | 2014 | Yes | ND | | | |
| | 2015 | Surface | Yes | Yes | Yes | ND | | | |
| | 2013 | Bottom | Yes | Yes | Yes | ND | | | |
| | 2016 | Surface | Yes | Yes | Yes | Yes | | | |
| | 2016 | Bottom | Yes | 2016 | Yes | 2016 | | | |
| | 2017 | Surface | Yes | Yes | Yes | 2017 | | | |
| | 2017 | Bottom | Yes | 2017 | Yes | 2017 | | | |
| | 2018 | Surface | Yes | Yes | Yes | Yes | | | |
| | 2010 | Bottom | Yes | 2018 | Yes | 2018 | | | |
| | 2019 | Surface | Yes | Yes | Yes | Yes | | | |
| | 2019 | Bottom | Yes | 2019 | Yes | 2019 | | | |

- 1. SP = spring; SU = summer; FA = fall; WI = winter; DO = dissolved oxygen; MWQSOG = Manitoba Water Quality Standards, Objectives, and Guidelines; PAL = Protection of Aquatic Life.
- 2. ND = No data
- 3. DO concentrations were compared to the most stringent MWQSOGs instantaneous minimum PAL objectives for each season; i.e., 5 mg/L for cool-water early life for the open-water season and 8 mg/L for cold-water early life the ice-cover season.
- 4. Cells with a year indicated denote instances of stratification or non-compliance with MWQSOGs instantaneous minimum objectives.
- 5. Sampling did not occur.



Table 3.2-4. 2008-2019 Off-system sites DO, water depth, and ice thickness summary statistics.

| | | | | | Dissolved | l Oxygen (D | O) | | | Water | | Ice |
|------|----------------|------------------------|-------|-----------------------|-----------|--------------------------------|------|-------------------------------|------|----------------------|------|--------------------------|
| Site | Statistic | DO – Surface (mg/L) | | DO – Bottom (mg/L) | | DO Saturation - Surface (%) | | DO Saturation - Bottom (%) | | Depth at Site (m) | | Thickness at Site (m) |
| | | ow | IC | ow | IC | ow | IC | ow | IC | ow | IC | IC |
| | Mean | 9.47 | 10.16 | 6.84 | 3.01 | 98.5 | 73.6 | 68.5 | 23.1 | 9.2 | 8.9 | 0.75 |
| | Median | 9.31 | 9.95 | 8.76 | 2.42 | 97.2 | 74.5 | 88.4 | 19.4 | 9.7 | 9.9 | 0.76 |
| | Minimum | 7.94 | 7.80 | 0.00 | 0.00 | 86.6 | 56.1 | 0.0 | 0.0 | 4.1 | 3.9 | 0.63 |
| | Maximum | 11.40 | 12.82 | 10.54 | 7.54 | 121.3 | 89.1 | 103.0 | 56.7 | 11.6 | 11.0 | 0.85 |
| | SD | 0.870 | 1.59 | 3.73 | 2.71 | 7.75 | 10.1 | 37.0 | 20.2 | 1.81 | 2.22 | 0.067 |
| LEFT | SE | 0.159 | 0.60 | 0.692 | 1.02 | 1.41 | 3.81 | 6.86 | 7.6 | 0.31 | 0.67 | 0.020 |
| | Lower Quartile | 8.84 | 9.37 | 2.84 | 1.44 | 93.3 | 70.0 | 32.1 | 11.5 | 8.6 | 8.8 | 0.71 |
| | Upper Quartile | 10.07 | 10.92 | 9.54 | 4.11 | 101.9 | 77.9 | 93.3 | 31.5 | 10.5 | 10.0 | 0.78 |
| | n | 30 | 7 | 29 | 7 | 30 | 7 | 29 | 7 | 33 | 11 | 11 |
| | % Detections | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | - | - | - |



^{1.} OW = Open-water season; IC = Ice-cover season.

^{2.} SD = standard deviation; SE = standard error; n = number of samples.

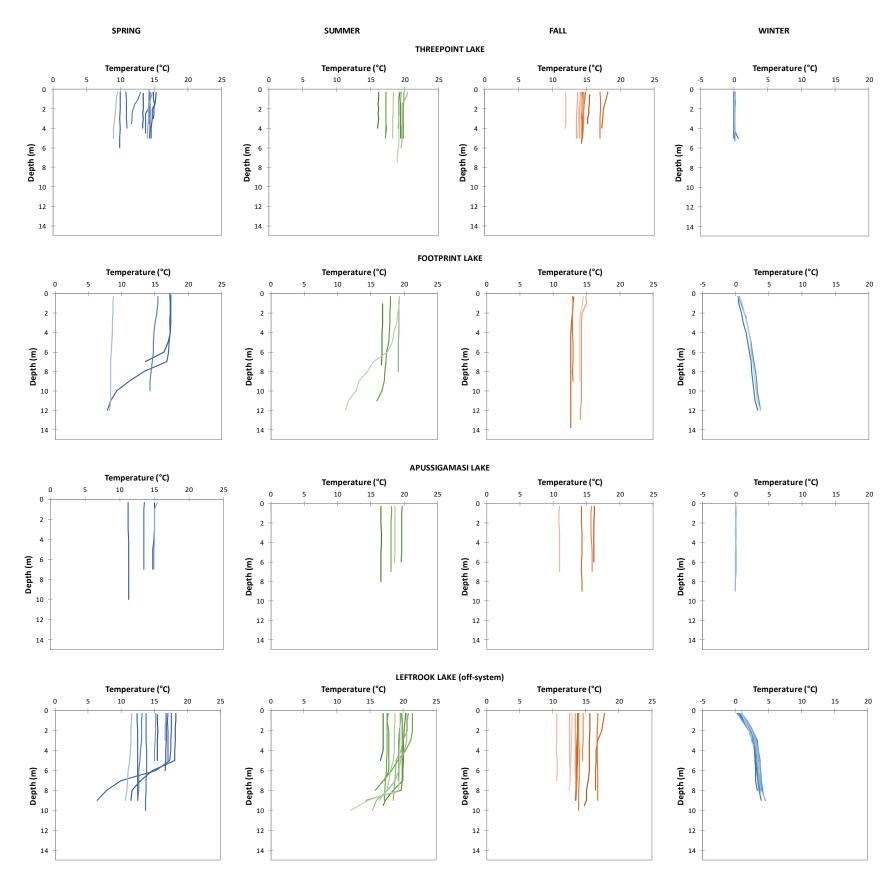


Figure 3.2-1. 2008-2019 On-system and off-system water temperature depth profiles.



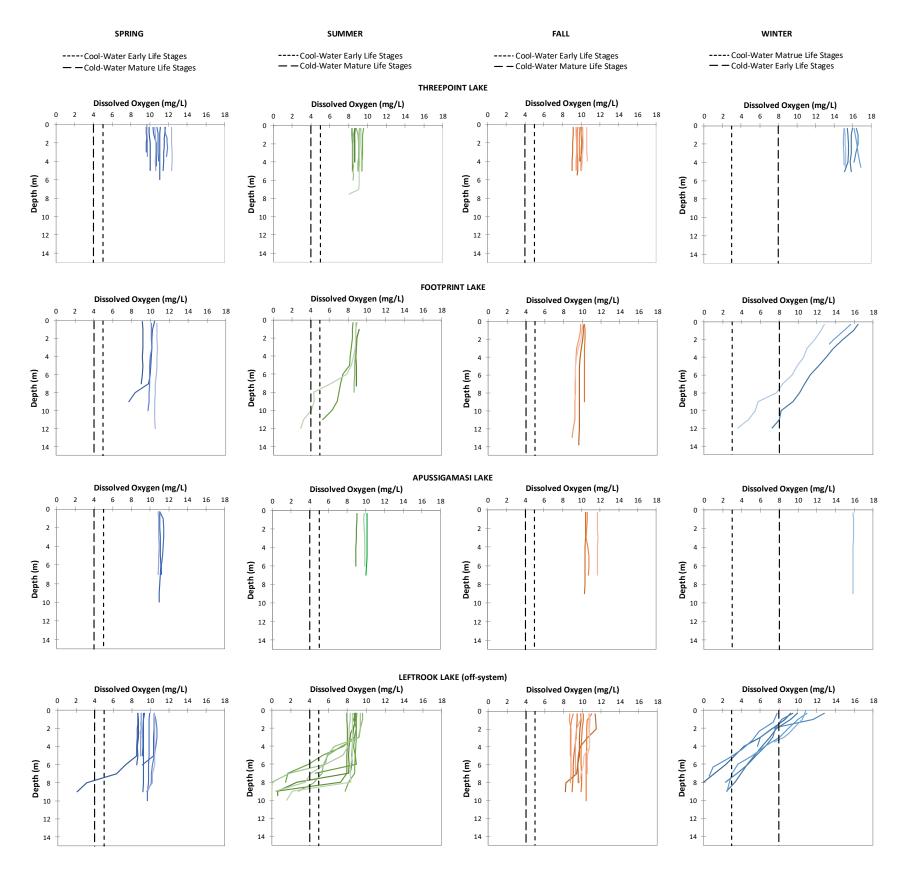


Figure 3.2-2. 2008-2019 On-system and off-system dissolved oxygen depth profiles and comparison to instantaneous minimum objectives for the protection of aquatic life.



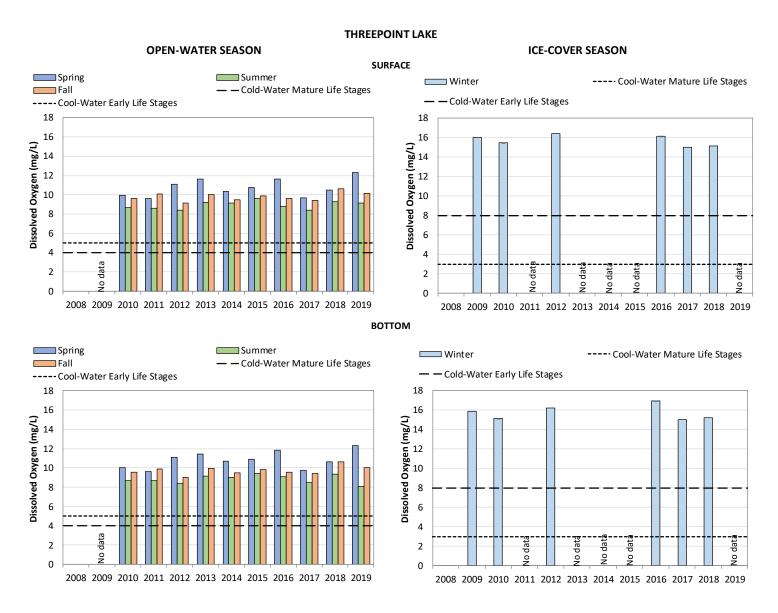
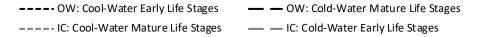
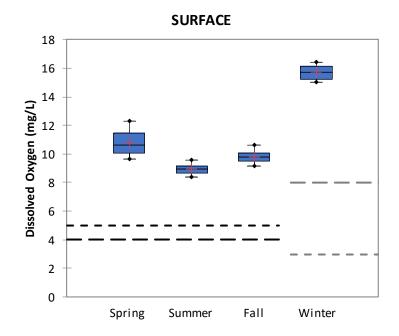


Figure 3.2-3. 2008-2019 Threepoint Lake surface and bottom dissolved oxygen concentrations with comparison to instantaneous minimum objectives for the protection of aquatic life.



THREEPOINT LAKE





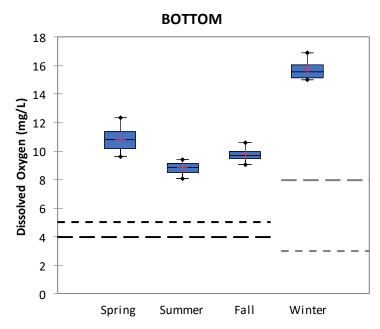


Figure 3.2-4. 2008-2019 On-system seasonal surface and bottom dissolved oxygen concentrations with comparison to instantaneous minimum objectives for the protection of aquatic life.



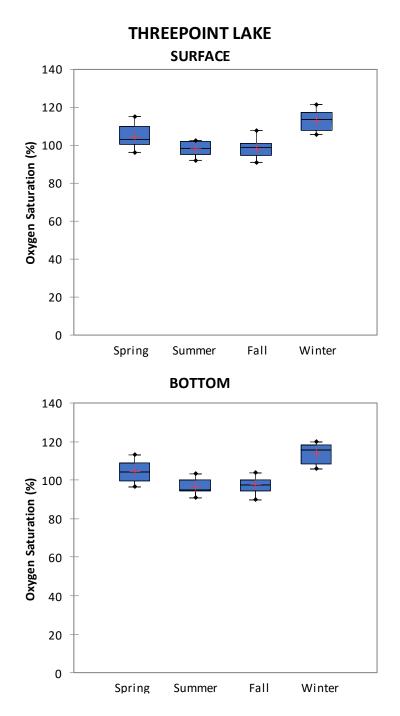


Figure 3.2-5. 2008-2019 On-system seasonal surface and bottom dissolved oxygen saturation.



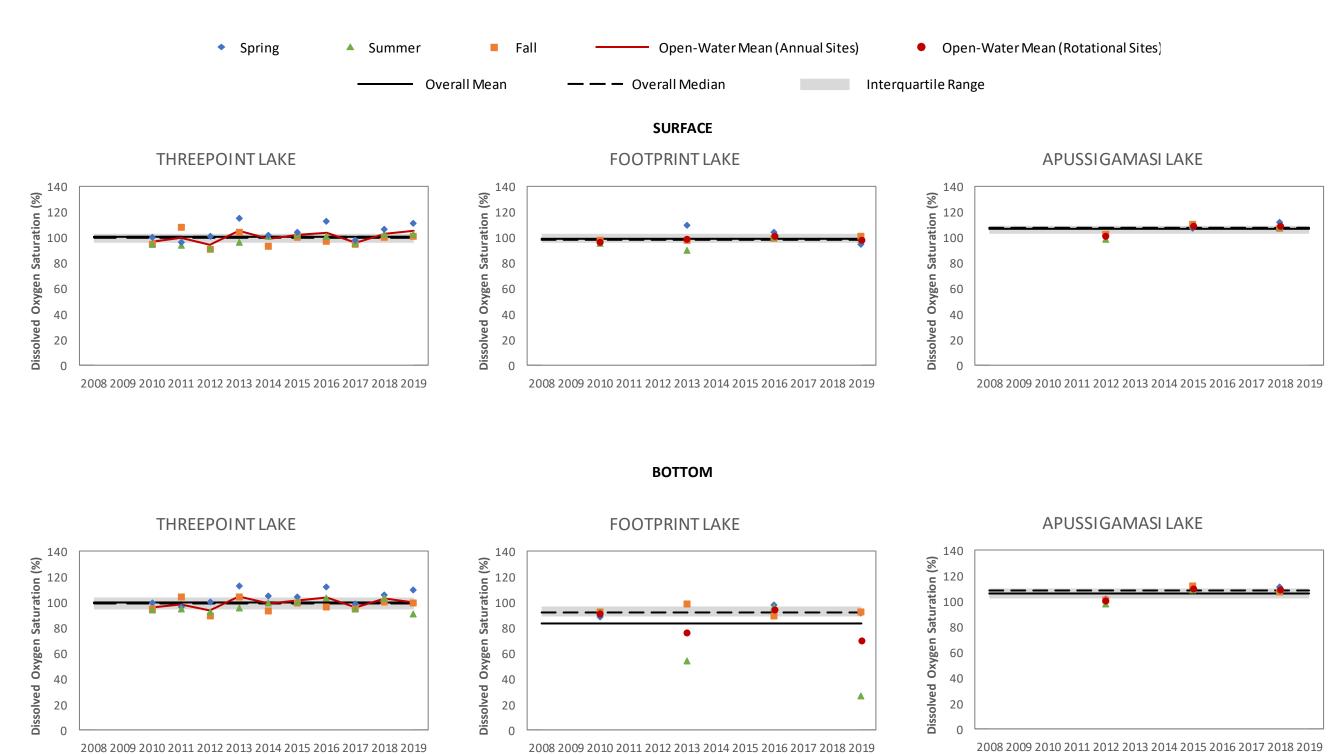


Figure 3.2-6. 2008-2019 On-system open-water season surface and bottom dissolved oxygen saturation.



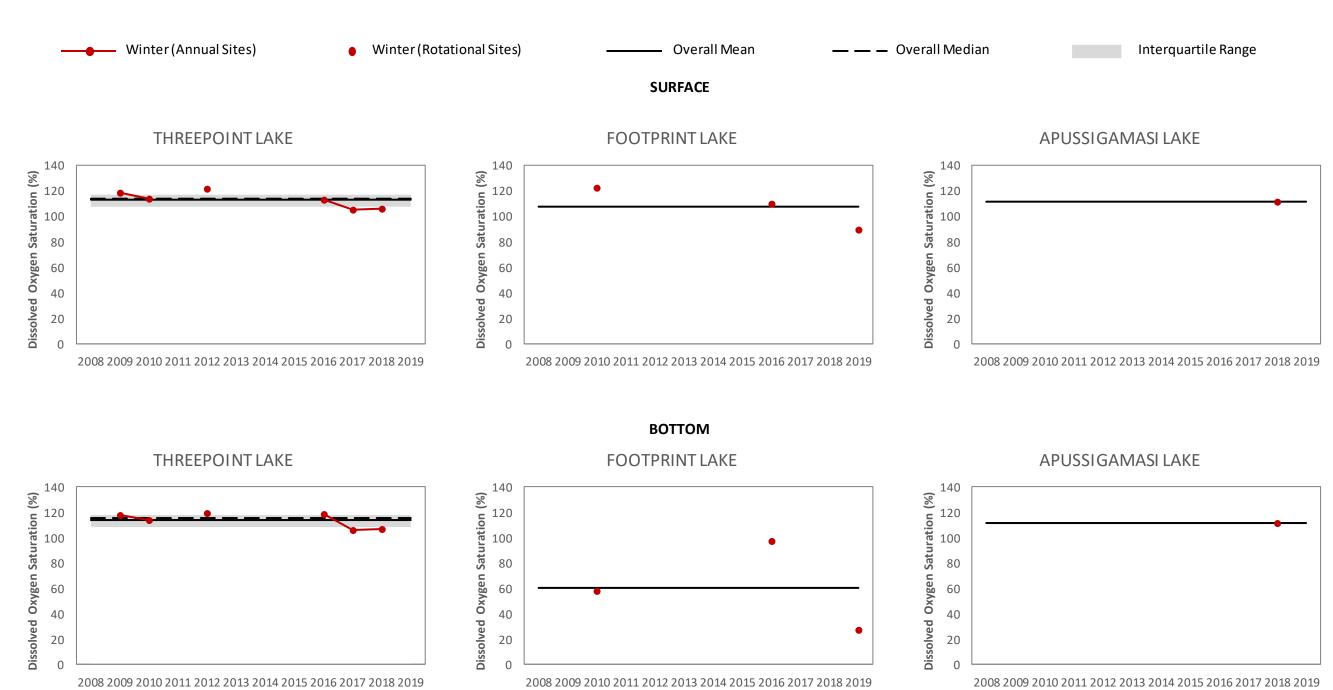


Figure 3.2-7. 2008-2019 On-system ice-cover season surface and bottom dissolved oxygen saturation.



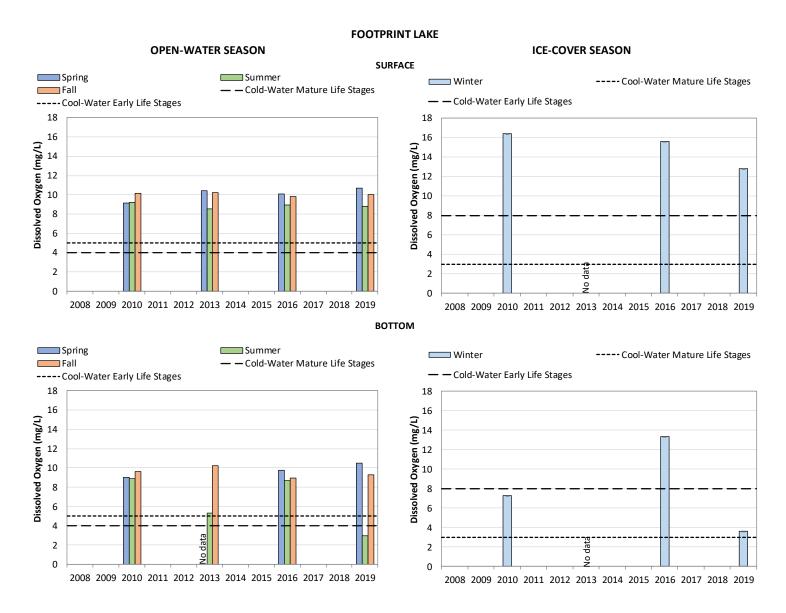


Figure 3.2-8. 2008-2019 Footprint Lake surface and bottom dissolved oxygen concentrations with comparison to instantaneous minimum objectives for the protection of aquatic life.



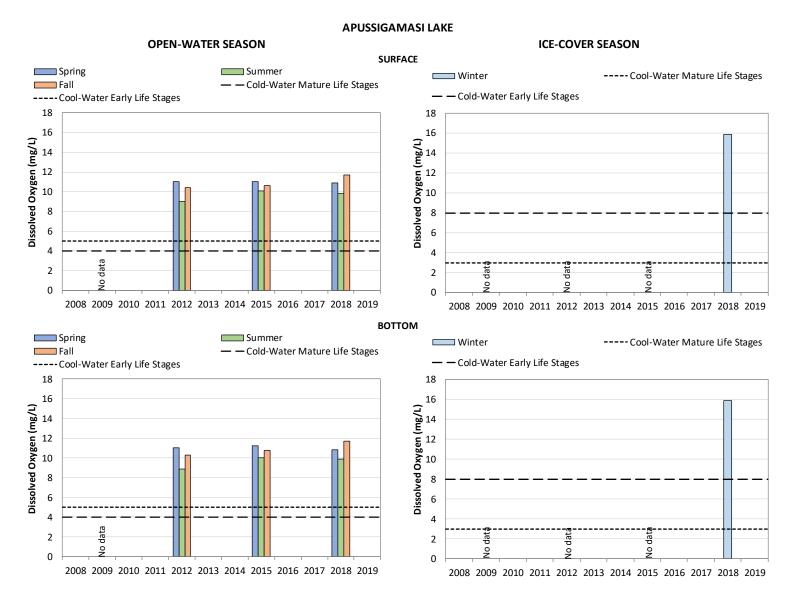


Figure 3.2-9. 2008-2019 Apussigamasi Lake surface and bottom dissolved oxygen concentrations with comparison to instantaneous minimum objectives for the protection of aquatic life.



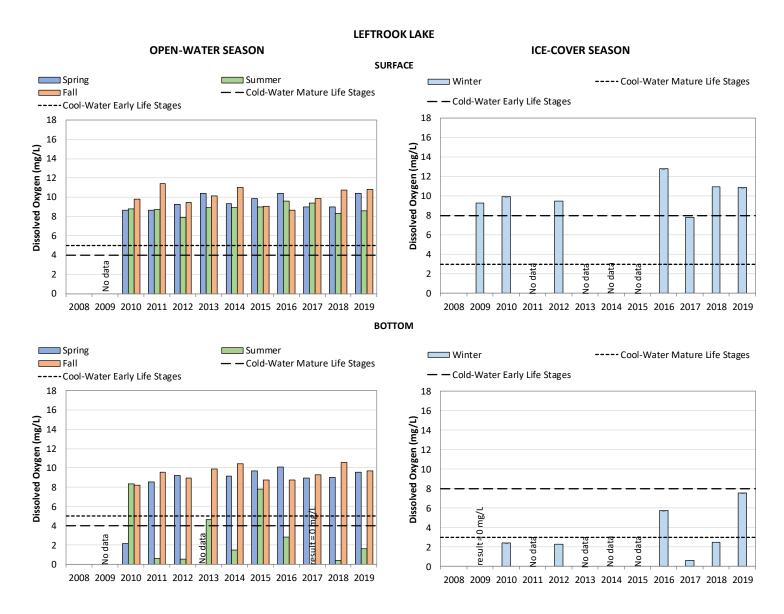
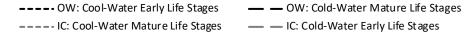
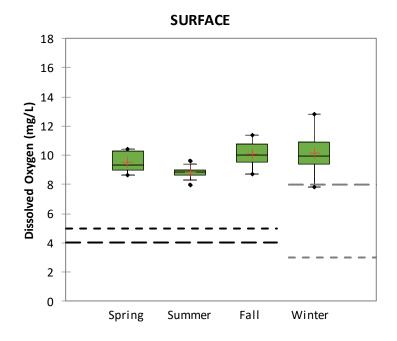


Figure 3.2-10. 2008-2019 Leftrook Lake surface and bottom dissolved oxygen concentrations with comparison to instantaneous minimum objectives for the protection of aquatic life.



LEFTROOK LAKE





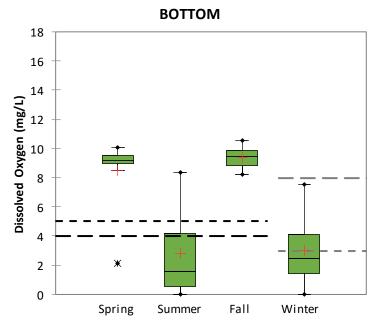
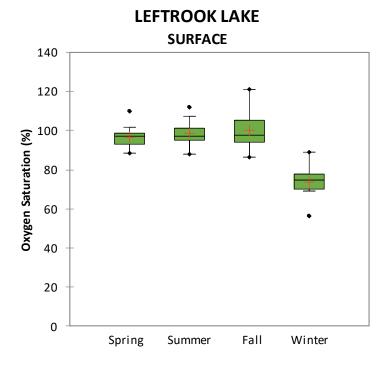


Figure 3.2-11. 2008-2019 Off-system seasonal surface and bottom dissolved oxygen concentrations with comparison to instantaneous minimum objectives for the protection of aquatic life.





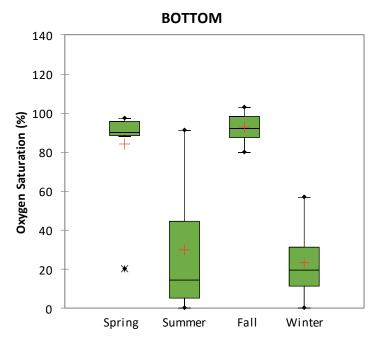


Figure 3.2-12. 2008-2019 Off-system seasonal surface and bottom dissolved oxygen saturation.



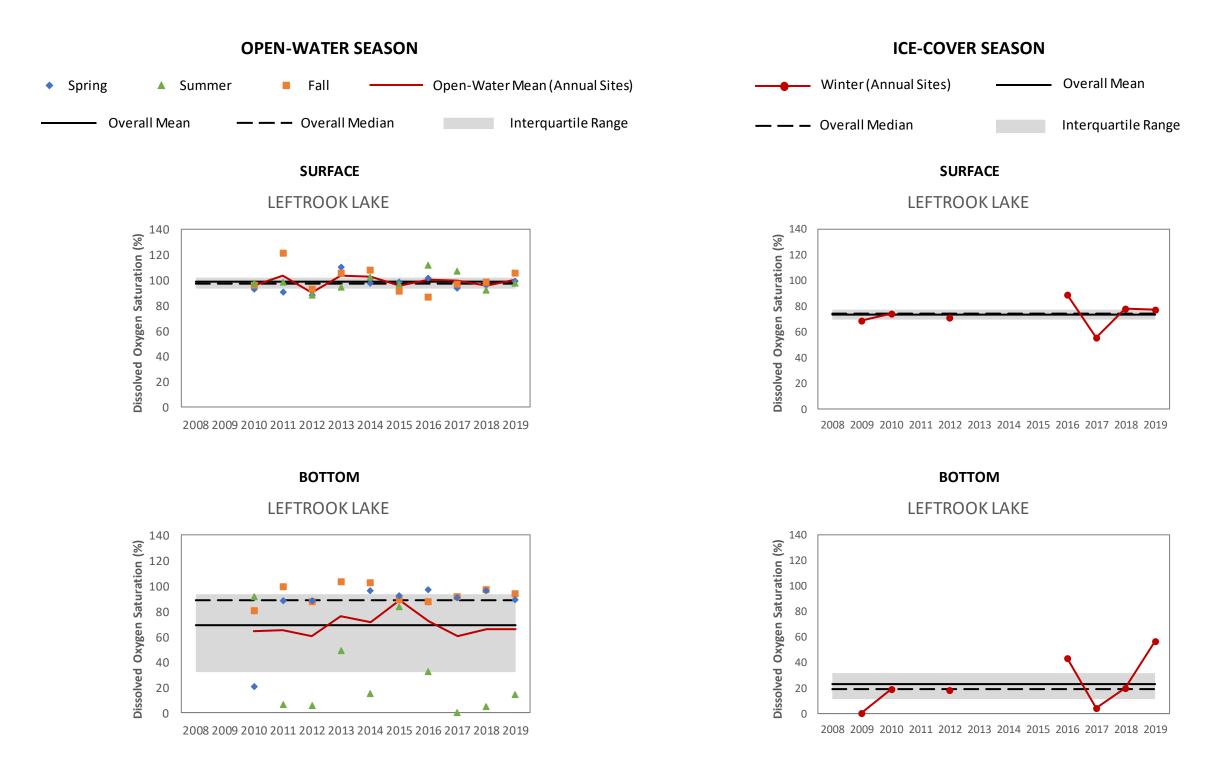


Figure 3.2-13. 2008-2019 Off-system open-water and ice-cover season surface and bottom dissolved oxygen saturation.



3.3 WATER CLARITY

3.3.1 SECCHI DISK DEPTH

3.3.1.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Secchi disk depth in Threepoint Lake ranged from 0.30 to 1.00 m during the open-water season. The mean and median measurements for the 11 years of monitoring were 0.55 and 0.53 m, respectively. Mean annual Secchi disk depths ranged from 0.43 to 0.70 m and were within the IQR (0.45 to 0.59 m) in seven of the 11 years of monitoring. Mean Secchi disk depths were below the IQR in 2014 and above the IQR in 2009, 2012, and 2019 (Table 3.3-1 and Figure 3.3-1).

No clear seasonality was observed for Secchi disk depth in Threepoint Lake over the 11 years of monitoring. However, the mean Secchi disk depth was highest in spring (0.60 m) and lowest in summer and fall (0.52 m for both; Figure 3.3-2).

ROTATIONAL SITES

Footprint Lake

Secchi disk depth in Footprint Lake ranged from 0.65 to 1.90 m during the open-water season. The mean was 1.16 m, the median was 1.05 m, and the IQR was 0.97 to 1.22 m for the four years of monitoring. Mean annual Secchi disk depths ranged from 0.88 to 1.45 m and were within the IQR in 2010, below the IQR in 2016, and above the IQR in 2013 and 2019 (Table 3.3-1 and Figure 3.3-1).

Apussigamasi Lake

Secchi disk depth in Apussigamasi Lake ranged from 0.27 to 0.48 m during the open-water season. The mean and median were both 0.34 m and the IQR was 0.30 to 0.35 m for the four years of monitoring. Mean annual Secchi disk depths ranged from 0.31 to 0.39 m and were within the IQR in 2012, 2015, and 2018 but above the IQR in 2009 (Table 3.3-1 and Figure 3.3-1).



3.3.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

Secchi disk depth in Leftrook Lake ranged from 0.80 to 2.80 m during the open-water season. The mean and median measurements for the 11 years of monitoring were 1.63 and 1.55 m, respectively. Mean annual Secchi disk depths ranged from 1.21 to 2.03 m and were within the IQR (1.25 to 1.95 m) in nine of the 11 years of monitoring. Mean Secchi disk depths were below the IQR in 2013 and above the IQR in 2019 (Table 3.3-2 and Figure 3.3-3).

On average, Secchi disk depths were higher in the spring (2.13 m) than in the summer and fall (1.43 and 1.34 m, respectively; Figure 3.3-4).

ROTATIONAL SITES

There are no off-system rotational sites in this region.



Table 3.3-1. 2008-2019 On-system sites water clarity summary statistics.

| C:Lo | Chatiatia | Secchi Disk | Depth (m) | Turbidit | ty (NTU) | TSS (mg/L) | | |
|------|----------------|-------------|-----------|----------|----------|------------|-------|--|
| Site | Statistic | ow | IC | ow | IC | ow | IC | |
| | Mean | 0.55 | - | 17.9 | 10.5 | 7.5 | <2.0 | |
| | Median | 0.53 | - | 18.2 | 10.5 | 7.0 | <2.0 | |
| | Minimum | 0.30 | - | 8.22 | 7.52 | 2.9 | <2.0 | |
| | Maximum | 1.00 | - | 41.5 | 13.0 | 18.6 | 3.5 | |
| | SD | 0.137 | - | 5.42 | 1.82 | 3.03 | 1.04 | |
| 3PT | SE | 0.024 | - | 0.943 | 0.576 | 0.527 | 0.330 | |
| | Lower Quartile | 0.45 | - | 16.2 | 9.09 | 5.6 | <2.0 | |
| | Upper Quartile | 0.59 | - | 19.7 | 11.8 | 8.6 | 2.8 | |
| | n | 33 | - | 33 | 10 | 33 | 10 | |
| | % Detections | 100 | - | 100 | 100 | 100 | 50 | |
| | Mean | 1.16 | - | 7.33 | 11.4 | 5.1 | <2.0 | |
| | Median | 1.05 | - | 7.36 | - | 4.9 | - | |
| | Minimum | 0.65 | - | 5.30 | 6.98 | 2.9 | <2.0 | |
| | Maximum | 1.90 | - | 11.5 | 15.1 | 7.0 | 2.4 | |
| | SD | 0.368 | - | 1.80 | 3.45 | 1.27 | 0.712 | |
| FOOT | SE | 0.106 | - | 0.520 | 1.72 | 0.368 | 0.356 | |
| | Lower Quartile | 0.97 | - | 5.75 | - | 4.5 | - | |
| | Upper Quartile | 1.22 | - | 8.08 | - | 5.8 | - | |
| | n | 12 | - | 12 | 4 | 12 | 4 | |
| | % Detections | 100 | - | 100 | 100 | 100 | 50 | |
| | Mean | 0.34 | - | 29.5 | 16.4 | 13.6 | 8.2 | |
| | Median | 0.34 | - | 30.8 | - | 12.7 | - | |
| | Minimum | 0.27 | - | 21.0 | 14.1 | 7.6 | 6.5 | |
| | Maximum | 0.48 | ı | 35.0 | 19.4 | 24.4 | 9.6 | |
| | SD | 0.052 | ı | 4.47 | 2.69 | 5.23 | 1.55 | |
| APU | SE | 0.015 | - | 1.29 | 1.55 | 1.51 | 0.897 | |
| | Lower Quartile | 0.30 | - | 27.1 | - | 9.4 | - | |
| | Upper Quartile | 0.35 | - | 32.1 | - | 15.5 | - | |
| | n | 12 | - | 12 | 3 | 12 | 3 | |
| | % Detections | 100 | - | 100 | 100 | 100 | 100 | |



^{1.} OW = Open-water season; IC = Ice-cover season.

^{2.} SD = standard deviation; SE = standard error; n = number of samples.

Table 3.3-2. 2008-2019 Off-system sites water clarity metric summary statistics.

| Site | Statistic | Secchi Disk | Depth (m) | Turbidit | y (NTU) | TSS (mg/L) | | |
|------|----------------|-------------|-----------|----------|---------|------------|------|--|
| Site | Statistic | ow | IC | ow | IC | ow | IC | |
| | Mean | 1.63 | - | 3.25 | 0.90 | 4.3 | <2.0 | |
| | Median | 1.55 | - | 2.58 | 0.85 | 3.9 | <2.0 | |
| | Minimum | 0.80 | - | 1.00 | 0.42 | <2.0 | <2.0 | |
| | Maximum | 2.80 | - | 6.55 | 1.53 | 16.8 | <2.0 | |
| | SD | 0.574 | - | 1.55 | 0.357 | 3.15 | - | |
| LEFT | SE | 0.100 | - | 0.274 | 0.108 | 0.548 | - | |
| | Lower Quartile | 1.25 | - | 2.06 | 0.71 | 2.4 | <2.0 | |
| | Upper Quartile | 1.95 | - | 4.38 | 1.12 | 4.8 | <2.0 | |
| | n | 33 | - | 32 | 11 | 33 | 11 | |
| | % Detections | 100 | - | 100 | 100 | 82 | 0 | |

- 1. OW = Open-water season; IC = Ice-cover season.
- 2. SD = standard deviation; SE = standard error; n = number of samples.
- 3. Turbidity statistics for LEFT exclude suspect value of 19.5 NTU from summer 2011.



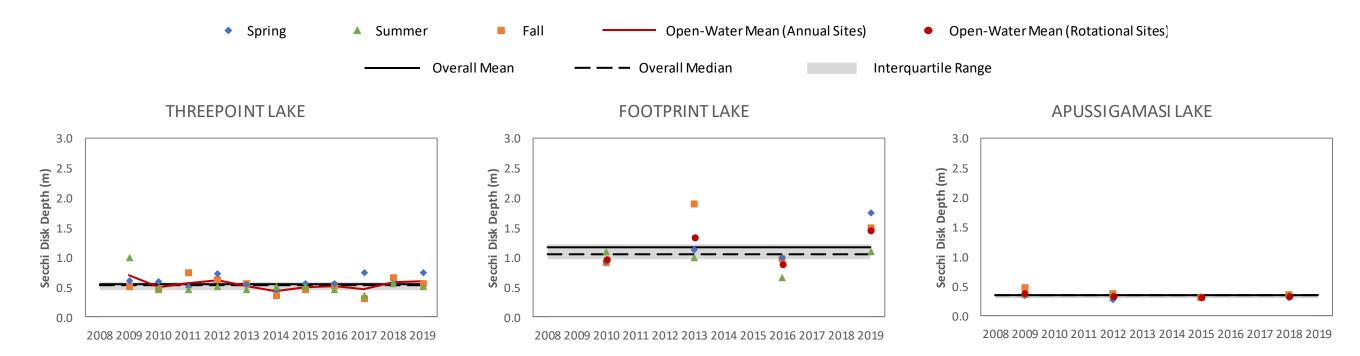


Figure 3.3-1. 2008-2019 On-system open-water season Secchi disk depths.



THREEPOINT LAKE

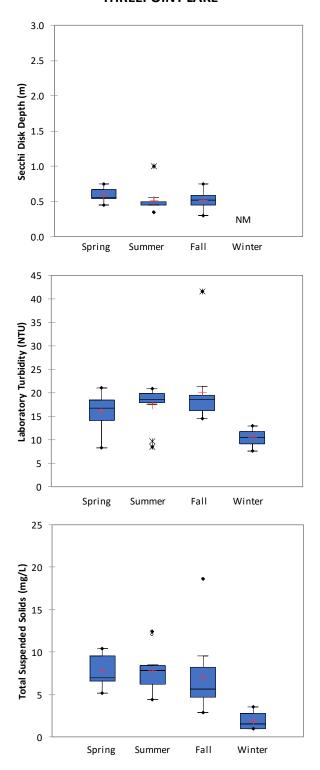


Figure 3.3-2. 2008-2019 On-system seasonal Secchi disk depth, turbidity, and TSS concentrations.



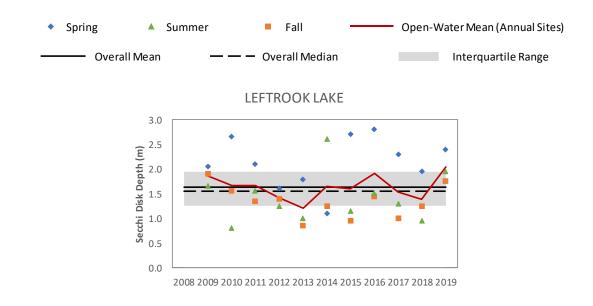
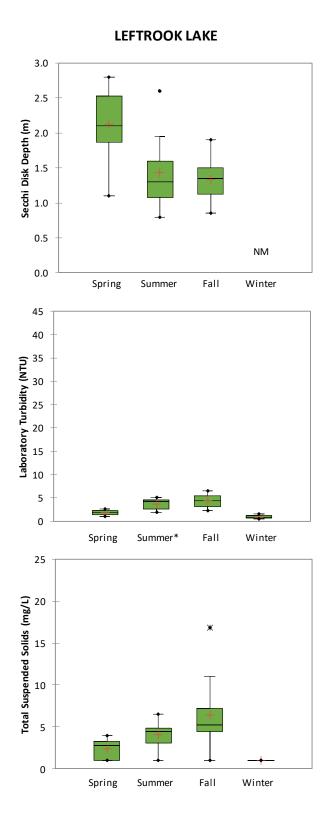


Figure 3.3-3. 2008-2019 Off-system open-water season Secchi disk depths.



2024



^{*}Excludes suspect turbidity value of 19.5 NTU from summer 2011.

Figure 3.3-4. 2008-2019 Off-system seasonal Secchi disk depth, turbidity, and TSS concentrations.



3.3.2 TURBIDITY

3.3.2.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Turbidity in Threepoint Lake ranged from 8.22 to 41.5 NTU during the open-water season. The mean and median turbidity for the 11 years of monitoring were 17.9 and 18.2 NTU, respectively. Open-water season mean annual turbidity ranged from 14.0 to 25.5 NTU and was within the IQR (16.2 to 19.7 NTU) in six of the 11 years. Mean turbidity was below the IQR in 2010, 2011, and 2012 and above the IQR in 2016 and 2017 (Table 3.3-1 and Figure 3.3-5).

Turbidity in the ice-cover season ranged from 7.52 to 13.0 NTU, with both a mean and median of 10.5 NTU for the 10 years of monitoring. The IQR was 9.09 to 11.8 NTU (Table 3.3-1 and Figure 3.3-5).

Turbidity was lower in winter (mean = 10.5 NTU) than the open-water season over the 11 years of monitoring. No clear seasonality was observed for turbidity during the open-water season; however, the lowest mean turbidity occurred in spring (16.2 NTU) and the highest in fall (19.9 NTU; Figure 3.3-2).

ROTATIONAL SITES

Footprint Lake

Turbidity in Footprint Lake ranged from 5.30 to 11.5 NTU during the open-water season. The mean was 7.33 NTU, the median was 7.36 NTU, and the IQR was 5.75 to 8.08 NTU for the four years of monitoring. Mean annual turbidity in the open-water season ranged from 6.51 to 8.90 NTU and was within the IQR in 2010, 2013, and 2019 but above the IQR in 2016 (Table 3.3-1 and Figure 3.3-5).

During the ice-cover season, turbidity ranged from 6.98 to 15.1 NTU, with a mean of 11.4 NTU (Table 3.3-1 and Figure 3.3-5).



Apussigamasi Lake

Turbidity in Apussigamasi Lake ranged from 21.0 to 35.0 NTU during the open-water season. The mean was 29.5 NTU, the median was 30.8 NTU, and the IQR was 27.1 to 32.1 NTU for the four years of monitoring. Mean annual turbidity in the open-water season ranged from 26.7 to 31.3 NTU and was within the IQR in 2009, 2015, and 2018 but below the IQR in 2012 (Table 3.3-1 and Figure 3.3-5).

During the ice-cover season, turbidity ranged from 14.1 to 19.4 NTU, with a mean of 16.4 NTU (Table 3.3-1 and Figure 3.3-5).

3.3.2.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

Turbidity in Leftrook Lake ranged from 1.00 to 6.55 NTU during the open-water season. The mean and median turbidity for the 11 years of monitoring were 3.25 and 2.58 NTU, respectively. Openwater season mean annual turbidity ranged from 2.14 to 4.53 NTU and was within the IQR (2.06 to 4.38 NTU) in 10 of the 11 years. The exception was 2015 when mean turbidity was above the IQR (Table 3.3-2 and Figure 3.3-6). ¹

Turbidity in the ice-cover season ranged from 0.42 to 1.53 NTU, with a mean of 0.90 NTU and median of 0.85 NTU for the 11 years of monitoring. The IQR was 0.71 to 1.12 NTU (Table 3.3-2 and Figure 3.3-6).

Seasonally, turbidity was lower in winter (mean = 0.90 NTU) than during the open-water season over the 11 years of monitoring. Additionally, mean turbidity was lower in spring (1.84 NTU) than in summer and fall (3.67 and 4.28 NTU, respectively; Figure 3.3-4).

¹ A suspect value of 19.5 NTU from summer 2011 has been excluded from the data reported for the open-water season. This value is unusually high for Leftrook Lake and does not correspond to the *in situ* turbidity measurement nor the TSS concentration reported for the sample.

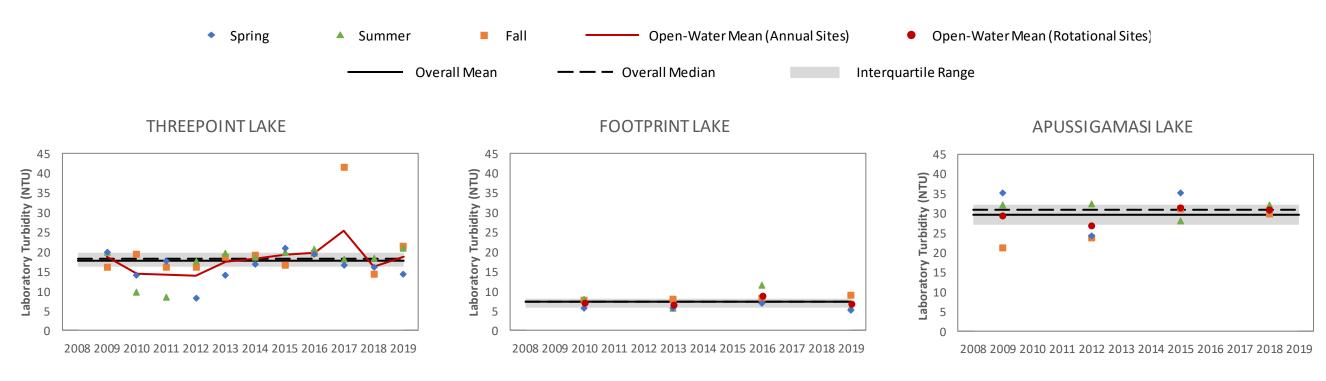


ROTATIONAL SITES

There are no off-system rotational sites in this region.



OPEN-WATER SEASON





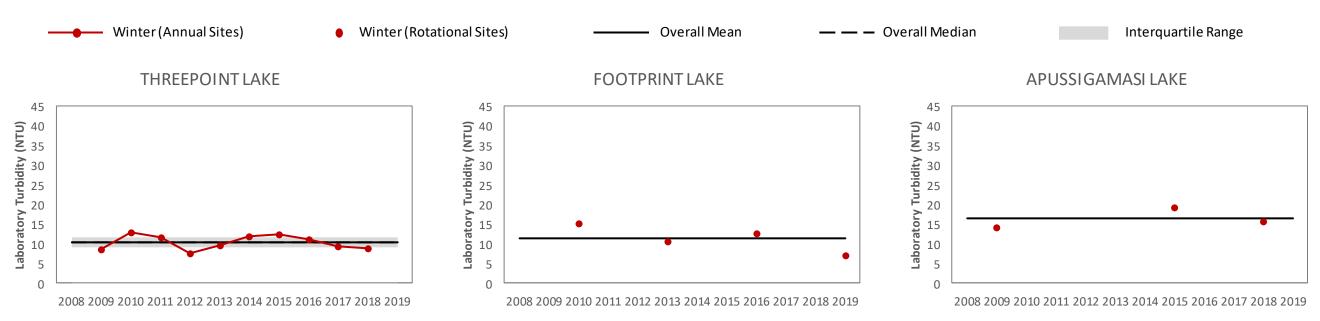
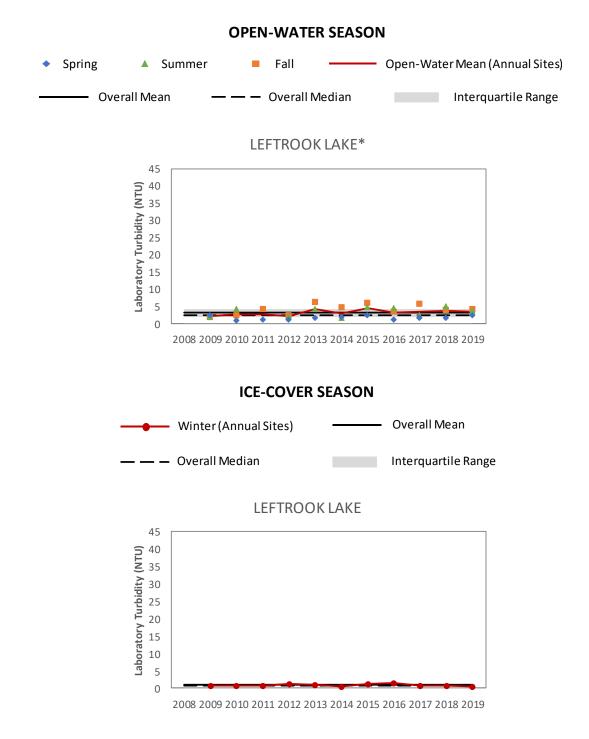


Figure 3.3-5. 2008-2019 On-system open-water and ice-cover season turbidity levels.





^{*}Excludes suspect value of 19.5 NTU from summer 2011.

Figure 3.3-6. 2008-2019 Off-system open-water and ice-cover season turbidity levels.



3.3.3 TOTAL SUSPENDED SOLIDS

3.3.3.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

TSS concentrations in Threepoint Lake ranged from 2.9 to 18.6 mg/L during the open-water season. The mean and median concentrations for the 11 years of monitoring were 7.5 and 7.0 mg/L, respectively. Open-water season mean annual TSS concentrations ranged from 5.4 to 10.7 mg/L and were within the IQR (5.6 to 8.6 mg/L) in seven of the 11 years. Mean TSS concentrations were below the IQR in 2019 and above the IQR in 2013, 2014, and 2017. TSS concentrations were consistently above the detection limit (DL; 2.0 mg/L) during the open-water season (percent detections = 100; Table 3.3-1 and Figure 3.3-7).

TSS concentrations in the ice-cover season ranged from <2.0 to 3.5 mg/L, both the mean and median were <2.0 mg/L, and the IQR was <2.0 to 2.8 mg/L for the ten years of monitoring. TSS concentrations were below the DL (2.0 mg/L) in half the samples collected during the ice-cover season (percent detections = 50; Table 3.3-1 and Figure 3.3-7).

TSS concentrations in Threepoint Lake were lower in winter (mean <2.0 mg/L), and more frequently below the DL, compared to the open-water season. No clear seasonality was observed for TSS concentrations in the open-water season; however, the lowest mean TSS concentration occurred in fall (7.0 mg/L) and the highest in spring (7.9 mg/L; Figure 3.3-2).

ROTATIONAL SITES

Footprint Lake

TSS concentrations in Footprint Lake ranged from 2.9 to 7.0 mg/L during the open-water season. The mean was 5.1 mg/L and median was 4.9 mg/L, and the IQR was 4.5 to 5.8 mg/L for the four years of monitoring. Mean annual TSS concentrations in the open-water season ranged from 3.6 to 6.5 mg/L and were within the IQR in 2013 and 2016 but below the IQR in 2019 and above the IQR in 2010. TSS concentrations were consistently above the DL (2.0 mg/L) during the open-water season (Table 3.3-1 and Figure 3.3-7).



During the ice-cover season, TSS concentrations ranged from <2.0 to 2.4 mg/L, with a mean of <2.0 mg/L. TSS concentrations were below the DL (2.0 mg/L) in two of four samples collected in winter (i.e., percent detections = 50; Table 3.3-1 and Figure 3.3-7).

Apussigamasi Lake

TSS concentrations in Apussigamasi Lake ranged from 7.6 to 24.4 mg/L during the open-water season. The mean was 13.6 mg/L and median was 12.7 mg/L, and the IQR was 9.4 to 15.5 mg/L for the four years of monitoring. Mean annual TSS concentrations in the open-water season ranged from 10.3 to 20.8 mg/L and were within the IQR in 2009, 2015, and 2018 but above the IQR in 2012 (Table 3.3-1 and Figure 3.3-7).

During the ice-cover season, TSS concentrations ranged from 6.5 to 9.6 mg/L, with a mean of 8.2 mg/L (Table 3.3-1 and Figure 3.3-7).

3.3.3.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

TSS concentrations in Leftrook Lake ranged from <2.0 to 16.8 mg/L during the open-water season. The mean and median concentrations for the 11 years of monitoring were 4.3 and 3.9 mg/L, respectively. Open-water season mean annual TSS concentrations ranged from <2.0 to 7.1 mg/L and were within the IQR (2.4 to 4.8 mg/L) in seven of the 11 years. Mean TSS concentrations were below the IQR in 2010 and above the IQR in 2013, 2014, and 2017. TSS concentrations were occasionally below, but typically above, the DL (2.0 mg/L) during the open-water season (percent detections = 82; Table 3.3-2 and Figure 3.3-8).

TSS concentrations in the ice-cover season were consistently below the DL (2.0 mg/L) over the 11 years of monitoring (Table 3.3-2 and Figure 3.3-8).

Over the 12-year period, TSS concentrations in Leftrook Lake were lower in winter (mean <2.0 mg/L) than the open-water season. During the open-water season, mean TSS concentrations were lowest in spring (2.4 mg/L) and highest in fall (6.4 mg/L; Figure 3.3-4).

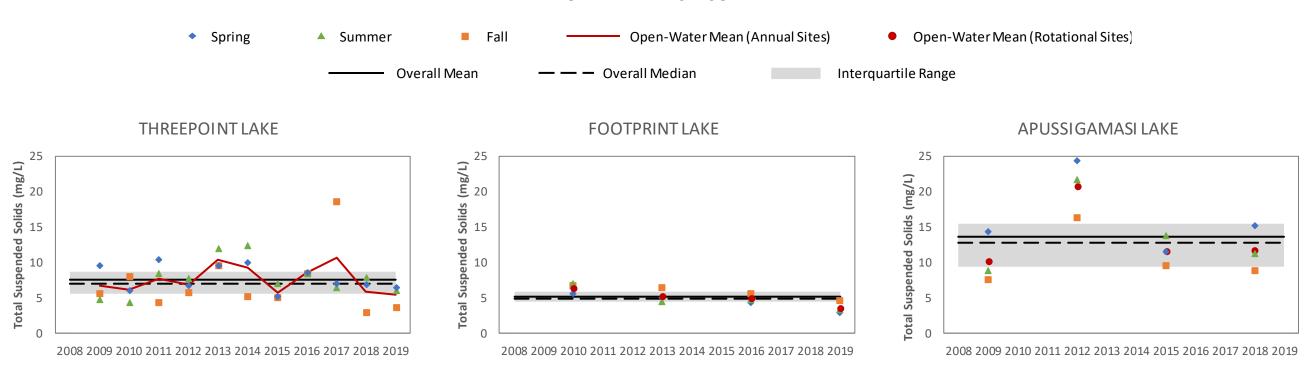


ROTATIONAL SITES

There are no off-system rotational sites in this region.



OPEN-WATER SEASON



ICE-COVER SEASON

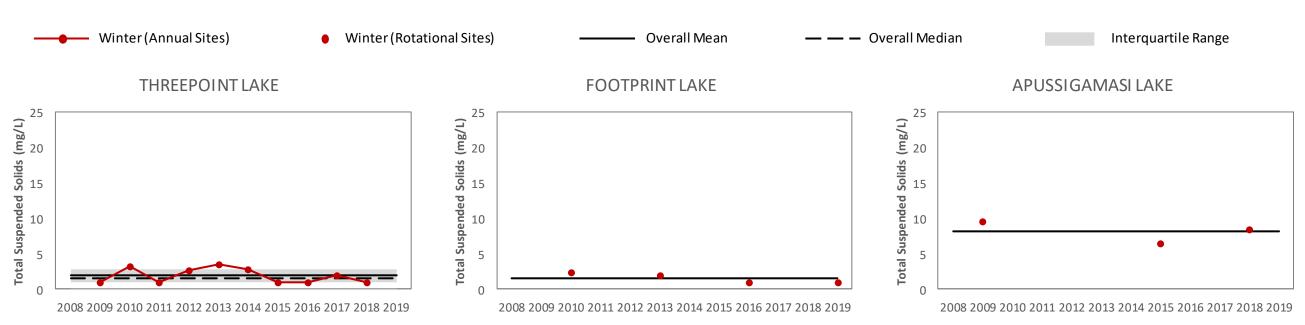


Figure 3.3-7. 2008-2019 On-system open-water and ice-cover season TSS concentrations.



OPEN-WATER SEASON Spring Summer Fall Open-Water Mean (Annual Sites) Interquartile Range Overall Mean Overall Median **LEFTROOK LAKE** 25 0 2008 2009 2010 2011 2012 2013 2014 2015 2016 2017 2018 2019 **ICE-COVER SEASON** Overall Mean Winter (Annual Sites) **–** Overall Median Interquartile Range **LEFTROOK LAKE** 25 Total Suspended Solids (mg/L) 20 15 10 5

Figure 3.3-8. 2008-2019 Off-system open-water and ice-cover season TSS concentrations.

2008 2009 2010 2011 2012 2013 2014 2015 2016 2017 2018 2019



3.4 NUTRIENTS AND TROPHIC STATUS

3.4.1 TOTAL PHOSPHORUS

3.4.1.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

TP concentrations in Threepoint Lake ranged from 0.010 to 0.047 mg/L during the open-water season. The mean and median concentrations for the 11 years of monitoring were both 0.024 mg/L. Open-water season mean annual TP concentrations ranged from 0.016 to 0.035 mg/L and were within the IQR (0.017 to 0.030 mg/L) in eight of the 11 years. Mean TP concentrations were below the IQR in 2013 and above the IQR in 2016 and 2017 (Table 3.4-1 and Figure 3.4-1).

TP concentrations in the ice-cover season ranged from 0.021 to 0.028 mg/L, with a mean of 0.024 mg/L and a median of 0.025 mg/L for the 10 years of monitoring. The IQR was 0.022 to 0.026 mg/L (Table 3.4-1 and Figure 3.4-1).

No clear seasonality was observed for TP in Threepoint Lake over the 11 years of monitoring. Mean TP concentrations were identical in spring, summer, and winter (0.024 mg/L for each season) and the mean TP concentration was only slighter higher in fall (0.025 mg/L; Figure 3.4-2).

Threepoint Lake was meso-eutrophic (0.020 to 0.035 mg/L) on the basis of the 2009-2019 mean open-water season TP concentration (0.024 mg/L). Mean annual TP concentrations (0.016 to 0.035 mg/L) in the open-water season were within the meso-eutrophic range (0.020 to 0.035 mg/L) in eight of the 11 years of monitoring. Mean TP concentrations were in the mesotrophic range (0.010 to 0.020 mg/L) in 2011, 2012, and 2013 (Table 3.4-2).

ROTATIONAL SITES

Footprint Lake

TP concentrations in Footprint Lake ranged from 0.010 to 0.035 mg/L during the open-water season. The mean was 0.023 mg/L, the median was 0.025 mg/L, and the IQR was 0.017 to 0.029 mg/L for the four years of monitoring. Mean annual TP concentrations in the open-water



season ranged from 0.019 to 0.027 mg/L and were within the IQR in all four years (Table 3.4-1 and Figure 3.4-1).

During the ice-cover season, TP concentrations ranged from 0.022 to 0.033 mg/L, with a mean of 0.028 mg/L (Table 3.4-1 and Figure 3.4-1).

Footprint Lake was meso-eutrophic (0.020 to 0.035 mg/L) based on the mean of the open-water season TP concentrations for the four years of monitoring (0.023 mg/L). Open-water season mean annual TP concentrations (0.019 to 0.027 mg/L) were also within the meso-eutrophic range in three of the four years of monitoring; however, the open-water mean TP concentration was within the mesotrophic range (0.010 to 0.020 mg/L) in 2013 (Table 3.4-2).

Apussigamasi Lake

TP concentrations in Apussigamasi Lake ranged from 0.023 to 0.044 mg/L during the open-water season. The mean was 0.033 mg/L, the median was 0.034 mg/L, and the IQR was 0.032 to 0.035 mg/L for the four years of monitoring. Mean annual TP concentrations in the open-water season ranged from 0.030 to 0.035 mg/L and were within the IQR in 2009, 2015, and 2018 but below the IQR in 2012 (Table 3.4-1 and Figure 3.4-1).

During the ice-cover season, TP concentrations ranged from 0.027 to 0.032 mg/L, with a mean of 0.030 mg/L for the three years of monitoring (Table 3.4-1 and Figure 3.4-1).

Apussigamasi Lake was meso-eutrophic (0.020 to 0.035 mg/L) based on the mean of the openwater season TP concentrations for the four years of monitoring (0.033 mg/L). Open-water season mean annual TP concentrations (0.030 to 0.035 mg/L) were also within the meso-eutrophic range in each year sampled (Table 3.4-2).

3.4.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

TP concentrations in Leftrook Lake ranged from 0.010 to 0.062 mg/L during the open-water season. The mean and median concentrations for the 11 years of monitoring were 0.028 mg/L and 0.024 mg/L, respectively. Open-water season mean annual TP concentrations ranged from



0.019 to 0.038 mg/L and were within the IQR (0.019 to 0.036 mg/L) in 10 of the 11 years. Mean TP concentrations were above the IQR in 2013 (Table 3.4-3 and Figure 3.4-3).

TP concentrations in the ice-cover season ranged from 0.012 to 0.023 mg/L, with a mean of 0.018 mg/L and a median of 0.019 mg/L for the 11 years of monitoring. The IQR was 0.014 to 0.021 mg/L (Table 3.4-3 and Figure 3.4-3).

On average, TP concentrations were lower in spring (0.018 mg/L) and winter (0.018 mg/L) than summer (0.029 mg/L) or fall (0.036 mg/L) over the 11 years of monitoring (Figure 3.4-4).

Leftrook Lake was meso-eutrophic (0.020 to 0.035 mg/L) on the basis of the 2009-2019 mean open-water season TP concentration (0.028 mg/L). Mean annual TP concentrations (0.019 to 0.038 mg/L) in the open-water season were also within the meso-eutrophic range (0.020 to 0.035 mg/L) in eight of the 11 years of monitoring. However, mean annual TP concentrations were within the mesotrophic range (0.010 to 0.020 mg/L) in 2011 and in the eutrophic range (0.035 to 0.100 mg/L) in 2010 and 2013 (Table 3.4-4).

ROTATIONAL SITES

There are no off-system rotational sites in this region.



Table 3.4-1. 2008-2019 On-system sites TP, TN, and chlorophyll *a* summary statistics.

| Cito | Statistic | TP (mg/L) | | TN (mg/L) | | Chlorophyll a (µg/L) | |
|----------|----------------|-----------|--------|-----------|--------------|----------------------|-------|
| Site | Statistic | ow | IC | ow | IC | ow | IC |
| | Mean | 0.024 | 0.024 | 0.34 | 0.37 | 3.28 | <0.60 |
| | Median | 0.024 | 0.025 | 0.33 | 0.37 | 3.05 | <0.60 |
| 3PT FOOT | Minimum | 0.010 | 0.021 | <0.20 | 0.28 | 0.93 | <0.60 |
| | Maximum | 0.047 | 0.028 | 0.70 | 0.37 0.37 | 9.16 | 0.67 |
| 2DT | SD | 0.0083 | 0.0024 | 0.107 | 0.076 | 1.61 | - |
| 361 | SE | 0.0014 | 0.0008 | 0.019 | 0.024 | 0.280 | - |
| | Lower Quartile | 0.017 | 0.022 | 0.30 | 0.31 | 2.48 | <0.60 |
| | Upper Quartile | 0.030 | 0.026 | 0.39 | 0.44 | 3.82 | <0.60 |
| | n | 33 | 10 | 32 | 10 | 33 | 10 |
| | % Detections | 100 | 100 | 100 | 100 | 100 | 20 |
| | Mean | 0.023 | 0.028 | 0.51 | 0.41 | 8.14 | <0.60 |
| | Median | 0.025 | - | 0.48 | - | 7.88 | - |
| | Minimum | 0.010 | 0.022 | 0.29 | 0.34 | 1.55 | <0.60 |
| | Maximum | 0.035 | 0.033 | 1.00 | 0.48 | 23.8 | <0.60 |
| | SD | 0.0081 | 0.0043 | 0.198 | 0.061 | 5.87 | - |
| FUUT | SE | 0.0023 | 0.0022 | 0.057 | | 1.69 | - |
| | Lower Quartile | 0.017 | - | 0.40 | - | 4.25 | - |
| | Upper Quartile | 0.029 | - | 0.54 | - | 9.54 | - |
| | n | 12 | 4 | 12 | 4 | 12 | 4 |
| | % Detections | 100 | 100 | 100 | 100 | 100 | 25 |
| | Mean | 0.033 | 0.030 | 0.36 | 0.45 | 3.12 | <0.60 |
| APU | Median | 0.034 | 1 | 0.34 | - | 2.92 | - |
| | Minimum | 0.023 | 0.027 | 0.24 | 0.38 | 2.30 | <0.60 |
| | Maximum | 0.044 | 0.032 | 0.47 | 0.54 | 4.42 | <0.60 |
| | SD | 0.0052 | 0.0026 | 0.076 | 0.082 | 0.703 | - |
| AFU | SE | 0.0015 | 0.0015 | 0.022 | 0.047 | 0.203 | - |
| | Lower Quartile | 0.032 | - | 0.33 | - | 2.61 | - |
| | Upper Quartile | 0.035 | - | 0.44 | - | 3.76 | - |
| | n | 12 | 3 | 12 | 3 | 12 | 3 |
| | % Detections | 100 | 100 | 100 | 100 | 100 | 67 |

- 1. OW = Open-water season; IC = Ice-cover season.
- 2. SD = standard deviation; SE = standard error; n = number of samples.
- 3. TN statistics for 3PT exclude suspect value of 3.06 mg/L from spring 2014.



Table 3.4-2. 2008-2019 On-system trophic status based on TP, TN, and chlorophyll *a* openwater season mean concentrations.

| Trophic Categories | Total Phosphorus (mg/L) | | Total Nitrogen (mg/L) | | Chlorophyll a (μg/L) | | | | |
|---------------------|------------------------------|-------------|-----------------------|-----------------|----------------------|------|-------------|------|------|
| Ultra-oligotrophic | | <0.004 | | | | | | | |
| Oligotrophic | | 0.004-0.010 |) | <0.350 | | <2.5 | | | |
| Mesotrophic | | 0.010-0.020 |) | 0.350-0.650 | | | 2.5-8 | | |
| Meso-eutrophic | 0.020-0.035 | | | | | | | | |
| Eutrophic | 0.035-0.100 | | | 0.651-1.20 | | | 8-25 | | |
| Hypereutrophic | > 0.100 | | | >1.20 | | | >25 | | |
| References | CCME (1999; updated to 2024) | | | Nürnberg (1996) | | | OECD (1982) | | |
| Sampling Year | 3РТ | FOOT | APU | 3PT | FOOT | APU | 3РТ | FOOT | APU |
| 2008 | - | - | - | - | _ | - | - | - | - |
| 2009 | 0.023 | - | 0.035 | 0.35 | - | 0.43 | 2.40 | - | 2.70 |
| 2010 | 0.028 | 0.026 | - | 0.33 | 0.61 | - | 1.37 | 3.33 | - |
| 2011 | 0.020 | - | - | 0.37 | - | - | 2.42 | - | - |
| 2012 | 0.017 | - | 0.030 | 0.33 | - | 0.31 | 3.48 | - | 2.75 |
| 2013 | 0.016 | 0.019 | - | 0.24 | 0.33 | - | 4.98 | 8.20 | - |
| 2014 | 0.028 | - | - | 0.42 | - | - | 2.76 | - | - |
| 2015 | 0.024 | - | 0.033 | 0.40 | - | 0.38 | 3.03 | - | 2.94 |
| 2016 | 0.032 | 0.027 | - | 0.32 | 0.52 | - | 3.69 | 7.84 | - |
| 2017 | 0.035 | - | - | 0.38 | - | - | 2.53 | - | - |
| 2018 | 0.020 | - | 0.035 | 0.28 | - | 0.33 | 4.19 | - | 4.07 |
| 2019 | 0.025 | 0.022 | - | 0.35 | 0.58 | - | 5.19 | 13.2 | - |
| Overall (2008-2019) | 0.024 | 0.023 | 0.033 | 0.34 | 0.51 | 0.36 | 3.28 | 8.14 | 3.12 |

- 1. CCME = Canadian Council of Ministers of the Environment.
- 2. OECD = Organization for Economic Cooperation and Development.
- 3. TN values for 3PT exclude suspect value of 3.06 mg/L from spring 2014.



Table 3.4-3. 2008-2019 Off-system sites TP, TN and chlorophyll *a* summary statistics.

| Site | Statistic | TP (mg/L) | | TN (mg/L) | | Chlorophyll α (μg/L) | |
|------|----------------|-----------|--------|-----------|-------|----------------------|-------|
| | | ow | IC | ow | IC | ow | IC |
| | Mean | 0.028 | 0.018 | 0.57 | 0.59 | 13.4 | 1.03 |
| | Median | 0.024 | 0.019 | 0.58 | 0.60 | 10.3 | 1.15 |
| | Minimum | 0.010 | 0.012 | 0.24 | 0.52 | 0.78 | <0.60 |
| | Maximum | 0.062 | 0.023 | 0.89 | 0.66 | 34.0 | 1.53 |
| LEET | SD | 0.0121 | 0.0041 | 0.156 | 0.048 | 9.61 | 0.500 |
| LEFT | SE | 0.0021 | 0.0012 | 0.028 | 0.015 | 1.67 | 0.151 |
| | Lower Quartile | 0.019 | 0.014 | 0.46 | 0.55 | 5.35 | 0.85 |
| | Upper Quartile | 0.036 | 0.021 | 0.67 | 0.63 | 21.3 | 1.42 |
| | n | 33 | 11 | 32 | 11 | 33 | 11 |
| | % Detections | 100 | 100 | 100 | 100 | 100 | 82 |

- 1. OW = Open-water season; IC = Ice-cover season.
- 2. SD = standard deviation; SE = standard error; n = number of samples.
- 3. TN statistics for LEFT exclude suspect value of 4.33 mg/L from spring 2014.

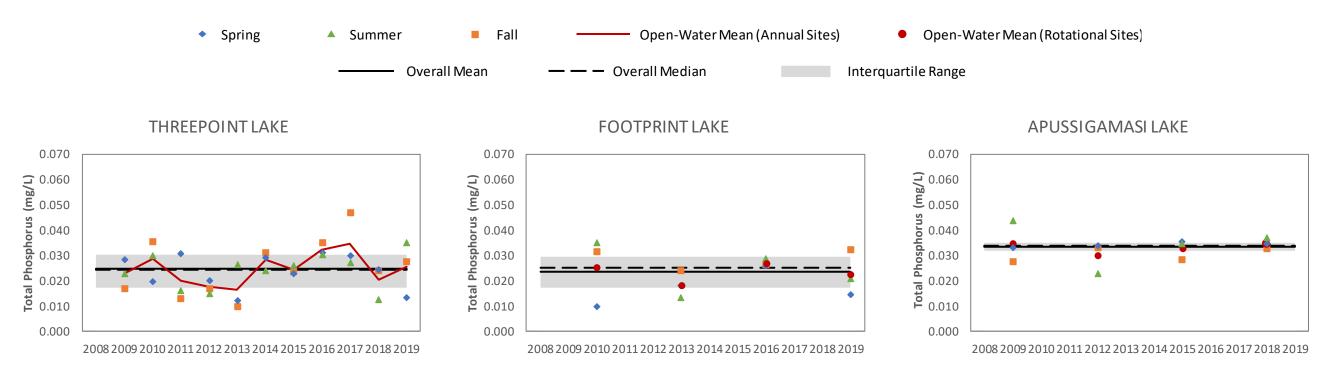


Table 3.4-4. 2008-2019 Off-system trophic status based on TP, TN, and chlorophyll *a* openwater season mean concentrations.

| Trophic Categories | Total Phosphorus (mg/L) Total Nitrogen (mg/L) | | Chlorophyll a (μg/L) | |
|---------------------|---|-------------------|----------------------|--|
| Ultra-oligotrophic | <0.004 | | | |
| Oligotrophic | 0.004-0.010 | <0.350 | <2.5 | |
| Mesotrophic | 0.010-0.020 | 0.350-0.650 | 2.5-8 | |
| Meso-eutrophic | 0.020-0.035 | | | |
| Eutrophic | 0.035-0.100 | 0.651-1.20 | 8-25 | |
| Hypereutrophic | > 0.100 | >1.20 | >25 | |
| References | CCME (1999; updated to 2024) | Nürnberg (1996) | OECD (1982) | |
| Sampling Year | LEFT | LEFT | LEFT | |
| 2008 | - | - | - | |
| 2009 | 0.021 | 0.50 | 6.37 | |
| 2010 | 0.036 | 0.61 | 9.06 | |
| 2011 | 0.019 | 0.66 | 16.0 | |
| 2012 | 0.024 | 0.53 | 13.6 | |
| 2013 | 0.038 | 0.63 | 20.6 | |
| 2014 | 0.025 | 0.60 ¹ | 12.8 | |
| 2015 | 0.030 | 0.58 | 14.2 | |
| 2016 | 0.028 | 0.55 | 7.77 | |
| 2017 | 0.030 | 0.58 | 14.0 | |
| 2018 | 0.030 | 0.55 | 17.4 | |
| 2019 | 0.023 | 0.52 | 15.6 | |
| Overall (2008-2019) | 0.028 | 0.57 ¹ | 13.4 | |

- 1. CCME = Canadian Council of Ministers of the Environment.
- 2. OECD = Organization for Economic Cooperation and Development.
- 3. TN values for LEFT exclude suspect value of 4.33 mg/L from spring 2014.







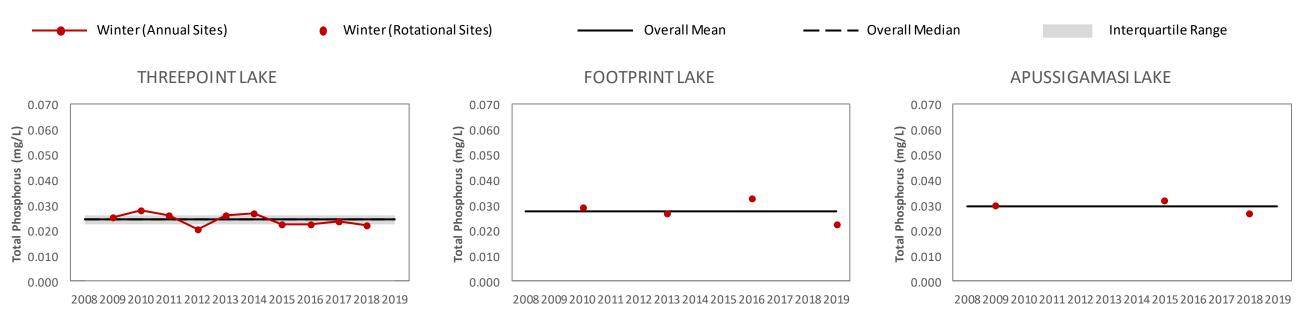
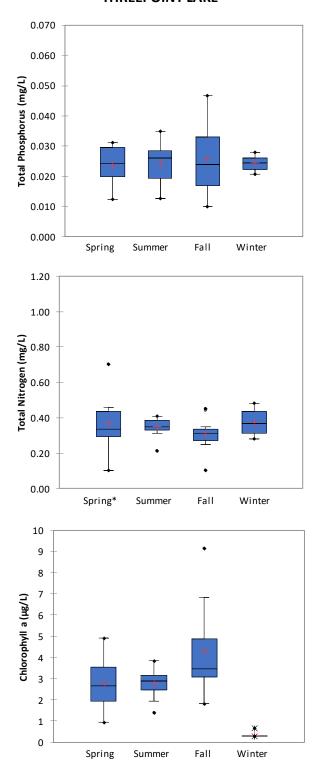


Figure 3.4-1. 2008-2019 On-system open-water and ice-cover season TP concentrations.



THREEPOINT LAKE



^{*}Excludes suspect TN value of 3.06 mg/L from spring 2014.

Figure 3.4-2. 2008-2019 On-system seasonal total phosphorus, total nitrogen, and chlorophyll *a* concentrations.





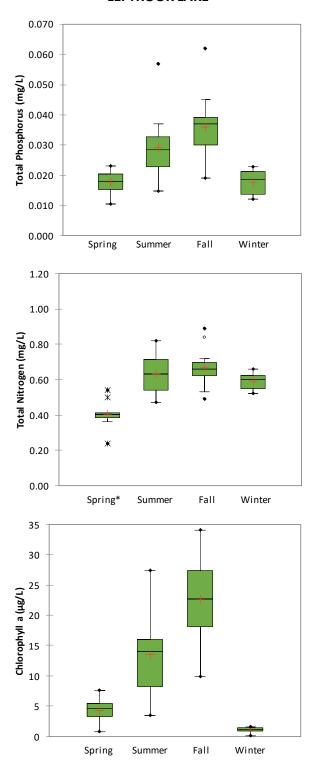
Figure 3.4-3. 2008-2019 Off-system open-water and ice-cover season TP concentrations.

2008 2009 2010 2011 2012 2013 2014 2015 2016 2017 2018 2019

0.000







^{*}Excludes suspect TN value of 4.33 mg/L from spring 2014.

Figure 3.4-4. 2008-2019 Off-system seasonal total phosphorus, total nitrogen, and chlorophyll *a* concentrations.



3.4.2 TOTAL NITROGEN

3.4.2.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

TN concentrations in Threepoint Lake ranged from <0.20 to 0.70 mg/L during the open-water season. The mean and median concentrations for the 11 years of monitoring were 0.34 mg/L and 0.33 mg/L, respectively. Open-water season mean annual TN concentrations ranged from 0.24 to 0.42 mg/L and were within the IQR (0.30 to 0.39 mg/L) in seven of the 11 years. Mean TN concentrations were below the IQR in 2013 and 2018 and above the IQR in 2014 and 2015 (Table 3.4-1 and Figure 3.4-5).²

TN concentrations in the ice-cover season ranged from 0.28 to 0.48 mg/L, with both a mean and median of 0.37 mg/L for the 10 years of monitoring. The IQR was 0.31 to 0.44 mg/L (Table 3.4-1 and Figure 3.4-5).

No clear seasonality was observed for TN in Threepoint Lake over the 11 years of monitoring. However, mean TN concentrations were lowest in fall (0.35 mg/L) and highest in spring and winter 0.37 mg/L for both; Figure 3.4-2).

Threepoint Lake was oligotrophic (<0.350 mg/L) on the basis of the 2009-2019 mean open-water season TN concentration (0.34 mg/L). Mean annual TN concentrations (0.24 to 0.42 mg/L) in the open-water season were also within the oligotrophic range (<0.350 mg/L) in five of the 11 years of monitoring. Mean annual TN concentrations were within the mesotrophic range (0.350 to 0.650 mg/L) in 2009, 2011, 2014, 2015, 2017, and 2019 (Table 3.4-4).

ROTATIONAL SITES

Footprint Lake

TN concentrations in Footprint Lake ranged from 0.29 to 1.00 mg/L during the open-water season. The mean was 0.51 mg/L, the median was 0.48 mg/L, and the IQR was 0.40 to 0.54 mg/L for the



² A suspect TN value of 3.06 mg/L from spring 2014 has been excluded from the data reported for the open-water season.

four years of monitoring. Mean annual TN concentrations in the open-water season ranged from 0.33 to 0.61 mg/L and were within the IQR in 2016 but were below the IQR in 2013 and above the IQR in 2010 and 2019 (Table 3.4-1 and Figure 3.4-5).

During the ice-cover season, TN concentrations ranged from 0.34 to 0.48 mg/L, with a mean of 0.41 mg/L (Table 3.4-1 and Figure 3.4-5).

Footprint Lake was mesotrophic (0.350 to 0.650 mg/L) based on the mean of the open-water season TN concentrations for the four years of monitoring (0.51 mg/L). Open-water season mean annual TN concentrations (0.33 to 0.61 mg/L) were also within the mesotrophic range in three of the four years of monitoring; however, the open-water mean TN concentration was within the oligotrophic range (<0.350 mg/L) in 2013 (Table 3.4-2).

Apussigamasi Lake

TN concentrations in Apussigamasi Lake ranged from 0.24 to 0.47 mg/L during the open-water season. The mean was 0.36 mg/L, the median was 0.34 mg/L, and the IQR was 0.33 to 0.44 mg/L for the four years of monitoring. Mean annual TN concentrations in the open-water season ranged from 0.31 to 0.43 mg/L and were within the IQR in three of the four years. The mean annual TN concentration was below the IQR in 2012 (Table 3.4-1 and Figure 3.4-5).

During the ice-cover season, TN concentrations ranged from 0.38 to 0.54 mg/L, with a mean of 0.45 mg/L for the three years of monitoring (Table 3.4-1 and Figure 3.4-5).

Apussigamasi Lake was mesotrophic (0.350 to 0.650 mg/L) based on the mean of the open-water season TN concentrations for the four years of monitoring (0.36 mg/L). Open-water season mean annual TN concentrations (0.31 to 0.43 mg/L) were also within the mesotrophic range in 2009 and 2015; however, the open-water mean TN concentrations were within the oligotrophic range (<0.350 mg/L) in 2012 and 2018 (Table 3.4-2).

3.4.2.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

TN concentrations in Leftrook Lake ranged from 0.24 to 0.89 mg/L during the open-water season. The mean and median concentrations for the 11 years of monitoring were 0.57 mg/L and



0.58 mg/L, respectively. Open-water season mean annual TN concentrations ranged from 0.50 to 0.66 mg/L and were within the IQR (0.46 to 0.67 mg/L) in all years (Table 3.4-3 and Figure 3.4-6).³

TN concentrations in the ice-cover season ranged from 0.52 to 0.66 mg/L, with a mean of 0.59 mg/L and a median of 0.60 mg/L for the 11 years of monitoring. The IQR was 0.55 to 0.63 mg/L (Table 3.4-3 and Figure 3.4-6).

On average, TN concentrations were lower in spring (0.40 mg/L) than in summer (0.63 mg/L), fall (0.67 mg/L) or winter (0.59 mg/L) over the 11 years of monitoring Figure 3.4-4).

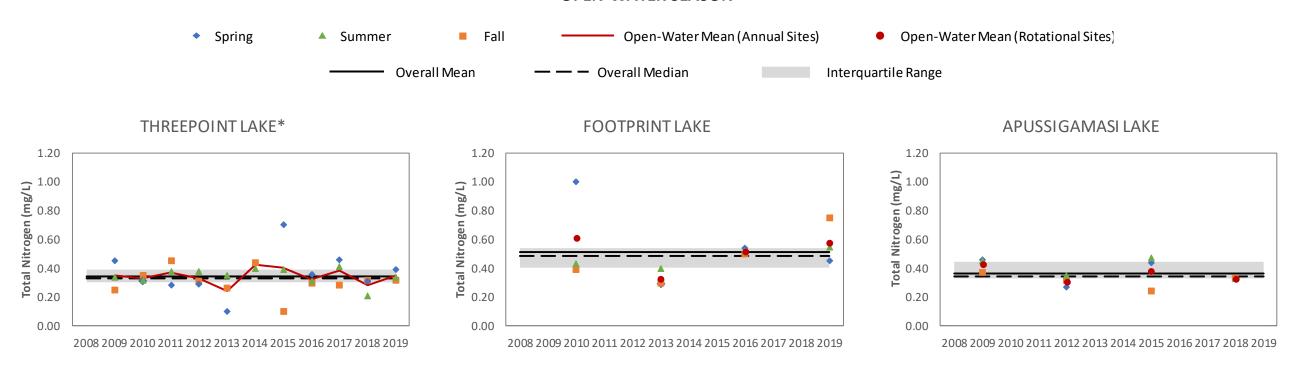
Leftrook Lake was mesotrophic (0.350 to 0.650 mg/L) on the basis of the 2009-2019 mean openwater season TN concentration (0.57 mg/L). Mean annual TN concentrations (0.50 to 0.66 mg/L) in the open-water season were also within the meso-eutrophic range (0.020 to 0.035 mg/L) in 10 of the 11 years of monitoring; however, the mean annual TN concentration was within the eutrophic range (0.651 to 1.20 mg/L) in 2011 (Table 3.4-4).

ROTATIONAL SITES

There are no off-system rotational sites in this region.

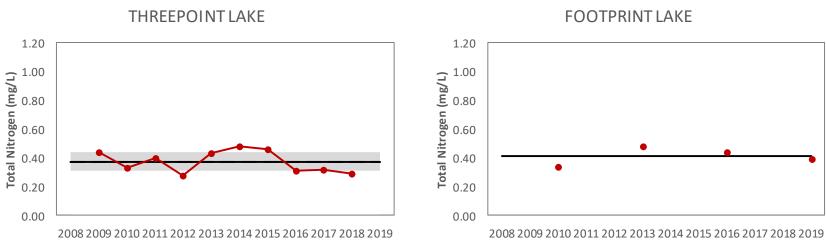


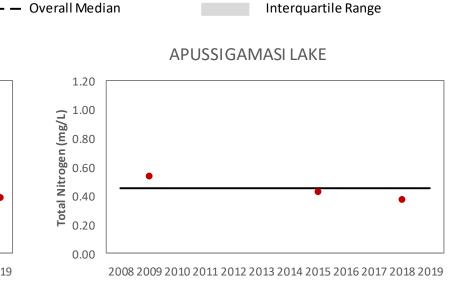
³ A suspect TN value of 4.33 mg/L from spring 2014 has been excluded from the data reported for the open-water season.



ICE-COVER SEASON

Winter (Annual Sites) ■ Winter (Rotational Sites) — Overall Mean

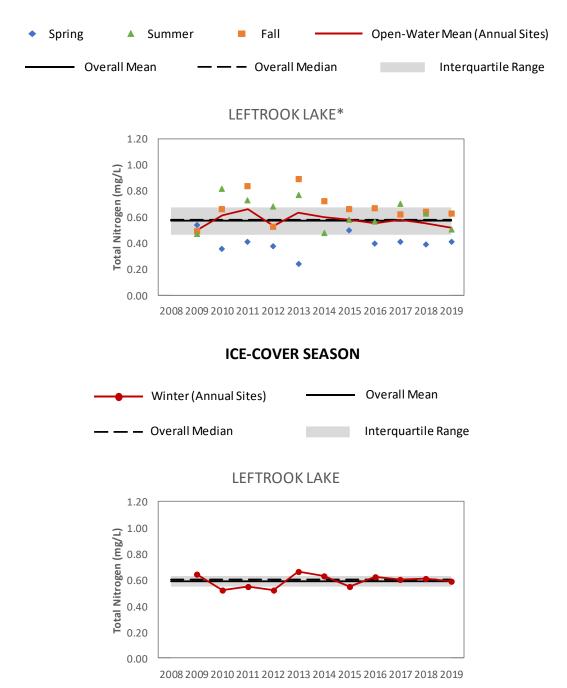




*Excludes suspect value of 3.06 mg/L at 3PT from spring 2014.

Figure 3.4-5. 2008-2019 On-system open-water and ice-cover season TN concentrations.





^{*}Excludes suspect value of 4.33 mg/L from spring 2014.

Figure 3.4-6. 2008-2019 Off-system open-water and ice-cover season TN concentrations.



3.4.3 CHLOROPHYLL A

3.4.3.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Chlorophyll a concentrations in Threepoint Lake ranged from 0.93 to 9.16 μ g/L during the openwater season. The mean and median for the 11 years of monitoring were 3.28 μ g/L and 3.05 μ g/L, respectively. Open-water season mean annual chlorophyll a concentrations ranged from 1.37 to 5.19 μ g/L and were within the IQR (2.48 to 3.82 μ g/L) in five of the 11 years. Mean chlorophyll a concentrations were below the IQR in 2009, 2010, and 2011 and above the IQR in 2013, 2018, and 2019 (Table 3.4-1 and Figure 3.4-7).

Chlorophyll a concentrations in the ice-cover season ranged from <0.60 to 0.67 μ g/L, with both a mean and median of <0.60 μ g/L for the 10 years of monitoring. The IQR was below the analytical DL of 0.60 μ g/L (Table 3.4-1 and Figure 3.4-7).

Chlorophyll a concentrations were lower in the winter and frequently below the DL (0.040-0.60 μ g/L; percent detection = 20), compared to the open-water season (Table 3.4-1). On average, chlorophyll a concentrations during the open-water season were lowest in spring and summer (2.77 and 2.78 μ g/L, respectively) and highest in fall (4.28 μ g/L; Figure 3.4-2).

Threepoint Lake was mesotrophic (2.5 to 8 μ g/L) on the basis of the 2009-2019 mean open-water season chlorophyll a concentration (3.28 μ g/L). Mean annual chlorophyll a concentrations (1.37 to 5.19 μ g/L) in the open-water season were also within the mesotrophic range in eight of the 11 years. However, the mean chlorophyll a concentration was within the oligotrophic range (<2.5 μ g/L) in 2009, 2010, and 2011 (Table 3.4-2).

ROTATIONAL SITES

Footprint Lake

Chlorophyll a concentrations in Footprint Lake ranged from 1.55 to 23.8 μ g/L during the openwater season. The mean was 8.14 μ g/L, the median was 7.88 μ g/L, and the IQR was 4.25 to 9.54 μ g/L for the four years of monitoring. Mean annual chlorophyll a concentrations in the open-



water season ranged from 3.33 to 13.2 μ g/L and were within the IQR in 2013 and 2016 but were below the IQR in 2010 and above the IQR in 2019 (Table 3.4-1 and Figure 3.4-7).

Chlorophyll a concentrations were typically below the DL (0.10-0.60 μ g/L) during the ice-cover season (percent detection = 25) and all concentrations were less than 0.60 μ g/L. The mean chlorophyll a concentration for the four years of monitoring was <0.60 μ g/L (Table 3.4-1 and Figure 3.4-7).

Footprint Lake was eutrophic (8-25 μ g/L) based on the mean of the open-water season chlorophyll a concentrations for the four years of monitoring (8.14 μ g/L). Open-water season mean annual chlorophyll a concentrations (3.33 to 13.2 μ g/L) were also within the eutrophic range in 2013 and 2019 but were within the mesotrophic range (2.5 to 8 μ g/L) in 2010 and 2016 (Table 3.4-2).

Apussigamasi Lake

Chlorophyll a concentrations in Apussigamasi Lake ranged from 2.30 to 4.42 μ g/L during the open-water season. The mean was 3.12 μ g/L, the median was 2.92 μ g/L, and the IQR was 2.61 to 3.76 μ g/L for the four years of monitoring. Mean annual chlorophyll a concentrations in the open-water season ranged from 2.70 to 4.07 μ g/L and were within the IQR in three of the four years but were above the IQR in 2018 (Table 3.4-1 and Figure 3.4-7).

Chlorophyll a concentrations were near or below the DL (0.050-0.60 μ g/L) during the ice-cover season (percent detection = 67) and all concentrations were less than 0.60 μ g/L. The mean chlorophyll a concentration for the three years of monitoring was <0.60 μ g/L (Table 3.4-1 and Figure 3.4-7).

Apussigamasi Lake was mesotrophic (2.5 to 8 μ g/L) based on the mean of the open-water season chlorophyll a concentrations for the four years of monitoring (3.12 μ g/L). Open-water season mean annual chlorophyll a concentrations (2.70 to 4.07 μ g/L) were also within the mesotrophic range in each year of monitoring (Table 3.4-2).



3.4.3.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

Chlorophyll a concentrations in Leftrook Lake ranged from 0.78 to 34.0 μ g/L during the openwater season. The mean and median for the 11 years of monitoring were 13.4 μ g/L and 10.3 μ g/L, respectively. Open-water season mean annual chlorophyll a concentrations ranged from 6.37 to 20.6 μ g/L and were within the IQR (5.35 to 21.3 μ g/L) in all years (Table 3.4-3 and Figure 3.4-8).

Chlorophyll a concentrations in the ice-cover season ranged from <0.60 to 1.53 μ g/L, with a mean of 1.03 μ g/L and median of 1.15 μ g/L for the 11 years of monitoring. The IQR was 0.85 to 1.42 μ g/L (Table 3.4-3 and Figure 3.4-8).

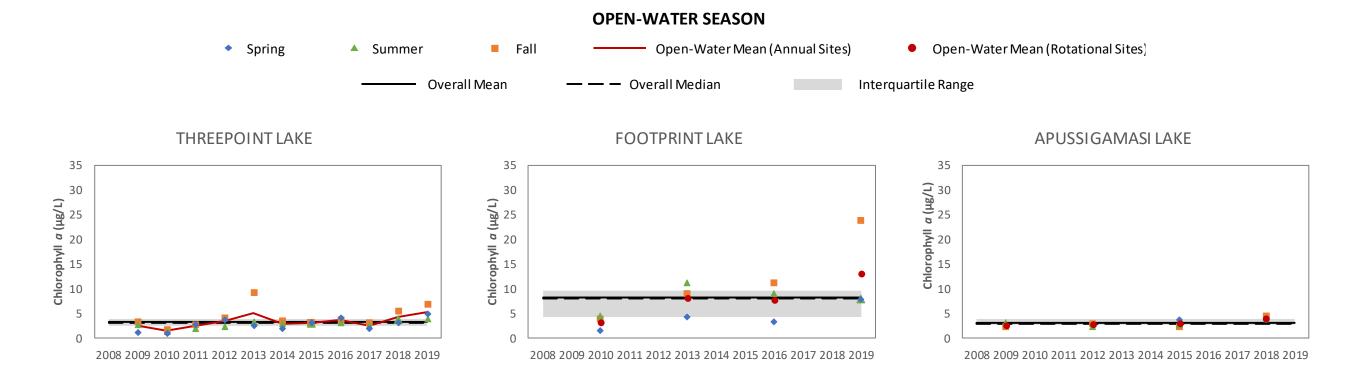
On average, chlorophyll a concentrations were lower in winter (1.03 μ g/L) and spring (4.30 μ g/L) than in summer (13.4 μ g/L) and fall (22.5 μ g/L; Figure 3.4-4).

Leftrook Lake was eutrophic (8 to 25 μ g/L) on the basis of the 2009-2019 mean open-water season chlorophyll a concentration (13.4 μ g/L). Mean annual chlorophyll a concentrations (6.37 to 20.6 μ g/L) in the open-water season were also within the mesotrophic range in nine of the 11 years. However, the mean chlorophyll a concentration was within the mesotrophic range (2.5 to 8 μ g/L) in 2009 and 2016 (Table 3.4-4).

ROTATIONAL SITES

There are no off-system rotational sites in this region.





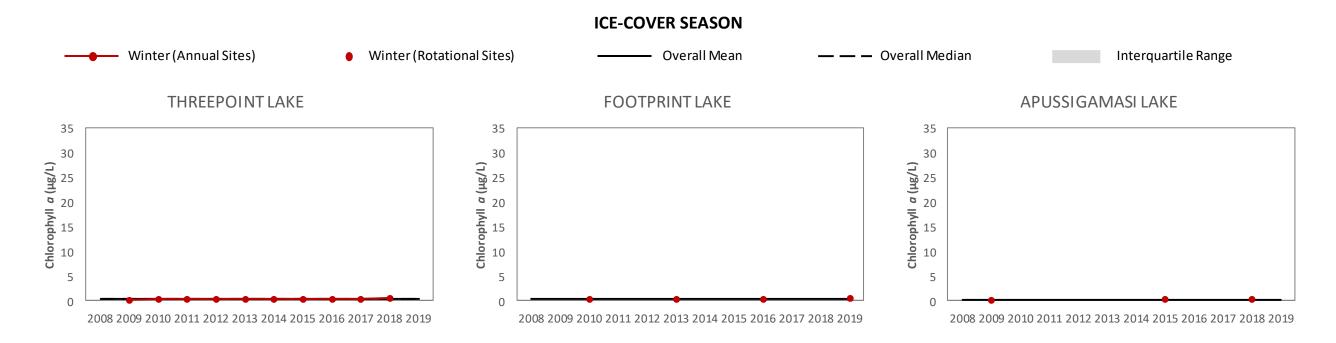


Figure 3.4-7. 2008-2019 On-system open-water and ice-cover season chlorophyll *a* concentrations.



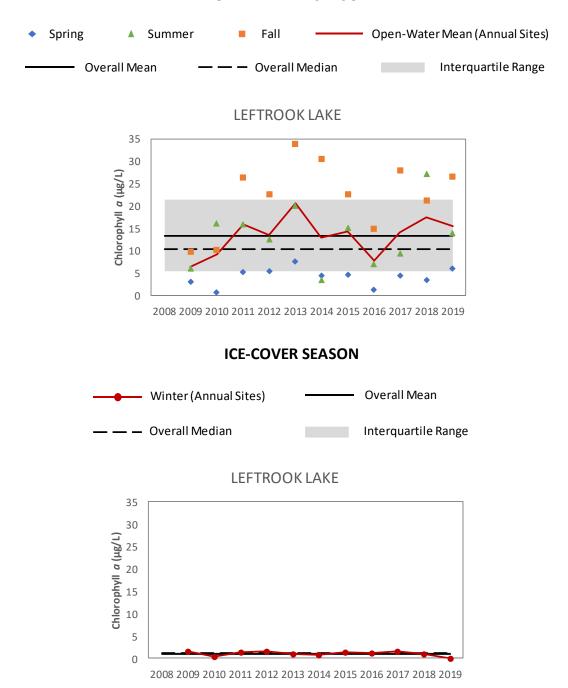


Figure 3.4-8. 2008-2019 Off-system open-water and ice-cover season chlorophyll *a* concentrations.



APPENDIX 3-1. WATER QUALITY SAMPLING SITES: 2008-2019



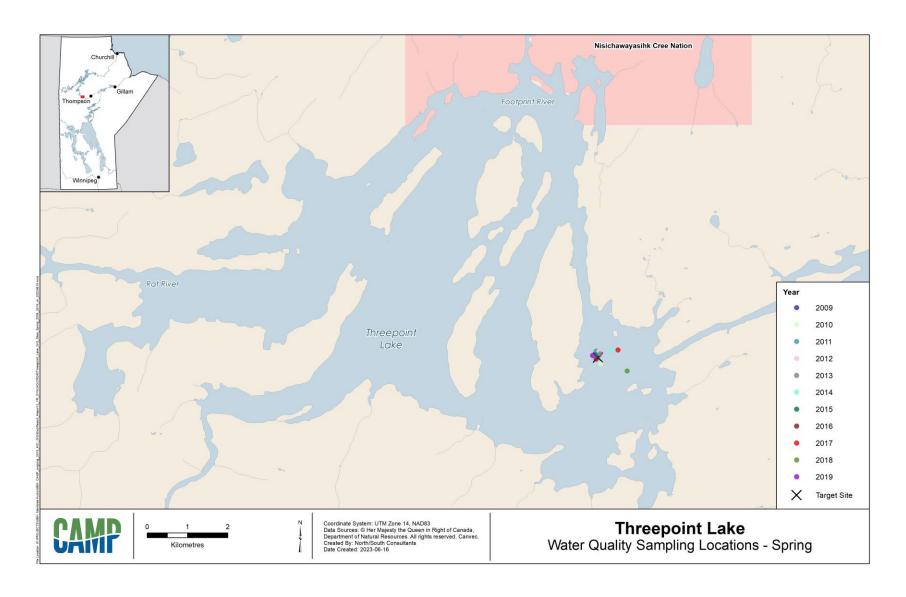


Figure A3-1-1. Spring water quality sampling locations: Threepoint Lake.



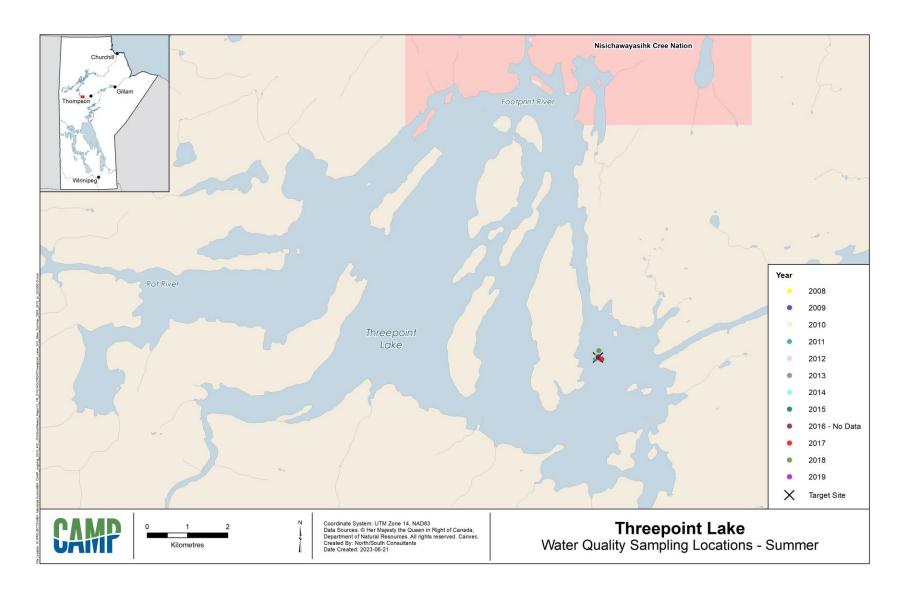


Figure A3-1-2. Summer water quality sampling locations: Threepoint Lake.



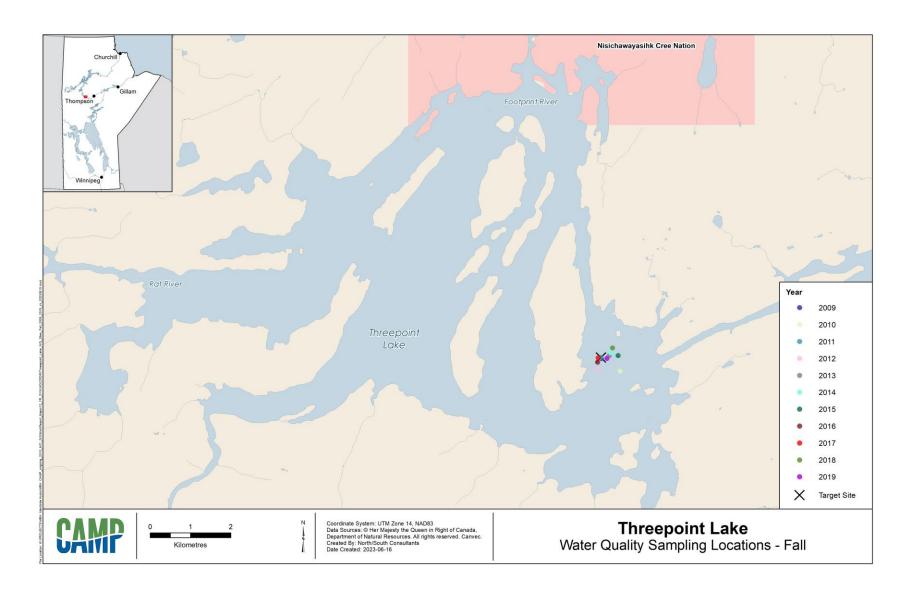


Figure A3-1-3. Fall water quality sampling locations: Threepoint Lake.



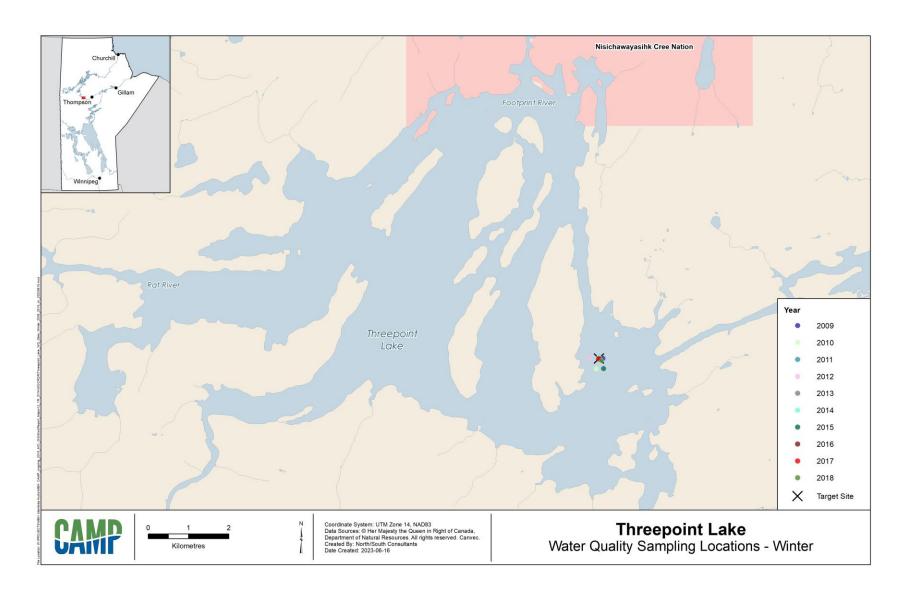


Figure A3-1-4. Winter water quality sampling locations: Threepoint Lake.



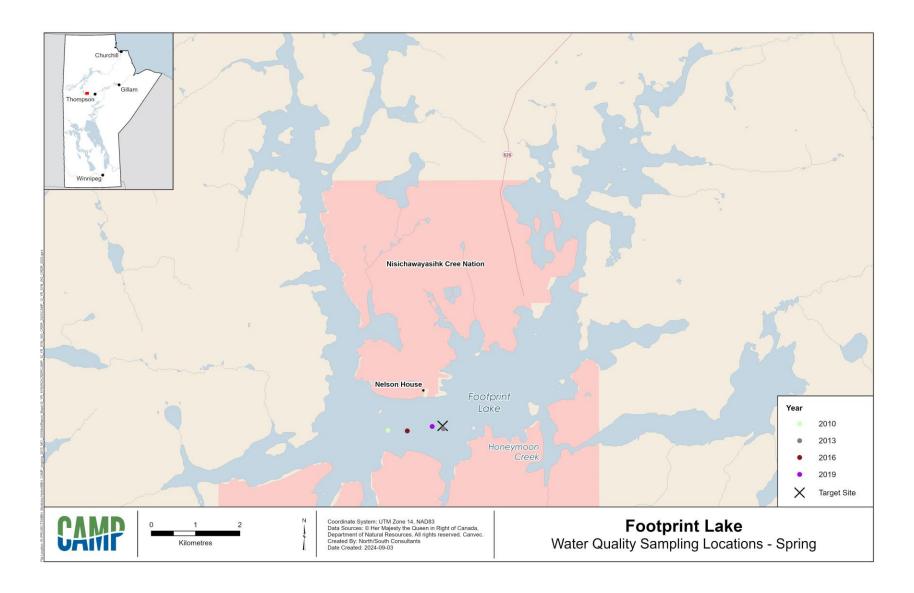


Figure A3-1-5. Spring water quality sampling locations: Footprint Lake.



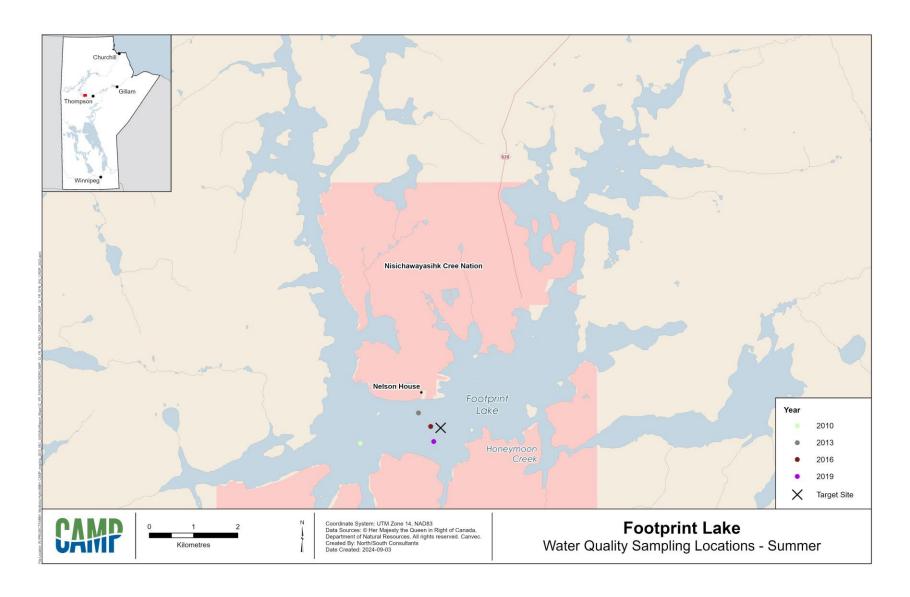


Figure A3-1-6. Summer water quality sampling locations: Footprint Lake.



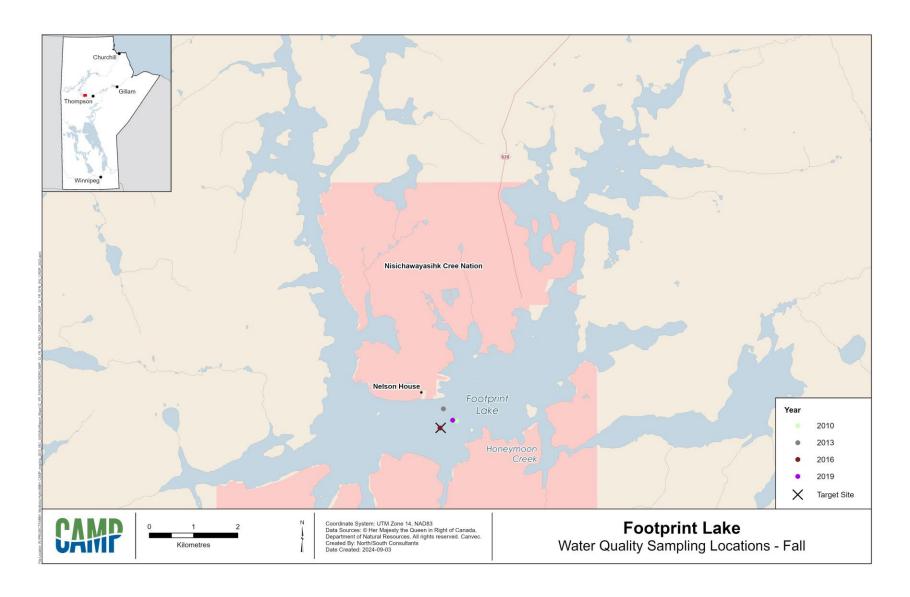


Figure A3-1-7. Fall water quality sampling locations: Footprint Lake.



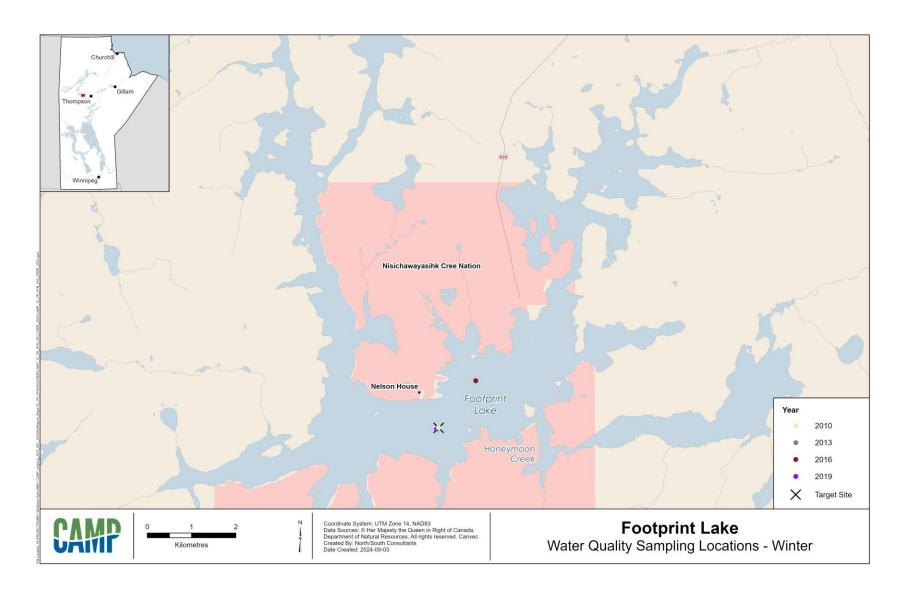


Figure A3-1-8. Winter water quality sampling locations: Footprint Lake.



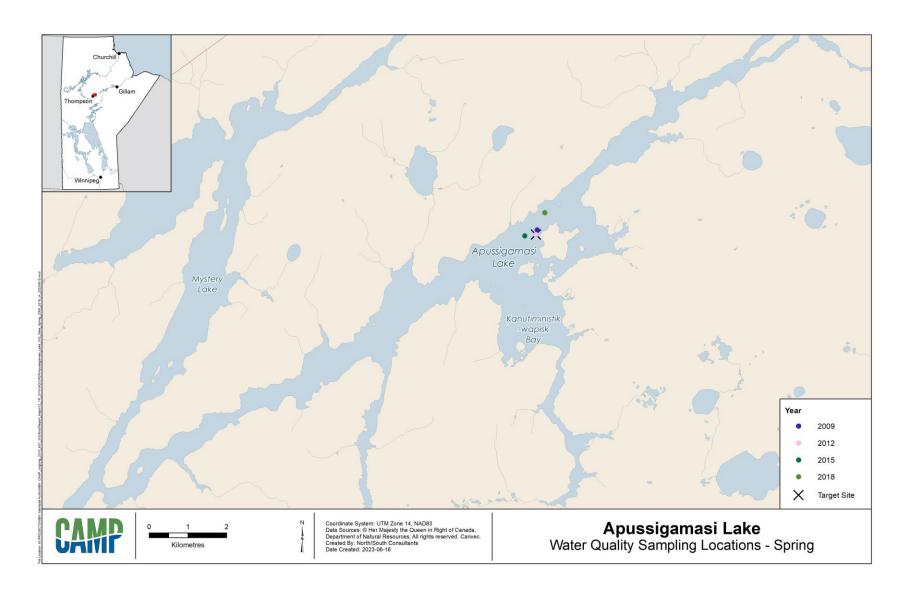


Figure A3-1-9. Spring water quality sampling locations: Apussigamasi Lake.



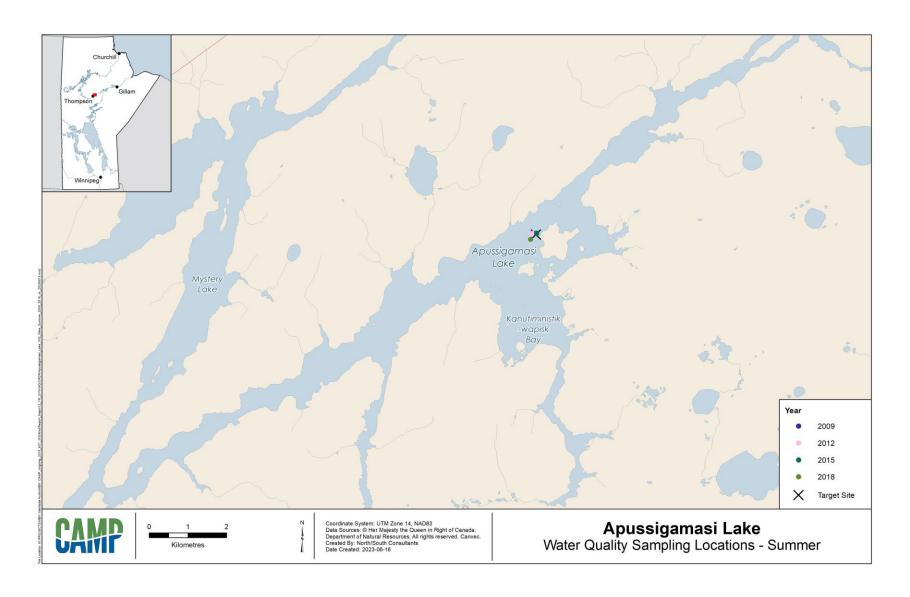


Figure A3-1-10. Summer water quality sampling locations: Apussigamasi Lake.



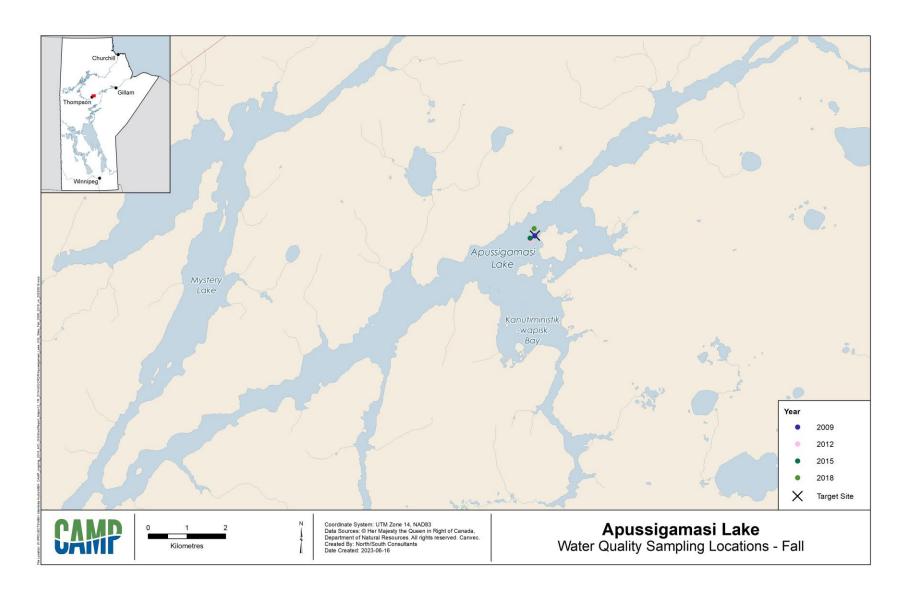


Figure A3-1-11. Fall water quality sampling locations: Apussigamasi Lake.



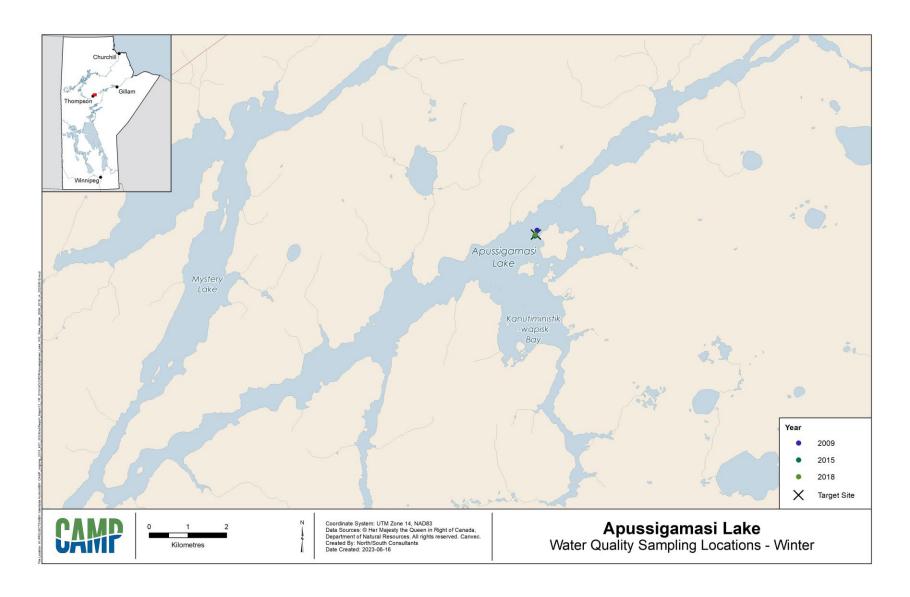


Figure A3-1-12. Winter water quality sampling locations: Apussigamasi Lake.



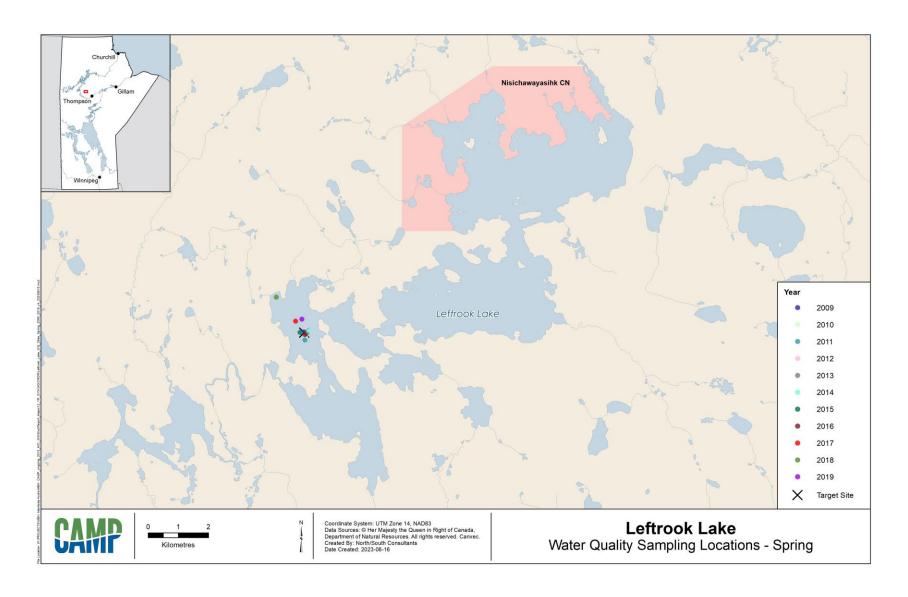


Figure A3-1-13. Spring water quality sampling locations: Leftrook Lake.



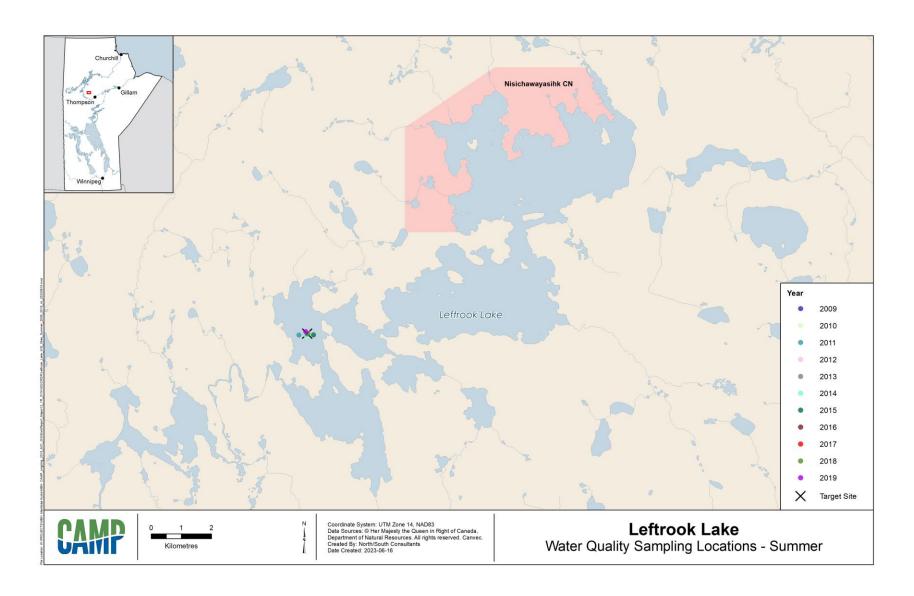


Figure A3-1-14. Summer water quality sampling locations: Leftrook Lake.



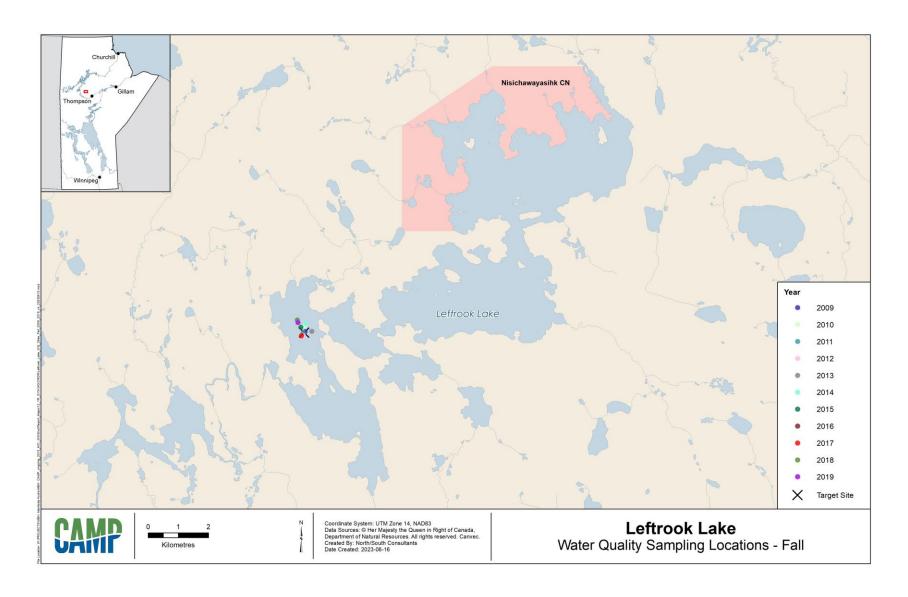


Figure A3-1-15. Fall water quality sampling locations: Leftrook Lake.



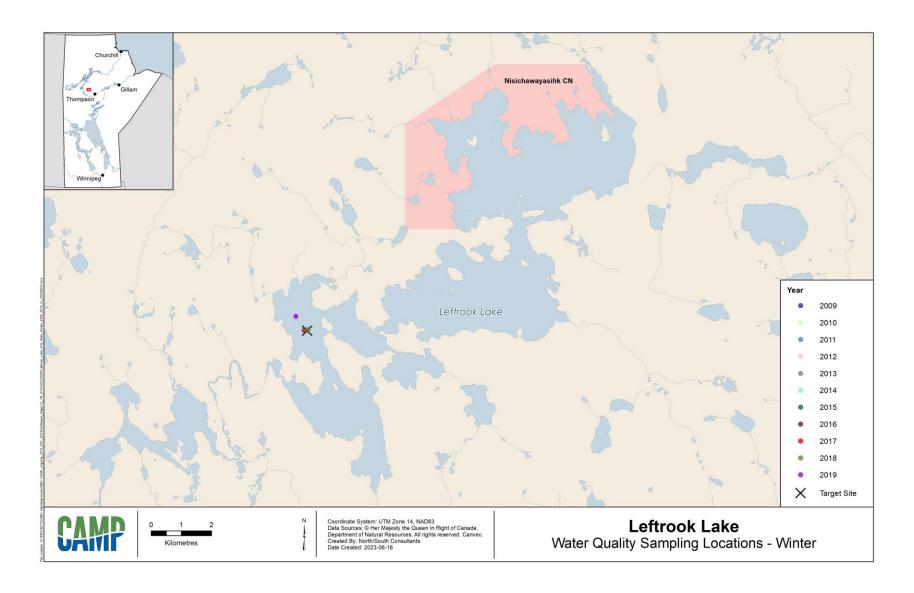


Figure A3-1-16. Winter water quality sampling locations: Leftrook Lake.



4.0 BENTHIC INVERTEBRATES

4.1 INTRODUCTION

The following presents the results of the benthic invertebrate community monitoring conducted from 2010-2019 in the Churchill River Diversion Region. Annual monitoring in this region began in 2009; the 2009 benthic invertebrate dataset was excluded due to a significant change in the sampling design in 2010.

Four waterbodies were monitored in the Churchill River Diversion Region: one on-system annual site (Threepoint Lake), two on-system rotational sites (Footprint Lake and Apussigamasi Lake), and one off-system annual site (Leftrook Lake; Table 4.1-1 and Figure 4.1-1).

Two sampling polygons (nearshore [NS] and offshore [OS]) defined by water depth, flow, and substrate composition were sampled in each waterbody in late summer/fall per year (Appendix 4-1). Five benthic invertebrate samples were collected in each polygon for a total of ten invertebrate samples per waterbody per year. Five sediment samples were also collected in each polygon (where possible) to provide supporting information on substrate composition, total organic carbon (TOC), and texture. Dominant substrate type(s) and sediment analysis results are presented in Appendix 4-2. Sampling was completed at all sites as planned over the period of 2010-2019.

Four benthic invertebrate indicators (abundance, community composition, taxonomic richness, and diversity) were selected for detailed reporting (Table 4.1-2). Metrics for these indicators that are presented herein include: total invertebrate abundance or total invertebrate density; the Ephemeroptera, Plecoptera, and Trichoptera (EPT) Index; the Oligochaeta and Chironomidae (O+C) Index; total taxa richness; EPT taxa richness; and Hill's effective richness (Hill's Index). A detailed description of these indicators is provided in CAMP (2024).

A detailed description of the program design and sampling methods are provided Technical Document 1, Section 2.4.



Table 4.1-1. 2010 to 2019 Benthic invertebrate sampling inventory.

| Waterbody/ | Sampling Year | | | | | | | | | | | |
|------------|---------------|--------|------|------|------|------|------|------|------|------|------|------|
| Area | 2008 | 2009 ¹ | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
| 3РТ | | - | • | • | • | • | • | • | • | • | • | • |
| FOOT | | | • | | | • | | | • | | | • |
| APU | | - | | | • | | | • | | | • | |
| LEFT | | - | • | • | • | • | • | • | • | • | • | • |

Notes:

Table 4.1-2. Benthic invertebrate indicators and metrics.

| Indicator | Metric | Units | | | | |
|--------------------------|--|---------------------------|--|--|--|--|
| Abundance | Total Invertebrate Abundance | Number (no.) per sample | | | | |
| Abundance | Total Invertebrate Density | no. per square metre (m²) | | | | |
| | Relative Proportions of Major Invertebrate Group | | | | | |
| Community Composition | EPT Index | % | | | | |
| Composition | O+C Index | % | | | | |
| Taxonomic | Total Taxa Richness | no. of families | | | | |
| Richness | EPT Taxa Richness | no. of families | | | | |
| Diversity | Hill's Effective Richness (Hill's Index) | value | | | | |



^{1.} Dataset excluded from analysis and reporting due to change in sampling design in 2010.

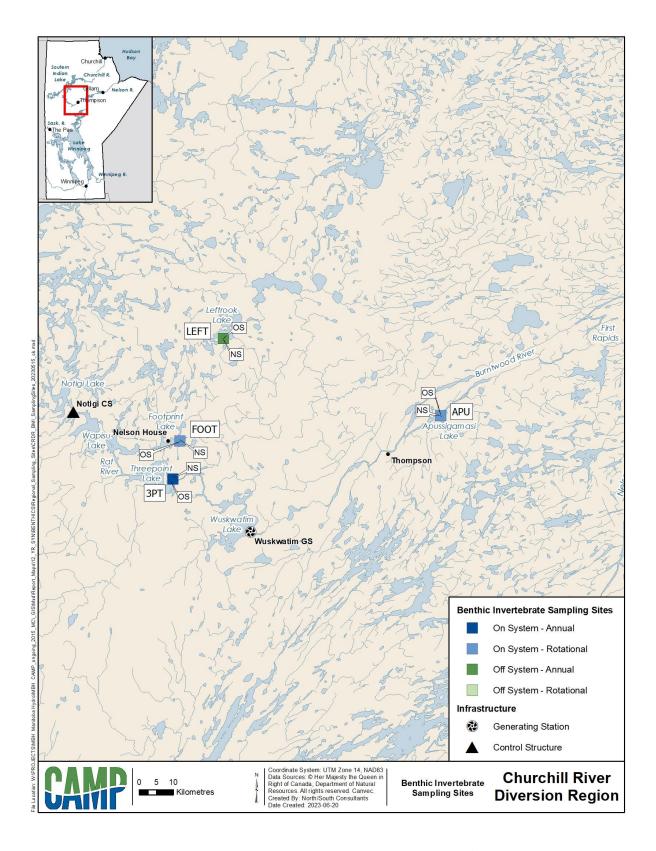


Figure 4.1-1. 2010-2019 Benthic invertebrate nearshore (NS) and offshore (OS) sampling sites.



4.2 ABUNDANCE

4.2.1 TOTAL INVERTEBRATE ABUNDANCE

4.2.1.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Nearshore Habitat

Annual mean abundance over the ten years of monitoring ranged from 138 invertebrates per sample (2010) to 387 invertebrates per sample (2016; Figure 4.2-1). The overall mean abundance was 247 invertebrates per sample, the overall median abundance was 214 invertebrates per sample, and the IQR was 131 to 334 invertebrates per sample. Annual means were within the IQR, except for 2015 and 2016 (above).

Offshore Habitat

Annual mean abundance (density) over the ten years of monitoring ranged from 848 invertebrates per m² (2011) to 7,846 invertebrates per m² (2014; Figure 4.2-2). The overall mean abundance was 4,777 invertebrates per m², the overall median abundance was 5,158 invertebrates per m², and the IQR was 2,889 to 6,925 invertebrates per m². Annual means were below the IQR in 2010, 2011, and 2012, and above the IQR in 2014, 2015, and 2019.

ROTATIONAL SITES

Footprint Lake

Nearshore Habitat

Annual mean abundance over the four years of monitoring ranged from 104 invertebrates per sample (2010) to 795 invertebrates per sample (2013; Figure 4.2-1). The overall mean abundance was 463 invertebrates per sample, the overall median abundance was 395 invertebrates per sample, and the IQR was 177 to 861 invertebrates per sample. Annual means were within the IQR, except for 2010 (below).



Offshore Habitat

Annual mean abundance (density) over the four years of monitoring ranged from 678 invertebrates per m² (2010) to 2,897 invertebrates per m² (2019; Figure 4.2-2). The overall mean abundance was 1,802 invertebrates per m², the overall median abundance was 1,666 invertebrates per m², and the IQR was 927 to 2,543 invertebrates per m². Annual means were below the IQR in 2010, and above the IQR in 2013 and 2019.

Apussigamasi Lake

Nearshore Habitat

Annual mean abundance over the three years of monitoring ranged from 87 invertebrates per sample (2012) to 435 invertebrates per sample (2018; Figure 4.2-1). The overall mean abundance was 230 invertebrates per sample, the overall median abundance was 165 invertebrates per sample, and the IQR was 96 to 319 invertebrates per sample. Annual means were below the IQR in 2012, and above the IQR in 2018.

Offshore Habitat

Annual mean abundance (density) over the three years of monitoring ranged from 1,803 invertebrates per m² (2015) to 5,944 invertebrates per m² (2018; Figure 4.2-2). The overall mean abundance was 3,473 invertebrates per m², the overall median abundance was 2,611 invertebrates per m², and the IQR was 779 to 5,749 invertebrates per m². Annual means were within the IQR, except for 2018 (above).

4.2.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

Nearshore Habitat

Annual mean abundance over the ten years of monitoring ranged from 453 invertebrates per sample (2010) to 6,323 invertebrates per sample (2013; Figure 4.2-1). The overall mean abundance was 2,498 invertebrates per sample, the overall median abundance was 1,803 invertebrates per sample, and the IQR was 860 to 3,695 invertebrates per sample. Annual means were below the IQR in 2010, and above the IQR in 2013 and 2018.



Offshore Habitat

Annual mean abundance (density) over the ten years of monitoring ranged from 1,627 invertebrates per m² (2012) to 13,109 invertebrates per m² (2015; Figure 4.2-2). The overall mean abundance was 6,650 invertebrates per m², the overall median abundance was 5,497 invertebrates per m², and the IQR was 3,275 to 9,263 invertebrates per m². Annual means were below the IQR in 2010 and 2012, and above the IQR in 2013, 2015, and 2018.



LEFT

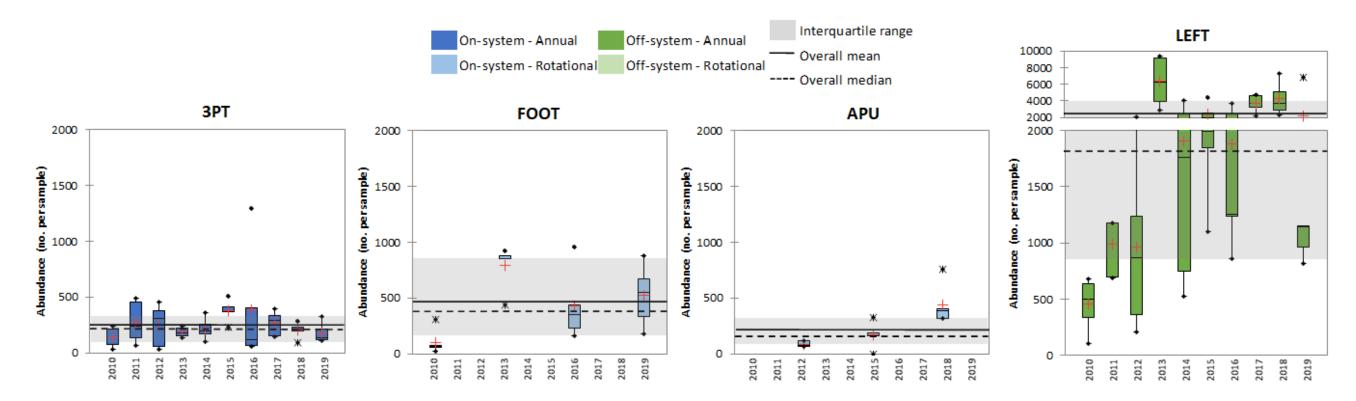


Figure 4.2-1. 2010 to 2019 Nearshore benthic invertebrate abundance (total no. per sample).

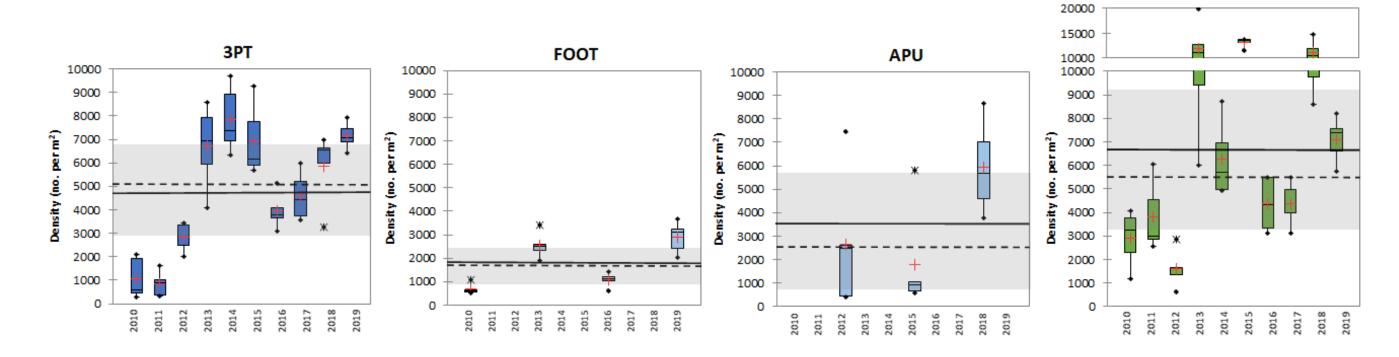


Figure 4.2-2. 2010 to 2019 Offshore benthic invertebrate abundance (density; total no. per m²).



4.3 COMMUNITY COMPOSITION

4.3.1 RELATIVE ABUNDANCE

4.3.1.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Nearshore Habitat

Benthic invertebrate community composition varied over the ten years of monitoring (2010 to 2019; Table 4.3-1). Corixidae (water boatmen) was the dominant taxon in 2010 (54%), 2015 (48%), and 2018 (41%). Chironomidae (non-biting midges, 29%) and Corixidae (31%) were the dominant groups in 2011. Oligochaeta (aquatic segmented worms, 37%) and Corixidae (26%) dominated in 2012. Oligochaeta (27%), Amphipoda (freshwater shrimps, mainly Hyalellidae, 27%) and Chironomidae (26%) were nearly co-dominant in 2013. Amphipoda (mainly Hyalellidae) was the most abundant taxon in 2014 (73%), 2016 (40%), and 2019 (59%). Oligochaeta (28%) and Chironomidae (31%) were nearly co-dominant in 2017.

Offshore Habitat

Bivalvia (mainly Sphaeriidae, fingernail clams) dominated the benthic invertebrate community over the ten years of monitoring (2010 to 2019; Table 4.3-2). Among those years, mean annual relative abundance of Bivalvia ranged from 48% (2011) to 69% (2019). Aside from 2018, Amphipoda (freshwater shrimps, mainly Pontoporeiidae) was the second most abundant group, with mean relative abundance ranging between 15% (2017) to 39% (2015). The second most abundant taxon in 2018 was Gastropoda (snails, mainly Hydrobiidae, 17%).

ROTATIONAL SITES

Footprint Lake

Nearshore Habitat

Benthic invertebrate community composition varied over the four years of monitoring (2010, 2013, 2016, and 2019; Table 4.3-3). Corixidae (water boatmen) was the dominant taxon in 2010 (49%) and 2019 (44%). Chironomidae (non-biting midges) was the dominant taxon in 2013 (38%).



Amphipoda (freshwater shrimps, mainly Hyalellidae, 26%) and Chironomidae (31%) dominated the invertebrate community in 2016.

Offshore Habitat

Benthic invertebrate community composition varied over the four years of monitoring (2010, 2013, 2016, and 2019; Table 4.3-4). One taxon within the "Other Diptera" category (Chaoboridae, phantom midges, 54%) was most abundant in 2010. Amphipoda (freshwater shrimps, Pontoporeiidae) was dominant in 2013 (68%). Amphipoda (41%) and Chironomidae (36%) (non-biting midges) were dominated in 2016. Amphipoda (46%) and Chironomidae (44%) were nearly co-dominant in 2019.

Apussigamasi Lake

Nearshore Habitat

Benthic invertebrate community composition varied over the three years of monitoring (2012, 2015, and 2018; Table 4.3-5). Amphipoda (freshwater shrimps, mainly Hyalellidae) was dominant in 2012 (36%) and 2018 (40%). Oligochaeta (aquatic segmented worms) dominated in 2015 (26%).

Offshore Habitat

Bivalvia (Sphaeriidae, fingernail clams) dominated the benthic invertebrate community in all three years of monitoring (2012, 2015, and 2018; Table 4.3-6). Among those years, mean annual relative abundance of Bivalvia ranged from 60% (2012) to 74% (2018). Amphipoda (freshwater shrimps, Pontoporeiidae) was the second most relatively abundant taxon in 2012 (21%), 2015 (26%), and 2018 (13%).

4.3.1.2 OFF-SYSTEM SITES

ANNUAL SITES

<u>Leftrook Lake</u>

Nearshore Habitat

Benthic invertebrate community composition varied over the ten years of monitoring (2010 to 2019; Table 4.3-7). Amphipoda (freshwater shrimps, Hyalellidae) dominated in 2010 (55%). Oligochaeta (aquatic segmented worms) was dominant in 2011 (44%), 2012 (27%), 2014 (37%), and 2018 (49%). Corixidae (water boatmen) dominated in 2013 (28%). Gastropoda (snails, 28%,



mainly Lymnaeidae) and Corixidae (25%) were nearly co-dominant in 2015. Ephemeroptera (mayflies, 42%, mainly Caenidae) was the dominant taxon in 2016. Oligochaeta (22%), Amphipoda (mainly Hyalellidae, 17%), and Bivalvia (mainly Sphaeriidae, 16%) dominated in 2017. Gastropoda was the dominant taxon in 2019 (52%, mainly Lymnaeidae and Planorbidae).

Offshore Habitat

Bivalvia (Sphaeriidae, fingernail clams) dominated the benthic invertebrate community in nine of the ten years of monitoring (2010 to 2014 and 2016 to 2019; Table 4.3-8). Of those years, mean annual relative abundance of Bivalvia ranged from 55% (2013 and 2019) to 74% (2011). Bivalvia (Sphaeriidae, 40%) and Chironomidae (non-biting midges, 38%) were nearly co-dominant in 2015.



2024

Table 4.3-1. 2010 to 2019 Threepoint Lake nearshore benthic invertebrate relative abundance.

| Invertebrate | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
|-----------------|------|------|------|------|------|------|------|------|------|------|
| Таха | | | | | | | | | | |
| Oligochaeta | 3% | 18% | 37% | 27% | 6% | 10% | 17% | 28% | 19% | 5% |
| Amphipoda | 5% | 12% | 21% | 27% | 73% | 12% | 40% | 24% | 24% | 59% |
| Bivalvia | 3% | 0% | <1% | <1% | 1% | <1% | 0% | <1% | <1% | 0% |
| Gastropoda | 10% | <1% | <1% | 1% | 2% | 2% | 1% | <1% | <1% | 2% |
| Ceratopogonidae | <1% | 1% | 1% | <1% | 1% | 1% | <1% | 0% | <1% | <1% |
| Chironomidae | 16% | 29% | 6% | 26% | 5% | 19% | 25% | 31% | 7% | 4% |
| Other Diptera | 1% | 1% | 1% | <1% | <1% | 2% | <1% | 1% | 1% | 1% |
| Ephemeroptera | 5% | 6% | 4% | 10% | 5% | 1% | 5% | 5% | 3% | 3% |
| Plecoptera | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Trichoptera | 1% | <1% | 1% | <1% | <1% | 1% | <1% | 1% | 1% | 1% |
| Corixidae | 54% | 31% | 26% | 7% | 5% | 48% | 10% | 11% | 41% | 22% |
| Coleoptera | 1% | 2% | 1% | 1% | 1% | 2% | 1% | <1% | 2% | 1% |
| All other taxa | <1% | <1% | 1% | 1% | <1% | 1% | <1% | <1% | 1% | 1% |

Table 4.3-2. 2010 to 2019 Threepoint Lake offshore benthic invertebrate relative abundance.

| Invertebrate Taxa | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
|----------------------|------|------|------|------|------|------|------|------|------|------|
| Oligochaeta | 2% | 3% | 1% | 2% | 1% | 1% | 1% | 1% | <1% | 1% |
| Amphipoda | 29% | 24% | 16% | 31% | 33% | 39% | 36% | 15% | 9% | 21% |
| Bivalvia | 52% | 48% | 62% | 51% | 61% | 55% | 51% | 65% | 65% | 69% |
| Gastropoda | 1% | 1% | 13% | 7% | 1% | <1% | 1% | 7% | 17% | 5% |
| Ceratopogonidae | 1% | 1% | <1% | <1% | <1% | <1% | <1% | <1% | 1% | <1% |
| Chironomidae | 7% | 9% | 4% | 7% | 2% | 2% | 5% | 6% | 5% | 2% |
| Other Diptera | 0% | 0% | 0% | 0% | 0% | <1% | 0% | 0% | 0% | 0% |
| Ephemeroptera | 4% | 12% | 2% | 2% | 2% | 1% | 6% | 5% | 2% | 1% |
| Plecoptera | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | <1% | 0% |
| Trichoptera | 4% | 3% | 2% | <1% | <1% | <1% | 1% | 1% | 2% | <1% |
| Corixidae | 0% | 0% | <1% | 0% | <1% | 0% | <1% | <1% | 0% | <1% |
| Coleoptera | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| All other taxa | 1% | 0% | <1% | 0% | <1% | <1% | <1% | 0% | <1% | <1% |



Table 4.3-3. 2010 to 2019 Footprint Lake nearshore benthic invertebrate relative abundance.

| Invertebrate Taxa | 2010 | 2013 | 2016 | 2019 |
|----------------------|------|------|------|------|
| Oligochaeta | 8% | 13% | 14% | 10% |
| Amphipoda | 23% | 17% | 26% | 20% |
| Bivalvia | 1% | <1% | 0% | <1% |
| Gastropoda | 2% | 1% | 1% | 1% |
| Ceratopogonidae | 1% | 1% | 1% | 4% |
| Chironomidae | 7% | 38% | 31% | 11% |
| Other Diptera | <1% | 2% | <1% | 2% |
| Ephemeroptera | 5% | 4% | 1% | 1% |
| Plecoptera | 0% | 0% | 0% | 0% |
| Trichoptera | 1% | 1% | 1% | 2% |
| Corixidae | 49% | 12% | 23% | 44% |
| Coleoptera | 2% | 3% | 1% | 1% |
| All other taxa | 1% | 7% | 1% | 3% |

Table 4.3-4. 2010 to 2019 Footprint Lake offshore benthic invertebrate relative abundance.

| Invertebrate Taxa | 2010 | 2013 | 2016 | 2019 |
|----------------------|------|------|------|------|
| Oligochaeta | <1% | 4% | 5% | 1% |
| Amphipoda | 15% | 68% | 41% | 46% |
| Bivalvia | 2% | 4% | 10% | 5% |
| Gastropoda | 0% | <1% | 1% | <1% |
| Ceratopogonidae | 4% | <1% | <1% | 1% |
| Chironomidae | 20% | 18% | 36% | 44% |
| Other Diptera | 54% | 1% | 2% | 3% |
| Ephemeroptera | 5% | 4% | 1% | 1% |
| Plecoptera | 0% | 0% | 0% | 0% |
| Trichoptera | 0% | <1% | 1% | 0% |
| Corixidae | 0% | 0% | 0% | 0% |
| Coleoptera | 0% | 0% | 0% | 0% |
| All other taxa | 0% | <1% | 3% | <1% |



Table 4.3-5. 2010 to 2019 Apussigamasi Lake nearshore benthic invertebrate relative abundance.

| 0% <1% to 15% >15% to 25% >25% to 50% >50% | 0% | <1% to 15% | >15% to 25% | >25% to 50% | >50% |
|--|----|------------|-------------|-------------|------|
|--|----|------------|-------------|-------------|------|

| Invertebrate Taxa | 2012 | 2015 | 2018 |
|----------------------|------|------|------|
| Oligochaeta | 10% | 26% | 14% |
| Amphipoda | 36% | 12% | 40% |
| Bivalvia | 0% | <1% | 0% |
| Gastropoda | 3% | 3% | 2% |
| Ceratopogonidae | <1% | 6% | 2% |
| Chironomidae | 14% | 18% | 12% |
| Other Diptera | 0% | <1% | <1% |
| Ephemeroptera | 4% | 12% | 13% |
| Plecoptera | 0% | <1% | 0% |
| Trichoptera | 3% | 1% | 3% |
| Corixidae | 23% | 12% | 13% |
| Coleoptera | 3% | 3% | <1% |
| All other taxa | 5% | 6% | 1% |

Table 4.3-6. 2010 to 2019 Apussigamasi Lake offshore benthic invertebrate relative abundance.

| Invertebrate Taxa | 2012 | 2015 | 2018 |
|----------------------|------|------|------|
| Oligochaeta | 0% | 0% | 0% |
| Amphipoda | 21% | 26% | 13% |
| Bivalvia | 60% | 63% | 74% |
| Gastropoda | 7% | 1% | 4% |
| Ceratopogonidae | 1% | 1% | <1% |
| Chironomidae | 4% | 3% | 2% |
| Other Diptera | 0% | 0% | 0% |
| Ephemeroptera | 5% | 5% | 3% |
| Plecoptera | 0% | 0% | 0% |
| Trichoptera | 3% | 1% | 2% |
| Corixidae | 0% | 0% | 0% |
| Coleoptera | 0% | 0% | 0% |
| All other taxa | <1% | 0% | <1% |



Table 4.3-7. 2010 to 2019 Leftrook Lake nearshore benthic invertebrate relative abundance.

| 0% | <1% to 15% | >15% to 25% | >25% to 50% | >50% |
|----|------------|-------------|-------------|------|
| | | | | |

| Invertebrate Taxa | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
|----------------------|------|------|------|------|------|------|------|------|------|------|
| Oligochaeta | 12% | 44% | 27% | 13% | 37% | 21% | 13% | 22% | 49% | 7% |
| Amphipoda | 55% | 14% | 13% | 18% | 17% | 9% | 17% | 17% | 4% | 11% |
| Bivalvia | 6% | 1% | 3% | 6% | 2% | 1% | 4% | 16% | 4% | 6% |
| Gastropoda | 3% | 9% | 17% | 12% | 10% | 28% | 4% | 11% | 8% | 52% |
| Ceratopogonidae | 0% | <1% | <1% | 0% | <1% | 0% | 0% | <1% | 0% | 0% |
| Chironomidae | 5% | 12% | 22% | 6% | 15% | 9% | 9% | 13% | 7% | 7% |
| Other Diptera | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Ephemeroptera | 4% | 4% | 3% | 6% | 1% | 1% | 42% | 8% | 12% | 1% |
| Plecoptera | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Trichoptera | 8% | 4% | 4% | 9% | 2% | 4% | 5% | 12% | 8% | 6% |
| Corixidae | 2% | 9% | 10% | 28% | 15% | 25% | 2% | 1% | 5% | 8% |
| Coleoptera | 1% | <1% | 1% | <1% | <1% | <1% | <1% | <1% | <1% | <1% |
| All other taxa | 5% | 1% | 1% | 1% | 1% | 1% | 2% | 1% | 4% | 2% |

Table 4.3-8. 2010 to 2019 Leftrook Lake offshore benthic invertebrate relative abundance.

| Invertebrate Taxa | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
|----------------------|------|------|------|------|------|------|------|------|------|------|
| Oligochaeta | 7% | 11% | 6% | 6% | 14% | 21% | 8% | 10% | 12% | 13% |
| Amphipoda | <1% | 0% | 0% | <1% | <1% | <1% | 0% | <1% | 0% | 0% |
| Bivalvia | 71% | 74% | 57% | 55% | 72% | 40% | 66% | 70% | 65% | 55% |
| Gastropoda | 1% | 1% | 1% | 2% | 2% | 1% | 3% | 4% | 4% | 3% |
| Ceratopogonidae | <1% | 0% | 0% | <1% | 0% | 0% | 0% | <1% | <1% | <1% |
| Chironomidae | 20% | 13% | 35% | 36% | 12% | 38% | 19% | 15% | 17% | 28% |
| Other Diptera | 0% | 0% | 0% | <1% | <1% | 0% | 0% | 0% | 0% | 0% |
| Ephemeroptera | 1% | 1% | <1% | 1% | 0% | 0% | 3% | 0% | <1% | <1% |
| Plecoptera | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Trichoptera | 0% | 0% | 0% | <1% | 0% | 0% | 1% | 0% | <1% | <1% |
| Corixidae | 0% | 0% | 1% | 0% | 0% | 0% | <1% | 0% | 0% | <1% |
| Coleoptera | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| All other taxa | 1% | <1% | <1% | <1% | <1% | <1% | <1% | <1% | 1% | <1% |



4.3.2 EPT INDEX

4.3.2.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Nearshore Habitat

Annual mean EPT Index over the ten years of monitoring ranged from 2% (2015) to 9% (2013; Figure 4.3-1). The overall mean was 5%, the overall median was 3%, and the IQR was less than 2% to less than 7%. Annual means were within the IQR, except for 2013 (above).

Offshore Habitat

Annual mean EPT Index over the ten years of monitoring ranged from 2% (2014, 2015, and 2019) to 15% (2011; Figure 4.3-2). The overall mean was 5%, the overall median was 3%, and the IQR was 2% to less than 7%. Annual means were below the IQR in 2015 and 2019, and above the IQR in 2011 and 2016.

ROTATIONAL SITES

Footprint Lake

Nearshore Habitat

Annual mean EPT Index over the four years of monitoring ranged from 2% (2016 and 2019) to 4% (2013; Figure 4.3-1). The overall mean was 3%, the overall median was 2%, and IQR was less than 1% to 5%. Annual means were within the IQR in all years.

Offshore Habitat

Annual mean EPT Index over the four years of monitoring ranged from 1% (2016 and 2019) to 5% (2010; Figure 4.3-2). The overall mean was 3%, the overall median was 2%, and the IQR was less than 1% to 5%. Annual means were within the IQR in all years.



Nearshore Habitat

Annual mean EPT Index over the three years of monitoring ranged from 7% (2012) to 31% (2015; Figure 4.3-1). The overall mean was 19%, the overall median was 15%, and the IQR was less than 8% to 19%. Annual means were below the IQR in 2012, and above the IQR in 2015.

Offshore Habitat

Annual mean EPT Index over the three years of monitoring ranged from 6% (2018) to 12% (2012; Figure 4.3-2). The overall mean was 9%, the overall median was 7%, and the IQR was 5% to 12%. Annual means were within the IQR in all years.

4.3.2.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

Nearshore Habitat

Annual mean EPT Index over the ten years of monitoring ranged from 4% (2014) to 37% (2016; Figure 4.3-1). The overall mean was 14%, the overall median was 11%, and the IQR was 6% to less than 18%. Annual means were below the IQR in 2014 and 2015, and above the IQR in 2016 and 2017.

Offshore Habitat

Annual mean EPT Index over the ten years of monitoring ranged from 0% (2014, 2015 and 2017) to 3% (2016; Figure 4.3-2). The overall mean and median were less than 1% and IQR was 0% to slightly above 1%. Annual means were above the IQR in 2011 and 2016.



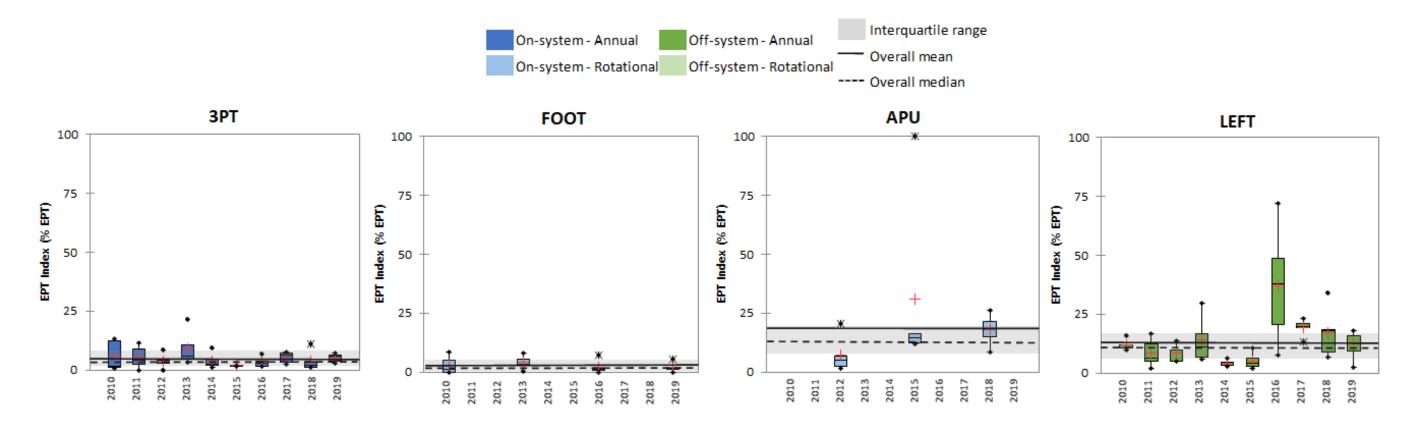


Figure 4.3-1. 2010 to 2019 Nearshore benthic invertebrate EPT Index.

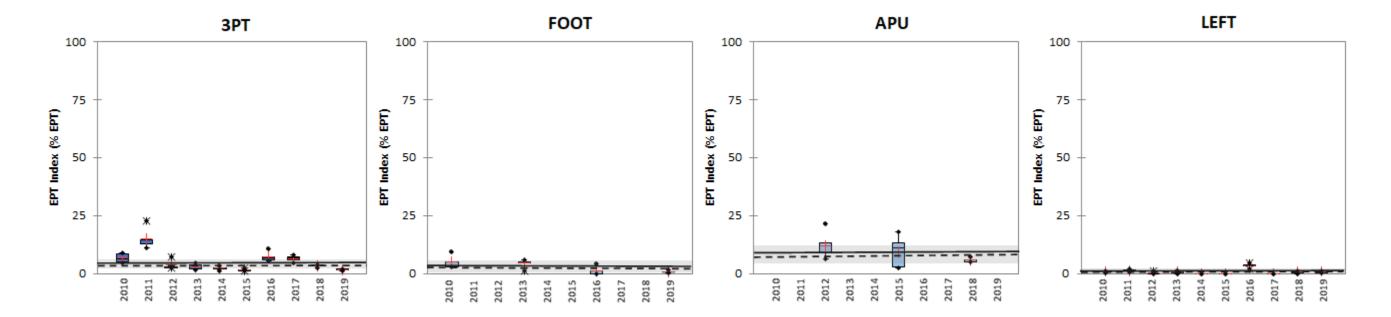


Figure 4.3-2. 2010 to 2019 Offshore benthic invertebrate EPT Index.



4.3.3 O+C INDEX

4.3.3.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Nearshore Habitat

Annual mean O+C Index over the ten years of monitoring ranged from 10% (2019) to 61% (2017; Figure 4.3-3). The overall mean was 35%, the overall median was 34%, and the IQR was 13% to less than 54%. Annual means were below the IQR in 2014 and 2019, and above the IQR in 2013 and 2017.

Offshore Habitat

Annual mean O+C Index over the ten years of monitoring ranged from 3% (2014, 2015, and 2019) to 16% (2011; Figure 4.3-4). The overall mean was 7%, the overall median was 5%, and the IQR was slightly above 3% to less than 8%. Annual means were below the IQR in 2014, 2015 and 2019, and above the IQR in 2010, 2011, and 2013.

ROTATIONAL SITES

Footprint Lake

Nearshore Habitat

Annual mean O+C Index over the four years of monitoring ranged from 21% (2019) to 54% (2013; Figure 4.3-3). The overall mean was 37%, the overall median was 39%, and the IQR was less than 22% to less than 48%. Annual means were below the IQR in 2019, and above the IQR in 2013 and 2016.

Offshore Habitat

Annual mean O+C Index over the four years of monitoring ranged from 23% (2010 and 2013) to 43% (2019; Figure 4.3-4). The overall mean was 33%, the overall median was 29%, and the IQR was 24% to 44%. Annual means were below the IQR in 2010 and 2013.



Nearshore Habitat

Annual mean O+C Index over the three years of monitoring ranged from 20% (2012) to 35% (2015; Figure 4.3-3). The overall mean was 28%, the overall median was 30%, and the IQR was less than 12% to less than 42%. Annual means were within the IQR in all years.

Offshore Habitat

Annual mean O+C Index over the three years of monitoring ranged from 2% (2018) to 6% (2012 and 2015; Figure 4.3-4). The overall mean was 4%, the overall median was 2%, and IQR was less than 2% to 4%. Annual means were above the IQR in 2012 and 2015.

4.3.3.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

Nearshore Habitat

Annual mean O+C Index over the ten years of monitoring ranged from 16% (2019) to 55% (2011 and 2018; Figure 4.3-3). The overall mean and median were 36%, and the IQR was less than 20% to less than 49%. Annual means were below the IQR in 2013 and 2019, and above the IQR in 2011 and 2018.

Offshore Habitat

Annual mean O+C Index over the ten years of monitoring ranged from 24% (2011) to 59% (2015; Figure 4.3-4). The overall mean was 35%, the overall median was 31%, and the IQR was less than 25% to less than 42%. Annual means were below the IQR in 2011 and 2017, and above the IQR in 2012, 2013, and 2015.



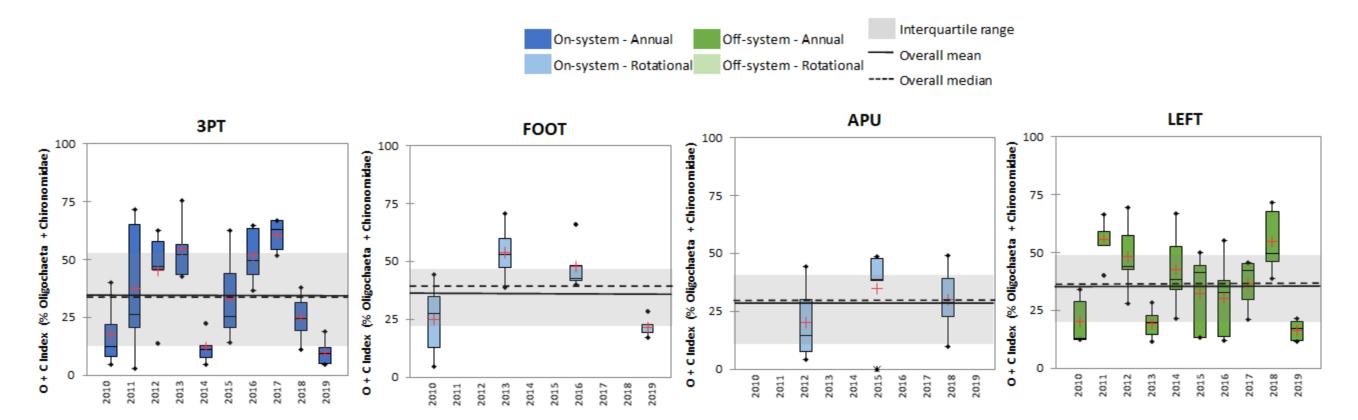


Figure 4.3-3. 2010 to 2019 Nearshore benthic invertebrate O+C Index.

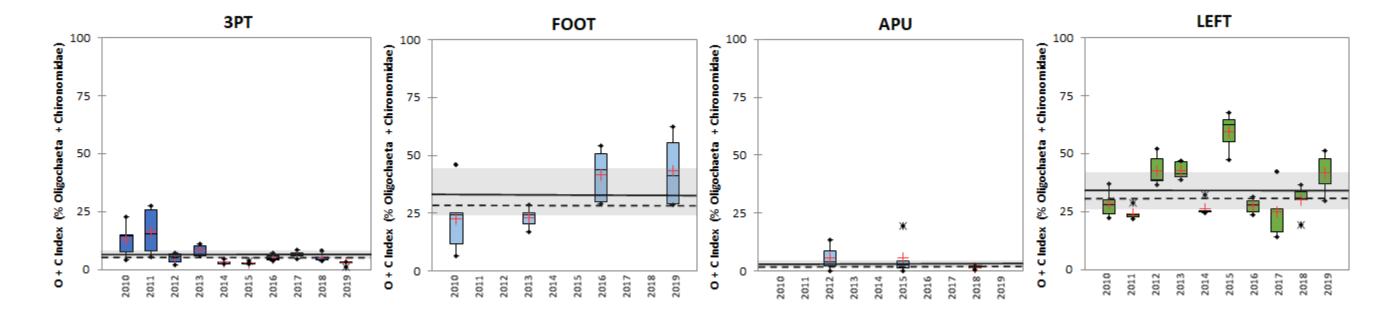


Figure 4.3-4. 2010 to 2019 Offshore benthic invertebrate O+C Index.



4.4 RICHNESS

4.4.1 TOTAL TAXA RICHNESS

4.4.1.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Nearshore Habitat

Annual mean total taxa richness over the ten years of monitoring ranged from ten families (2011, 2012, 2016, and 2017) to 17 families (2015; Figure 4.4-1). The overall mean and median were 12 families, and the IQR was 9 to less than 14 families. Annual means were within the IQR, except for 2015 (above).

Offshore Habitat

Annual mean total taxa richness over the ten years of monitoring ranged from six families (2011) to 11 families (2018; Figure 4.4-2). The overall mean and median were nine families, and the IQR was 8 to 10 families. Annual means were below the IQR in 2010 and 2011, and above the IQR in 2018.

ROTATIONAL SITES

Footprint Lake

Nearshore Habitat

Annual mean total taxa richness over the four years of monitoring ranged from nine families (2010) to 17 families (2013; Figure 4.4-1). The overall mean and median were 13 families, and the IQR was 8 to 17 families. Annual means were within the IQR in all years.

Offshore Habitat

Annual mean total taxa richness over the four years of monitoring ranged from less than six families (2010) to eight families (2019; Figure 4.4-2). The overall mean was seven families, the overall median was six families, and the IQR was 6 to 8 families. Annual means were within the IQR, except in 2010 (below).



Nearshore Habitat

Annual mean total taxa richness over the three years of monitoring ranged from nine families (2012) to 18 families (2018; Figure 4.4-1). The overall mean was 14 families, the overall median was 16 families, and the IQR was less than 10 to 17 families. Annual means were below the IQR in 2012, and above the IQR in 2018.

Offshore Habitat

Annual mean total taxa richness over the three years of monitoring ranged from six families (2012 and 2015) to ten families (2018; Figure 4.4-2). The overall mean and median were seven families, and the IQR was 6 to 9 families. Annual means were within the IQR, except in 2018 (above).

4.4.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

Nearshore Habitat

Annual mean total taxa richness over the ten years of monitoring ranged from 17 families (2012 and 2016) to 21 families (2018; Figure 4.4-1). The overall mean and median were 18 families, and the IQR was 17 to 20 families. Annual means were below the IQR in 2010, 2013, and 2015, and above the IQR in 2017 and 2018.

Offshore Habitat

Annual mean total taxa richness over the ten years of monitoring ranged from less than five families (2012 and 2015) to more than seven families (2013, 2016 and 2018; Figure 4.4-2). The overall mean and median were six families, and the IQR was 5 to 7 families. Annual means were below the IQR in 2012 and 2015, and above the IQR in 2013, 2016, and 2018.



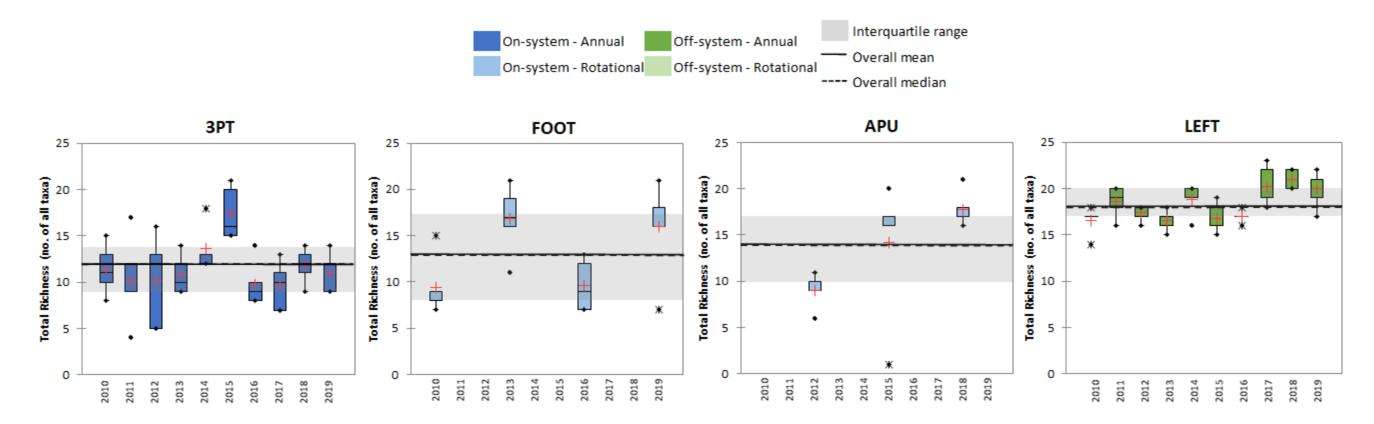


Figure 4.4-1. 2010 to 2019 Nearshore benthic invertebrate total richness (family level).

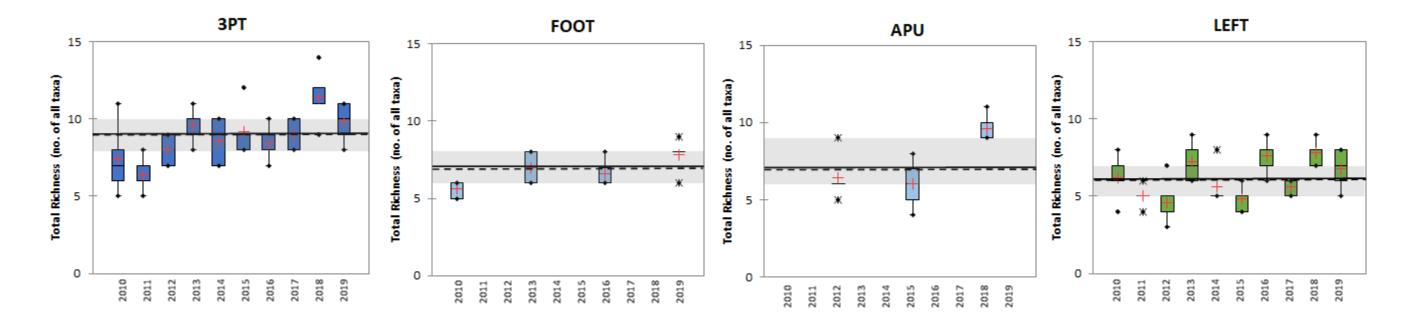


Figure 4.4-2. 2010 to 2019 Offshore benthic invertebrate total richness (family-level).



4.4.2 EPT TAXA RICHNESS

4.4.2.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Nearshore Habitat

Annual mean EPT taxa richness over the ten years of monitoring ranged from two families (2010, 2012, 2013, 2016, and 2017) to four families (2019; Figure 4.4-3). The overall mean and median were three families, and the IQR was 2 to 4 families. Annual means were within the IQR in all years.

Offshore Habitat

Annual mean EPT taxa richness over the ten years of monitoring ranged from one family (2013) to three families (2017, 2018, and 2019; Figure 4.4-4). The overall mean and median were two families, and the IQR was 2 to 3 families. Annual means were below the IQR in 2011 and 2013, and above the IQR in 2018.

ROTATIONAL SITES

Footprint Lake

Nearshore Habitat

Annual mean EPT taxa richness over the four years of monitoring ranged from one family (2010) to five families (2013; Figure 4.4-3). The overall mean and median were three families, and the IQR was 1 to 4 families. Annual means were within the IQR, except in 2013 (above).

Offshore Habitat

Annual mean EPT taxa richness over all four years of monitoring ranged between less than one and slightly greater than one family (2010, 2013, 2016, and 2019; Figure 4.4-4). The overall mean and median were one family, and the IQR was within one family. Annual means were below the IQR in 2016, and above the IQR in 2013.



Nearshore Habitat

Annual mean EPT taxa richness over the three years of monitoring ranged from two families (2012) to seven families (2018; Figure 4.4-3). The overall mean was five families, the overall median was six families, and the IQR was 2 to less than 7 families. Annual means were below the IQR in 2012, and above the IQR in 2018.

Offshore Habitat

Annual mean EPT taxa richness over the three years of monitoring ranged from two families (2012 and 2015) to less than three families (2018; Figure 4.4-4). The overall mean and median were two families, and the IQR was less than 2 to 3 families. Annual means were within the IQR in all years.

4.4.2.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

Nearshore Habitat

Annual mean EPT taxa richness over the ten years of monitoring ranged from five families (2012, 2013, and 2015) to eight families (2017, 2018, and 2019; Figure 4.4-3). The overall mean and median were six families, and the IQR was 5 to 8 families. Annual means were below the IQR in 2015, and above the IQR in 2019.

Offshore Habitat

Annual mean EPT taxa richness over the ten years of monitoring ranged from zero families (2014, 2015, and 2017) to less than two families (2016; Figure 4.4-4). The overall mean and median were one family, and the IQR was between zero and 1 family. Annual means were above the IQR in 2013, 2016, 2018, and 2019.



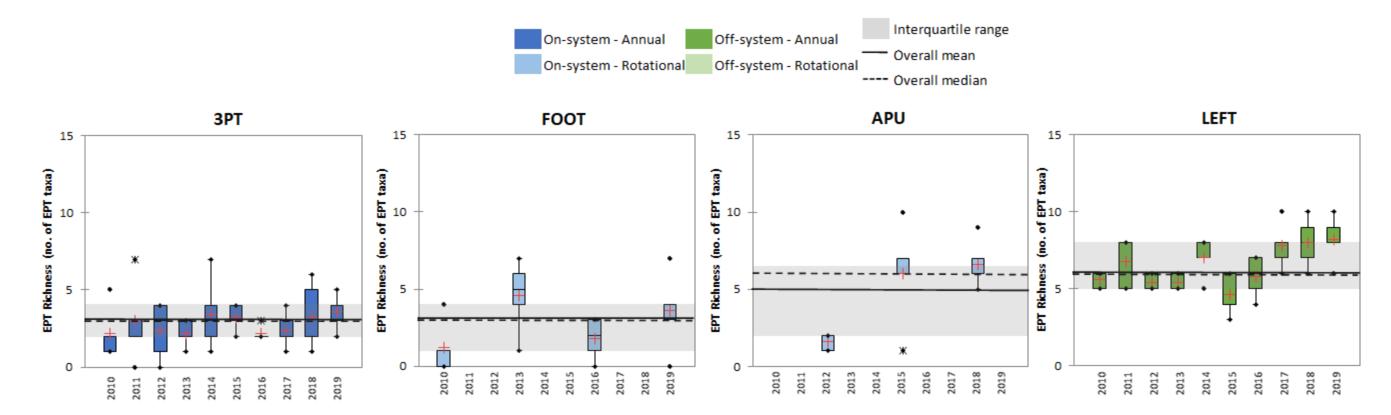


Figure 4.4-3. 2010 to 2019 Nearshore benthic invertebrate EPT richness (family level).

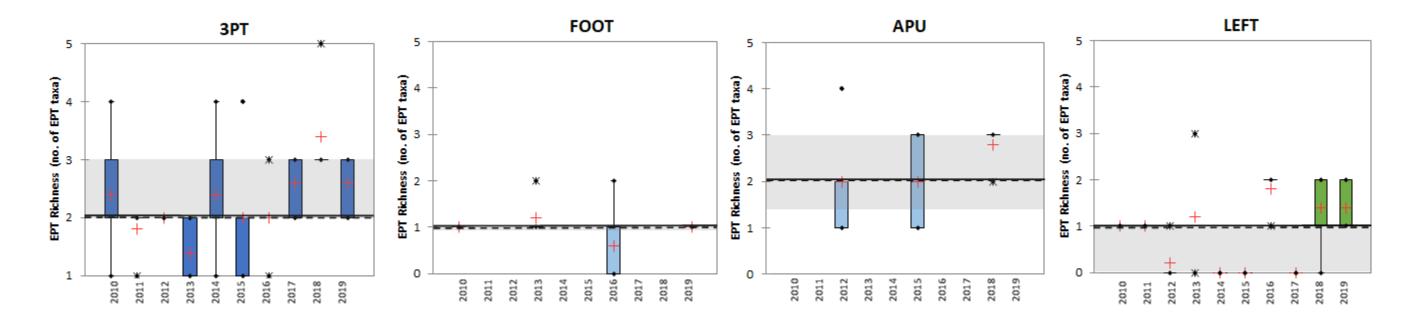


Figure 4.4-4. 2010 to 2019 Offshore benthic invertebrate EPT richness (family level).



4.5 DIVERSITY

4.5.1 HILL'S EFFECTIVE RICHNESS

4.5.1.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Nearshore Habitat

Annual mean Hill's effective richness (Hill's index) over the ten years of monitoring ranged from three (2014) to five (2010, 2013, 2015, 2016, and 2017; Figure 4.5-1). The overall mean and median were four, and the IQR was less than 4 to less than 5. Annual means were within the IQR, except in 2014 (below).

Offshore Habitat

Annual mean Hill's index over the ten years of monitoring ranged from two (2019) to greater than four (2011; Figure 4.5-2). The overall mean and median were three, and the IQR was less than 3 to less than 4. Annual means were below the IQR in 2014, 2015, and 2019, and above the IQR in 2010, 2011, and 2013.

ROTATIONAL SITES

Footprint Lake

Nearshore Habitat

Annual mean Hill's effective richness (Hill's index) over the four years of monitoring ranged from four (2010) to six (2013; Figure 4.5-1). The overall mean and median were five, and the IQR was less than 4 to less than 6. Annual means were within the IQR, except in 2013 (above).

Offshore Habitat

Annual mean Hill's index over the four years of monitoring ranged from three (2013 and 2019) to four (2016; Figure 4.5-2). The overall mean and median were three, and the IQR was between less than 3 and greater than 3. Annual means were above the IQR in 2010 and 2016.



Nearshore Habitat

Annual mean Hill's effective richness (Hill's index) over the three years of monitoring ranged from five (2012) to seven (2015 and 2018; Figure 4.5-1). The overall mean was six, the overall median was eight, and the IQR was less than less than 5 to less than 8. Annual means were within the IQR in all years.

Offshore Habitat

Annual mean Hill's index over the three years of monitoring ranged from less than three (2018) to four (2012; Figure 4.5-2). The overall mean and median were three, and the IQR was between less than 3 and greater than 3. Annual means were within the IQR, except in 2012 (above).

4.5.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

Nearshore Habitat

Annual mean Hill's effective richness (Hill's index) over the ten years of monitoring ranged from six (2011, 2012, 2014, and 2015) to ten (2019; Figure 4.5-1). The overall mean and median were seven, and the IQR was less than 6 to 8. Annual means were above the IQR in 2017 and 2019.

Offshore Habitat

Annual mean Hill's index over the ten years of monitoring ranged from two (all years except 2015) to slightly more than three (2015; Figure 4.5-2). The overall mean and median were less than three, and the IQR was less than 3 to 3. Annual means were with in the IQR, except in 2011 (slightly below).



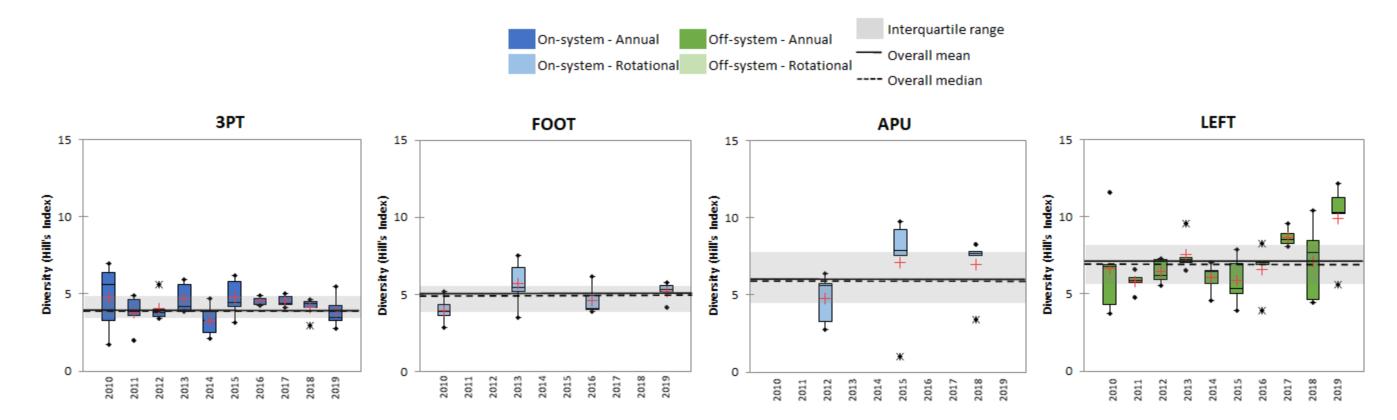


Figure 4.5-1. 2010 to 2019 Nearshore benthic invertebrate diversity (family level).

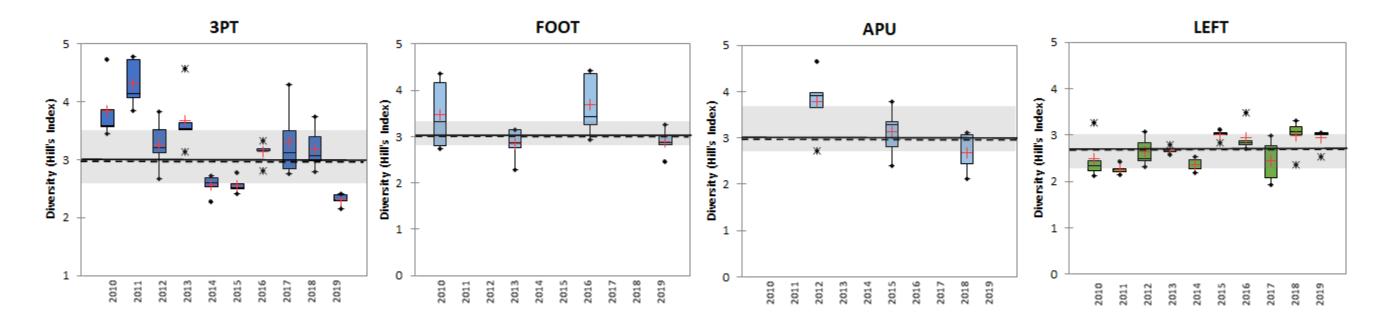


Figure 4.5-2. 2010 to 2019 Offshore benthic invertebrate diversity (family level).



APPENDIX 4-1. BENTHIC INVERTEBRATE NEARSHORE AND OFFSHORE SAMPLING SITES: 2008-2019



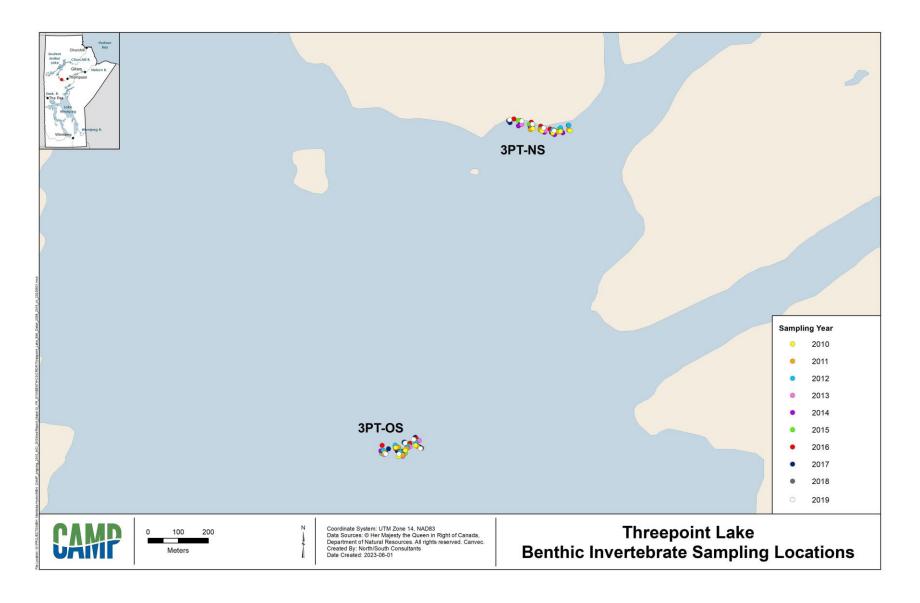


Figure A4-1-1. 2010 to 2019 Threepoint Lake nearshore (NS) and offshore (OS) benthic invertebrate sampling sites.



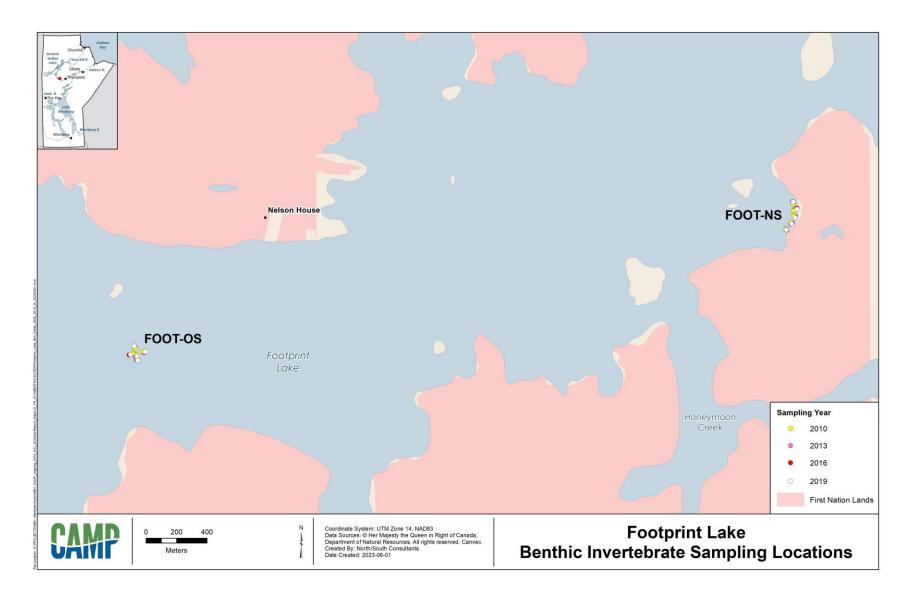


Figure A4-1-2. 2010 to 2019 Footprint Lake nearshore (NS) and offshore (OS) benthic invertebrate sampling sites.



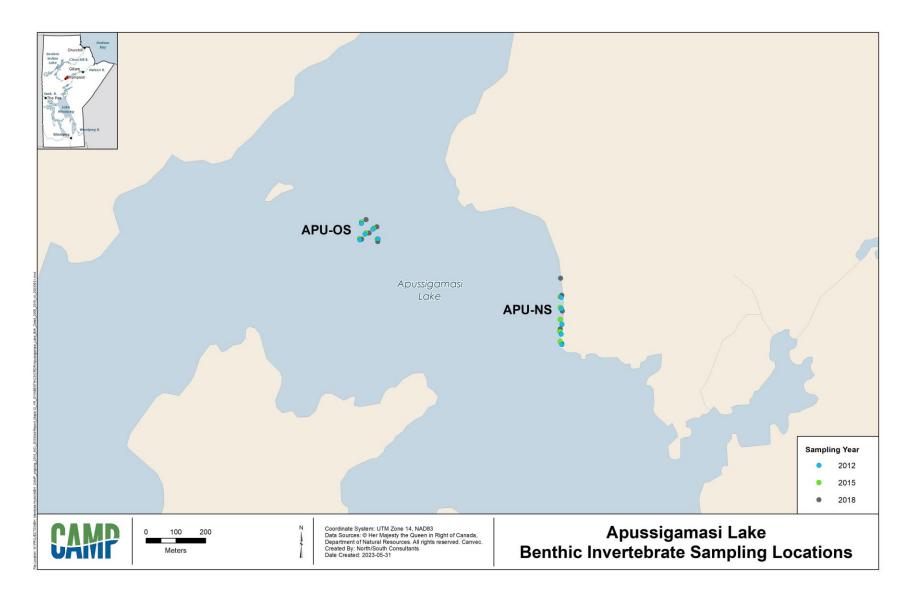


Figure A4-1-3. 2010 to 2019 Apussigamasi Lake nearshore (NS) and offshore (OS) benthic invertebrate sampling sites.



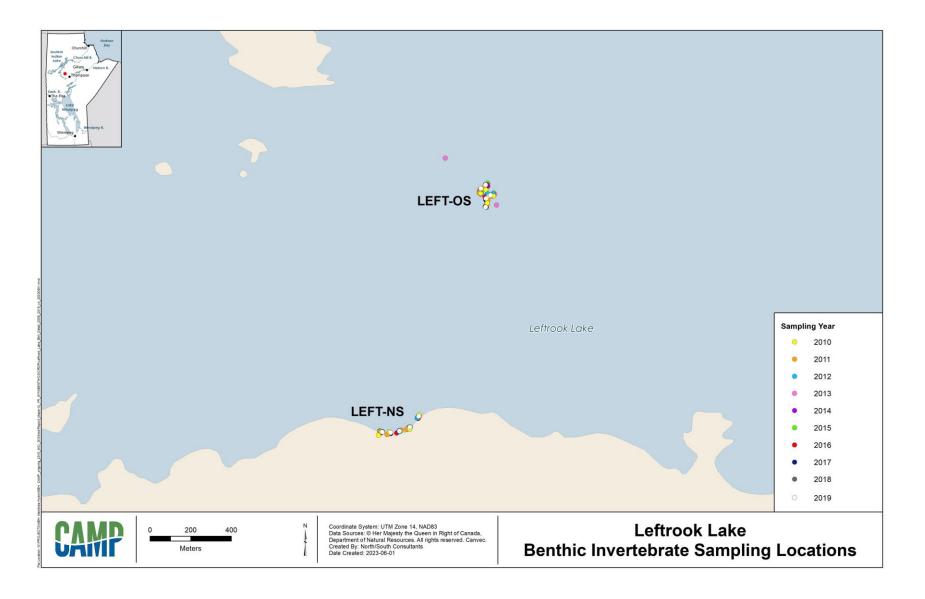


Figure A4-1-4. 2010 to 2019 Leftrook Lake nearshore (NS) and offshore (OS) benthic invertebrate sampling sites.



APPENDIX 4-2. BENTHIC INVERTEBRATE NEARSHORE AND OFFSHORE SUPPORTING SUBSTRATE DATA BY YEAR



Table A4-2-1. 2010 to 2019 Threepoint Lake nearshore supporting benthic substrate data.

| | | Sample | | Supp | orting Sub | bstrate Analysis | | | | | |
|------|-----------------------------|-----------|------|-------------|------------|------------------|-----------------|--|--|--|--|
| Year | Dominant Substrate | Water | Mean | Particle Si | ze (%) | Mean TOC | | | | | |
| | Substrate | Depth (m) | Sand | Silt | Clay | (%) | Texture | | | | |
| 2010 | fines and organics | 0.4 | 8.0 | 45.6 | 46.5 | 0.7 | Clay | | | | |
| 2011 | fines | 0.7 | 11.1 | 18.1 | 70.8 | 0.5 | Clay | | | | |
| 2012 | fines | 0.6 | 24.7 | 39.3 | 35.9 | 0.4 | Loam | | | | |
| 2013 | fines | 0.9 | 10.5 | 22.3 | 67.1 | 1.0 | Clay | | | | |
| 2014 | fines and organics | 0.4 | 5.0 | 31.5 | 63.6 | 1.4 | Clay | | | | |
| 2015 | fines, coarse, and organics | 0.4 | 41.5 | 44.8 | 13.7 | 0.3 | Loam | | | | |
| 2016 | fines | 0.9 | 10.6 | 28.7 | 60.7 | 0.8 | Clay | | | | |
| 2017 | fines | 0.6 | 2.8 | 45.8 | 51.8 | - | Silty clay | | | | |
| 2018 | fines | 0.3 | 18.0 | 62.8 | 19.3 | 0.5 | Silt loam | | | | |
| 2019 | fines | 0.6 | 9.3 | 48.2 | 42.5 | 0.4 | Silty clay loam | | | | |

1. TOC = Total organic carbon.

Table A4-2-2. 2010 to 2019 Threepoint Lake offshore supporting benthic substrate data.

| | | Sample | | Sup | porting Sub | strate Analysis | 5 |
|------|-----------------------|-----------|------|-------------|-------------|-----------------|-----------------|
| Year | Dominant Substrate | Water | Mean | Particle Si | ze (%) | Mean TOC | Tt |
| | Substrate | Depth (m) | Sand | Silt | Clay | (%) | Texture |
| 2010 | fines | 4.6 | 9.2 | 55.9 | 34.9 | 1.6 | Silty clay loam |
| 2011 | fines | 4.2 | 13.1 | 50.8 | 36.1 | 1.5 | Silty clay loam |
| 2012 | fines | 5.9 | 17.7 | 52.0 | 30.3 | 1.5 | Silty clay loam |
| 2013 | fines | 4.8 | 12.3 | 54.2 | 33.6 | 1.6 | Silty clay loam |
| 2014 | fines | 3.9 | 10.7 | 53.6 | 35.8 | 1.7 | Silty clay loam |
| 2015 | fines | 5.3 | 13.3 | 54.3 | 32.3 | 1.3 | Silty clay loam |
| 2016 | fines | 5.9 | 16.6 | 67.8 | 15.6 | 1.5 | Silt loam |
| 2017 | fines | 4.5 | 14.5 | 60.4 | 25.0 | 1.5 | Silty clay loam |
| 2018 | fines | 6.0 | 17.5 | 65.4 | 21.4 | 1.6 | Silt loam |
| 2019 | fines and organics | 5.9 | 16.2 | 55.2 | 28.6 | 1.5 | Silt loam |

Notes:

1. TOC = Total organic carbon.



2010 to 2019 Footprint Lake nearshore supporting benthic substrate data. Table A4-2-3.

| | | Sample | Supporting Substrate Analysis | | | | | | | | | | | |
|------|-----------------------------|-----------|-------------------------------|------------|-------|----------|-----------------|--|--|--|--|--|--|--|
| Year | Dominant Substrate | Water | Comp | osition (m | ean%) | тос | | | | | | | | |
| | Substrate | Depth (m) | Sand | Silt | Clay | (mean %) | Texture | | | | | | | |
| 2010 | fines, coarse, and organics | 0.5 | 11.1 | 44.9 | 43.9 | 0.9 | Silty/Clay | | | | | | | |
| 2013 | fines | 0.4 | 27.9 | 19.9 | 52.2 | 0.6 | Clay loam | | | | | | | |
| 2016 | coarse and fines | 0.7 | 32.4 | 46.3 | 32.0 | 0.9 | Silty clay loam | | | | | | | |
| 2019 | fines, coarse, and organics | 0.4 | 41.7 | 41.4 | 25.1 | 0.3 | Silt loam | | | | | | | |

Table A4-2-4. 2010 to 2019 Footprint Lake offshore supporting benthic substrate data.

| | | Sample | Supporting Substrate Analysis | | | | | | | | | | | |
|------|-----------------------|-----------|-------------------------------|------------|-------|----------|-----------------|--|--|--|--|--|--|--|
| Year | Dominant Substrate | Water | Comp | osition (m | ean%) | тос | _ | | | | | | | |
| | Substrate | Depth (m) | Sand | Silt | Clay | (mean %) | Texture | | | | | | | |
| 2010 | fines | 6.3 | 40.1 | 29.0 | 30.8 | 1.2 | Clay loam | | | | | | | |
| 2013 | fines and coarse | 6.5 | 47.1 | 18.1 | 34.9 | 1.1 | Clay | | | | | | | |
| 2016 | fines | 7.7 | 61.4 | 23.4 | 15.2 | 1.2 | Sandy loam | | | | | | | |
| 2019 | fines and coarse | 7.7 | 50.9 | 17.3 | 31.8 | 1.0 | Sandy clay loam | | | | | | | |

Notes:

2010 to 2019 Apussigamasi Lake nearshore supporting benthic substrate data. Table A4-2-5.

| | | Sample | Supporting Substrate Analysis | | | | | | | | | | | |
|------|-----------------------|-------------|-------------------------------|--------------|------|----------|------------|--|--|--|--|--|--|--|
| Year | Dominant Substrate | Water Depth | Comp | oosition (me | TOC | | | | | | | | | |
| | Substrate | (m) | Sand | Silt | Clay | (mean %) | Texture | | | | | | | |
| 2012 | fines | 0.7 | 1.1 | 14.7 | 84.2 | 1.3 | Clay | | | | | | | |
| 2015 | fines | 0.9 | 0.8 | 53.6 | 45.6 | 0.6 | Silty clay | | | | | | | |
| 2018 | fines | 0.5 | 1 | 53.6 | 45.8 | 0.9 | Silty clay | | | | | | | |

Notes:

2010 to 2019 Apussigamasi Lake offshore supporting benthic substrate data. Table A4-2-6.

| | | Sample | | Sup | porting Sub | strate Analysis | 1 | |
|------|-----------------------|-----------|------|-------------|-------------|-----------------|-----------|--|
| Year | Dominant Substrate | Water | Comp | osition (me | ean%) | тос | Texture | |
| | Substrate | Depth (m) | Sand | Silt | Clay | (mean %) | | |
| 2012 | fines | 7.3 | 9.4 | 73.0 | 17.5 | 0.8 | Silt loam | |
| 2015 | fines | 6.8 | 5.7 | 85.0 | 9.4 | 0.7 | Silt | |
| 2018 | fines | 7.3 | 8.7 | 78.4 | 14.6 | 1.2 | Silt loam | |

Notes:

1. TOC = Total organic carbon.



^{1.} TOC = Total organic carbon.

^{1.} TOC = Total organic carbon.

^{1.} TOC = Total organic carbon.

Table A4-2-7. 2010 to 2019 Leftrook Lake nearshore supporting benthic substrate data.

| | | Sample | Supporting Substrate Analysis | | | | | | | | | | |
|------|---------------------------|------------------|-------------------------------|--------------|-------|----------|---------|--|--|--|--|--|--|
| Year | Dominant Substrate | Water Depth | Comp | oosition (me | ean%) | тос | | | | | | | |
| | Substrate | (m) | Sand | Silt | Clay | (mean %) | Texture | | | | | | |
| 2010 | hard and coarse | no sample | | | | | | | | | | | |
| 2011 | hard and coarse | no sample | | | | | | | | | | | |
| 2012 | hard and coarse | no sample | | | | | | | | | | | |
| 2013 | hard | no sample | | | | | | | | | | | |
| 2014 | hard and coarse | no sample | | | | | | | | | | | |
| 2015 | hard, coarse and organics | no sample | | | | | | | | | | | |
| 2016 | hard | 0.6 ² | 98.9 | | | 0.5 | Sand | | | | | | |
| 2017 | coarse | no sample | | | | | | | | | | | |
| 2018 | coarse and hard | no sample | | | - | | | | | | | | |
| 2019 | hard and coarse | no sample | | | - | | | | | | | | |

- 1. TOC = Total organic carbon.
- 2. Surface/interstitial sediment was submitted for analysis.

Table A4-2-8. 2010 to 2019 Leftrook Lake offshore supporting benthic substrate data.

| | | Sample | | Supp | porting Sub | strate Analysis | 3 |
|------|-----------|-----------|------|-------------|-------------|-----------------|-----------------|
| Year | Dominant | Water | Comp | osition (me | ean%) | тос | |
| | Substrate | Depth (m) | Sand | Silt | Clay | (mean %) | Texture |
| 2010 | fines | 7.9 | 0.2 | 26.2 | 73.7 | 5.1 | Clay |
| 2011 | fines | 8.7 | 0.2 | 26.2 | 73.6 | 4.6 | Clay |
| 2012 | fines | 8.2 | 2.1 | 47.8 | 50.1 | 4.9 | Clay |
| 2013 | fines | 8.3 | 0.9 | 49.2 | 50.0 | 5.5 | Silty clay |
| 2014 | fines | 8.3 | 0.1 | 40.0 | 59.8 | 5.1 | Silty clay |
| 2015 | fines | 8.2 | 0.1 | 42.7 | 57.2 | 4.7 | Silty clay |
| 2016 | fines | 8.4 | | 65.7 | 34.0 | 4.8 | Silty clay loam |
| 2017 | fines | 8.1 | 3.6 | 63.6 | 34.7 | 4.7 | Silty clay |
| 2018 | fines | 8.3 | | 36.0 | 63.8 | 4.9 | Silty clay |
| 2019 | fines | 8.1 | 1.8 | 47.0 | 52.4 | 4.8 | Silty clay |

Notes:

1. TOC = Total organic carbon.



5.0 FISH COMMUNITY

5.1 INTRODUCTION

The following presents the results of fish community monitoring conducted from 2008 to 2019 in the Churchill River Diversion Region. Four waterbodies were monitored in the Churchill River Diversion Region starting in 2009: one on-system annual site (Threepoint Lake); two on-system rotational sites (Footprint and Apussigamasi Lake) and one off-system annual site (Leftrook Lake; Table 5.1-1 and Figure 5.1-1). There were no departures from the planned field sampling during the 11-year period.

Monitoring targeted both small-bodied fish species (i.e., forage fish) and large-bodied fish species (e.g., fish targeted in subsistence, commercial, and/or recreational fisheries). Within a given waterbody, sampling was conducted at approximately the same time of year during each year of monitoring. Standard gang index gill nets (GN; 51, 76, 95, 108, and 127 mm stretched mesh panels) were set at each site and a small mesh index gillnet gang (SN; 16, 20, and 25 mm bar measure panels) was attached to the end of the standard gang at approximately every third site (Appendix 5-1). Gill nets were set for approximately 24 hours (h). All fish captured at each site were counted by mesh size and species. Individual metrics (e.g., length, weight, deformities, erosion, lesions, and tumours [DELTs], sex and maturity, and age) were collected for species of management interest (i.e., "target" species). These include: Lake Whitefish (*Coregonus clupeaformis*), Walleye (*Sander vitreus*) and Northern Pike (*Esox lucius*) from all waterbodies in all years; Sauger (*Sander canadensis*) from all waterbodies starting in 2017, except Leftrook Lake where the species has not been caught; and White Sucker (*Catostomus commersonii*) from all waterbodies starting in 2010. All other species were bulk weighed.

Five fish community indicators (abundance, condition, growth, recruitment, and community diversity) were selected for detailed reporting (Table 5.1-2). Metrics for these indicators that are presented herein include: catch-per-unit-effort (CPUE); Fulton's condition factor (KF); relative weight (Wr); fork length-at-age (FLA); relative year-class strength (RYCS); Hill's effective species richness (Hill's index); and relative species abundance (RSA; Table 5.1-2).

A detailed description of the program design and sampling methods is provided in Technical Document 1, Section 2.5.



Table 5.1-1. 2008-2019 Inventory of fish community sampling.

| | | Sampling Year | | | | | | | | | | | | | | |
|----------------|------|---------------|------|------|------|------|------|------|------|------|------|------|--|--|--|--|
| Waterbody/Area | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | | | | |
| 3PT | | • | • | • | • | • | • | • | • | • | • | • | | | | |
| FOOT | | | • | | | • | | | • | | | • | | | | |
| APU | | • | | | • | | | • | | | • | | | | | |
| LEFT | | • | • | • | • | • | • | • | • | • | • | • | | | | |

Table 5.1-2. Fish community indicators and metrics.

| Indicator | Metric | Units |
|-------------|---|--|
| Abundance | Catch-Per-Unit-Effort (CPUE) | # fish/30 m/24 hour (h) # fish/100 m/24 h |
| Condition | Fulton's Condition Factor (KF) | - |
| Condition | Relative Weight (Wr) | - |
| Growth | Fork Length-At-Age (FLA) | mm |
| Recruitment | Relative Year-Class Strength (RYCS) | - |
| Divorcity | Hill's Effective Species Richness | species |
| Diversity | Relative Species Abundance (RSA) ¹ | % |

1. Supporting metric.



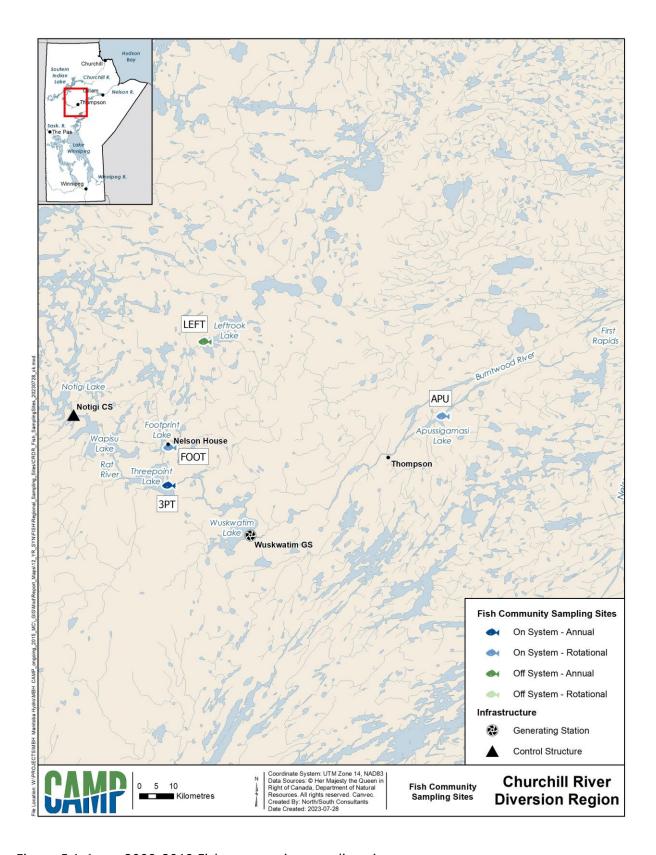


Figure 5.1-1. 2008-2019 Fish community sampling sites.



5.2 ABUNDANCE

5.2.1 CPUE

5.2.1.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Standard Gang Index Gill Nets

The annual mean CPUE varied over the 11 years of monitoring from a low of 19.2 in 2010 to a maximum of 42.6 fish/100 m/24 h in 2016 (Table 5.2-1; Figure 5.2-1).

The overall mean CPUE was 31.9, the median was 30.5, and the IQR was 28.4-35.4 fish/100 m/24 h (Figure 5.2-1). The annual mean CPUE fell within the overall IQR except in 2010, 2011 and 2012 when it was below the IQR and in 2014, 2016 and 2018 when it was above the IQR.

Small Mesh Index Gill Nets

The annual mean CPUE in the small mesh gangs over the 11 years of monitoring was more variable than in the standard gangs, with the mean ranging from a low of 25.9 in 2012 to a high of 448.6 fish/30 m/24 h in 2017 (Table 5.2-1; Figure 5.2-2).

The overall mean CPUE was 128.5, the median was 83.9, and the IQR was 50.0-154.8 fish/30 m/24 h (Figure 5.2-2). The annual mean CPUE fell within the overall IQR except in 2009, 2010, and 2012 when it was below the IQR and in 2015, 2017, and 2018 when it was above the IQR.

Lake Whitefish

Catches of Lake Whitefish were relatively low in Threepoint Lake over the 11 years of monitoring, with the annual mean CPUE ranging from a low of 0.1 in 2010 to a high of 1.5 fish/100 m/24 h in 2013 and 2017 (Table 5.2-1; Figure 5.2-3).

The overall mean CPUE was 1.0, the median was 1.1, and the IQR was 0.6-1.3 fish/100 m/24 h (Figure 5.2-3). The annual mean CPUE fell within the overall IQR except in 2009, 2010, and 2016 when it was below the IQR and in 2017 when it was above the IQR.



Northern Pike

Catches of Northern Pike were relatively consistent in Threepoint Lake over the 11 years of monitoring, with the annual mean CPUE ranging from a low of 1.4 in 2016 to a high of 4.6 fish/100 m/24 h in 2009 (Table 5.2-1; Figure 5.2-4).

The overall mean CPUE was 2.5, the median was 2.3, and the IQR was 1.9-3.0 fish/100 m/24 h (Figure 5.2-4). The annual mean CPUE fell within the overall IQR except in 2016, 2018, and 2019 when it was below the IQR and in 2009, 2014, and 2015 when it was above the IQR.

Sauger

The annual mean CPUE over the 11 years of monitoring varied up to about 12-fold from year-to-year, ranging from a low of 2.5 in 2010 to a high of 14.1 fish/100 m/24 h in 2016 (Table 5.2-1; Figure 5.2-5).

The overall mean CPUE was 6.0, the median was 5.2, and the IQR was 4.2-6.6 fish/100 m/24 h (Figure 5.2-5). The annual mean CPUE fell within the overall IQR except in 2010, 2012, and 2013 when it was below the IQR and in 2014, 2016, and 2017 when it was above the IQR.

Walleye

The annual mean CPUE over the 11 years of monitoring varied by up to about two-fold from year-to-year, ranging from a low of 5.7 in 2010 to a high of 12.5 fish/100 m/24 h in 2013 (Table 5.2-1; Figure 5.2-6).

The overall mean and median CPUE were 9.7, and the IQR was 9.0-11.0 fish/100 m/24 h (Figure 5.2-6). The annual mean CPUE fell within the overall IQR except in 2010, 2015, and 2018 when it was below the IQR and 2014, 2016, and 2017 when it was above the IQR.

White Sucker

The annual mean CPUE over the 11 years of monitoring varied up to about two-fold from year-to-year, with the mean ranging from a low of 7.1 in 2010 to a high of 12.6 fish/100 m/24 h in 2014 (Table 5.2-1; Figure 5.2-7).

The overall mean CPUE was 9.9, the median was 10.6, and the IQR was 8.2-11.3 fish/100 m/24 h (Figure 5.2-7). The annual mean CPUE fell within the overall IQR except in 2010, 2011, and 2015 when it was below the IQR and in 2014, 2016, and 2018 when it was above the IQR.



ROTATIONAL SITES

Footprint Lake

Standard Gang Index Gill Nets

The annual mean CPUE varied over the four years of monitoring from a low of 42.1 in 2010 to a maximum of 58.8 fish/100 m/24 h in 2019 (Table 5.2-1; Figure 5.2-1).

The overall mean CPUE was 53.9, the median was 57.3, and the IQR was 56.7-58.2 fish/100 m/24 h (Figure 5.2-1). The annual mean CPUE fell within the overall IQR except in 2010 when it was below the IQR and in 2019 when it was above the IQR.

Small Mesh Index Gill Nets

The annual mean CPUE in the small mesh gangs over the four years of monitoring was more variable than in the standard gangs, with the mean ranging from a low of 24.7 in 2010 to a high of 182.1 fish/30 m/24 h in 2016 (Table 5.2-1; Figure 5.2-2).

The overall mean CPUE was 99.8, the median was 96.1, and the IQR was 51.3-157.6 fish/30 m/24 h (Figure 5.2-2). The annual mean CPUE fell within the overall IQR in 2019, below the IQR in 2010 and 2013, and above the IQR in 2016.

Lake Whitefish

Catches of Lake Whitefish were relatively low in Footprint Lake over the four years of monitoring, with the annual mean CPUE ranging from a low of 0.1 in 2013 and 206 to a high of 0.3 fish/100 m/24 h in 2019 (Table 5.2-1; Figure 5.2-3).

The overall mean and median CPUE were 0.2, and the IQR was 0.1-0.2 fish/100 m/24 h (Figure 5.2-3). The annual mean CPUE fell within the overall IQR except in 2013 and 2016 when it was below the IQR and in 2019 when it was above the IQR.

Northern Pike

Catches of Northern Pike were relatively low in Footprint Lake over the four years of monitoring, with the annual mean CPUE ranging from a low of 0.4 in 2016 to a high of 5.8 fish/100 m/24 h in 2013 (Table 5.2-1; Figure 5.2-4).

The overall mean and median CPUE were 3.1, and the IQR was 2.3-4.4 fish/100 m/24 h (Figure 5.2-4). The annual mean CPUE fell within the overall IQR in 2010, below the IQR in 2016 and 2019, and above the IQR in 2013.



Sauger

The annual mean CPUE over the four years of monitoring varied up to about four-fold from year-to-year, ranging from a low of 1.9 in 2010 to a high of 7.1 fish/100 m/24 h in 2016 (Table 5.2-1; Figure 5.2-5).

The overall mean CPUE was 4.8, the median was 5.1, and the IQR was 4.2-5.7 fish/100 m/24 h (Figure 5.2-5). The annual mean CPUE fell within the overall IQR except in 2010 when it was below the IQR and in 2016 when it was above the IQR.

Walleye

The annual mean CPUE over the four years of monitoring varied by up to about two-fold from year-to-year, ranging from a low of 14.9 in 2010 to a high of 31.1 fish/100 m/24 h in 2016 (Table 5.2-1; Figure 5.2-6).

The overall mean CPUE was 23.3, the median was 23.7, and the IQR was 23.4-25.8 fish/100 m/24 h (Figure 5.2-6). The annual mean CPUE fell within the overall IQR except in 2010 when it was below the IQR and in 2016 when it was above the IQR.

White Sucker

The annual mean CPUE over the four years of monitoring varied up to about three-fold from year-to-year, with the mean ranging from a low of 6.6 in 2013 to a high of 14.6 fish/100 m/24 h in 2010 (Table 5.2-1; Figure 5.2-7).

The overall mean CPUE was 10.2, the median was 9.9, and the IQR was 7.9-12.8 fish/100 m/24 h (Figure 5.2-7). The annual mean CPUE fell within the overall IQR except in 2013 and 2016 when it was below the IQR and in 2010 when it was above the IQR.

Apussigamasi Lake

Standard Gang Index Gill Nets

The annual mean CPUE varied over the four years of monitoring from a low of 36.1 in 2012 to a maximum of 44.4 fish/100 m/24 h in 2009 (Table 5.2-1; Figure 5.2-1).

The overall mean CPUE was 41.2, the median was 42.2, and the IQR was 40.4-43.0 fish/100 m/24 h (Figure 5.2-1). The annual mean CPUE fell within the overall IQR except in 2012 when it was below the IQR and in 2009 when it was above the IQR.



Small Mesh Index Gill Nets

The annual mean CPUE in the small mesh gangs over the four years of monitoring was more variable than in the standard gangs, with the mean ranging from a low of 42.6 in 2009 to a high of 65.5 fish/30 m/24 h in 2018 (Table 5.2-1; Figure 5.2-2).

The overall mean CPUE was 58.6, the median was 63.2, and the IQR was 58.0-63.8 fish/30 m/24 h (Figure 5.2-2). The annual mean CPUE fell within the overall IQR except in 2009 when it was below the IQR and in 2018 when it was above the IQR.

Lake Whitefish

Catches of Lake Whitefish were relatively low in Apussigamasi Lake over the four years of monitoring, with the annual mean CPUE ranging from a low of 2.3 in 2012 to a high of 4.0 fish/100 m/24 h in 2009 (Table 5.2-1; Figure 5.2-3).

The overall mean and median CPUE were 3.1, and the IQR was 2.6-3.7 fish/100 m/24 h (Figure 5.2-3). The annual mean CPUE fell within the overall IQR except in 2012 when it was below the IQR and in 2009 when it was above the IQR.

Northern Pike

Catches of Northern Pike were relatively low in the Apussigamasi Lake over the four years of monitoring, with the annual mean CPUE ranging from a low of 0.5 in 2012 to a high of 2.1 fish/100 m/24 h in 2009 (Table 5.2-1; Figure 5.2-4).

The overall mean CPUE was 1.3, the median was 1.2, and the IQR was 0.7-1.8 fish/100 m/24 h (Figure 5.2-4). The annual mean CPUE fell within the overall IQR except in 2012 when it was below the IQR and in 2009 when it was above the IQR.

Sauger

The annual mean CPUE over the four years of monitoring varied little from year-to-year, ranging from a low of 7.6 in 2009 to a high of 9.0 fish/100 m/24 h in 2015 (Table 5.2-1; Figure 5.2-5).

The overall mean and median CPUE were 8.2, and the IQR was 7.8-8.7 fish/100 m/24 h (Figure 5.2-5). The annual mean CPUE fell within the overall IQR in 2012, below the IQR in 2009 and 2018, and above the IQR in 2015.



Walleye

The annual mean CPUE over the four years of monitoring varied almost two-fold from year-to-year, ranging from a low of 8.5 in 2018 to a high of 15.6 fish/100 m/24 h in 2009 (Table 5.2-1; Figure 5.2-6).

The overall mean CPUE was 11.8, the median was 11.5, and the IQR was 10.1-13.1 fish/100 m/24 h (Figure 5.2-6). The annual mean CPUE fell within the overall IQR except in 2018 when it was below the IQR and 2009 when it was above the IQR.

White Sucker

The annual mean CPUE over the four years of monitoring varied up to about three-fold from year-to-year, with the mean ranging from a low of 7.4 in 2009 to a high of 11.9 fish/100 m/24 h in 2018 (Table 5.2-1; Figure 5.2-7).

The overall mean CPUE was 9.2, the median was 8.8, and the IQR was 7.9-10.2 fish/100 m/24 h (Figure 5.2-7). The annual mean CPUE fell within the overall IQR except in 2009 when it was below the IQR and in 2018 when it was above the IQR.

5.2.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

Standard Gang Index Gill Nets

The annual mean CPUE over the 11 years of monitoring varied by almost two-fold, with the mean ranging from a low of 66.9 in 2010 to a high of 114.1 fish/100 m/24 h in 2012 (Table 5.2-1; Figure 5.2-1).

The overall mean CPUE was 91.9, the median was 94.4, and the IQR was 93.0-97.8 fish/100 m/24 h (Figure 5.2-1). The annual mean CPUE fell within the overall IQR except in in 2010, 2011, 2013, 2016, and 2019 when it was below the IQR and in 2012, 2014, and 2018 when it was above the IQR.



Small Mesh Index Gill Nets

The annual mean CPUE in the small mesh gangs over the 11 years of monitoring was more variable than in the standard gangs, with the mean ranging from a low of 93.2 in 2011 to a high of 1046.9 fish/30 m/24 h in 2014 (Table 5.2-1; Figure 5.2-2).

The overall mean CPUE was 297.4, the median was 169.9, and the IQR was 162.6-404.5 fish/30 m/24 h (Figure 5.2-2). The annual mean CPUE fell within the overall IQR except in 2009, 2011, 2013, 2018, and 2019 when it was below the IQR and in 2012, 2014, and 2017 when it was above the IQR.

Lake Whitefish

Catches of Lake Whitefish were relatively high in Leftrook Lake over the 11 years of monitoring, with the annual mean CPUE ranging from a low of 7.1 in 2011 to a high of 17.8 fish/100 m/24 h in 2012 (Table 5.2-1; Figure 5.2-3).

The overall mean CPUE was 11.0, the median was 10.0 and the IQR was 9.6-13.0 fish/100 m/24 h (Figure 5.2-3). The annual mean CPUE fell within the overall IQR except in 2010, 2011, 2016, 2018, and 2019 when it was below the IQR and in 2012-2014 when it was above the IQR.

Northern Pike

Catches of Northern Pike were relatively high in Leftrook Lake over the 11 years of monitoring, with the annual mean CPUE ranging from a low of 7.0 in 2016 to a high of 16.6 fish/100 m/24 h in 2014 (Table 5.2-1; Figure 5.2-4).

The overall mean CPUE was 10.9, median was 10.4 and the IQR was 10.4-12.4 fish/100 m/24 h (Figure 5.2-4). The annual mean CPUE fell within the overall IQR except in 2009, 2010, 2013, 2016, and 2017 when it was below the IQR and in 2011, 2012, and 2014 when it was above the IQR.

Sauger

Sauger were not captured in Leftrook Lake over the 11 years of monitoring (Table 5.2-1).

Walleye

The annual mean CPUE over the 11 years of monitoring varied by up to about three-fold, with the mean ranging from a low of 13.7 in 2010 to a high of 41.9 fish/100 m/24 h in 2017 (Table 5.2-1; Figure 5.2-6).



The overall mean CPUE was 29.7, the median was 29.5, and the IQR was 29.4-34.2 fish/100 m/24 h (Figure 5.2-6). The annual mean CPUE fell within the overall IQR except in in 2009-2011, 2013, and 2019 when it was below the IQR and in 2012, 2017, and 2018 when it was above the IQR.

White Sucker

The annual mean CPUE over the 11 years of monitoring varied by almost two-fold, with the mean ranging from a low of 19.5 in 2017 to a high of 35.0 fish/100 m/24 h in 2011 (Table 5.2-1; Figure 5.2-7).

The overall mean CPUE was 28.1, the median was 28.0, and the IQR was 28.0-31.1 fish/100 m/24 h (Figure 5.2-7). The annual mean CPUE fell within the overall IQR except in 2010, 2012, 2014, 2017, and 2019 when it was below the IQR and 2009, 2011, and 2013 when it was above the IQR.



Table 5.2-1. 2009-2019 Catch-per-unit-effort.

| Maria da ad | V | | Small N | Mesh Catch ¹ | l | | Tota | al Catch ² | | | LKWH | | | NRPK | | | SAUG | | | WALL | | | WHSC | |
|-------------|----------|-----------------|-----------------------------|-------------------------|--------|----|----------------|-----------------------|------|----------------|------|------|----------------|------|-----|----------------|------|------|----------------|------|------|----------------|------|------|
| Waterbody | Year | ns ³ | n _F ⁴ | Mean | SE⁵ | ns | n _F | Mean | SE | n _F | Mean | SE | n _F | Mean | SE | n _F | Mean | SE | n _F | Mean | SE | n _F | Mean | SE |
| 3PT | 2009 | 2 | 77 | 39.3 | 33.3 | 9 | 366 | 35.2 | 15.7 | 7 | 0.6 | 0.5 | 50 | 4.6 | 3.4 | 56 | 5.6 | 4.0 | 103 | 9.3 | 8.8 | 106 | 10.7 | 4.3 |
| | 2010 | 3 | 120 | 43.4 | 16.0 | 9 | 263 | 19.2 | 7.1 | 2 | 0.1 | 0.3 | 32 | 2.4 | 1.9 | 31 | 2.5 | 1.9 | 91 | 5.7 | 4.0 | 87 | 7.1 | 3.6 |
| | 2011 | 1 | 210 | 116.0 | - | 9 | 327 | 28.1 | 8.0 | 15 | 1.1 | 1.4 | 24 | 1.9 | 1.7 | 47 | 4.2 | 3.1 | 111 | 10.2 | 5.6 | 96 | 8.1 | 2.3 |
| | 2012 | 3 | 50 | 25.9 | 5.6 | 9 | 233 | 25.9 | 13.1 | 7 | 0.8 | 1.1 | 17 | 1.9 | 1.4 | 37 | 4.1 | 4.2 | 85 | 9.7 | 8.8 | 76 | 8.3 | 3.9 |
| | 2013 | 3 | 239 | 83.8 | 9.4 | 9 | 327 | 33.5 | 13.3 | 15 | 1.5 | 1.5 | 22 | 2.3 | 2.2 | 40 | 4.1 | 4.5 | 122 | 12.5 | 10.9 | 104 | 10.6 | 5.0 |
| | 2014 | 3 | 403 | 151.6 | 77.4 | 9 | 395 | 42.0 | 14.0 | 10 | 1.1 | 1.3 | 39 | 4.1 | 2.2 | 84 | 9.2 | 5.7 | 120 | 12.3 | 8.2 | 117 | 12.6 | 9.8 |
| | 2015 | 3 | 593 | 210.7 | 160.2 | 9 | 291 | 30.5 | 15.8 | 13 | 1.4 | 1.6 | 32 | 3.4 | 3.8 | 63 | 6.6 | 4.5 | 85 | 8.9 | 7.7 | 70 | 7.3 | 5.6 |
| | 2016 | 3 | 159 | 56.6 | 21.1 | 9 | 409 | 42.6 | 13.4 | 5 | 0.5 | 0.5 | 14 | 1.4 | 1.4 | 134 | 14.1 | 10.8 | 113 | 11.6 | 6.5 | 114 | 11.9 | 6.6 |
| | 2017 | 3 | 1239 | 448.6 | 324.9 | 9 | 351 | 35.6 | 13.7 | 15 | 1.5 | 2.1 | 25 | 2.5 | 1.5 | 65 | 6.7 | 6.2 | 102 | 10.3 | 5.0 | 104 | 10.6 | 5.5 |
| | 2018 | 3 | 460 | 157.9 | 12.2 | 9 | 278 | 28.8 | 12.2 | 6 | 0.6 | 0.7 | 18 | 1.9 | 1.4 | 50 | 5.2 | 3.5 | 73 | 7.5 | 5.6 | 120 | 12.5 | 6.7 |
| | 2019 | 3 | 220 | 79.6 | 105.2 | 9 | 286 | 29.3 | 13.0 | 12 | 1.2 | 0.9 | 15 | 1.5 | 1.6 | 42 | 4.3 | 2.0 | 88 | 9.1 | 8.8 | 96 | 9.8 | 4.2 |
| FOOT | 2010 | 3 | 72 | 24.7 | 12.6 | 9 | 480 | 42.1 | 17.6 | 2 | 0.2 | 0.4 | 42 | 3.9 | 3.1 | 20 | 1.9 | 2.0 | 182 | 14.9 | 8.5 | 155 | 14.6 | 8.6 |
| | 2013 | 3 | 127 | 42.7 | 20.0 | 9 | 565 | 56.6 | 27.5 | 1 | 0.1 | 0.3 | 57 | 5.8 | 6.2 | 51 | 4.9 | 9.1 | 237 | 24.0 | 21.0 | 66 | 6.6 | 5.0 |
| | 2016 | 3 | 526 | 182.1 | 29.7 | 9 | 563 | 58.0 | 26.2 | 1 | 0.1 | 0.3 | 4 | 0.4 | 8.0 | 69 | 7.1 | 5.0 | 301 | 31.1 | 23.6 | 73 | 7.5 | 4.8 |
| | 2019 | 3 | 434 | 149.5 | 49.5 | 9 | 578 | 58.8 | 20.2 | 3 | 0.3 | 0.5 | 22 | 2.2 | 1.9 | 51 | 5.2 | 5.4 | 230 | 23.3 | 13.7 | 122 | 12.2 | 9.2 |
| APU | 2009 | 3 | 136 | 42.6 | 5.9 | 9 | 465 | 44.4 | 19.6 | 41 | 4.0 | 5.0 | 22 | 2.1 | 1.3 | 80 | 7.6 | 3.8 | 165 | 15.6 | 13.3 | 77 | 7.4 | 4.4 |
| | 2012 | 3 | 168 | 63.3 | 47.5 | 9 | 341 | 36.1 | 14.7 | 22 | 2.3 | 1.6 | 5 | 0.5 | 1.0 | 80 | 8.6 | 2.9 | 99 | 10.7 | 8.5 | 77 | 8.1 | 3.5 |
| | 2015 | 3 | 191 | 63.1 | 25.5 | 9 | 448 | 42.5 | 11.3 | 38 | 3.6 | 3.0 | 17 | 1.6 | 1.4 | 95 | 9.0 | 7.5 | 129 | 12.3 | 5.8 | 101 | 9.6 | 5.3 |
| | 2018 | 3 | 206 | 65.5 | 13.6 | 9 | 446 | 41.8 | 23.5 | 31 | 2.7 | 3.9 | 8 | 0.8 | 1.0 | 82 | 7.7 | 5.4 | 95 | 8.5 | 8.7 | 123 | 11.9 | 6.6 |
| LEFT | 2009 | 3 | 329 | 114.3 | 23.8 | 9 | 983 | 94.4 | 27.1 | 118 | 11.4 | 9.6 | 108 | 10.3 | 4.0 | 0 | - | - | 299 | 29.1 | 16.5 | 356 | 33.7 | 18.0 |
| | 2010 | 3 | 406 | 169.9 | 132.3 | 9 | 544 | 66.9 | 33.4 | 88 | 8.8 | 15.0 | 71 | 7.6 | 4.5 | 0 | - | - | 108 | 13.7 | 7.6 | 191 | 24.8 | 19.3 |
| | 2011 | 3 | 250 | 93.2 | 83.9 | 9 | 739 | 86.5 | 49.3 | 59 | 7.1 | 8.3 | 112 | 12.9 | 9.0 | 0 | - | - | 220 | 25.3 | 16.6 | 295 | 35.0 | 28.8 |
| | 2012 | 3 | 816 | 424.8 | 504.1 | 9 | 824 | 114.1 | 46.7 | 124 | 17.8 | 11.3 | 116 | 14.9 | 5.8 | 0 | - | - | 273 | 38.3 | 20.9 | 192 | 26.9 | 20.3 |
| | 2013 | 3 | 403 | 145.5 | 138.4 | 9 | 823 | 89.8 | 44.2 | 141 | 15.3 | 7.0 | 87 | 9.4 | 6.1 | 0 | - | - | 219 | 23.9 | 16.4 | 292 | 32.0 | 22.0 |
| | 2014 | 3 | 2530 | 1046.9 | 1448.7 | 9 | 882 | 99.6 | 40.4 | 128 | 14.5 | 6.7 | 150 | 16.6 | 6.7 | 0 | - | - | 256 | 29.5 | 21.1 | 204 | 23.2 | 17.9 |
| | 2015 | 3 | 560 | 217.4 | 181.0 | 9 | 852 | 96.0 | 36.6 | 89 | 10.0 | 5.6 | 102 | 11.9 | 5.0 | 0 | - | - | 283 | 32.1 | 12.9 | 252 | 28.0 | 14.1 |
| | 2016 | 3 | 1052 | 384.3 | 203.7 | 9 | 809 | 87.6 | 22.9 | 69 | 7.8 | 7.1 | 66 | 7.0 | 5.1 | 0 | - | - | 307 | 33.4 | 15.8 | 265 | 28.2 | 13.6 |
| | 2017 | 3 | 1160 | 440.6 | 173.0 | 9 | 855 | 94.6 | 40.2 | 102 | 10.9 | 9.4 | 72 | 7.8 | 4.3 | 0 | - | - | 379 | 41.9 | 27.8 | 173 | 19.5 | 14.8 |
| | 2018 | 3 | 393 | 134.1 | 36.9 | 9 | 1027 | 102.4 | 35.7 | 85 | 8.6 | 10.1 | 104 | 10.4 | 5.1 | 0 | - | - | 354 | 35.1 | 20.1 | 301 | 30.2 | 16.2 |
| | 2019 | 3 | 289 | 100.8 | 38.9 | 9 | 803 | 79.5 | 38.3 | 92 | 8.8 | 11.3 | 109 | 10.9 | 5.1 | 0 | - | - | 251 | 24.9 | 15.1 | 279 | 27.9 | 15.9 |

- 1. fish/30 m/24 h.
- 2. fish/100 m/24 h.
- 3. n_S = number of sites fished (excludes sets >36 h).
- 4. $n_F = number of fish caught$.
- 5. SE = standard error.



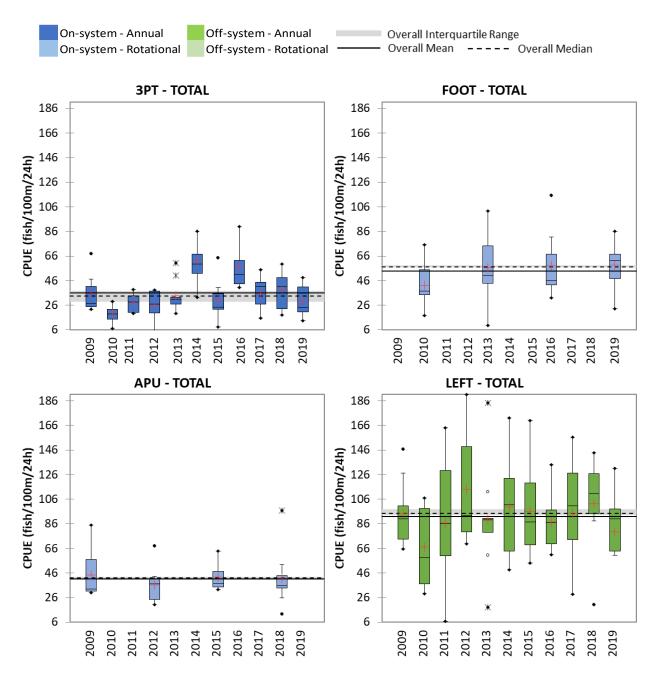


Figure 5.2-1. 2009-2019 Catch-per-unit-effort (CPUE) of standard gang index gill nets.



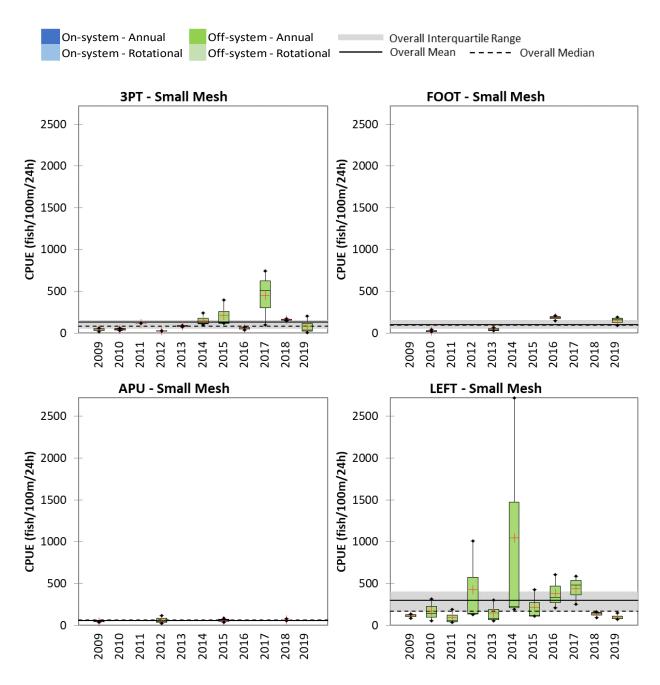
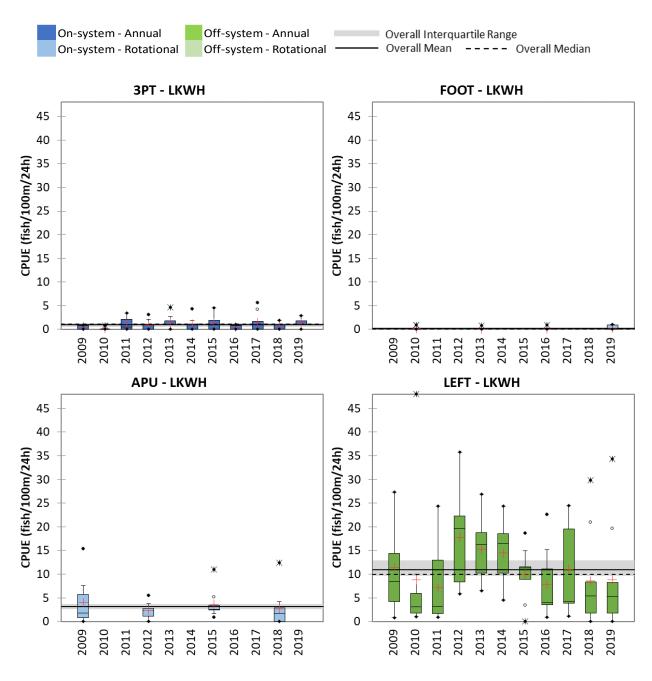


Figure 5.2-2. 2009-2019 Catch-per-unit-effort (CPUE) of small mesh index gill nets.





2009-2019 Catch-per-unit-effort (CPUE) of Lake Whitefish. Figure 5.2-3.



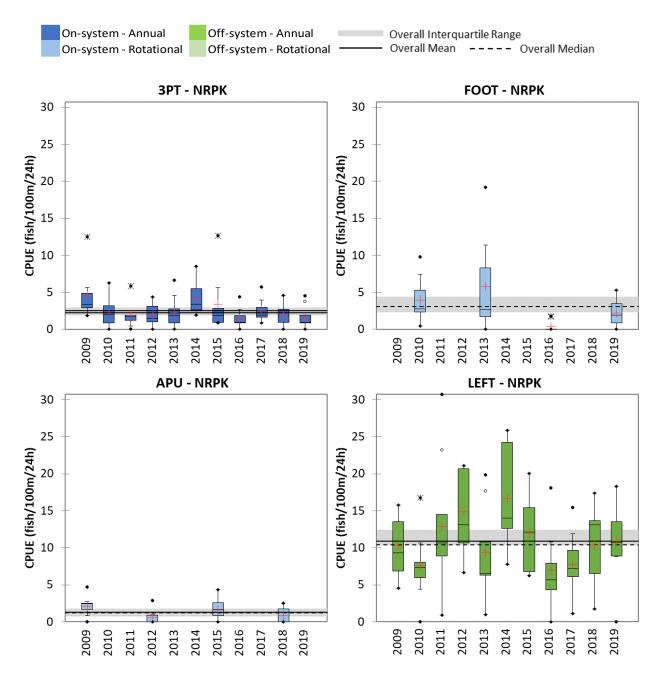


Figure 5.2-4. 2009-2019 Catch-per-unit-effort (CPUE) of Northern Pike.



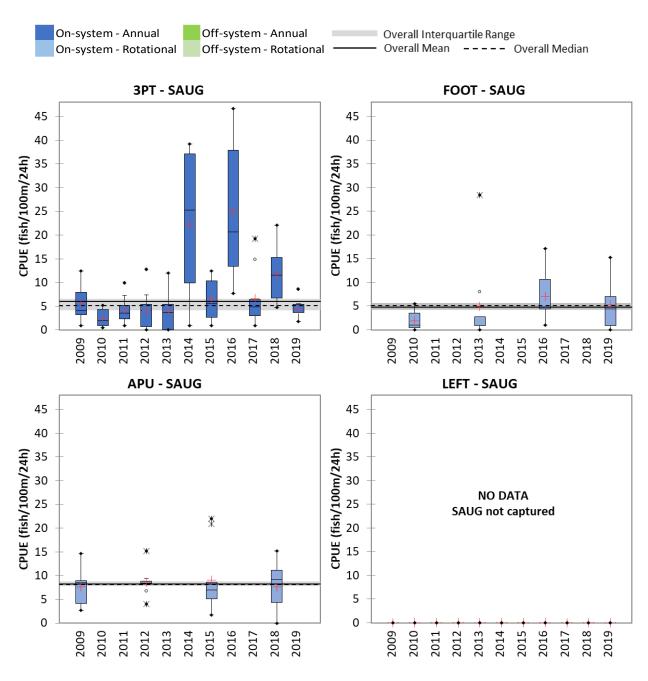


Figure 5.2-5. 2009-2019 Catch-per-unit-effort (CPUE) of Sauger.



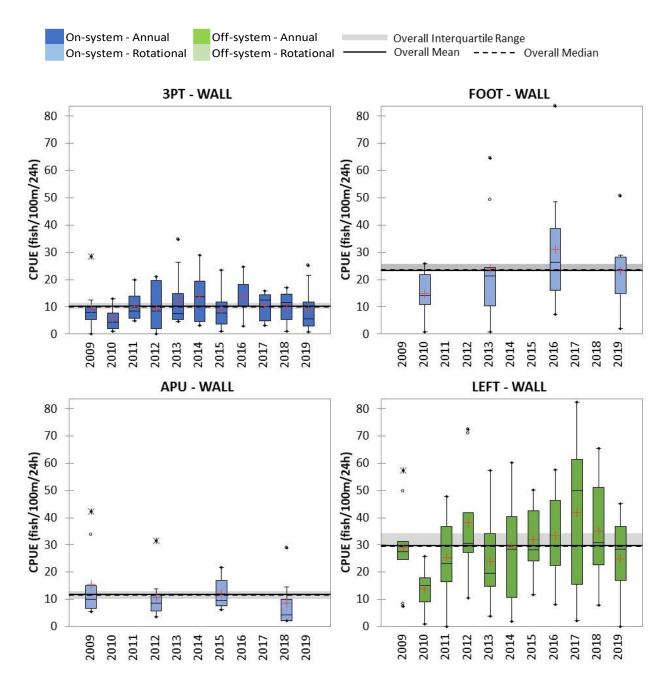


Figure 5.2-6. 2009-2019 Catch-per-unit-effort (CPUE) of Walleye.



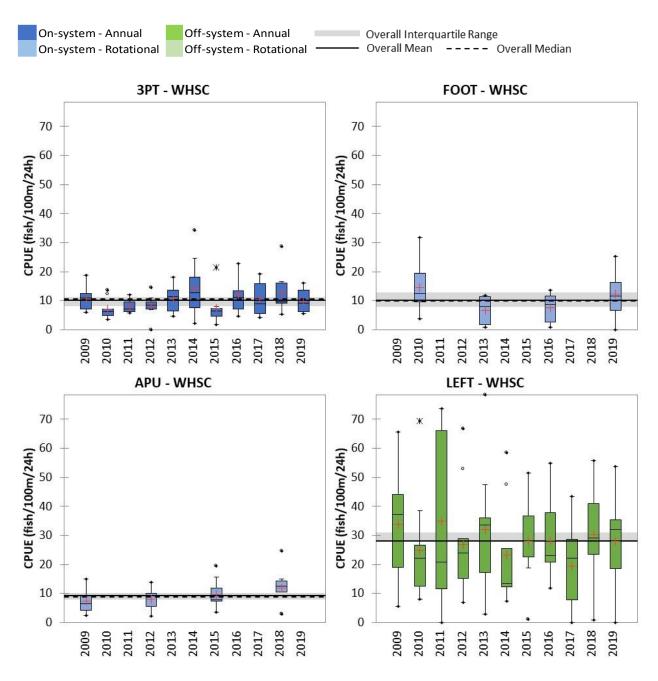


Figure 5.2-7. 2009-2019 Catch-per-unit-effort (CPUE) of White Sucker.



5.3 CONDITION

5.3.1 FULTON'S CONDITION FACTOR

5.3.1.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Lake Whitefish

The annual mean KF of Lake Whitefish between 400 and 499 mm in fork length over the 11 years of monitoring ranged from a low of 1.31 in 2017 to a high of 1.63 in 2009 (Table 5.3-1; Figure 5.3-1).

The overall mean KF was 1.43, the median was 1.40, and the IQR was 1.39-1.47 (Figure 5.3-1). The annual mean KF fell within the overall IQR except in 2011, 2013, and 2017 when it was below the IQR and in 2009, 2014, and 2018 when it was above the IQR.

Northern Pike

The annual mean KF of Northern Pike between 400 and 699 mm in fork length over the 11 years of monitoring ranged from a low of 0.64 in 2012 to a high of 0.70 in 2017 (Table 5.3-1; Figure 5.3-2).

The overall mean and median KF was 0.67, and the IQR was 0.66-0.69 (Figure 5.3-2). The annual mean KF fell within the overall IQR except in 2012, 2013, and 2019 when it was below the IQR and in 2009, 2015, and 2017 when it was above the IQR.

Sauger

Sauger was not a target species in Threepoint Lake until 2017; the annual mean KF of Sauger between 200 and 349 mm in fork length over the three years of monitoring ranged from a low of 0.85 in 2019 to a high of 0.91 in 2017 (Table 5.3-1; Figure 5.3-3).

The overall mean KF was 0.87, the median was 0.86 and the IQR was 0.86-0.88 (Figure 5.3-3). The annual mean KF fell within the overall IQR in 2018, below the IQR in 2019 and in above the IQR in 2017.



Walleye

The annual mean KF of Walleye between 300 and 499 mm in fork length over the 11 years of monitoring ranged from a low of 1.00 in 2012 to a high of 1.11 in 2009 (Table 5.3-1; Figure 5.3-4).

The overall mean and median KF were 1.06 and the IQR was 1.05-1.07 (Figure 5.3-4). The annual mean KF fell within the overall IQR except in 2012 and 2016 when it was below the IQR and in 2009, 2013, and 2014 when it was above the IQR.

White Sucker

White Sucker was not a target species in Threepoint until 2010; the annual mean KF of White Sucker between 300 and 499 mm in fork length over the 10 years of monitoring ranged from a low of 1.45 in 2019 to a high of 1.54 in 2010 (Table 5.3-1; Figure 5.3-5).

The overall mean and median KF was 1.50, and the IQR was 1.48-1.51 (Figure 5.3-5). The annual mean KF fell within the overall IQR except in 2011, 2012, and 2019 when it was below the IQR and in 2010 and 2013 when it was above the IQR.

ROTATIONAL SITES

Footprint Lake

Lake Whitefish

The annual mean KF of Lake Whitefish between 400 and 499 mm in fork length over the four years of monitoring ranged from a low of 1.46 in 2016 to a high of 1.58 in 2019 (Table 5.3-1; Figure 5.3-1).

The overall mean KF was 1.53, the median was 1.52 and the IQR was 1.51-1.55 (Figure 5.3-1). The annual mean KF fell within the overall IQR except in 2016 when it was below the IQR and in 2019 when it was above the IQR.

Northern Pike

The annual mean KF of Northern Pike between 400 and 699 mm in fork length over the four years of monitoring ranged from a low of 0.61 in 2019 to a high of 0.69 in 2013 (Table 5.3-1; Figure 5.3-2).



The overall mean KF was 0.66, the median was 0.67 and the IQR was 0.66-0.68 (Figure 5.3-2). The annual mean KF fell within the overall IQR except in 2019 when it was below the IQR and in 2013 when it was above the IQR.

Sauger

Sauger was not a target species in Footprint Lake until 2017 and KF is only available for 2019 (Table 5.3-1). The annual mean KF of Sauger between 200 and 349 mm in fork length in 2019 was 0.90 (Table 5.3-1; Figure 5.3-3).

Walleye

The annual mean KF of Walleye between 300 and 499 mm in fork length over the four years of monitoring ranged from a low of 1.03 in 2016 to a high of 1.09 in 2019 (Table 5.3-1; Figure 5.3-4).

The overall mean and median KF were 1.06 and the IQR was 1.06-1.07 (Figure 5.3-4). The annual mean KF fell within the overall IQR except in 2010 and 2016 when it was below the IQR and in 2019 when it was above the IQR.

White Sucker

The annual mean KF of White Sucker between 300 and 499 mm in fork length over the four years ranged from a low of 1.48 in 2016 to a high of 1.50 in 2013 (Table 5.3-1; Figure 5.3-5).

The overall mean and median KF was 1.49, and the IQR was 1.49-1.50 (Figure 5.3-5). The annual mean KF fell within the overall IQR except in 2016 and 2019 when it was below the IQR, and 2013 when it was above the IQR.

Apussigamasi Lake

Lake Whitefish

The annual mean KF of Lake Whitefish between 400 and 499 mm in fork length over the four years of monitoring ranged from a low of 1.59 in 2012 to a high of 1.74 in 2009 (Table 5.3-1; Figure 5.3-1).

The overall mean KF was 1.68, the median was 1.70, and the IQR was 1.67-1.71 (Figure 5.3-1). The annual mean KF fell within the overall IQR except in 2012 when it was below the IQR and in 2009 when it was above the IQR.



Northern Pike

The annual mean KF of Northern Pike between 400 and 699 mm in fork length over the four years of monitoring ranged from a low of 0.60 in 2012 to a high of 0.71 in 2009 (Table 5.3-1; Figure 5.3-2).

The overall mean KF was 0.68, the median was 0.71 and the IQR was 0.68-0.71 (Figure 5.3-2). The annual mean KF fell within the overall IQR except in 2012 when it was below the IQR and in 2009 when it was above the IQR.

Sauger

Sauger was not a target species in Apussigamasi Lake until 2017 and KF is only available for is only available for 2018 (Table 5.3-1). The annual mean KF of Sauger between 200 and 349 mm in fork length in 2018 was 0.87 (Table 5.3-1; Figure 5.3-3).

Walleye

The annual mean KF of Walleye between 300 and 499 mm in fork length over the four years of monitoring ranged from a low of 1.09 in 2012 and 2018 to a high of 1.13 in 2015 (Table 5.3-1; Figure 5.3-4).

The overall mean and median KF were 1.11 and the IQR was 1.09-1.12 (Figure 5.3-4). The annual mean KF fell within the overall IQR except in 2018 when it was below the IQR and in 2015 when it was above the IQR.

White Sucker

White Sucker was not a target species in Apussigamasi Lake until 2010; the annual mean KF of White Sucker between 300 and 499 mm in fork length over the three years of monitoring ranged from a low of 1.54 in 2015 to a high of 1.59 in 2018 (Table 5.3-1; Figure 5.3-5).

The overall mean KF was 1.56, the median was 1.55, and the IQR was 1.54-1.57 (Figure 5.3-5). The annual mean KF fell within the overall IQR except in 2015 when it was below the IQR and in 2018 when it was above the IQR.



5.3.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

Lake Whitefish

The annual mean KF of Northern Pike between 400 and 499 mm in fork length over the 11 years of monitoring ranged from a low of 1.42 in 2010 to a high of 1.63 in 2014 (Table 5.3-1; Figure 5.3-1).

The overall mean KF was 1.51, the median was 1.50, and the IQR was 1.50-1.56 (Figure 5.3-1). The annual mean KF fell within the overall IQR except in 2010 and 2016-2019 when it was below the IQR and in 2009, 2011, and 2014 when it was above the IQR.

Northern Pike

The annual mean KF of Northern Pike between 400 and 699 mm in fork length over the 11 years of monitoring ranged from a low of 0.62 in 2010 and 2015 to a high of 0.68 in 2011 (Table 5.3-1; Figure 5.3-2).

The overall mean and median KF was 0.64, and the IQR was 0.64-0.65 (Figure 5.3-2). The annual mean KF fell within the overall IQR except in 2010, 2013, and 2015 when it was below the IQR and in 2011, 2012, 2014, 2016, and 2017 when it was above the IQR.

Sauger

Sauger became a target species in 2017; it was not captured in Leftrook Lake over the three years of monitoring (Table 5.3-1).

Walleye

The annual mean KF of Walleye between 300 and 499 mm in fork length over the 11 years of monitoring ranged from a low of 0.98 in 2019 to a high of 1.10 in 2011 (Table 5.3-1; Figure 5.3-4).

The overall mean and median KF was 1.03, and the IQR was 1.03-1.04 (Figure 5.3-4). The annual mean KF fell within or was equal to the overall IQR except in 2010, 2012, 2013, 2015, and 2019 when it was below the IQR, and in 2009, 2011, and 2018 when it was above the IQR.



White Sucker

White Sucker became a targe species in 2010; the annual mean KF of White Sucker between 300 and 499 mm in fork length over the ten years of monitoring was 1.42 in 2019 and 1.53 in 2017 (Table 5.3-1; Figure 5.3-5).

The overall mean and median KF was 1.48, and the IQR was 1.48-1.49 (Figure 5.3-5). The annual mean KF fell within or was equal to the overall IQR except in 2010, 2012, 2015, and 2019 when it was below the IQR, and in 2011 and 2017 when it was above the IQR.



Table 5.3-1. 2009-2019 Fulton's condition factor of target species.

| Matarbady | Voor | | LKWH | | NRPK | | | | SAUG | ì | | WALL | | WHSC | | | |
|-----------|------|-----------------------------|------|-----------------|----------------|------|------|----------------|------|------|----------------|------|------|----------------|------|------|--|
| Waterbody | rear | n _F ¹ | Mean | SE ² | n _F | Mean | SE | |
| 3PT | 2009 | 5 | 1.63 | 0.06 | 25 | 0.69 | 0.02 | | | | 80 | 1.11 | 0.01 | | | | |
| | 2010 | 2 | 1.40 | 0.09 | 21 | 0.67 | 0.02 | | | | 76 | 1.06 | 0.01 | 82 | 1.54 | 0.01 | |
| | 2011 | 14 | 1.39 | 0.03 | 15 | 0.67 | 0.01 | | | | 87 | 1.06 | 0.01 | 96 | 1.47 | 0.01 | |
| | 2012 | 6 | 1.39 | 0.04 | 10 | 0.64 | 0.02 | | | | 68 | 1.00 | 0.01 | 72 | 1.47 | 0.01 | |
| | 2013 | 14 | 1.38 | 0.04 | 8 | 0.65 | 0.04 | | | | 117 | 1.10 | 0.01 | 98 | 1.52 | 0.01 | |
| | 2014 | 9 | 1.50 | 0.04 | 38 | 0.67 | 0.02 | | | | 94 | 1.07 | 0.01 | 101 | 1.50 | 0.01 | |
| | 2015 | 13 | 1.41 | 0.03 | 15 | 0.69 | 0.05 | | | | 69 | 1.05 | 0.01 | 69 | 1.49 | 0.01 | |
| | 2016 | 4 | 1.39 | 0.05 | 18 | 0.68 | 0.01 | | | | 88 | 1.04 | 0.01 | 112 | 1.50 | 0.01 | |
| | 2017 | 13 | 1.31 | 0.06 | 16 | 0.70 | 0.02 | 69 | 0.91 | 0.01 | 80 | 1.06 | 0.01 | 95 | 1.51 | 0.02 | |
| | 2018 | 6 | 1.49 | 0.08 | 19 | 0.67 | 0.02 | 99 | 0.86 | 0.01 | 61 | 1.06 | 0.01 | 112 | 1.51 | 0.01 | |
| | 2019 | 10 | 1.46 | 0.05 | 10 | 0.64 | 0.02 | 48 | 0.85 | 0.01 | 70 | 1.05 | 0.01 | 85 | 1.45 | 0.02 | |
| FOOT | 2010 | 2 | - | - | 22 | 0.66 | 0.02 | | | | 151 | 1.05 | 0.01 | 150 | 1.49 | 0.01 | |
| | 2013 | 1 | 1.53 | - | 37 | 0.69 | 0.01 | | | | 212 | 1.07 | 0.00 | 62 | 1.50 | 0.01 | |
| | 2016 | 1 | 1.46 | - | 4 | 0.67 | 0.02 | | | | 242 | 1.03 | 0.00 | 69 | 1.48 | 0.01 | |
| | 2019 | 2 | 1.58 | 0.02 | 9 | 0.61 | 0.02 | 58 | 0.90 | 0.01 | 213 | 1.09 | 0.01 | 119 | 1.49 | 0.02 | |
| APU | 2009 | 32 | 1.74 | 0.03 | 9 | 0.71 | 0.02 | | | | 153 | 1.12 | 0.01 | | | | |
| | 2012 | 14 | 1.59 | 0.04 | 2 | 0.60 | 0.02 | | | | 84 | 1.09 | 0.01 | 59 | 1.55 | 0.02 | |
| | 2015 | 37 | 1.69 | 0.03 | 5 | 0.70 | 0.04 | | | | 115 | 1.13 | 0.01 | 80 | 1.54 | 0.02 | |
| | 2018 | 28 | 1.70 | 0.03 | 8 | 0.71 | 0.02 | 91 | 0.87 | 0.01 | 72 | 1.09 | 0.01 | 113 | 1.59 | 0.01 | |
| LEFT | 2009 | 109 | 1.60 | 0.01 | 118 | 0.64 | 0.01 | | | | 348 | 1.04 | 0.00 | | | | |
| | 2010 | 83 | 1.42 | 0.02 | 66 | 0.62 | 0.01 | | | | 104 | 1.02 | 0.01 | 190 | 1.46 | 0.01 | |
| | 2011 | 61 | 1.59 | 0.02 | 100 | 0.68 | 0.01 | | | | 174 | 1.10 | 0.01 | 292 | 1.51 | 0.01 | |
| | 2012 | 120 | 1.51 | 0.01 | 95 | 0.67 | 0.01 | | | | 227 | 1.02 | 0.00 | 185 | 1.45 | 0.01 | |
| | 2013 | 140 | 1.50 | 0.01 | 79 | 0.63 | 0.01 | | | | 183 | 1.01 | 0.01 | 283 | 1.49 | 0.01 | |
| | 2014 | 128 | 1.63 | 0.01 | 141 | 0.66 | 0.01 | | | | 225 | 1.03 | 0.01 | 200 | 1.48 | 0.01 | |
| | 2015 | 87 | 1.54 | 0.02 | 95 | 0.62 | 0.01 | | | | 293 | 1.00 | 0.00 | 248 | 1.45 | 0.01 | |
| | 2016 | 71 | 1.50 | 0.01 | 53 | 0.64 | 0.01 | | | | 247 | 1.03 | 0.00 | 261 | 1.48 | 0.01 | |
| | 2017 | 100 | 1.46 | 0.01 | 79 | 0.64 | 0.01 | - | - | - | 336 | 1.03 | 0.00 | 161 | 1.53 | 0.01 | |
| | 2018 | 83 | 1.44 | 0.02 | 98 | 0.64 | 0.01 | - | | - | 360 | 1.04 | 0.00 | 278 | 1.49 | 0.01 | |
| | 2019 | 94 | 1.44 | 0.01 | 97 | 0.65 | 0.01 | - | - | - | 243 | 0.98 | 0.00 | 257 | 1.42 | 0.01 | |

- 1. n_F = number of fish measured for length and weight.
- 2. SE = standard error.
- 3. Grey shading indicates a species was not a target species in that year.



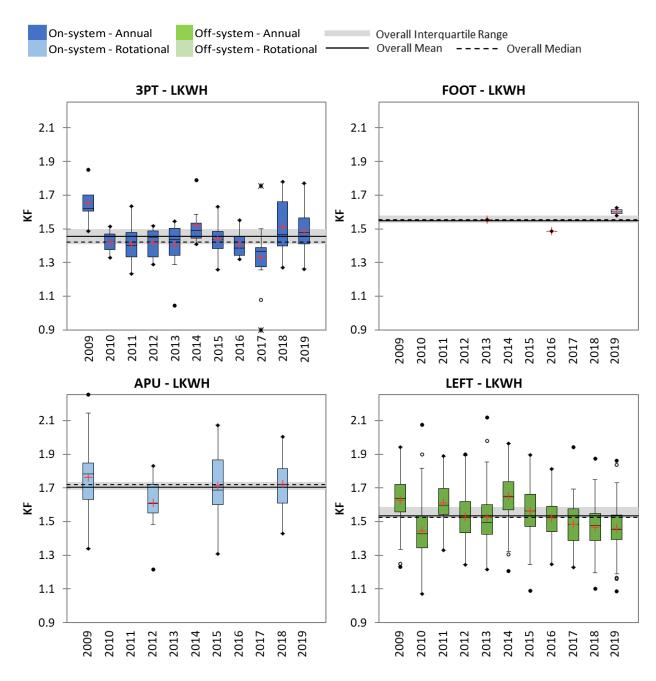


Figure 5.3-1. 2009-2019 Fulton's condition factor (KF) of Lake Whitefish.



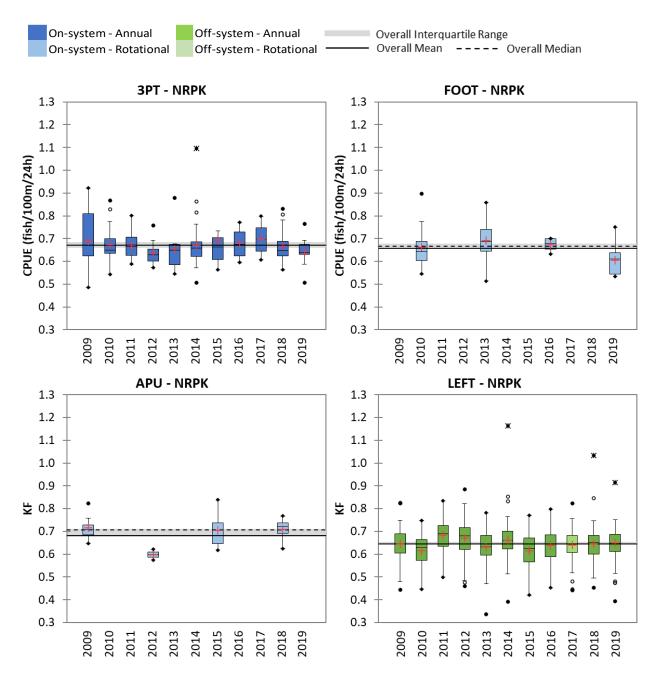


Figure 5.3-2. 2009-2019 Fulton's condition factor (KF) of Northern Pike.



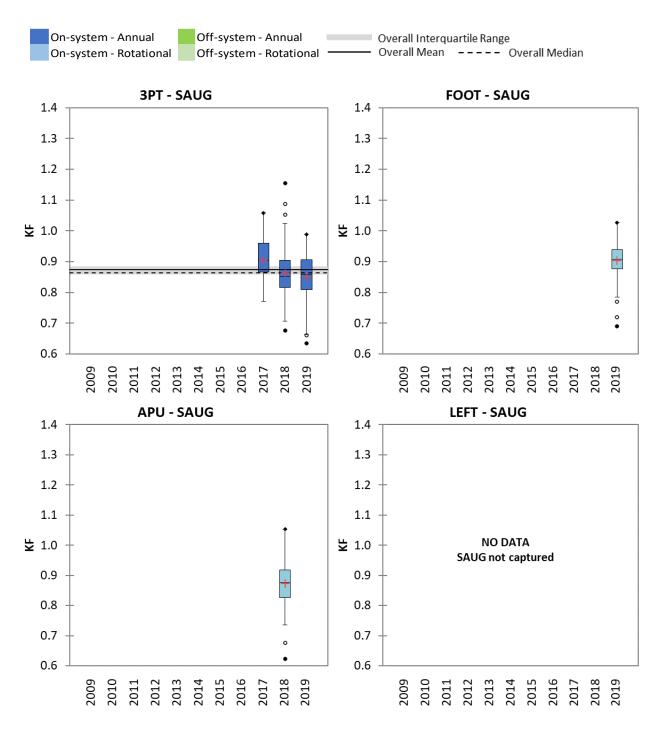


Figure 5.3-3. 2017-2019 Fulton's condition factor (KF) of Sauger.



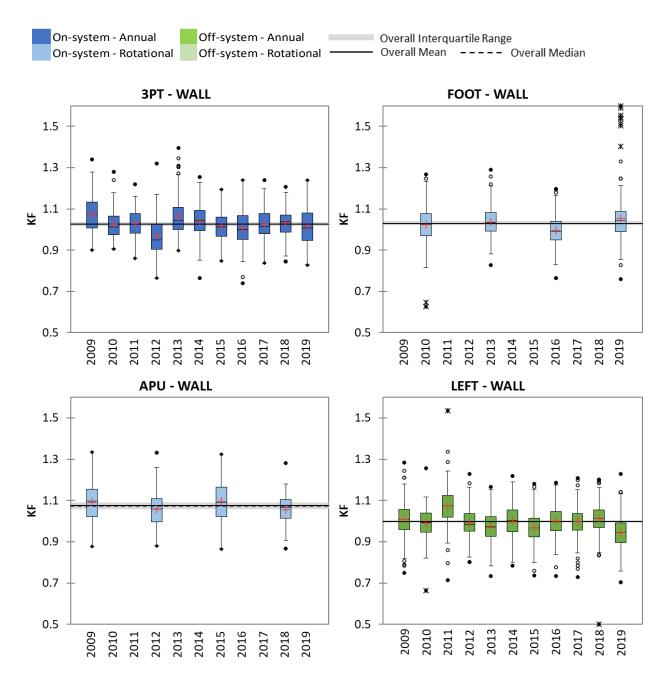


Figure 5.3-4. 2009-2019 Fulton's condition factor (KF) of Walleye.



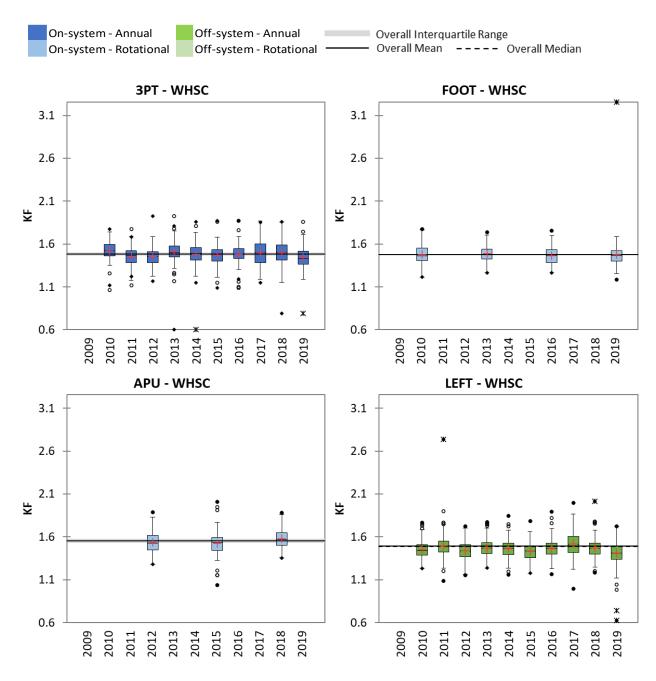


Figure 5.3-5. 2010-2019 Fulton's condition factor (KF) of White Sucker.



5.3.2 RELATIVE WEIGHT

5.3.2.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Lake Whitefish

The annual mean Wr of Lake Whitefish greater than 99 mm and less than 701 mm in total length over the 11 years of monitoring ranged from a low of 91 in 2017 to a high of 118 in 2009 (Table 5.3-2; Figure 5.3-6).

The overall mean Wr was 100, the median was 98, and the IQR was 97-103 (Figure 5.3-6). The annual mean Wr fell within the overall IQR except in 2016 and 2017 when it was below the IQR and in 2009, 2014, and 2018 when it was above the IQR.

Northern Pike

The annual mean Wr of Northern Pike greater than 99 mm in total length over the 11 years of monitoring ranged from a low of 80 in 2012 to a high of 91 in 2016 (Table 5.3-2; Figure 5.3-7).

The overall mean Wr was 84, the median was 83, and the IQR was 81-84 (Figure 5.3-7). The annual mean Wr fell within the overall IQR except in 2012 when it was below the IQR and in 2010, 2016, and 2017 when it was above the IQR.

Sauger

Sauger was not a target species in Threepoint Lake until 2017; the annual mean Wr of Sauger greater than 69 mm in total length over the three years of monitoring ranged from a low of 82 in 2019 to a high of 85 in 2017 (Table 5.3-2; Figure 5.3-8).

The overall mean and median Wr were 83, and the IQR was 82-84 (Figure 5.3-8). The annual mean Wr fell within the overall IQR except in 2009 when it was above the IQR.

Walleye

The annual mean Wr of Walleye greater than 29 mm in total length over the 11 years of monitoring ranged from a low of 78 in 2012 to a high of 88 in 2009, 2013, and 2017 (Table 5.3-2; Figure 5.3-9).



The overall mean and median Wr were 85 and the IQR was 84-87 (Figure 5.3-9). The annual mean Wr fell within the overall IQR except in 2012 and 2015 when it was below the IQR and in 2009, 2013, and 2017 when it was above the IQR.

White Sucker

White Sucker was not a target species in Threepoint until 2010; the annual mean Wr of White Sucker greater than 99 mm in total length over the ten years of monitoring ranged from a low of 93 in 2019 to a high of 99 in 2010 and 2013 (Table 5.3-2; Figure 5.3-10).

The overall mean and median Wr were 97, and the IQR was 95-98 (Figure 5.3-10). The annual mean Wr fell within the overall IQR except in 2019 when it was below the IQR and in 2010 and 2013 when it was above the IQR.

ROTATIONAL SITES

Fooprint Lake

Lake Whitefish

The annual mean Wr of Lake Whitefish greater than 99 mm and less than 701 mm in total length over the four years of monitoring ranged from a low of 100 in 2016 to a high of 110 in 2019 (Table 5.3-2; Figure 5.3-6).

The overall mean and median Wr were 105, and the IQR was 104-107 (Figure 5.3-6). The annual mean Wr fell within the overall IQR except in 2016 when it was below the IQR and in 2019 when it was above the IQR.

Northern Pike

The annual mean Wr of Northern Pike greater than 99 mm in total length over the four years of monitoring ranged from a low of 81 in 2016 to a high of 87 in 2013 (Table 5.3-2; Figure 5.3-7).

The overall mean and median Wr were 84, and the IQR was 83-85 (Figure 5.3-7). The annual mean Wr fell within the overall IQR except in 2016 when it was below the IQR and in 2013 when it was above the IQR.



Sauger

Sauger was not a target species in Footprint Lake until 2017; Wr is only available for 2019 (Table 5.3-1). The annual mean Wr of Sauger greater than 69 mm in total length in 2019 was 86 (Table 5.3-1; Figure 5.3-8).

Walleye

The annual mean Wr of Walleye greater than 29 mm in total length over the four years of monitoring ranged from a low of 83 in 2016 to a high of 88 in 2019 (Table 5.3-2; Figure 5.3-9).

The overall mean and median Wr were 86, and the IQR was 86-87 (Figure 5.3-9). The annual mean Wr fell within the overall IQR except in 2016 when it was below the IQR and in 2019 when it was above the IQR.

White Sucker

The annual mean Wr of White Sucker greater than 99 mm in total length ranged from a low of 95 in 2016 and 2019 to a high of 97 in 2013 (Table 5.3-2; Figure 5.3-10).

The overall mean and median Wr were 96, and the IQR was 95-96 (Figure 5.3-10). The annual mean Wr fell within the overall IQR except in 2013 when it was above the IQR.

Apussigamasi Lake

Lake Whitefish

The annual mean Wr of Lake Whitefish greater than 99 mm and less than 701 mm in total length over the four years of monitoring ranged from a low of 109 in 2012 to a high of 121 in 2009 (Table 5.3-2; Figure 5.3-6).

The overall mean Wr was 116, the median was 117, and the IQR was 115-118 (Figure 5.3-6). The annual mean Wr fell within the overall IQR except in 2012 when it was below the IQR and in 2009 when it was above the IQR.

Northern Pike

The annual mean Wr of Northern Pike greater than 99 mm in total length over the four years of monitoring ranged from a low of 79 in 2012 to a high of 92 in 2009 (Table 5.3-2; Figure 5.3-7).



The overall mean Wr was 87, the median was 88, and the IQR was 85-89 (Figure 5.3-7). The annual mean Wr fell within the overall IQR except in 2012 when it was below the IQR and in 2009 when it was above the IQR.

Sauger

Sauger was not a target species in Apussigamasi Lake until 2017; Wr is only available for 2018 (Table 5.3-2). The annual mean Wr of Sauger greater than 69 mm in total length in 2018 was 83 (Table 5.3-2; Figure 5.3-8).

Walleye

The annual mean Wr of Walleye greater than 29 mm in total length over the four years of monitoring ranged from a low of 87 in 2012 to a high of 91 in 2015 (Table 5.3-2; Figure 5.3-9).

The overall mean and median Wr were 89, and the IQR was 87-90 (Figure 5.3-9). The annual mean Wr fell within the overall IQR except in 2015 when it was above the IQR.

White Sucker

White sucker was not a target species in Apussigamasi Lake until 2010; the annual mean Wr of White Sucker greater than 99 mm in total length over the three years of monitoring ranged from a low of 97 in 2012 to a high of 103 in 2018 (Table 5.3-2; Figure 5.3-10).

The overall mean Wr was 100, the median was 99, and the IQR was 98-101 (Figure 5.3-10). The annual mean Wr fell within the overall IQR in 2015, below the IQR in 2012, and above the IQR in 2018.

5.3.2.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

Lake Whitefish

The annual mean Wr of Lake Whitefish greater than 99 mm and less than 701 mm in total length over the 11 years of monitoring ranged from a low of 98 in 2010 to a high of 112 in 2014 (Table 5.3-2; Figure 5.3-6).



The overall mean Wr was 104, the median was 103, and the IQR was 103-107 (Figure 5.3-6). The annual mean Wr fell within the overall IQR except in 2010 and 2017-2019 when it was below the IQR and in 2009, 2011, and 2014 when it was above the IQR.

Northern Pike

The annual mean Wr of Northern Pike greater than 99 mm in total length over the 11 years of monitoring ranged from a low of 77 in 2010 to a high of 87 in 2011 (Table 5.3-2; Figure 5.3-7).

The overall mean and median Wr were 81, and the IQR was 81-83 (Figure 5.3-7). The annual mean Wr fell within the overall IQR except in 2010, 2013, and 2015 when it was below the IQR and in 2011 and 2012 when it was above the IQR.

Sauger

Sauger was a not a target species in Leftrook Lake until 2017; Sauger were not captured in Leftrook Lake over the three years of monitoring (Table 5.2-2).

Walleye

The annual mean Wr of Walleye greater than 29 mm in total length over the 11 years of monitoring ranged from a low of 80 in 2019 to a high of 91 in 2011 (Table 5.3-2; Figure 5.3-9).

The overall mean and median Wr were 83, and the IQR was 83-84 (Figure 5.3-9). The annual mean Wr fell within the overall IQR except in 2010, 2013, 2015, and 2019 when it was below the IQR and in 2011 and 2018 when it was above the IQR.

White Sucker

White sucker was not a target species in Leftrook Lake until 2010; the annual mean Wr of White Sucker greater than 99 mm in total length over the ten years of monitoring ranged from a low of 92 in 2019 to a high of 98 in 2017 (Table 5.3-2; Figure 5.3-10).

The overall mean and median Wr was 95, and the IQR was 95-96 (Figure 5.3-10). The annual mean Wr fell within the overall IQR except in 2010, 2012, 2015, and 2019 when it was below the IQR and in 2011 and 2017 when it was above the IQR.



Table 5.3-2. 2009-2019 relative weight of target species.

| 14/ataulaada | V | LKWH | | | NRPK | | | SAUG | | | | WALL | | | WHSC | |
|--------------|------|-----------------------------|------|-----------------|----------------|------|-----|----------------|------|-----|----------------|------|-----|----------------|------|-----|
| Waterbody | Year | n _F ¹ | Mean | SE ² | n _F | Mean | SE |
| 3PT | 2009 | 7 | 118 | 4.7 | 50 | 84 | 1.9 | | | | 103 | 88 | 1.2 | 10 | 96 | 1.5 |
| | 2010 | 2 | 98 | 4.2 | 32 | 85 | 2.6 | | | | 91 | 85 | 0.6 | 87 | 99 | 0.8 |
| | 2011 | 15 | 98 | 2.9 | 24 | 81 | 1.5 | | | | 111 | 86 | 0.9 | 96 | 95 | 0.8 |
| | 2012 | 7 | 97 | 2.9 | 18 | 80 | 2.0 | | | | 109 | 78 | 0.8 | 76 | 95 | 0.8 |
| | 2013 | 15 | 97 | 2.7 | 26 | 84 | 2.5 | | | | 127 | 88 | 0.7 | 104 | 99 | 0.8 |
| | 2014 | 10 | 105 | 2.7 | 66 | 83 | 1.3 | | | | 127 | 85 | 0.8 | 113 | 96 | 0.9 |
| | 2015 | 13 | 98 | 2.1 | 32 | 81 | 3.0 | | | | 91 | 83 | 0.7 | 70 | 96 | 0.9 |
| | 2016 | 6 | 95 | 2.8 | 29 | 91 | 9.8 | | | | 142 | 84 | 0.6 | 115 | 97 | 0.8 |
| | 2017 | 15 | 91 | 3.7 | 26 | 86 | 1.6 | 77 | 85 | 1.0 | 120 | 88 | 1.6 | 104 | 98 | 0.9 |
| | 2018 | 7 | 104 | 5.2 | 21 | 81 | 2.0 | 118 | 83 | 0.7 | 100 | 86 | 0.7 | 123 | 97 | 0.8 |
| | 2019 | 12 | 102 | 3.2 | 15 | 82 | 2.2 | 60 | 82 | 1.0 | 96 | 85 | 0.7 | 96 | 93 | 1.0 |
| FOOT | 2010 | 0 | - | 1 | 42 | 83 | 1.2 | | | | 182 | 86 | 0.5 | 155 | 96 | 0.6 |
| | 2013 | 1 | 105 | | 63 | 87 | 1.2 | | | | 251 | 87 | 0.4 | 66 | 97 | 0.7 |
| | 2016 | 1 | 100 | | 12 | 81 | 1.5 | | | | 331 | 83 | 0.3 | 73 | 95 | 0.8 |
| | 2019 | 3 | 110 | 0.5 | 28 | 85 | 5.7 | 70 | 86 | 0.8 | 259 | 88 | 0.8 | 121 | 95 | 0.6 |
| APU | 2009 | 41 | 121 | 1.9 | 25 | 92 | 2.1 | 127 | 83 | 0.5 | 179 | 90 | 0.5 | 77 | 103 | 0.8 |
| | 2012 | 22 | 109 | 2.5 | 7 | 79 | 4.9 | | | | 113 | 87 | 0.7 | 78 | 97 | 1.2 |
| | 2015 | 38 | 117 | 1.9 | 17 | 88 | 3.0 | | | | 156 | 91 | 0.8 | 101 | 99 | 1.1 |
| | 2018 | 31 | 117 | 2.0 | 9 | 88 | 2.2 | 103 | 83 | 0.7 | 117 | 88 | 1.2 | 123 | 103 | 0.7 |
| LEFT | 2009 | 122 | 109 | 8.0 | 126 | 80 | 8.0 | | | | 393 | 84 | 0.3 | 322 | 96 | 0.4 |
| | 2010 | 88 | 98 | 1.2 | 71 | 77 | 1.0 | | | | 108 | 82 | 0.6 | 191 | 94 | 0.5 |
| | 2011 | 65 | 109 | 1.1 | 112 | 87 | 1.1 | | | | 220 | 91 | 0.7 | 295 | 97 | 0.5 |
| | 2012 | 124 | 104 | 0.9 | 120 | 85 | 0.9 | | | | 287 | 83 | 0.4 | 192 | 94 | 0.5 |
| | 2013 | 143 | 103 | 8.0 | 106 | 80 | 0.9 | | | | 245 | 81 | 0.4 | 292 | 96 | 0.4 |
| | 2014 | 132 | 112 | 8.0 | 170 | 83 | 0.7 | | | | 277 | 83 | 0.4 | 204 | 95 | 0.4 |
| | 2015 | 91 | 106 | 1.0 | 127 | 79 | 0.9 | | | | 354 | 81 | 0.3 | 252 | 93 | 0.4 |
| | 2016 | 72 | 103 | 1.0 | 78 | 81 | 1.1 | | | | 344 | 84 | 0.4 | 265 | 96 | 0.4 |
| | 2017 | 102 | 100 | 0.8 | 94 | 81 | 0.9 | 0 | - | - | 419 | 84 | 0.3 | 173 | 98 | 0.8 |
| | 2018 | 85 | 99 | 1.0 | 128 | 82 | 0.8 | 0 | - | - | 434 | 85 | 0.3 | 301 | 96 | 0.4 |
| | 2019 | 95 | 99 | 1.0 | 112 | 81 | 0.8 | 0 | - | | 313 | 80 | 0.4 | 280 | 92 | 0.6 |

Notes:

- 1. n_F = number of fish measured for length and weight.
- 2. SE = standard error.
- 3. Grey shading indicates a species was not a target species in that year.



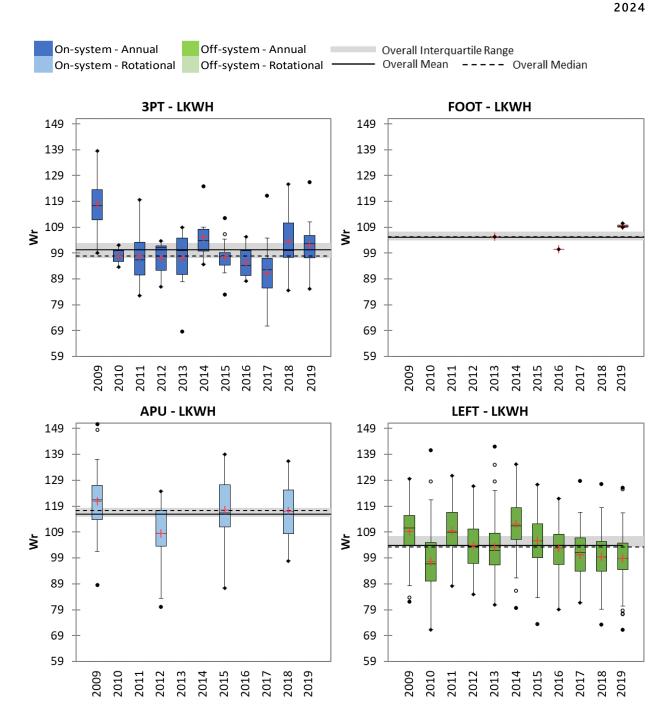


Figure 5.3-6. Relative weight (Wr) of Lake Whitefish.



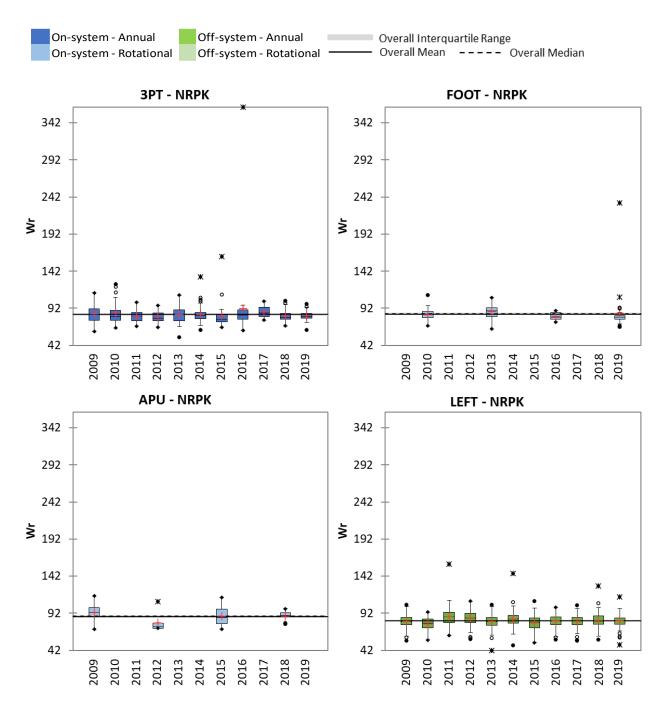


Figure 5.3-7. Relative weight (Wr) of Northern Pike.



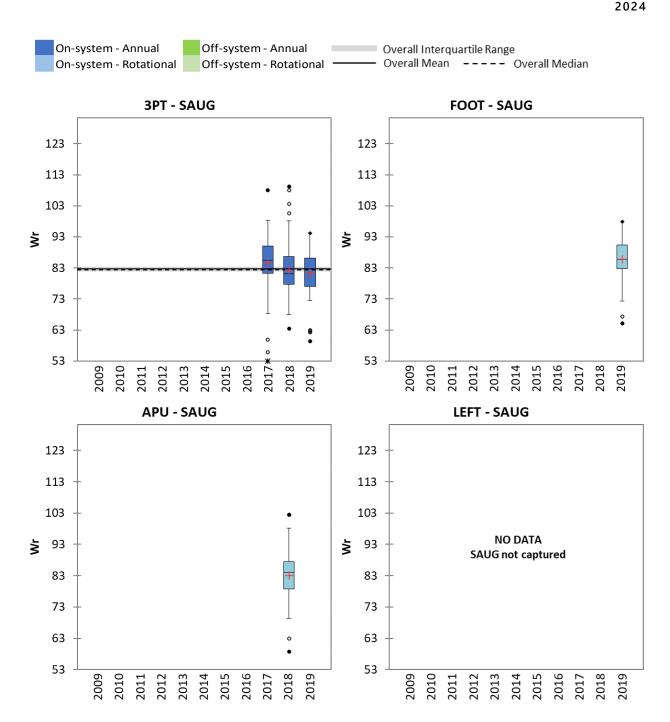


Figure 5.3-8. Relative weight (Wr) of Sauger.



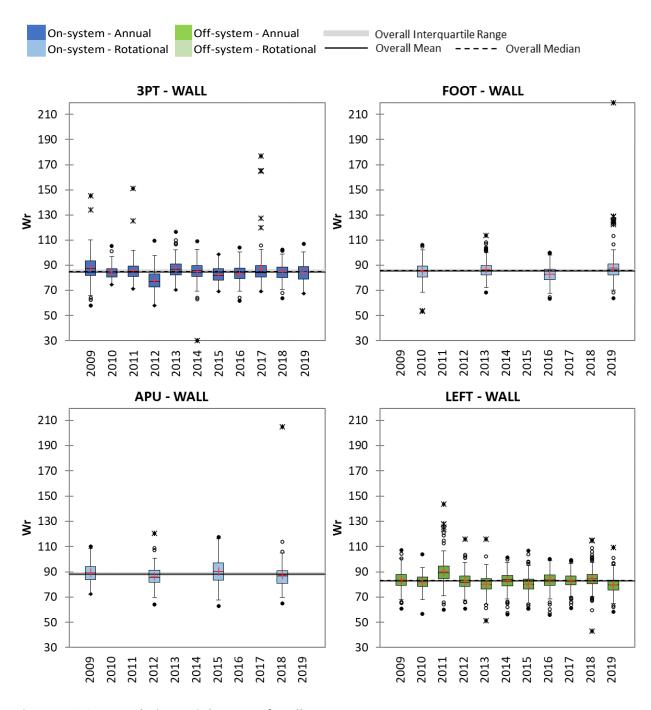


Figure 5.3-9. Relative weight (Wr) of Walleye.



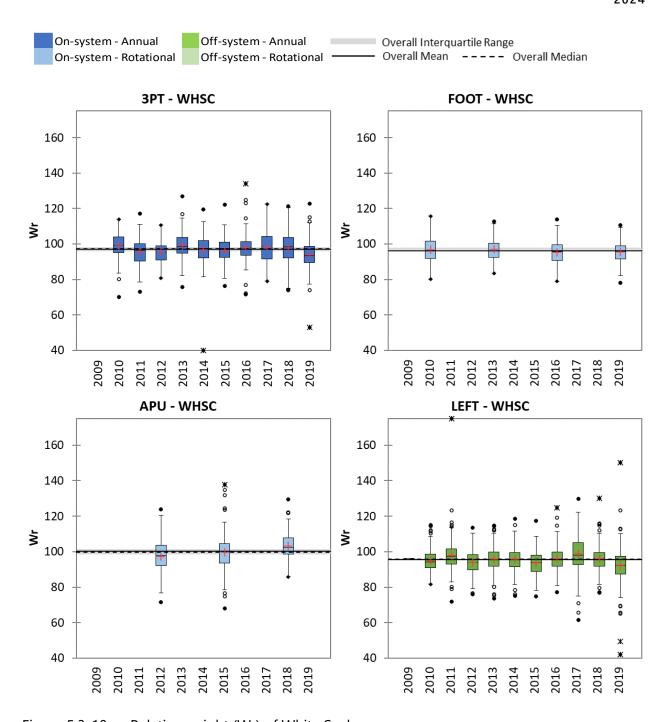


Figure 5.3-10. Relative weight (Wr) of White Sucker.



5.4 GROWTH

5.4.1 LENGTH-AT-AGE

5.4.1.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Lake Whitefish

The annual mean FLA of 4-year-old Lake Whitefish over the 11 years of monitoring ranged from a low of 200 in 2017 to a high of 372 mm in 2015 (Table 5.4-1; Figure 5.4-1).

The overall mean FLA was 299, the median was 309, and the IQR was 270-329 mm (Figure 5.4-1). The annual mean FLA fell within the overall IQR except in 2009 and 2017 when it was below the IQR and in 2013-2015 when it was above the IQR.

Northern Pike

The annual mean FLA of 4-year-old Northern Pike over the 11 years of monitoring ranged from a low of 361 in 2011 to a high of 476 mm in 2019 (Table 5.4-1; Figure 5.4-2).

The overall mean FLA was 418, the median was 429, and the IQR was 396-438 mm (Figure 5.4-2). The annual mean FLA fell within the overall IQR except in 2009, 2011, and 2013 when it was below the IQR and in 2015, 2018, and 2019 when it was above the IQR.

Sauger

Individual Sauger were measured for length and aged at Threepoint Lake starting in 2017 (Table 5.3-1). The annual mean FLA of 3-year-old Sauger over the three years of monitoring ranged from a low of 170 in 2017 to a high of 186 mm in 2018 (Table 5.4-1; Figure 5.4-3).

The overall mean FLA was 179, the median was 181, and the IQR was 175-183 mm (Figure 5.4-3). The annual mean FLA fell within the overall IQR except in 2017 when it was below the IQR and in 2018 when it was above the IQR.

Walleye

The annual mean FLA of 3-year-old Walleye over the 11 years of monitoring ranged from a low of 193 in 2012 to a high of 250 mm in 2013 (Table 5.4-1; Figure 5.4-4).



The overall mean FLA was 227, the median was 230, and the IQR was 221-237 mm (Figure 5.4-4). The annual mean FLA fell within the overall IQR except in 2012 and 2018 when it was below the IQR and in 2013 and 2016 when it was above the IQR.

ROTATIONAL SITES

Footprint Lake

Lake Whitefish

Age-4 Lake Whitefish were not captured in Footprint Lake over the four years of monitoring (Table 5.2-1).

Northern Pike

The annual mean FLA of 4-year-old Northern Pike over the four years of monitoring ranged from a low of 370 in 2010 to a high of 437 mm in 2019 (Table 5.4-1; Figure 5.4-2).

The overall mean FLA was 402, the median was 401, and the IQR was 380-423 mm (Figure 5.4-2). The annual mean FLA fell within the overall IQR except in 2010 when it was below the IQR and in 2019 when it was above the IQR.

Sauger

Sauger was not a target species in Footprint Lake until 2017 and FLA is only available for 2019 (Table 5.3-1). The annual mean FLA of 3-year-old Sauger in 2019 was 297 mm (Table 5.4-1; Figure 5.4-3).

Walleye

The annual mean FLA of 3-year-old Walleye over the four years of monitoring ranged from a low of 217 in 2010 to a high of 250 mm in 2013 (Table 5.4-1; Figure 5.4-4).

The overall mean FLA was 227, the median was 221, and the IQR was 220-229 mm (Figure 5.4-4). The annual mean FLA fell within the overall IQR except in 2010 when it was below the IQR and in 2013 when it was above the IQR.



Apussigamasi Lake

Lake Whitefish

The annual mean FLA of 4-year-old Lake Whitefish over the four years of monitoring ranged from a low of 267 in 2018 to a high of 361 mm in 2015 (Table 5.4-1; Figure 5.4-1).

The overall mean FLA was 318, the median was 321, and the IQR was 301-338 mm (Figure 5.4-2). The annual mean FLA fell within the overall IQR except in 2018 when it was below the IQR and in 2015 when it was above the IQR.

Northern Pike

The annual mean FLA of 4-year-old Northern Pike over the four years of monitoring ranged from a low of 414 in 2012 to a high of 564 mm in 2018 (Table 5.4-1; Figure 5.4-2). No age-4 Northern Pike were captured in 2015.

The overall mean FLA was 468, the median was 426, and the IQR was 423-495 mm (Figure 5.4-2). The annual mean FLA fell within the overall IQR except in 2012 when it was below the IQR and in 2018 when it was above the IQR.

Sauger

Individual Sauger were measured for length and aged at Apussigamasi Lake starting in 2017 when it became a target species. (Table 5.3-1). The annual mean FLA of 3-year-old Sauger in 2018 was 182 mm (Table 5.3-1; Figure 5.3-3).

Walleye

The annual mean FLA of 3-year-old Walleye over the four years of monitoring ranged from a low of 209 in 2009 to a high of 244 mm in 2015 (Table 5.4-1; Figure 5.4-4).

The overall mean FLA was 228, the median was 229, and the IQR was 222-235 mm (Figure 5.4-4). The annual mean FLA fell within the overall IQR except in 2009 when it was below the IQR and in 2015 when it was above the IQR.



5.4.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

Lake Whitefish

The annual mean FLA of 4-year-old Lake Whitefish over the 11 years of monitoring ranged from a low of 297 in 2010 to a high of 406 mm in 2016 (Table 5.4-1; Figure 5.4-1).

The overall mean FLA was 349, the median was 350, and the IQR was 349-371 mm (Figure 5.4-1). The annual mean FLA fell within the overall IQR except in 2009-2011, 2015, and 2019 when it was below the IQR and in 2013, 2016, and 2017 when it was above the IQR.

Northern Pike

The annual mean FLA of 4-year-old Northern Pike over the 11 years of monitoring ranged from a low of 422 in 2012 to a high of 477 mm in 2019 (Table 5.4-1; Figure 5.4-2).

The overall mean FLA was 445, the median was 444, and the IQR was 443-454 mm (Figure 5.4-2). The annual mean FLA fell within the overall IQR except in 2010-2012 and 2015 when it was below the IQR and in 2013, 2018, and 2019 when it was above the IQR.

Sauger

Age-3 Sauger were not captured in Leftrook Lake over the 11 years of monitoring (Table 5.2-1).

Walleye

The annual mean FLA of 3-year-old Walleye over the 11 years of monitoring ranged from a low of 191 in 2009 to a high of 245 mm in 2016 (Table 5.4-1; Figure 5.4-4).

The overall mean FLA was 230, the median was 235, and the IQR was 235-237 mm (Figure 5.4-4). The annual mean FLA fell within the overall IQR except in 2009, 2011, 2012 and 2014 when it was below the IQR and in 2015, 2016, and 2018 when it was above the IQR.



Table 5.4-1. 2009-2019 Fork length-at-age of target species.

| Mataubad. | Vaar | | LKWH | | | NRPK | | | SAUG | | | WALL | |
|-----------|------|-----------------------------|------|-----------------|----------------|------|----|----------------|------|----|----------------|------|----|
| Waterbody | Year | n _F ¹ | Mean | SE ² | n _F | Mean | SE | n _F | Mean | SE | n _F | Mean | SE |
| 3PT | 2009 | 1 | 260 | - | 15 | 377 | 7 | | | | 0 | - | - |
| | 2010 | 0 | - | - | 5 | 411 | 8 | | | | 2 | 224 | 16 |
| | 2011 | 1 | 307 | - | 7 | 361 | 13 | | | | 0 | - | - |
| | 2012 | 0 | - | - | 5 | 399 | 12 | | | | 4 | 193 | 26 |
| | 2013 | 1 | 273 | - | 4 | 393 | 12 | | | | 2 | 250 | 6 |
| | 2014 | 3 | 343 | 16 | 18 | 432 | 11 | | | | 7 | 232 | 16 |
| | 2015 | 3 | 372 | 20 | 4 | 439 | 37 | | | | 8 | 228 | 6 |
| | 2016 | 1 | 325 | - | 10 | 437 | 19 | | | | 17 | 239 | 6 |
| | 2017 | 1 | 200 | - | 5 | 429 | 44 | 5 | 170 | 6 | 4 | 236 | 17 |
| | 2018 | 0 | - | - | 5 | 445 | 36 | 12 | 186 | 4 | 7 | 212 | 13 |
| | 2019 | 1 | 310 | - | 2 | 476 | 36 | 6 | 181 | 8 | 0 | - | - |
| FOOT | 2010 | 0 | - | - | 11 | 370 | 13 | | | | 2 | 217 | 1 |
| | 2013 | 0 | - | - | 22 | 418 | 12 | | | | 14 | 250 | 5 |
| | 2016 | 0 | - | - | 6 | 383 | 15 | | | | 28 | 222 | 2 |
| | 2019 | 0 | - | - | 9 | 437 | 27 | 2 | 197 | 24 | 1 | 220 | - |
| APU | 2009 | 3 | 331 | 33 | 1 | 426 | - | | | | 2 | 209 | 9 |
| | 2012 | 3 | 312 | 27 | 2 | 414 | 63 | | | | 9 | 226 | 2 |
| | 2015 | 3 | 361 | 14 | 0 | - | - | | | | 5 | 244 | 9 |
| | 2018 | 1 | 267 | - | 1 | 564 | - | 3 | 182 | 7 | 7 | 232 | 11 |
| LEFT | 2009 | 4 | 338 | 21 | 11 | 444 | 8 | | | | 6 | 191 | 13 |
| | 2010 | 2 | 297 | 81 | 15 | 436 | 8 | | | | 0 | - | - |
| | 2011 | 7 | 315 | 13 | 23 | 424 | 6 | | | | 9 | 229 | 3 |
| | 2012 | 7 | 352 | 8 | 30 | 422 | 7 | | | | 10 | 228 | 5 |
| | 2013 | 7 | 375 | 5 | 24 | 461 | 8 | | | | 31 | 236 | 3 |
| | 2014 | 14 | 350 | 6 | 44 | 452 | 6 | | | | 12 | 226 | 6 |
| | 2015 | 2 | 318 | 3 | 18 | 433 | 12 | | | | 20 | 238 | 4 |
| | 2016 | 2 | 406 | 6 | 26 | 443 | 10 | | | | 21 | 245 | 3 |
| | 2017 | 6 | 380 | 8 | 34 | 450 | 7 | 0 | - | - | 16 | 235 | 4 |
| | 2018 | 1 | 367 | - | 42 | 456 | 5 | 0 | - | - | 24 | 241 | 4 |
| | 2019 | 1 | 348 | - | 50 | 477 | 6 | 0 | - | - | 6 | 236 | 6 |

Notes:

1. n_F = number of fish measured for length and weight.

2. SE = standard error.

3. Grey shading indicates that a species was not a target species in that year.



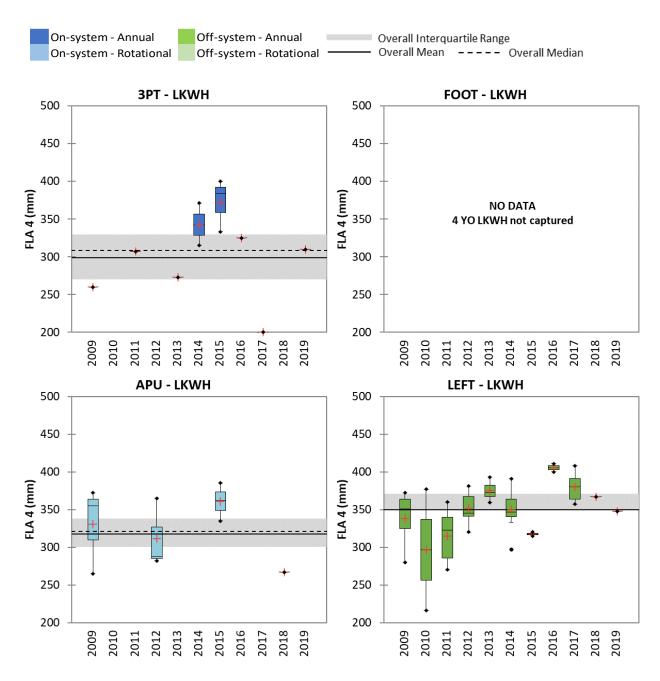


Figure 5.4-1. 2009-2019 Fork length-at-age (FLA) 4 of Lake Whitefish.



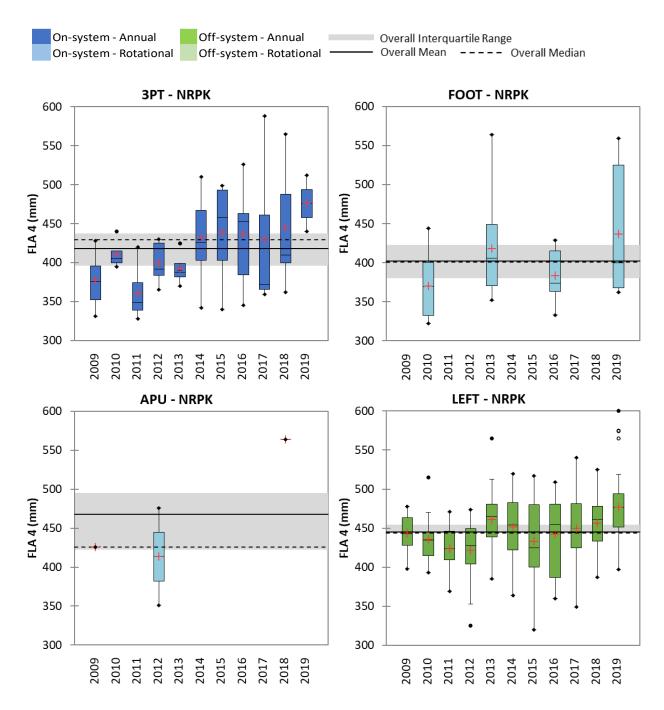


Figure 5.4-2. 2009-2019 Fork length-at-age (FLA) 4 of Northern Pike.



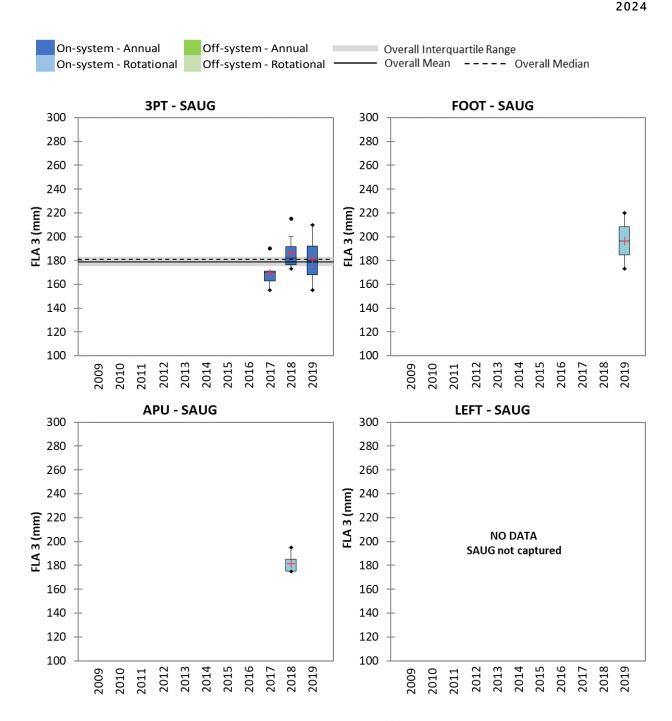


Figure 5.4-3. 2017-2019 Fork length-at-age (FLA) 3 of Sauger.



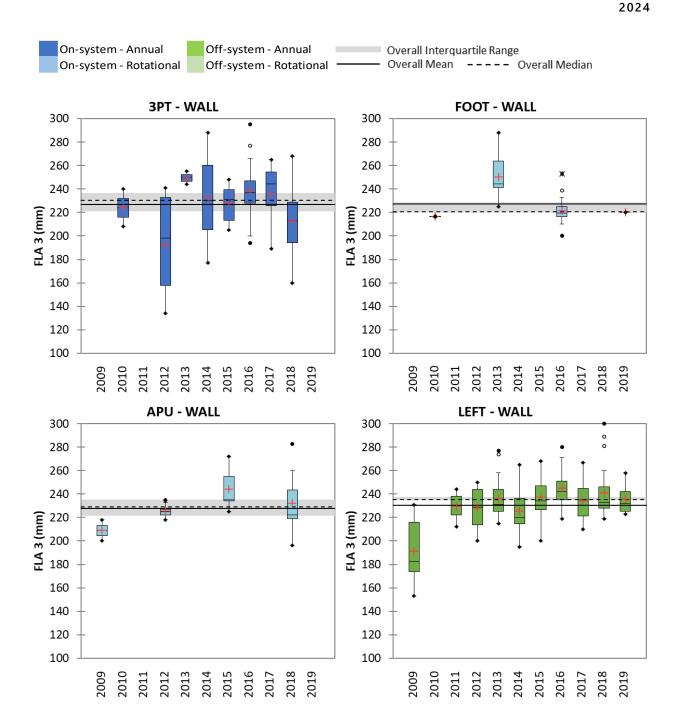


Figure 5.4-4. 2009-2019 Fork length-at-age (FLA) 3 of Walleye.



5.5 RECRUITMENT

5.5.1 RELATIVE YEAR-CLASS STRENGTH

5.5.1.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Lake Whitefish

The RYCS of Northern Pike over the 11 years of monitoring ranged from a low of 0 for the 2000 and 2014 cohort to a high of 238 for the 1999 cohort (Figure 5.5-1). There were four missing cohorts (1997, 1998, 2000 and 2014) from 1997-2014. Particularly weak cohorts (<50) occurred in 2007, 2009, and 2013.

Northern Pike

The RYCS of Northern Pike over the 11 years of monitoring ranged from a low of 23 for the 2002 cohort to a high of 199 for the 2001 cohort (Figure 5.5-2). There were no missing cohorts from 1999-2014. Particularly weak cohorts (<50) occurred in 1999, 2000, and 2002.

Sauger

Individual Sauger from Threepoint Lake were aged starting in 2017 when it became a target species (Table 5.3-1). The RYCS of Sauger over the three years of monitoring ranged from a low of 4 for the 2005 cohort to a high of 41 for the 2007 cohort (Figure 5.5-3). There were no missing cohorts from 2005-2014, but particularly weak cohorts (<50) occurred in all years.

Walleye

The RYCS of Walleye over the 11 years of monitoring ranged from a low of 50 for the 2009 cohort to a high of 168 for the 2013 cohort (Figure 5.5-4). There were no missing cohorts from 1997-2014. There were no particularly weak cohorts (<50), but strong cohorts (>100) were produced in 2001-2003, 2006, and 2010-2014.

ROTATIONAL SITES

RYCS analysis requires data be collected in at least three consecutive years and therefore cannot be conducted for rotational waterbodies.



5.5.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

Lake Whitefish

The RYCS of Northern Pike over the 11 years of monitoring ranged from a low of 29 for the 2012 cohort to a high of 215 for the 2007 cohort (Figure 5.5-1). There were no missing cohorts from 1997-2014. Particularly weak cohorts (<50) occurred in 2004, 2012, and 2013.

Northern Pike

The RYCS of Northern Pike over the 11 years of monitoring ranged from a low of 0 for the 1998 cohort to a high of 164 for the 1997 cohort (Figure 5.5-1). The 1998 cohort was the only missing cohort from 2002-2014. Particularly weak cohorts (<50) occurred in 1999, 2001, and 2002.

Sauger

Sauger were not captured in Leftrook Lake over the 11 years of monitoring (Table 5.2-1).

Walleye

The RYCS of Walleye over the 11 years of monitoring ranged from a low of 33 for the 2004 cohort to a high of 189 for the 1999 cohort (Figure 5.5-3). There were no missing cohorts from 1997-2014. Particularly weak cohorts (<50) occurred in 2004 and 2005 and strong cohorts (>100) were produced in 1997-2002, 2008, and 2010-2012.



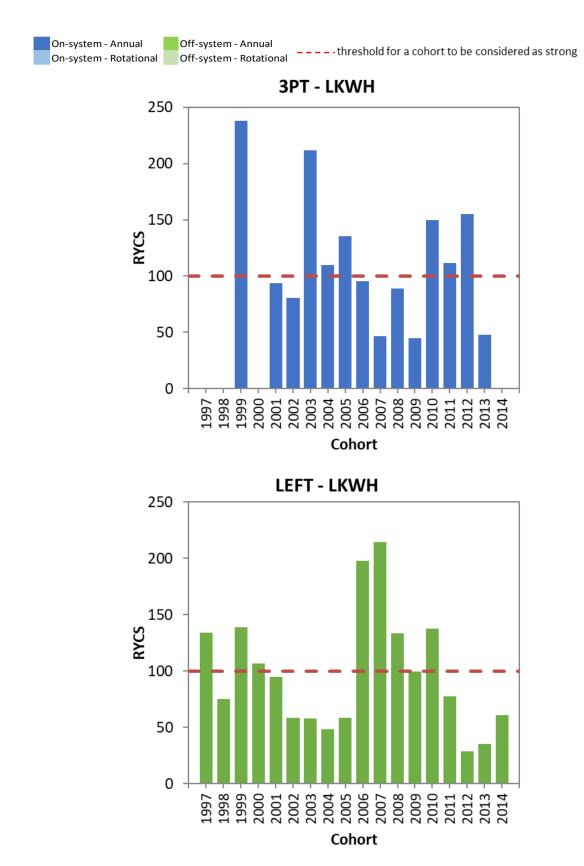


Figure 5.5-1. Relative year-class strength (RYCS) of Lake Whitefish.



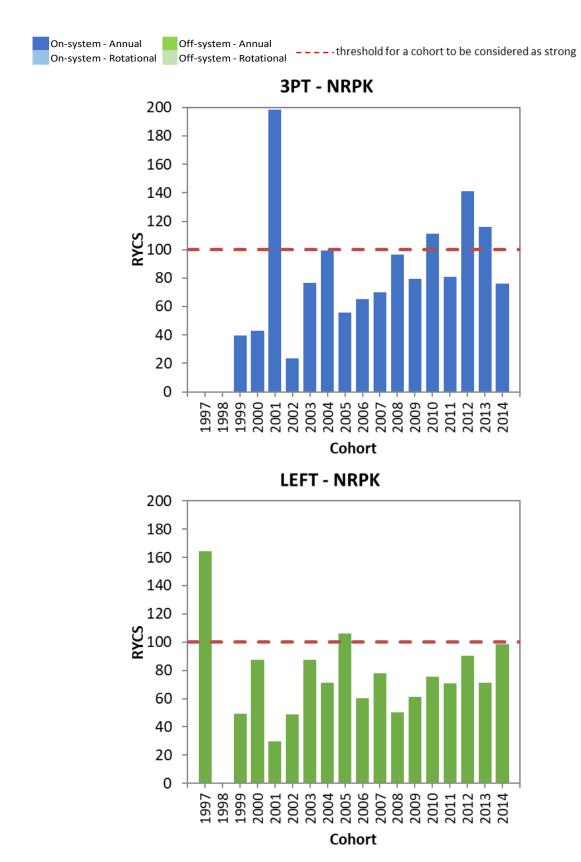


Figure 5.5-2. Relative year-class strength (RYCS) of Northern Pike.



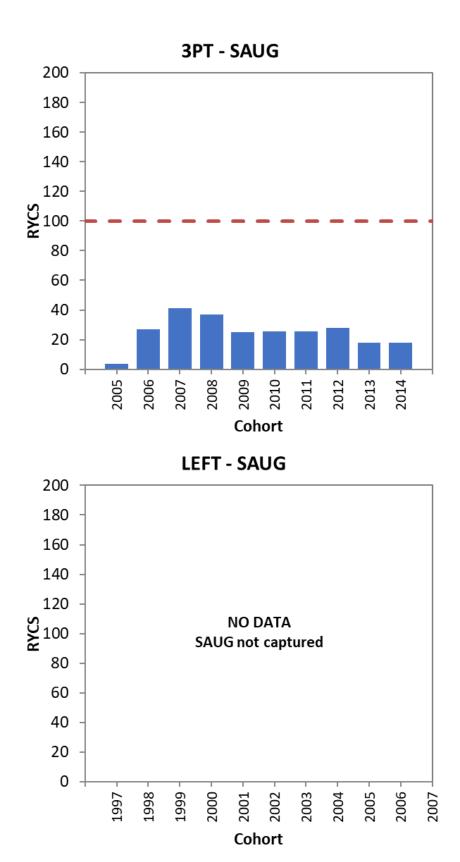


Figure 5.5-3 Relative year-class strength (RYCS) of Sauger.



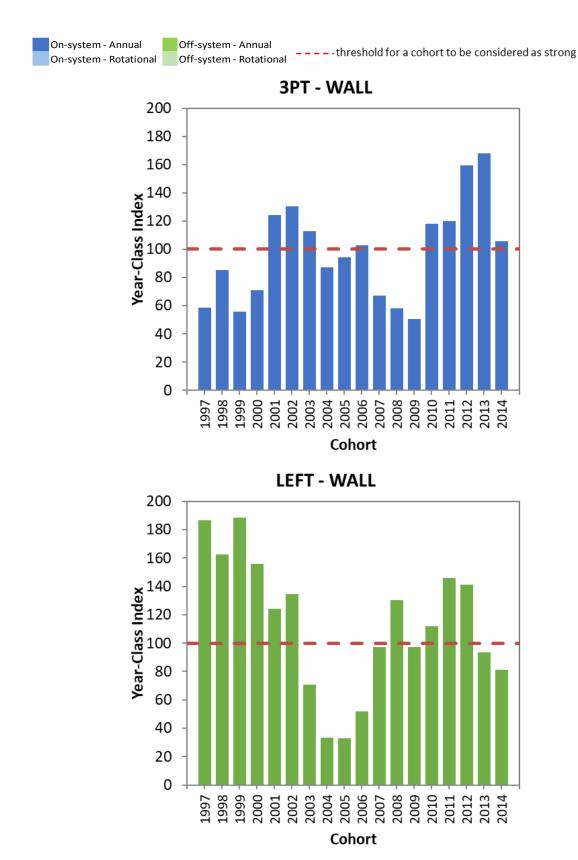


Figure 5.5-4. Relative year-class strength (RYCS) of Walleye.



5.6 DIVERSITY

5.6.1 RELATIVE SPECIES ABUNDANCE

5.6.1.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

A total of 13 fish species were captured in the combined standard and small mesh gangs at Threepoint Lake over 11 years of monitoring (Table 5.6-1) with the number of species caught each year ranging from 11-13 species (Tables 5.6-2 and 5.6-3).

Standard Gang Index Gill Nets

White Sucker and Walleye were the most frequently captured species at Threepoint Lake over 11 years of monitoring, each accounting for an average of >30% of the catch (Table 5.6-2). The annual RSA for White Sucker ranged from a low of 24% in 2015 to a high of 43% in 2018. The annual RSA for Walleye ranged from a low of 26% in 2018 to a high of 37% in 2013. Sauger accounted for >25% in one year (2016).

Small Mesh Index Gill Nets

The most common species captured in Threepoint Lake over 11 years of monitoring was Emerald Shiner (*Notropis atherinoides*), which accounted for an average of >25% of the catch (Table 5.6-3). The annual RSA for Emerald Shiner ranged from a low of 0% in 2012 to a high of 68% in 2015. Three other species accounted for >25% of the catch in some years: Sauger in 2009, 2012, and 2016; Walleye in 2012; and Spottail Shiner (*Notropis hudsonius*) in 2010, 2013, 2014, and 2017.

ROTATIONAL SITES

Footprint Lake

A total of 13 fish species were captured in the combined standard and small mesh gangs at Footprint Lake over four years of monitoring (Table 5.6-1), with the number of species caught each year ranging from 11-12 species (Tables 5.6-4 and 5.6-5). In one case, a sculpin was not identified to species (unidentified [unid.] sculpin species [spp.]).



Standard Gang Index Gill Nets

The catch in standard gangs set in Footprint Lake over four years of monitoring was dominated by Walleye, which accounted for an average of \geq 38% of the catch (Table 5.6-4). The annual RSA of Walleye ranged from a low of 38% in 2010 to a high of 53% in 2016.

Small Mesh Index Gill Nets

The most common species captured in Footprint Lake over four years of monitoring was Spottail Shiner, which accounted for an average of >25% of the catch (Table 5.6-5). The annual RSA for Spottail Shiner ranged from a low of 28% in 2010 to a high of 60% in 2016. Two other species accounted for >25% of the catch in some years: Emerald Shiner in 2019; and Walleye in 2010.

Apussigamasi Lake

A total of 16 fish species were captured in the combined standard and small mesh gangs at Apussigamasi Lake over four years of monitoring (Table 5.6-1), with the number of species caught each year ranging from 13-16 species (Tables 5.6-6 and 5.6-7).

Standard Gang Index Gill Nets

The catch in standard gangs set in Apussigamasi Lake over four years of monitoring was predominantly Walleye, which accounted for an average of \geq 25% of the catch (Table 5.6-6). The annual RSA of Walleye ranged from a low of 21% in 2018 to a high of 35% in 2009. White Sucker also accounted for >25% of the catch in 2018.

Small Mesh Index Gill Nets

The two most common species captured in Apussigamasi Lake over four years of monitoring was Sauger and Spottail Shiner, which both accounted for an average of >25% of the catch (Table 5.6-7). The annual RSA for Sauger ranged from a low of 10% in 2018 to a high of 42% in 2012, while the annual RSA for Spottail Shiner ranged from a low of 13% in 2012 to a high of 45% in 2018. No other species accounted for >25% of the catch in any year.



5.6.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

A total of 11 fish species were captured in the combined standard and small mesh gangs at Leftrook Lake over 11 years of monitoring (Table 5.6-1), with the number of species caught each year ranging from 9-10 species (Tables 5.6-8 and 5.6-9). Sauger were not captured at Leftrook Lake.

Standard Gang Index Gill Nets

White Sucker and Walleye were the most frequently captured species at Leftrook Lake over 11 years of monitoring, accounting for an average of >30% of the catch (Table 5.6-8). The annual RSA for White Sucker ranged from a low of 20% in 2017 to a high of 40% in 2011. The annual RSA for Walleye ranged from a low of 20% in 2010 to a high of 44% in 2017.

Small Mesh Index Gill Nets

The most common species captured in Leftrook Lake over 11 years of monitoring was Spottail Shiner, which accounted for an average of >40% of the catch (Table 5.6-9). The annual RSA for Spottail Shiner ranged from a low of 1% in 2019 to a high of 78% in 2014. Three other species accounted for >25% of the catch in some years, Emerald Shiner in 2010, 2013, 2015, 2016, and 2019; Trout-perch (*Percopsis omiscomaycus*) in 2011 and 2019, and Walleye in 2009.



Table 5.6-1. Inventory of fish species.

| Family | Species | Abbreviation | Status ¹ | Target | 3РТ | FOOT | APU | LEFT |
|--------------|--------------------|--------------|---------------------|--------|-----|------|-----|------|
| Hiodontidae | Goldeye | GOLD | Native | | | | • | |
| | Mooneye | MOON | Native | | | | • | |
| Cyprinidae | Lake Chub | LKCH | Native | | | | • | |
| '' | Emerald Shiner | EMSH | Native | | • | • | • | • |
| | Spottail Shiner | SPSH | Native | | • | • | • | • |
| Catostomidae | Longnose Sucker | LNSC | Native | | • | | • | |
| | White Sucker | WHSC | Native | • | • | • | • | • |
| | Shorthead Redhorse | SHRD | Native | | • | • | • | |
| Esocidae | Northern Pike | NRPK | Native | • | • | • | • | • |
| Salmonidae | Cisco | CISC | Native | | • | • | • | • |
| | Lake Whitefish | LKWH | Native | • | • | • | • | • |
| Percopsidae | Trout-perch | TRPR | Native | | • | • | • | • |
| Gadidae | Burbot | BURB | Native | | • | • | • | • |
| Cottidae | unid. Sculpin spp. | - | Native | | | • | | • |
| Percidae | Yellow Perch | YLPR | Native | | • | • | • | • |
| | Sauger | SAUG | Native | • | • | • | • | |
| | Walleye | WALL | Native | • | • | • | • | • |

Notes:



^{1.} Assigned from Stewart and Watkinson (2004).

Table 5.6-2. 2009-2019 Relative species abundance in standard gang index gill nets in Threepoint Lake.

| | | 0% | >0- | 5% > | 5-10% | >10-25 | 5% >25 | -50% | >50% | | | | |
|-------------|--------------|------|------|------|-------|--------|---------------|------|------|------|------|------|-------|
| Group | Species Code | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | Mean |
| Target | LKWH | 2% | 1% | 5% | 3% | 5% | 3% | 4% | 1% | 4% | 2% | 4% | 3% |
| | NRPK | 14% | 12% | 7% | 7% | 7% | 10% | 11% | 3% | 7% | 6% | 5% | 8% |
| | SAUG | 15% | 12% | 14% | 16% | 12% | 21% | 22% | 33% | 19% | 18% | 15% | 18% |
| | WALL | 28% | 35% | 34% | 36% | 37% | 30% | 29% | 28% | 29% | 26% | 31% | 31% |
| | WHSC | 29% | 33% | 29% | 33% | 32% | 30% | 24% | 28% | 30% | 43% | 34% | 31% |
| Mooneyes | GOLD | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | MOON | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Minnows | LKCH | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | EMSH | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | SPSH | 0% | 0% | 0% | 0% | 0% | 0.3% | 0% | 0% | 0% | 0% | 0% | 0.02% |
| Suckers | LNSC | 1% | 1% | 3% | 1% | 0.3% | 1% | 2% | 1% | 2% | 0.4% | 2% | 1% |
| | SHRD | 2% | 0.4% | 2% | 1% | 3% | 1% | 4% | 3% | 5% | 1% | 6% | 3% |
| Coregonids | CISC | 2% | 5% | 3% | 1% | 2% | 1% | 2% | 1% | 3% | 2% | 2% | 2% |
| Trout-perch | TRPR | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Codfishes | BURB | 1% | 0.4% | 1% | 0.4% | 0.3% | 0% | 0.3% | 0% | 0.3% | 0% | 0% | 0.3% |
| Sculpins | Sculpin spp. | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Perch | YLPR | 6% | 1% | 2% | 1% | 2% | 3% | 1% | 1% | 2% | 1% | 1% | 2% |



Table 5.6-3. 2009-2019 Relative species abundance in small mesh index gill nets in Threepoint Lake.

| | | 0% | >0- | 5% | >5-10% | >10-25 | >25 | -50% | >50% | | | | _ |
|-------------|--------------|------|------|------|--------|--------|------|------|------|------|------|------|-------|
| Group | Species Code | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | Mean |
| Target | LKWH | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | NRPK | 4% | 1% | 0.5% | 2% | 2% | 0.5% | 0% | 3% | 0.1% | 0% | 0% | 1% |
| | SAUG | 36% | 23% | 20% | 42% | 24% | 19% | 9% | 42% | 1% | 7% | 11% | 21% |
| | WALL | 5% | 13% | 17% | 48% | 2% | 1% | 1% | 6% | 1% | 1% | 4% | 9% |
| | WHSC | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0.2% | 0% | 0.02% |
| Mooneyes | GOLD | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | MOON | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Minnows | LKCH | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | EMSH | 34% | 14% | 19% | 0% | 26% | 28% | 68% | 29% | 56% | 62% | 58% | 36% |
| | SPSH | 10% | 38% | 21% | 0% | 40% | 44% | 18% | 4% | 38% | 24% | 18% | 23% |
| Suckers | LNSC | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | SHRD | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0.5% | 0.04% |
| Coregonids | CISC | 1% | 4% | 10% | 0% | 1% | 3% | 1% | 0% | 1% | 2% | 1% | 2% |
| Trout-perch | TRPR | 9% | 7% | 14% | 8% | 5% | 3% | 3% | 17% | 2% | 3% | 8% | 7% |
| Codfishes | BURB | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Sculpins | Sculpin spp. | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Perch | YLPR | 0% | 0% | 0% | 0% | 0.4% | 0% | 0% | 0% | 0% | 1% | 0% | 0.1% |



Table 5.6-4. 2009-2019 Relative species abundance in standard gang index gill nets in Footprint Lake.

| 0% | >0-5% >5-10 | 0% >10 | 0-25% | >25-509 | % >50 |)% |
|-------------|--------------|--------|-------|---------|-------------|------|
| Group | Species Code | 2010 | 2013 | 2016 | 2019 | Mean |
| Target | LKWH | 0.4% | 0.2% | 0.2% | 1% | 0.3% |
| | NRPK | 9% | 10% | 1% | 4% | 6% |
| | SAUG | 4% | 9% | 12% | 9% | 9% |
| | WALL | 38% | 42% | 53% | 40% | 43% |
| | WHSC | 32% | 12% | 13% | 21% | 20% |
| Mooneyes | GOLD | 0% | 0% | 0% | 0% | 0% |
| | MOON | 0% | 0% | 0% | 0% | 0% |
| Minnows | LKCH | 0% | 0% | 0% | 0% | 0% |
| | EMSH | 0% | 0% | 0% | 0% | 0% |
| | SPSH | 0% | 0% | 0% | 0% | 0% |
| Suckers | LNSC | 0% | 0% | 0% | 0% | 0% |
| | SHRD | 0.2% | 0% | 0.2% | 2% | 0.5% |
| Coregonids | CISC | 12% | 19% | 18% | 20% | 17% |
| Trout-perch | TRPR | 0% | 0% | 0% | 0% | 0% |
| Codfishes | BURB | 0.2% | 0% | 0.4% | 0.2% | 0.2% |
| Sculpins | Sculpin spp. | 0% | 0% | 0% | 0% | 0% |
| Perch | YLPR | 4% | 8% | 2% | 5% | 5% |



Table 5.6-5. 2009-2019 Relative species abundance in small mesh index gill nets in Footprint Lake.

| 0% | >0-5% >5-10 | 0% >10 | 0-25% | >25-509 | % >50 |)% |
|-------------|--------------|--------|-------|---------|-------------|------|
| Group | Species Code | 2010 | 2013 | 2016 | 2019 | Mean |
| Target | LKWH | 0% | 0% | 0% | 0% | 0% |
| | NRPK | 4% | 5% | 2% | 1% | 3% |
| | SAUG | 15% | 23% | 5% | 4% | 12% |
| | WALL | 39% | 13% | 6% | 7% | 16% |
| | WHSC | 1% | 0% | 0% | 0% | 0.3% |
| Mooneyes | GOLD | 0% | 0% | 0% | 0% | 0% |
| | MOON | 0% | 0% | 0% | 0% | 0% |
| Minnows | LKCH | 0% | 0% | 0% | 0% | 0% |
| | EMSH | 4% | 1% | 23% | 40% | 17% |
| | SPSH | 28% | 39% | 60% | 43% | 43% |
| Suckers | LNSC | 0% | 0% | 0% | 0% | 0% |
| | SHRD | 0% | 0% | 0% | 0% | 0% |
| Coregonids | CISC | 6% | 9% | 3% | 0.2% | 5% |
| Trout-perch | TRPR | 3% | 3% | 1% | 3% | 2% |
| Codfishes | BURB | 0% | 0% | 0% | 0% | 0% |
| Sculpins | Sculpin spp. | 0% | 1% | 0% | 0% | 0.2% |
| Perch | YLPR | 0% | 6% | 0.4% | 1% | 2% |



Table 5.6-6. 2009-2019 Relative species abundance in standard gang index gill nets in Apussigamasi Lake.

| 0% | >0-5% >5-10 | 0% >10 | 0-25% | >25-509 | % >50 |)% |
|-------------|--------------|--------|-------|---------|-------|------|
| Group | Species Code | 2009 | 2012 | 2015 | 2018 | Mean |
| Target | LKWH | 9% | 6% | 8% | 7% | 8% |
| | NRPK | 5% | 1% | 4% | 2% | 3% |
| | SAUG | 17% | 23% | 21% | 18% | 20% |
| | WALL | 35% | 29% | 29% | 21% | 29% |
| | WHSC | 17% | 23% | 23% | 28% | 22% |
| Mooneyes | GOLD | 0% | 0% | 0% | 0% | 0% |
| | MOON | 5% | 1% | 5% | 9% | 5% |
| Minnows | LKCH | 0.4% | 0% | 0% | 0% | 0.1% |
| | EMSH | 0% | 0% | 0% | 0% | 0% |
| | SPSH | 0% | 0% | 0% | 0% | 0% |
| Suckers | LNSC | 3% | 3% | 3% | 7% | 4% |
| | SHRD | 2% | 7% | 3% | 5% | 4% |
| Coregonids | CISC | 3% | 3% | 3% | 2% | 3% |
| Trout-perch | TRPR | 0% | 0% | 0% | 0% | 0% |
| Codfishes | BURB | 1% | 1% | 0% | 1% | 1% |
| Sculpins | Sculpin spp. | 0% | 0% | 0% | 0% | 0% |
| Perch | YLPR | 3% | 1% | 1% | 0.2% | 1% |



Table 5.6-7. 2009-2019 Relative species abundance in small mesh index gill nets in Apussigamasi Lake.

| 0% | >0-5% >5-10 | 0% >10 | 0-25% | >25-509 | % >5(|)% |
|-------------|--------------|--------|-------|---------|-------|------|
| Group | Species Code | 2009 | 2012 | 2015 | 2018 | Mean |
| Target | LKWH | 0% | 0% | 0% | 0% | 0% |
| | NRPK | 2% | 1% | 0% | 0.5% | 1% |
| | SAUG | 40% | 42% | 27% | 10% | 30% |
| | WALL | 13% | 8% | 15% | 11% | 12% |
| | WHSC | 0% | 1% | 0% | 0% | 0.1% |
| Mooneyes | GOLD | 1% | 0% | 0% | 0% | 0.2% |
| | MOON | 5% | 1% | 4% | 7% | 4% |
| Minnows | LKCH | 0% | 0% | 0% | 0.5% | 0.1% |
| | EMSH | 4% | 2% | 8% | 17% | 8% |
| | SPSH | 21% | 13% | 41% | 45% | 30% |
| Suckers | LNSC | 0% | 0% | 0% | 0% | 0% |
| | SHRD | 0% | 0% | 0% | 0% | 0% |
| Coregonids | CISC | 1% | 10% | 4% | 3% | 5% |
| Trout-perch | TRPR | 13% | 22% | 2% | 6% | 10% |
| Codfishes | BURB | 0% | 0% | 0% | 0% | 0% |
| Sculpins | Sculpin spp. | 0% | 0% | 0% | 0% | 0% |
| Perch | YLPR | 0% | 1% | 0% | 0% | 0.1% |



Table 5.6-8. 2009-2019 Relative species abundance in standard gang index gill nets in Leftrook Lake.

| | | 0% | >0- | 5% > | 5-10% | 6 >10-25% >25-50% >5 | | | >50% | | | | |
|-------------|--------------|------|------|------|-------|--------------------------------------|------|------|------|------|------|------|------|
| Group | Species Code | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | Mean |
| Target | LKWH | 12% | 16% | 8% | 15% | 17% | 15% | 10% | 9% | 12% | 8% | 11% | 12% |
| | NRPK | 11% | 13% | 15% | 14% | 11% | 17% | 12% | 8% | 8% | 10% | 14% | 12% |
| | SAUG | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | WALL | 30% | 20% | 30% | 33% | 27% | 29% | 33% | 38% | 44% | 34% | 31% | 32% |
| | WHSC | 36% | 35% | 40% | 23% | 35% | 23% | 30% | 33% | 20% | 29% | 35% | 31% |
| Mooneyes | GOLD | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | MOON | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Minnows | LKCH | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | EMSH | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | SPSH | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Suckers | LNSC | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | SHRD | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Coregonids | CISC | 4% | 13% | 2% | 11% | 8% | 5% | 3% | 8% | 5% | 9% | 4% | 7% |
| Trout-perch | TRPR | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Codfishes | BURB | 0.3% | 0% | 0.1% | 0% | 0% | 0.1% | 0% | 0% | 0.1% | 0% | 0% | 0.1% |
| Sculpins | Sculpin spp. | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Perch | YLPR | 6% | 2% | 5% | 4% | 2% | 11% | 12% | 5% | 10% | 9% | 5% | 6% |



Table 5.6-9. 2009-2019 Relative species abundance in small mesh index gill nets in Leftrook Lake.

| | | 0% | >0- | 5% > | 5-10% | >10-25% >25-50% | | | >50% | | | | |
|-------------|--------------|------|------|------|-------|-----------------|------|------|------|------|------|------|------|
| Group | Species Code | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | Mean |
| Target | LKWH | 1% | 0% | 2% | 0% | 0.5% | 0.2% | 0.4% | 0.3% | 0% | 0% | 1% | 1% |
| | NRPK | 5% | 2% | 2% | 0.5% | 5% | 1% | 4% | 1% | 2% | 6% | 1% | 3% |
| | SAUG | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | WALL | 43% | 13% | 15% | 2% | 6% | 1% | 13% | 4% | 3% | 20% | 21% | 13% |
| | WHSC | 1% | 0.5% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 1% | 1% | 0.3% |
| Mooneyes | GOLD | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | MOON | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Minnows | LKCH | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | EMSH | 6% | 36% | 3% | 20% | 42% | 17% | 52% | 31% | 16% | 8% | 37% | 24% |
| | SPSH | 26% | 30% | 27% | 75% | 37% | 78% | 23% | 56% | 72% | 55% | 1% | 44% |
| Suckers | LNSC | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | SHRD | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Coregonids | CISC | 0% | 1% | 3% | 0% | 1% | 0% | 0% | 0% | 0% | 1% | 2% | 1% |
| Trout-perch | TRPR | 17% | 14% | 41% | 2% | 5% | 0.4% | 3% | 4% | 4% | 4% | 36% | 12% |
| Codfishes | BURB | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Sculpins | Sculpin spp. | 0% | 2% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0.2% |
| Perch | YLPR | 0.3% | 2% | 7% | 0% | 3% | 3% | 4% | 3% | 3% | 4% | 0.3% | 3% |



5.6.2 HILL'S EFFECTIVE RICHNESS

5.6.2.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Hill's effective species richness over the 11 years of monitoring ranged from a low of 4.8 in 2012 to a high of 8.3 species in 2011 (Table 5.6-10; Figure 5.6-1).

The overall mean Hill's index value was 6.6, the median was 6.9, and the IQR was 5.8-7.4 species (Figure 5.6-1). The annual mean Hill's index value fell within the overall IQR except in 2012, 2015, and 2017 when it was below the IQR and in 2011, 2013, and 2019 when it was above the IQR.

ROTATIONAL SITES

Footprint Lake

The Hill's effective species richness over the four years of monitoring ranged from a low of 5.3 in 2010 to a high of 7.2 species in 2019 (Table 5.6-10; Figure 5.6-1).

The overall mean Hill's index value was 6.1, the median was 5.9, and the IQR was 5.8-6.4 species (Figure 5.6-1). The annual mean Hill's index value was below the IQR in 2010 and was above the IQR in 2019.

<u>Apussigamasi Lake</u>

The Hill's effective species richness over the four years of monitoring ranged from a low of 7.7 in 2012 to a high of 9.4 species in 2018 (Table 5.6-10; Figure 5.6-1).

The overall mean Hill's index value was 8.3, the median was 8.1, and the IQR was 7.8-8.6 species (Figure 5.6-1). The annual mean Hill's index value was below the IQR in 2012 and was above the IQR in 2018.



5.6.2.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

Hill's effective species richness over the 11 years of monitoring ranged from a low of 4.3 in 2014 to a high of 8.1 species in 2010 (Table 5.6-10; Figure 5.6-1).

The overall mean and median Hill's index values were 6.4, and the IQR was 6.4-6.7 species (Figure 5.6-1). The annual mean Hill's index value fell within the overall IQR except in 2009, 2012, 2014, and 2017 when it was below the IQR and in 2010, 2013, and 2015 when it was above the IQR.



Table 5.6-10. 2009-2019 Hill's effective species richness.

| Waterbody | Year | n _F ¹ | n _{spp} ² | Value |
|-----------|------|-----------------------------|-------------------------------|-------|
| 3PT | 2009 | 443 | 13 | 7.3 |
| | 2010 | 383 | 13 | 6.9 |
| | 2011 | 537 | 13 | 8.3 |
| | 2012 | 283 | 11 | 4.8 |
| | 2013 | 566 | 13 | 7.5 |
| | 2014 | 798 | 12 | 7.2 |
| | 2015 | 884 | 13 | 5.7 |
| | 2016 | 568 | 12 | 6.0 |
| | 2017 | 1590 | 13 | 4.8 |
| | 2018 | 738 | 12 | 5.9 |
| | 2019 | 506 | 12 | 7.6 |
| FOOT | 2010 | 552 | 12 | 5.3 |
| | 2013 | 692 | 11 | 6.1 |
| | 2016 | 1089 | 12 | 5.8 |
| | 2019 | 1012 | 12 | 7.2 |
| APU | 2009 | 601 | 16 | 8.3 |
| | 2012 | 509 | 14 | 7.7 |
| | 2015 | 639 | 13 | 7.9 |
| | 2018 | 652 | 15 | 9.4 |
| LEFT | 2009 | 1312 | 10 | 6.1 |
| | 2010 | 950 | 10 | 8.1 |
| | 2011 | 989 | 10 | 6.4 |
| | 2012 | 1640 | 9 | 6.2 |
| | 2013 | 1226 | 9 | 7.2 |
| | 2014 | 3412 | 10 | 4.3 |
| | 2015 | 1412 | 9 | 6.9 |
| | 2016 | 1861 | 9 | 6.4 |
| | 2017 | 2015 | 10 | 5.8 |
| | 2018 | 1420 | 9 | 6.5 |
| | 2019 | 1092 | 9 | 6.4 |

Notes:

1. n_F = number of fish caught in standard and small mesh gill nets.

2. n_{spp} = number of species caught in standard and small mesh gill nets.





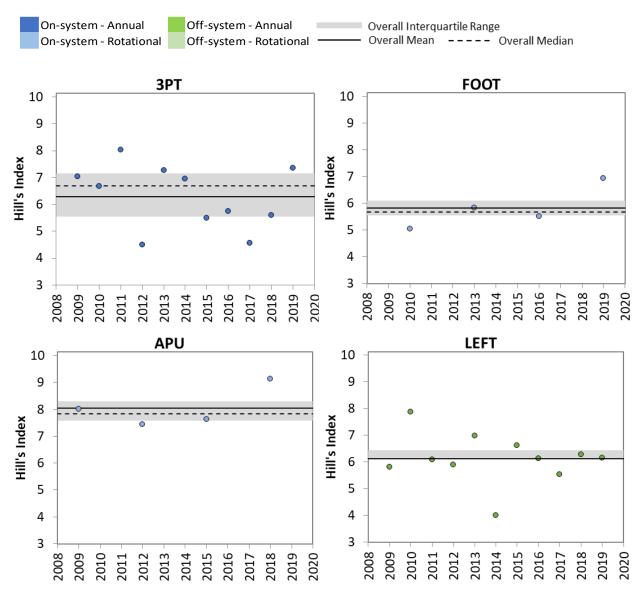


Figure 5.6-1. 2009-2019 Hill's effective species richness.



APPENDIX 5-1. GILLNETTING SITE INFORMATION AND LOCATIONS



The following is a summary of modifications and deviations in sampling locations over the 11 years of monitoring in the Churchill River Diversion Region:

Threepoint Lake

- Gill nets were set at the target locations in all 11 years with the following exceptions:
 - GN-01 was sampled in 2009 and 2011 but was discontinued after the Pilot Program, starting in 2012.
 - GN-09 was not sampled in 2009 or 2011 (the years GN-01 was sampled) but was sampled in 2010 (the year GN-01 was not sampled) and has been sampled every year after the Pilot Program, starting in 2012.
 - SN-02 was sampled in 2009, 2011, and 2012 but was discontinued starting in 2013.
 - SN-09 was not sampled in 2009, 2011, or 2012 (the years SN-02 was sampled) but was sampled in 2010 (the year SN-02 was not sampled) and has been sampled every year starting in 2013.
 - SN-16 was named SN-13 in 2010 but was set at the SN-16 target location.

Footprint Lake

Gill nets were set at the target locations in all four years.

Apussigamasi Lake

Gill nets were set at the target locations in all four years.

Leftrook Lake

Gill nets were set at the target locations in all 11 years.



Table A5-1-1. 2008-2019 Set information for gillnetting sites.

| Waterbody | Site | Set Date | Coordinates | | Duration (dec. hrs) | Wa Dep (m | oth | Set Water Temperature | |
|-----------|-------|-------------|-------------|---------|------------------------|-----------------|-------|--------------------------|------|
| | | | Zone | Easting | Northing | | Start | End | (°C) |
| 3PT | GN-01 | 18-Aug-09 | 14 | 504524 | 6175533 | 22.2 | 4.1 | 3.4 | 15.0 |
| | GN-02 | 16-Aug-09 | 14 | 503777 | 6173256 | 48.3 | 3.8 | 0.9 | 16.0 |
| | GN-04 | 16-Aug-09 | 14 | 501692 | 6174036 | 47.7 | 5.8 | 5.9 | 16.0 |
| | GN-05 | 15-Aug-09 | 14 | 501586 | 6174659 | 21.2 | 6.2 | 6.1 | 17.0 |
| | GN-06 | 15-Aug-09 | 14 | 500052 | 6174858 | 21.8 | 4.5 | 1.7 | 16.0 |
| | GN-13 | 19-Aug-09 | 14 | 508609 | 6170673 | 23.5 | 2.4 | 3.9 | 15.0 |
| | GN-15 | 19-Aug-09 | 14 | 507000 | 6169388 | 22.9 | 4.4 | 3.2 | 14.0 |
| | GN-16 | 19-Aug-09 | 14 | 507537 | 6169677 | 23.7 | 4.8 | 4.5 | 15.0 |
| | GN-17 | 18-Aug-09 | 14 | 505944 | 6174769 | 22.4 | 4.1 | 3.6 | 15.0 |
| | SN-02 | 16-Aug-09 | 14 | 503911 | 6173274 | 48.3 | 3.8 | 0.9 | 16.0 |
| | SN-15 | 19-Aug-09 | 14 | 506862 | 6169370 | 22.9 | 4.4 | 3.2 | 14.0 |
| | SN-16 | 19-Aug-09 | 14 | 507503 | 6169809 | 23.7 | 4.8 | 4.5 | 15.0 |
| | GN-02 | 25-Aug-10 | 14 | 503842 | 6172984 | 24.5 | 4.4 | 1.0 | 14.0 |
| | GN-04 | 23-Aug-10 | 14 | 502095 | 6174452 | 46.9 | 5.8 | 6.3 | 15.0 |
| | GN-05 | 23-Aug-10 | 14 | 501259 | 6174416 | 46.3 | 5.9 | 5.7 | 15.0 |
| | GN-06 | 23-Aug-10 | 14 | 500133 | 6174917 | 47.3 | 2.4 | 4.3 | 15.0 |
| | GN-09 | 25-Aug-10 | 14 | 503201 | 6169924 | 23.5 | 3.9 | 4.5 | 14.0 |
| | GN-13 | 26-Aug-10 | 14 | 508559 | 6170496 | 21.0 | 4.3 | 0.7 | 14.0 |
| | GN-15 | 26-Aug-10 | 14 | 507251 | 6169354 | 21.1 | 4.7 | 4.7 | 14.0 |
| | GN-16 | 26-Aug-10 | 14 | 507527 | 6169574 | 20.9 | 5.0 | 5.9 | 14.0 |
| | GN-17 | 25-Aug-10 | 14 | 505986 | 6174705 | 24.0 | 4.2 | 1.0 | 14.0 |
| | SN-09 | 25-Aug-10 | 14 | 503334 | 6169955 | 23.5 | 3.4 | 3.4 | 14.0 |
| | SN-15 | 26-Aug-10 | 14 | 507121 | 6169444 | 21.1 | 4.7 | 4.7 | 14.0 |
| | SN-16 | 26-Aug-10 | 14 | 507527 | 6169574 | 20.9 | 5.0 | 5.0 | 14.0 |
| | GN-01 | 18-Aug-11 | 14 | 504404 | 6175672 | 22.0 | 3.5 | 3.8 | 18.0 |
| | GN-02 | 18-Aug-11 | 14 | 503721 | 6173833 | 22.1 | 3.3 | 2.9 | 18.0 |
| | GN-04 | 19-Aug-11 | 14 | 501695 | 6174049 | 23.1 | 5.4 | 5.5 | 17.0 |
| | GN-05 | 19-Aug-11 | 14 | 501715 | 6174726 | 23.3 | 5.6 | 5.5 | 17.0 |
| | GN-06 | 19-Aug-11 | 14 | 500079 | 6174824 | 22.8 | 3.9 | 3.1 | 17.0 |
| | GN-13 | 16-Aug-11 | 14 | 507673 | 6169596 | 43.6 | 4.2 | 4.5 | 18.5 |
| | GN-15 | 16-Aug-11 | 14 | 506933 | 6169297 | 44.2 | 4.0 | 3.6 | 18.5 |
| | GN-16 | 16-Aug-11 | 14 | 507415 | 6169730 | 43.1 | 3.9 | 4.2 | 18.5 |



Table A5-1-1. continued.

| Waterbody | Site | Set Date | Coordinates | | Duration (dec. hrs) | Wa Dep (m | oth | Set Water Temperature | |
|-----------|-------|-------------|-------------|---------|------------------------|-----------------|-------|-----------------------|------|
| | | | Zone | Easting | Northing | | Start | End | (°C) |
| 3PT | GN-17 | 18-Aug-11 | 14 | 505969 | 6174633 | 21.9 | 2.5 | 3.9 | 18.0 |
| | SN-02 | 18-Aug-11 | 14 | 503721 | 6173833 | 22.1 | 3.3 | 3.3 | 18.0 |
| | SN-13 | 16-Aug-11 | 14 | 507673 | 6169596 | 43.6 | 4.2 | 4.2 | 18.5 |
| | SN-15 | 16-Aug-11 | 14 | 506933 | 6169297 | 43.4 | 4.0 | 4.0 | 18.5 |
| | GN-02 | 13-Aug-12 | 14 | 503956 | 6172998 | 19.1 | 2.8 | 5.7 | 19.5 |
| | GN-04 | 13-Aug-12 | 14 | 502081 | 6174452 | 19.5 | 7.1 | 7.0 | 19.5 |
| | GN-05 | 13-Aug-12 | 14 | 501191 | 6174346 | 19.7 | 7.1 | 7.7 | 19.5 |
| | GN-06 | 13-Aug-12 | 14 | 500138 | 6174942 | 19.8 | 1.9 | 5.6 | 19.5 |
| | GN-09 | 12-Aug-12 | 14 | 503374 | 6169910 | 27.2 | 3.9 | 5.2 | 19.5 |
| | GN-13 | 11-Aug-12 | 14 | 508542 | 6170523 | 20.4 | 5.3 | 2.6 | 21.5 |
| | GN-15 | 11-Aug-12 | 14 | 507214 | 6169353 | 20.1 | 6.0 | 5.8 | 21.5 |
| | GN-16 | 12-Aug-12 | 14 | 507644 | 6169668 | 27.4 | 7.1 | 7.7 | 19.5 |
| | GN-17 | 11-Aug-12 | 14 | 506032 | 6174440 | 20.8 | 2.3 | 5.7 | 21.5 |
| | SN-02 | 13-Aug-12 | 14 | 503968 | 6173003 | 19.1 | 1.2 | 1.2 | 19.5 |
| | SN-15 | 11-Aug-12 | 14 | 507231 | 6169352 | 20.1 | 6.0 | 6.0 | 21.5 |
| | SN-16 | 12-Aug-12 | 14 | 507638 | 6169662 | 27.4 | 6.8 | 6.8 | 19.5 |
| | GN-02 | 15-Aug-13 | 14 | 503958 | 6173021 | 23.0 | 1.3 | 4.7 | 20.0 |
| | GN-04 | 16-Aug-13 | 14 | 502034 | 6174433 | 22.2 | 6.0 | 6.0 | 19.5 |
| | GN-05 | 16-Aug-13 | 14 | 501264 | 6174406 | 22.6 | 7.1 | 6.6 | 19.5 |
| | GN-06 | 16-Aug-13 | 14 | 500140 | 6174933 | 22.2 | 2.4 | 4.5 | 20.0 |
| | GN-09 | 15-Aug-13 | 14 | 503183 | 6169906 | 22.7 | 3.4 | 4.3 | 19.5 |
| | GN-13 | 14-Aug-13 | 14 | 508663 | 6170506 | 23.2 | 4.3 | 5.2 | 19.0 |
| | GN-15 | 14-Aug-13 | 14 | 507064 | 6169324 | 22.7 | 4.7 | 4.9 | 19.0 |
| | GN-16 | 14-Aug-13 | 14 | 507700 | 6169773 | 23.0 | 5.4 | 5.7 | 19.0 |
| | GN-17 | 15-Aug-13 | 14 | 505965 | 6174634 | 23.2 | 4.5 | 3.1 | 20.0 |
| | SN-09 | 15-Aug-13 | 14 | 503343 | 6169938 | 22.7 | 3.2 | 3.4 | 19.5 |
| | SN-15 | 14-Aug-13 | 14 | 507232 | 6169354 | 22.7 | 4.9 | 4.7 | 19.0 |
| | SN-16 | 14-Aug-13 | 14 | 507615 | 6169615 | 23.0 | 5.2 | 5.4 | 19.0 |
| | GN-02 | 13-Aug-14 | 14 | 503958 | 6173026 | 24.6 | 0.8 | 3.7 | 19.0 |
| | GN-04 | 13-Aug-14 | 14 | 501956 | 6174330 | 23.8 | 5.2 | 5.2 | 19.0 |
| | GN-05 | 12-Aug-14 | 14 | 501241 | 6174417 | 21.9 | 5.8 | 4.4 | 19.0 |
| | GN-06 | 12-Aug-14 | 14 | 500160 | 6174931 | 21.4 | 1.8 | 3.9 | 19.0 |
| | GN-09 | 14-Aug-14 | 14 | 503214 | 6170095 | 23.2 | 3.7 | 2.3 | 20.0 |
| | GN-13 | 14-Aug-14 | 14 | 508528 | 6170485 | 19.1 | 3.6 | 0.8 | 20.0 |
| | GN-15 | 14-Aug-14 | 14 | 507184 | 6169354 | 19.6 | 4.3 | 3.9 | 20.0 |



Table A5-1-1. continued.

| Waterbody | Site | Set Date | | UTM Coordina | | Duration (dec. hrs) | Wa Der (n | oth | Set Water Temperature |
|-----------|-------|-------------|------|-----------------|----------|------------------------|-----------------|-----|--------------------------|
| | | | Zone | Easting | Northing | | Start | End | (°C) |
| 3PT | GN-16 | 14-Aug-14 | 14 | 507743 | 6169625 | 18.6 | 4.2 | 5.5 | 20.0 |
| | GN-17 | 14-Aug-14 | 14 | 505997 | 6174766 | 23.2 | 3.7 | 2.2 | 20.0 |
| | SN-09 | 14-Aug-14 | 14 | 503214 | 6170095 | 23.2 | 3.7 | 3.7 | 20.0 |
| | SN-15 | 14-Aug-14 | 14 | 507235 | 6169364 | 19.6 | 4.0 | 4.3 | 20.0 |
| | SN-16 | 14-Aug-14 | 14 | 507760 | 6169600 | 18.6 | 4.3 | 4.2 | 20.0 |
| | GN-02 | 16-Aug-15 | 14 | 503963 | 6172988 | 22.4 | 1.0 | 5.0 | 17.5 |
| | GN-04 | 15-Aug-15 | 14 | 502112 | 6174443 | 21.9 | 6.2 | 6.3 | 19.0 |
| | GN-05 | 15-Aug-15 | 14 | 501381 | 6174493 | 21.6 | 7.0 | 6.8 | 19.0 |
| | GN-06 | 15-Aug-15 | 14 | 500160 | 6174936 | 21.5 | 1.8 | 4.9 | 19.0 |
| | GN-09 | 16-Aug-15 | 14 | 503199 | 6169857 | 22.9 | 4.8 | 3.8 | 17.5 |
| | GN-13 | 17-Aug-15 | 14 | 508680 | 6170515 | 23.0 | 4.6 | 1.4 | 17.5 |
| | GN-15 | 17-Aug-15 | 14 | 507137 | 6169440 | 22.3 | 5.2 | 5.2 | 17.5 |
| | GN-16 | 17-Aug-15 | 14 | 507529 | 6169553 | 22.8 | 5.4 | 5.9 | 17.5 |
| | GN-17 | 16-Aug-15 | 14 | 506108 | 6174601 | 23.2 | 1.2 | 5.0 | 17.5 |
| | SN-09 | 16-Aug-15 | 14 | 503225 | 6169893 | 22.9 | 3.2 | 4.8 | 17.5 |
| | SN-15 | 17-Aug-15 | 14 | 507137 | 6169440 | 22.3 | 5.2 | 5.2 | 17.5 |
| | SN-16 | 17-Aug-15 | 14 | 507552 | 6169573 | 22.8 | 5.4 | 5.4 | 17.5 |
| | GN-02 | 11-Aug-16 | 14 | 503957 | 6173038 | 23.8 | 2.0 | 5.2 | 18.0 |
| | GN-04 | 10-Aug-16 | 14 | 502059 | 6174428 | 22.4 | 6.9 | 6.9 | 18.0 |
| | GN-05 | 10-Aug-16 | 14 | 501274 | 6174450 | 22.0 | 7.5 | 7.5 | 18.0 |
| | GN-06 | 10-Aug-16 | 14 | 500160 | 6174941 | 21.4 | 1.5 | 5.5 | 18.0 |
| | GN-09 | 11-Aug-16 | 14 | 503237 | 6170078 | 23.1 | 5.3 | 2.6 | 18.0 |
| | GN-13 | 12-Aug-16 | 14 | 508648 | 6170557 | 20.3 | 1.8 | 5.3 | 18.0 |
| | GN-15 | 12-Aug-16 | 14 | 507134 | 6169383 | 22.3 | 5.7 | 5.9 | 18.0 |
| | GN-16 | 12-Aug-16 | 14 | 507770 | 6169672 | 21.5 | 6.1 | 6.0 | 18.0 |
| | GN-17 | 11-Aug-16 | 14 | 505966 | 6174630 | 24.3 | 3.4 | 5.5 | 18.0 |
| | SN-09 | 11-Aug-16 | 14 | 503203 | 6170085 | 23.1 | 5.8 | 5.3 | 18.0 |
| | SN-15 | 12-Aug-16 | 14 | 507134 | 6169344 | 22.3 | 5.7 | 5.7 | 18.0 |
| | SN-16 | 12-Aug-16 | 14 | 507766 | 6169597 | 21.5 | 5.8 | 6.1 | 18.0 |
| | GN-02 | 16-Aug-17 | 14 | 503844 | 6172959 | 21.0 | 3.8 | 3.8 | 20.0 |
| | GN-04 | 17-Aug-17 | 14 | 502122 | 6174456 | 22.6 | 5.3 | 5.1 | 19.0 |
| | GN-05 | 17-Aug-17 | 14 | 501237 | 6174420 | 21.8 | 5.9 | 5.7 | 19.0 |
| | GN-06 | 17-Aug-17 | 14 | 500151 | 6174913 | 21.0 | 3.1 | 3.8 | 19.0 |
| | GN-09 | 16-Aug-17 | 14 | 503311 | 6169982 | 19.8 | 2.7 | 3.8 | 20.0 |
| | GN-13 | 15-Aug-17 | 14 | 508667 | 6170467 | 26.0 | 2.6 | 3.5 | 21.0 |



Table A5-1-1. continued.

| Waterbody | Site | Set Date | | UTM Coordina | | Duration (dec. hrs) | Wa Dep (m | oth | Set Water Temperature |
|-----------|-------|-------------|------|-----------------|----------|------------------------|-----------------|-----|-----------------------|
| | | | Zone | Easting | Northing | | Start | End | (°C) |
| 3PT | GN-15 | 15-Aug-17 | 14 | 507199 | 6169344 | 24.8 | 4.0 | 3.9 | 21.0 |
| | GN-16 | 15-Aug-17 | 14 | 507686 | 6169710 | 25.8 | 4.9 | 4.1 | 21.0 |
| | GN-17 | 16-Aug-17 | 14 | 505960 | 6174656 | 21.8 | 2.3 | 3.7 | 20.0 |
| | SN-09 | 16-Aug-17 | 14 | 503333 | 6169964 | 19.8 | 2.7 | 2.7 | 20.0 |
| | SN-15 | 15-Aug-17 | 14 | 507236 | 6169362 | 24.8 | 4.0 | 3.7 | 21.0 |
| | SN-16 | 15-Aug-17 | 14 | 507665 | 6169710 | 25.8 | 4.9 | 5.0 | 21.0 |
| | GN-02 | 22-Aug-18 | 14 | 503845 | 6172962 | 22.8 | 5.7 | 5.2 | 17.0 |
| | GN-04 | 20-Aug-18 | 14 | 502097 | 6174455 | 21.9 | 7.7 | 7.2 | 17.5 |
| | GN-05 | 20-Aug-18 | 14 | 501280 | 6174425 | 21.9 | 8.0 | 7.7 | 17.5 |
| | GN-06 | 20-Aug-18 | 14 | 500136 | 6174906 | 21.6 | 5.2 | 5.8 | 17.5 |
| | GN-09 | 22-Aug-18 | 14 | 503287 | 6169950 | 23.5 | 4.6 | 5.4 | 17.0 |
| | GN-13 | 21-Aug-18 | 14 | 508566 | 6170491 | 22.7 | 5.4 | 5.5 | 17.0 |
| | GN-15 | 22-Aug-18 | 14 | 507146 | 6169560 | 22.8 | 6.1 | 6.2 | 17.0 |
| | GN-16 | 21-Aug-18 | 14 | 507532 | 6169633 | 23.6 | 6.3 | 7.1 | 17.0 |
| | GN-17 | 21-Aug-18 | 14 | 505974 | 6174711 | 21.9 | 5.6 | 5.6 | 17.0 |
| | SN-09 | 22-Aug-18 | 14 | 503315 | 6169954 | 23.5 | 5.3 | 4.6 | 17.0 |
| | SN-15 | 22-Aug-18 | 14 | 507123 | 6169473 | 22.8 | 6.2 | 6.1 | 17.0 |
| | SN-16 | 21-Aug-18 | 14 | 507519 | 6169588 | 23.6 | 6.1 | 6.3 | 17.0 |
| | GN-02 | 15-Aug-19 | 14 | 503997 | 6172946 | 22.6 | 3.4 | 5.6 | 18.0 |
| | GN-04 | 14-Aug-19 | 14 | 502262 | 6174488 | 22.9 | 7.0 | 6.9 | 19.0 |
| | GN-05 | 14-Aug-19 | 14 | 501255 | 6174421 | 22.5 | 7.6 | 7.5 | 19.0 |
| | GN-06 | 13-Aug-19 | 14 | 500109 | 6174941 | 22.3 | 3.9 | 5.5 | 19.0 |
| | GN-09 | 15-Aug-19 | 14 | 503330 | 6169973 | 23.4 | 4.5 | 5.0 | 18.0 |
| | GN-13 | 13-Aug-19 | 14 | 508605 | 6170597 | 22.3 | 3.0 | 5.2 | 19.0 |
| | GN-15 | 13-Aug-19 | 14 | 507225 | 6169305 | 21.9 | 5.9 | 5.8 | 19.0 |
| | GN-16 | 14-Aug-19 | 14 | 507586 | 6169611 | 23.1 | 6.5 | 6.8 | 18.0 |
| | GN-17 | 15-Aug-19 | 14 | 506006 | 6174649 | 23.3 | 4.3 | 5.6 | 18.0 |
| | SN-09 | 15-Aug-19 | 14 | 503360 | 6169988 | 23.4 | 3.9 | 4.5 | 18.0 |
| | SN-15 | 13-Aug-19 | 14 | 507228 | 6169333 | 21.9 | 5.9 | 5.9 | 19.0 |
| | SN-16 | 14-Aug-19 | 14 | 507559 | 6169597 | 23.1 | 6.0 | 6.5 | 18.0 |
| | GN-02 | 12-Aug-20 | 14 | 503830 | 6172979 | 23.1 | 4.9 | 4.7 | 20.0 |
| | GN-04 | 13-Aug-20 | 14 | 502097 | 6174472 | 22.7 | 6.2 | 6.1 | 19.0 |
| | GN-05 | 13-Aug-20 | 14 | 501249 | 6174420 | 21.9 | 6.5 | 6.7 | 19.0 |
| | GN-06 | 13-Aug-20 | 14 | 500132 | 6174903 | 21.0 | 4.8 | 3.9 | 19.0 |
| | GN-09 | 12-Aug-20 | 14 | 503283 | 6170034 | 22.6 | 4.6 | 3.1 | 20.0 |



Table A5-1-1. continued.

| Waterbody | Site | Set Date | | UTM Coordina | | Duration (dec. hrs) | Wa Dep (m | oth | Set Water Temperature (°C) |
|-----------|-------|-------------|------|-----------------|----------|------------------------|-----------------|------|----------------------------------|
| | | | Zone | Easting | Northing | | Start | End | () |
| 3PT | GN-13 | 11-Aug-20 | 14 | 508573 | 6170515 | 25.2 | 3.3 | 4.4 | 19.0 |
| | GN-15 | 11-Aug-20 | 14 | 507247 | 6169395 | 24.5 | 5.0 | 5.0 | 19.0 |
| | GN-16 | 11-Aug-20 | 14 | 507696 | 6169640 | 24.9 | 5.2 | 5.3 | 19.0 |
| | GN-17 | 12-Aug-20 | 14 | 505993 | 6174733 | 23.5 | 4.7 | 4.7 | 20.0 |
| | SN-09 | 12-Aug-20 | 14 | 503306 | 6170015 | 22.6 | 3.1 | 2.7 | 20.0 |
| | SN-15 | 11-Aug-20 | 14 | 507249 | 6169357 | 24.5 | 5.0 | 5.0 | 19.0 |
| | SN-16 | 11-Aug-20 | 14 | 507722 | 6169651 | 24.9 | 5.3 | - | 19.0 |
| | GN-02 | 18-Aug-21 | 14 | 503825 | 6172988 | 23.1 | 5.6 | 5.7 | 18.0 |
| | GN-04 | 19-Aug-21 | 14 | 502088 | 6174443 | 24.6 | 7.0 | 7.2 | 17.0 |
| | GN-05 | 18-Aug-21 | 14 | 501259 | 6174401 | 23.8 | 7.7 | 7.5 | 18.0 |
| | GN-06 | 19-Aug-21 | 14 | 500133 | 6174898 | 23.1 | 5.2 | 5.8 | 17.0 |
| | GN-09 | 18-Aug-21 | 14 | 503302 | 6169987 | 22.8 | 4.7 | 5.6 | 18.0 |
| | GN-13 | 17-Aug-21 | 14 | 508607 | 6170508 | 21.3 | 5.7 | 5.6 | 18.0 |
| | GN-15 | 17-Aug-21 | 14 | 507216 | 6169351 | 22.8 | 5.9 | 5.9 | 18.0 |
| | GN-16 | 17-Aug-21 | 14 | 507664 | 6169747 | 22.2 | 7.7 | 5.9 | 18.0 |
| | GN-17 | 19-Aug-21 | 14 | 505987 | 6174703 | 23.1 | 5.6 | 5.7 | 17.0 |
| | SN-09 | 18-Aug-21 | 14 | 503338 | 6169983 | 22.8 | 4.2 | 4.7 | 18.0 |
| | SN-15 | 17-Aug-21 | 14 | 507244 | 6169347 | 22.8 | 5.9 | 5.9 | 18.0 |
| | SN-16 | 17-Aug-21 | 14 | 507657 | 6169705 | 22.2 | 7.7 | 7.7 | 18.0 |
| FOOT | GN-02 | 21-Aug-10 | 14 | 511178 | 6187513 | 44.6 | 3.8 | 4.6 | 15.0 |
| | GN-03 | 21-Aug-10 | 14 | 509817 | 6186051 | 43.6 | 4.8 | - | 15.0 |
| | GN-05 | 18-Aug-10 | 14 | 508121 | 6182721 | 23.9 | 11.2 | 12.0 | 15.0 |
| | GN-06 | 19-Aug-10 | 14 | 509547 | 6181358 | 23.5 | 4.8 | 3.7 | 14.5 |
| | GN-09 | 18-Aug-10 | 14 | 506107 | 6180601 | 23.0 | 6.2 | 6.2 | 15.0 |
| | GN-11 | 18-Aug-10 | 14 | 504994 | 6179901 | 21.8 | 7.7 | 7.5 | 15.0 |
| | GN-12 | 19-Aug-10 | 14 | 509639 | 6183391 | 23.7 | 11.2 | 6.7 | 15.0 |
| | GN-13 | 17-Aug-10 | 14 | 505945 | 6189559 | 22.7 | 2.9 | - | 14.0 |
| | GN-14 | 17-Aug-10 | 14 | 504360 | 6186324 | 23.7 | 3.7 | 5.4 | 14.0 |
| | SN-06 | 19-Aug-10 | 14 | 509547 | 6181358 | 23.5 | 4.8 | 4.8 | 14.5 |
| | SN-09 | 18-Aug-10 | 14 | 506107 | 6180601 | 23.0 | 6.2 | 6.2 | 15.0 |
| | SN-14 | 17-Aug-10 | 14 | 504360 | 6186324 | 23.7 | 3.7 | 3.7 | 14.0 |
| | GN-02 | 19-Aug-13 | 14 | 511155 | 6187603 | 22.1 | 3.9 | 4.8 | 21.0 |
| | GN-03 | 19-Aug-13 | 14 | 509929 | 6186020 | 23.0 | 1.2 | 5.1 | 21.0 |
| | GN-05 | 18-Aug-13 | 14 | 508151 | 6182756 | 23.7 | 13.0 | 12.5 | 20.0 |
| | GN-06 | 19-Aug-13 | 14 | 509520 | 6181353 | 24.3 | 5.0 | 2.9 | 19.0 |



Table A5-1-1. continued.

| Waterbody | Site | Set Date | | UTM Coordina | | Duration (dec. hrs) | Wa Dep (m | oth | Set Water Temperature |
|-----------|-------|-------------|------|-----------------|----------|------------------------|-----------------|------|-----------------------|
| | | | Zone | Easting | Northing | | Start | End | (°C) |
| FOOT | GN-09 | 18-Aug-13 | 14 | 506038 | 6180704 | 23.4 | 6.5 | 6.6 | 20.0 |
| | GN-11 | 17-Aug-13 | 14 | 504983 | 6179880 | 24.4 | 7.9 | 8.2 | 21.0 |
| | GN-12 | 18-Aug-13 | 14 | 509598 | 6183480 | 24.6 | 11.4 | 12.1 | 21.0 |
| | GN-13 | 17-Aug-13 | 14 | 505956 | 6189559 | 22.6 | 9.9 | 5.3 | 22.0 |
| | GN-14 | 17-Aug-13 | 14 | 504444 | 6186330 | 23.1 | 4.5 | 6.0 | 20.0 |
| | SN-06 | 19-Aug-13 | 14 | 509430 | 6181247 | 24.3 | 5.0 | 5.0 | 19.0 |
| | SN-09 | 18-Aug-13 | 14 | 506118 | 6180509 | 23.4 | 2.7 | 6.5 | 20.0 |
| | SN-14 | 17-Aug-13 | 14 | 504297 | 6186347 | 23.1 | 1.0 | 4.5 | 20.0 |
| | GN-02 | 15-Aug-16 | 14 | 511312 | 6187442 | 21.5 | 5.6 | 5.5 | 18.0 |
| | GN-03 | 15-Aug-16 | 14 | 509710 | 6186155 | 22.1 | 6.0 | 6.2 | 18.0 |
| | GN-05 | 14-Aug-16 | 14 | 508164 | 6182589 | 22.4 | 14.5 | 13.2 | 18.0 |
| | GN-06 | 15-Aug-16 | 14 | 509473 | 6181251 | 24.1 | 5.7 | 6.3 | 18.0 |
| | GN-09 | 14-Aug-16 | 14 | 506076 | 6180715 | 21.8 | 8.0 | 7.6 | 18.0 |
| | GN-11 | 13-Aug-16 | 14 | 505154 | 6179821 | 23.2 | 2.0 | 9.1 | 18.0 |
| | GN-12 | 14-Aug-16 | 14 | 509610 | 6183511 | 23.3 | 6.5 | 12.8 | 18.0 |
| | GN-13 | 13-Aug-16 | 14 | 505820 | 6189469 | 21.6 | 3.1 | 6.7 | 18.0 |
| | GN-14 | 13-Aug-16 | 14 | 504436 | 6186399 | 23.6 | 6.7 | 6.6 | 18.0 |
| | SN-06 | 15-Aug-16 | 14 | 509547 | 6181375 | 24.1 | 5.2 | 5.7 | 18.0 |
| | SN-09 | 14-Aug-16 | 14 | 506130 | 6180589 | 21.8 | 7.5 | 8.0 | 18.0 |
| | SN-14 | 13-Aug-16 | 14 | 504301 | 6186337 | 23.6 | 1.8 | 6.7 | 18.0 |
| | GN-02 | 18-Aug-19 | 14 | 511163 | 6187598 | 23.7 | 5.6 | 5.4 | 18.0 |
| | GN-03 | 18-Aug-19 | 14 | 509934 | 6186022 | 24.1 | 5.6 | 6.2 | 18.0 |
| | GN-05 | 16-Aug-19 | 14 | 508117 | 6182743 | 21.0 | 15.5 | 12.3 | 18.0 |
| | GN-06 | 19-Aug-19 | 14 | 509510 | 6181356 | 23.1 | 5.8 | 5.9 | 17.0 |
| | GN-09 | 18-Aug-19 | 14 | 506076 | 6180567 | 22.3 | 8.2 | 7.9 | 18.0 |
| | GN-11 | 16-Aug-19 | 14 | 505118 | 6179866 | 21.9 | 9.3 | 7.6 | 18.0 |
| | GN-12 | 19-Aug-19 | 14 | 509609 | 6183575 | 22.4 | 12.5 | 11.0 | 17.0 |
| | GN-13 | 17-Aug-19 | 14 | 505854 | 6189459 | 23.0 | 6.1 | 5.7 | 18.0 |
| | GN-14 | 17-Aug-19 | 14 | 504406 | 6186321 | 23.9 | 6.8 | 6.7 | 18.0 |
| | SN-06 | 19-Aug-19 | 14 | 509519 | 6181382 | 23.1 | 5.2 | 5.8 | 17.0 |
| | SN-09 | 18-Aug-19 | 14 | 506098 | 6180546 | 22.3 | 7.9 | 8.2 | 18.0 |
| | SN-14 | 17-Aug-19 | 14 | 504367 | 6186327 | 23.9 | 6.7 | 6.8 | 18.0 |
| APU | GN-01 | 30-Aug-09 | 14 | 582102 | 6186932 | 22.5 | 4.3 | 4.2 | 15.0 |
| | GN-02 | 30-Aug-09 | 14 | 582543 | 6186771 | 23.0 | 6.5 | 7.7 | 15.0 |
| | GN-03 | 31-Aug-09 | 14 | 589045 | 6191129 | 25.3 | 4.4 | 6.3 | 15.0 |



Table A5-1-1. continued.

| Waterbody | Site | Set Date | | UTM Coordina | | Duration (dec. hrs) | Wa Dep (m | oth | Set Water Temperature |
|-----------|-------|-------------|------|-----------------|----------|------------------------|-----------------|-----|-----------------------|
| | | | Zone | Easting | Northing | | Start | End | (°C) |
| APU | GN-04 | 31-Aug-09 | 14 | 589079 | 6190616 | 25.7 | 4.1 | 4.4 | 15.0 |
| | GN-05 | 1-Sep-09 | 14 | 587271 | 6189353 | 24.3 | 4.2 | 4.0 | 16.0 |
| | GN-06 | 1-Sep-09 | 14 | 587191 | 6188491 | 25.5 | 4.4 | 4.3 | 16.0 |
| | GN-07 | 1-Sep-09 | 14 | 585096 | 6188941 | 23.3 | 3.9 | 4.0 | 16.0 |
| | GN-08 | 2-Sep-09 | 14 | 584187 | 6188111 | 24.8 | 3.9 | 3.9 | 15.5 |
| | GN-09 | 2-Sep-09 | 14 | 583994 | 6187194 | 25.8 | 6.3 | 7.4 | 15.5 |
| | SN-03 | 31-Aug-09 | 14 | 589108 | 6191131 | 25.3 | 6.2 | 6.3 | 15.0 |
| | SN-06 | 1-Sep-09 | 14 | 587124 | 6188527 | 25.5 | 4.0 | 4.2 | 16.0 |
| | SN-09 | 2-Sep-09 | 14 | 583958 | 6187130 | 25.8 | 4.2 | 4.0 | 15.5 |
| | GN-01 | 24-Aug-12 | 14 | 581930 | 6187032 | 25.8 | 5.1 | 5.3 | 18.0 |
| | GN-02 | 24-Aug-12 | 14 | 582707 | 6186622 | 26.7 | 5.5 | 4.6 | 18.0 |
| | GN-03 | 27-Aug-12 | 14 | 589043 | 6191137 | 22.5 | 5.8 | 5.7 | 18.0 |
| | GN-04 | 27-Aug-12 | 14 | 589043 | 6190591 | 22.0 | 5.1 | 5.2 | 18.0 |
| | GN-05 | 26-Aug-12 | 14 | 587326 | 6189266 | 18.6 | 5.4 | 5.4 | 19.0 |
| | GN-06 | 26-Aug-12 | 14 | 587180 | 6188615 | 20.0 | 5.2 | 5.0 | 19.0 |
| | GN-07 | 26-Aug-12 | 14 | 585094 | 6189032 | 18.0 | 6.0 | 7.7 | 19.0 |
| | GN-08 | 25-Aug-12 | 14 | 583722 | 6187976 | 21.3 | 5.3 | 5.1 | 19.0 |
| | GN-09 | 25-Aug-12 | 14 | 583733 | 6187344 | 23.0 | 6.4 | 6.6 | 19.0 |
| | SN-03 | 27-Aug-12 | 14 | 589069 | 6191154 | 22.5 | 5.4 | 5.4 | 18.0 |
| | SN-06 | 26-Aug-12 | 14 | 587200 | 6188638 | 20.3 | 5.2 | 5.2 | 19.0 |
| | SN-09 | 25-Aug-12 | 14 | 583711 | 6187314 | 23.0 | 6.0 | 6.0 | 19.0 |
| | GN-01 | 27-Aug-15 | 14 | 581937 | 6187009 | 24.0 | 4.6 | 4.5 | 17.0 |
| | GN-02 | 27-Aug-15 | 14 | 582709 | 6186664 | 24.7 | 4.9 | 4.6 | 17.0 |
| | GN-03 | 25-Aug-15 | 14 | 589039 | 6191073 | 24.1 | 5.1 | 4.9 | 17.0 |
| | GN-04 | 25-Aug-15 | 14 | 589039 | 6190582 | 24.3 | 4.7 | 4.7 | 17.0 |
| | GN-05 | 25-Aug-15 | 14 | 587415 | 6189345 | 24.4 | 4.9 | 4.9 | 17.0 |
| | GN-06 | 26-Aug-15 | 14 | 587180 | 6188608 | 24.0 | 4.6 | 4.7 | 16.0 |
| | GN-07 | 26-Aug-15 | 14 | 585103 | 6189031 | 24.8 | 5.6 | 7.3 | 16.0 |
| | GN-08 | 26-Aug-15 | 14 | 583702 | 6187987 | 25.7 | 4.8 | 4.6 | 16.0 |
| | GN-09 | 27-Aug-15 | 14 | 583726 | 6187363 | 24.8 | 6.1 | 7.7 | 17.0 |
| | SN-03 | 25-Aug-15 | 14 | 589051 | 6191127 | 24.1 | 5.1 | 5.1 | 17.0 |
| | SN-06 | 26-Aug-15 | 14 | 587199 | 6188632 | 24.0 | 4.7 | 4.6 | 16.0 |
| | SN-09 | 27-Aug-15 | 14 | 583724 | 6187328 | 24.8 | 4.9 | 6.1 | 17.0 |
| | GN-01 | 19-Aug-18 | 14 | 581944 | 6187004 | 19.7 | 5.4 | 5.9 | 17.0 |
| | GN-02 | 19-Aug-18 | 14 | 582534 | 6186793 | 19.9 | 8.7 | 5.7 | 17.0 |



Table A5-1-1. continued.

| Waterbody | Site | Set Date | | UTM Coordina | | Duration (dec. hrs) | Wa Dep (m | oth | Set Water Temperature |
|-----------|-------|-------------|------|-----------------|----------|------------------------|-----------------|------|-----------------------|
| | | | Zone | Easting | Northing | | Start | End | (°C) |
| APU | GN-03 | 17-Aug-18 | 14 | 589084 | 6191115 | 24.6 | 5.6 | 6.2 | 17.0 |
| | GN-04 | 17-Aug-18 | 14 | 589125 | 6190730 | 24.2 | 5.5 | 5.3 | 17.0 |
| | GN-05 | 17-Aug-18 | 14 | 587429 | 6189350 | 25.3 | 5.6 | 6.1 | 17.0 |
| | GN-06 | 18-Aug-18 | 14 | 587169 | 6188550 | 29.0 | 5.5 | 5.5 | 17.0 |
| | GN-07 | 18-Aug-18 | 14 | 585027 | 6189049 | 28.7 | 5.6 | 7.9 | 17.0 |
| | GN-08 | 18-Aug-18 | 14 | 584033 | 6188114 | 28.7 | 7.3 | 7.5 | 17.0 |
| | GN-09 | 19-Aug-18 | 14 | 583977 | 6187159 | 20.5 | 5.5 | 5.8 | 17.0 |
| | SN-03 | 17-Aug-18 | 14 | 589121 | 6191122 | 24.6 | 5.2 | 5.6 | 17.0 |
| | SN-06 | 18-Aug-18 | 14 | 587133 | 6188533 | 29.0 | 5.2 | 5.5 | 17.0 |
| | SN-09 | 19-Aug-18 | 14 | 583970 | 6187131 | 20.5 | 5.8 | 5.5 | 17.0 |
| | GN-01 | 19-Aug-21 | 14 | 581932 | 6186901 | 24.0 | 5.9 | 5.4 | 16.0 |
| | GN-02 | 19-Aug-21 | 14 | 582637 | 6186864 | 23.0 | 8.8 | 9.0 | 16.0 |
| | GN-03 | 17-Aug-21 | 14 | 589075 | 6191159 | 20.8 | 6.3 | 7.6 | 17.0 |
| | GN-04 | 17-Aug-21 | 14 | 589079 | 6190617 | 21.4 | 5.3 | 5.5 | 17.0 |
| | GN-05 | 17-Aug-21 | 14 | 587344 | 6189455 | 22.0 | 5.5 | 5.8 | 17.0 |
| | GN-06 | 18-Aug-21 | 14 | 587153 | 6188481 | 20.9 | 5.2 | 5.2 | 17.0 |
| | GN-07 | 18-Aug-21 | 14 | 585108 | 6188952 | 21.4 | 7.7 | 8.2 | 17.0 |
| | GN-08 | 18-Aug-21 | 14 | 584085 | 6188065 | 21.4 | 7.5 | 7.4 | 17.0 |
| | GN-09 | 19-Aug-21 | 14 | 583941 | 6187190 | 22.6 | 5.2 | 5.4 | 16.0 |
| | SN-03 | 17-Aug-21 | 14 | 589106 | 6191149 | 20.8 | 4.6 | 6.3 | 17.0 |
| | SN-06 | 18-Aug-21 | 14 | 587110 | 6188519 | 20.9 | 5.0 | 5.2 | 17.0 |
| | SN-09 | 19-Aug-21 | 14 | 583927 | 6187160 | 22.6 | 5.2 | 5.2 | 16.0 |
| LEFT | GN-01 | 30-Jul-09 | 14 | 525846 | 6217158 | 24.8 | 4.0 | 6.3 | 17.0 |
| | GN-02 | 31-Jul-09 | 14 | 525373 | 6216074 | 24.8 | 3.5 | 2.3 | 16.0 |
| | GN-05 | 31-Jul-09 | 14 | 523426 | 6217475 | 23.0 | 4.0 | 2.7 | 16.0 |
| | GN-08 | 30-Jul-09 | 14 | 524030 | 6213574 | 25.4 | 9.8 | 10.1 | 17.0 |
| | GN-09 | 30-Jul-09 | 14 | 521582 | 6213736 | 26.2 | 4.4 | 3.6 | 17.0 |
| | GN-10 | 29-Jul-09 | 14 | 518299 | 6210078 | 25.3 | 4.1 | 6.3 | 17.0 |
| | GN-11 | 28-Jul-09 | 14 | 519799 | 6209109 | 22.9 | 12.7 | 7.3 | 17.0 |
| | GN-12 | 28-Jul-09 | 14 | 517146 | 6213104 | 21.5 | 5.2 | 4.7 | 17.0 |
| | GN-13 | 29-Jul-09 | 14 | 517898 | 6212949 | 24.7 | 9.4 | 8.4 | 17.0 |
| | SN-05 | 31-Jul-09 | 14 | 523453 | 6217489 | 23.0 | 3.9 | 4.0 | 16.0 |
| | SN-08 | 30-Jul-09 | 14 | 524004 | 6213611 | 25.4 | 8.8 | 9.8 | 17.0 |
| | SN-12 | 28-Jul-09 | 14 | 517142 | 6213075 | 21.5 | 4.8 | 5.2 | 17.0 |
| | GN-01 | 25-Jul-10 | 14 | 525927 | 6216983 | 14.1 | 4.5 | 5.5 | 21.0 |



Table A5-1-1. continued.

| Waterbody | Site | Set Date | | UTM Coordina | | Duration (dec. hrs) | Wa Dep (m | oth | Set Water Temperature (°C) |
|-----------|-------|-------------|------|-----------------|----------|------------------------|-----------------|------|----------------------------------|
| | | | Zone | Easting | Northing | | Start | End | () |
| LEFT | GN-02 | 25-Jul-10 | 14 | 525506 | 6216031 | 14.7 | 2.5 | 2.0 | 21.0 |
| | GN-05 | 25-Jul-10 | 14 | 523396 | 6217408 | 15.1 | 4.5 | 1.0 | 21.0 |
| | GN-08 | 26-Jul-10 | 14 | 523809 | 6213799 | 27.5 | 5.0 | 7.0 | 21.0 |
| | GN-09 | 27-Jul-10 | 14 | 521700 | 6213666 | 20.0 | 3.0 | - | 20.0 |
| | GN-10 | 27-Jul-10 | 14 | 518468 | 6209942 | 20.8 | 5.5 | 5.5 | 20.0 |
| | GN-11 | 26-Jul-10 | 14 | 519766 | 6209122 | 23.7 | 10.0 | 1.5 | 21.0 |
| | GN-12 | 27-Jul-10 | 14 | 517109 | 6213471 | 20.8 | 2.0 | 4.0 | 20.0 |
| | GN-13 | 26-Jul-10 | 14 | 517899 | 6213124 | 25.0 | 7.0 | 2.0 | 21.0 |
| | SN-05 | 25-Jul-10 | 14 | 523396 | 6217408 | 15.1 | 4.5 | 4.5 | 21.0 |
| | SN-08 | 26-Jul-10 | 14 | 523809 | 6213799 | 27.5 | 5.0 | 5.0 | 21.0 |
| | SN-12 | 27-Jul-10 | 14 | 517109 | 6213471 | 20.8 | 2.0 | 2.0 | 20.0 |
| | GN-01 | 22-Jul-11 | 14 | 525632 | 6217203 | 18.8 | 5.5 | 2.5 | 19.0 |
| | GN-02 | 22-Jul-11 | 14 | 525351 | 6216123 | 21.7 | 1.3 | 2.5 | 19.0 |
| | GN-05 | 22-Jul-11 | 14 | 523435 | 6217418 | 19.8 | 3.3 | 0.5 | 19.0 |
| | GN-08 | 23-Jul-11 | 14 | 524023 | 6213553 | 17.8 | 8.5 | 6.8 | 19.0 |
| | GN-09 | 23-Jul-11 | 14 | 521509 | 6213765 | 18.4 | 1.8 | 3.0 | 19.0 |
| | GN-10 | 24-Jul-11 | 14 | 518283 | 6210095 | 23.1 | 3.5 | 5.5 | 18.0 |
| | GN-11 | 24-Jul-11 | 14 | 519759 | 6209209 | 22.5 | 11.5 | 12.0 | 18.0 |
| | GN-12 | 24-Jul-11 | 14 | 517122 | 6213105 | 24.4 | 4.3 | 5.5 | 18.0 |
| | GN-13 | 23-Jul-11 | 14 | 517873 | 6212955 | 19.7 | 9.0 | 7.0 | 19.0 |
| | SN-05 | 22-Jul-11 | 14 | 523435 | 6217418 | 21.4 | 3.3 | 3.3 | 19.0 |
| | SN-08 | 23-Jul-11 | 14 | 524023 | 6213553 | 17.8 | 8.5 | 8.5 | 19.0 |
| | SN-12 | 24-Jul-11 | 14 | 517122 | 6213105 | 24.4 | 4.3 | 4.3 | 18.0 |
| | GN-01 | 24-Jul-12 | 14 | 525907 | 6216964 | 15.8 | 6.8 | 6.6 | 23.5 |
| | GN-02 | 24-Jul-12 | 14 | 525532 | 6216088 | 15.1 | 4.7 | 1.6 | 24.0 |
| | GN-05 | 24-Jul-12 | 14 | 523300 | 6217422 | 14.5 | 2.2 | 3.5 | 23.5 |
| | GN-08 | 25-Jul-12 | 14 | 523803 | 6213783 | 13.8 | 4.7 | 6.1 | 23.0 |
| | GN-09 | 25-Jul-12 | 14 | 521489 | 6213767 | 16.0 | 1.4 | 2.9 | 22.0 |
| | GN-10 | 26-Jul-12 | 14 | 518587 | 6210009 | 21.7 | 5.5 | 5.4 | 21.0 |
| | GN-11 | 26-Jul-12 | 14 | 519524 | 6209068 | 21.3 | 4.5 | 6.9 | 21.0 |
| | GN-12 | 26-Jul-12 | 14 | 517143 | 6213482 | 24.2 | 1.5 | 4.6 | 22.0 |
| | GN-13 | 25-Jul-12 | 14 | 517988 | 6213040 | 16.9 | 5.9 | 3.2 | 22.5 |
| | SN-05 | 24-Jul-12 | 14 | 523296 | 6217422 | 14.5 | 1.8 | 1.8 | 23.5 |
| | SN-08 | 25-Jul-12 | 14 | 523804 | 6213800 | 13.8 | 4.5 | 4.5 | 23.0 |
| | SN-12 | 26-Jul-12 | 14 | 517137 | 6213491 | 24.2 | 1.6 | 1.6 | 22.0 |



Table A5-1-1. continued.

| Waterbody | Site | Set Date | | UTM Coordin | | Duration (dec. hrs) | Wa De _l (n | oth | Set Water Temperature | |
|-----------|-------|-------------|------|----------------|----------|------------------------|-----------------------------|------|--------------------------|--|
| | | | Zone | Easting | Northing | (| Start | End | (°C) | |
| LEFT | GN-01 | 27-Jul-13 | 14 | 525881 | 6217063 | 19.4 | 5.5 | 5.6 | 22.0 | |
| | GN-02 | 27-Jul-13 | 14 | 525439 | 6216156 | 19.9 | 5.0 | 1.4 | 22.0 | |
| | GN-05 | 27-Jul-13 | 14 | 523439 | 6217477 | 19.0 | 3.2 | 3.1 | 22.0 | |
| | GN-08 | 28-Jul-13 | 14 | 523788 | 6213672 | 22.6 | 4.6 | 5.5 | 19.0 | |
| | GN-09 | 28-Jul-13 | 14 | 521596 | 6213742 | 23.4 | 2.9 | 3.0 | 19.0 | |
| | GN-10 | 29-Jul-13 | 14 | 518413 | 6210003 | 22.6 | 5.6 | 5.4 | 19.0 | |
| | GN-11 | 29-Jul-13 | 14 | 519673 | 6209066 | 22.2 | 11.2 | 4.8 | 18.0 | |
| | GN-12 | 29-Jul-13 | 14 | 517228 | 6213374 | 22.6 | 2.4 | 4.5 | 19.0 | |
| | GN-13 | 28-Jul-13 | 14 | 517995 | 6213168 | 24.3 | 7.3 | 3.7 | 20.0 | |
| | SN-05 | 27-Jul-13 | 14 | 523297 | 6217423 | 19.0 | 2.3 | 3.2 | 22.0 | |
| | SN-08 | 28-Jul-13 | 14 | 523796 | 6213818 | 22.6 | 4.3 | 4.6 | 19.0 | |
| | SN-12 | 29-Jul-13 | 14 | 517135 | 6213488 | 22.6 | 0.8 | 2.4 | 19.0 | |
| | GN-01 | 22-Jul-14 | 14 | 525944 | 6217018 | 18.8 | 6.2 | 5.9 | 20.0 | |
| | GN-02 | 22-Jul-14 | 14 | 525432 | 6216164 | 19.8 | 0.9 | 4.2 | 20.0 | |
| | GN-05 | 22-Jul-14 | 14 | 523396 | 6217408 | 19.0 | 3.0 | 3.7 | 20.0 | |
| | GN-08 | 23-Jul-14 | 14 | 523767 | 6213789 | 20.3 | 6.3 | 6.3 | 21.0 | |
| | GN-09 | 23-Jul-14 | 14 | 521598 | 6213751 | 20.4 | 3.5 | 3.4 | 21.0 | |
| | GN-10 | 24-Jul-14 | 14 | 518468 | 6209942 | 23.3 | 0.0 | 0.0 | - | |
| | GN-11 | 24-Jul-14 | 14 | 519766 | 6209122 | 22.4 | 0.0 | 0.0 | - | |
| | GN-12 | 24-Jul-14 | 14 | 517109 | 6213471 | 22.9 | 0.0 | 0.0 | - | |
| | GN-13 | 23-Jul-14 | 14 | 517899 | 6213124 | 21.6 | 0.0 | 0.0 | 21.0 | |
| | SN-05 | 22-Jul-14 | 14 | 523276 | 6217405 | 19.0 | 1.8 | 3.0 | 20.0 | |
| | SN-08 | 23-Jul-14 | 14 | 523782 | 6213822 | 20.3 | 5.0 | 5.0 | 21.0 | |
| | SN-12 | 24-Jul-14 | 14 | 517109 | 6213471 | 22.9 | 0.0 | 0.0 | - | |
| | GN-01 | 23-Jul-15 | 14 | 526040 | 6217019 | 20.1 | 6.8 | 6.9 | 20.0 | |
| | GN-02 | 23-Jul-15 | 14 | 523818 | 6213805 | 19.8 | 0.5 | 1.0 | 20.0 | |
| | GN-05 | 23-Jul-15 | 14 | 523300 | 6217428 | 18.3 | 3.2 | 1.3 | 20.0 | |
| | GN-08 | 22-Jul-15 | 14 | 523716 | 6213908 | 23.6 | 3.9 | 4.2 | 20.0 | |
| | GN-09 | 22-Jul-15 | 14 | 521646 | 6213690 | 22.5 | 3.0 | 3.0 | 20.0 | |
| | GN-10 | 21-Jul-15 | 14 | 518551 | 6209958 | 18.5 | 5.5 | 5.7 | 20.0 | |
| | GN-11 | 21-Jul-15 | 14 | 519661 | 6209062 | 17.8 | 8.1 | 11.1 | 20.0 | |
| | GN-12 | 21-Jul-15 | 14 | 517148 | 6213472 | 20.4 | 2.5 | 4.3 | 20.0 | |
| | GN-13 | 22-Jul-15 | 14 | 517818 | 6213217 | 23.4 | 7.5 | 4.1 | 21.0 | |
| | SN-05 | 23-Jul-15 | 14 | 523440 | 6217456 | 18.3 | 3.2 | 3.2 | 20.0 | |
| | SN-08 | 22-Jul-15 | 14 | 523724 | 6213915 | 23.6 | 4.2 | 3.9 | 20.0 | |



Table A5-1-1. continued.

| Waterbody | Site | Set Date | | UTM Coordina | | Duration (dec. hrs) | Wa Dep (m | oth | Set Water Temperature |
|-----------|-------|-------------|------|-----------------|----------|------------------------|-----------------|------|-----------------------|
| | | | Zone | Easting | Northing | | Start | End | (°C) |
| LEFT | SN-12 | 21-Jul-15 | 14 | 517139 | 6213489 | 20.4 | 1.4 | 2.5 | 20.0 |
| | GN-01 | 22-Jul-16 | 14 | 526040 | 6217019 | 19.5 | 5.9 | 5.3 | 18.5 |
| | GN-02 | 22-Jul-16 | 14 | 523818 | 6213805 | 20.8 | 4.9 | 1.7 | 19.0 |
| | GN-05 | 22-Jul-16 | 14 | 523300 | 6217428 | 19.4 | 2.7 | 3.4 | 18.0 |
| | GN-08 | 23-Jul-16 | 14 | 523782 | 6213838 | 22.2 | 4.5 | 4.2 | 18.0 |
| | GN-09 | 23-Jul-16 | 14 | 521700 | 6213662 | 22.6 | 3.2 | 3.0 | 19.0 |
| | GN-10 | 24-Jul-16 | 14 | 518386 | 6210016 | 21.1 | 5.9 | 5.6 | 18.0 |
| | GN-11 | 23-Jul-16 | 14 | 519781 | 6209076 | 23.5 | 11.7 | 4.9 | 19.0 |
| | GN-12 | 24-Jul-16 | 14 | 517125 | 6213454 | 23.2 | 2.7 | 4.9 | 18.0 |
| | GN-13 | 24-Jul-16 | 14 | 517899 | 6213127 | 23.8 | 6.7 | - | 19.0 |
| | SN-05 | 22-Jul-16 | 14 | 523440 | 6217456 | 19.4 | 2.2 | 2.7 | 18.0 |
| | SN-08 | 23-Jul-16 | 14 | 523774 | 6213796 | 22.2 | 5.1 | 4.5 | 18.0 |
| | SN-12 | 24-Jul-16 | 14 | 517111 | 6213483 | 23.2 | 2.4 | 2.7 | 18.0 |
| | GN-01 | 26-Jul-17 | 14 | 525932 | 6216992 | 21.7 | 6.1 | 6.2 | 20.0 |
| | GN-02 | 26-Jul-17 | 14 | 525442 | 6216154 | 23.1 | 0.9 | 4.3 | 21.0 |
| | GN-05 | 26-Jul-17 | 14 | 523319 | 6217419 | 19.3 | 3.0 | 3.4 | 20.0 |
| | GN-08 | 25-Jul-17 | 14 | 523756 | 6213783 | 23.2 | 4.7 | 4.9 | 20.0 |
| | GN-09 | 25-Jul-17 | 14 | 521713 | 6213651 | 21.7 | 3.0 | 3.2 | 19.0 |
| | GN-10 | 24-Jul-17 | 14 | 518466 | 6209963 | 19.8 | 5.9 | 6.5 | 20.0 |
| | GN-11 | 24-Jul-17 | 14 | 519765 | 6209125 | 19.1 | 11.9 | 12.1 | 20.0 |
| | GN-12 | 24-Jul-17 | 14 | 517125 | 6213449 | 21.6 | 3.2 | 4.6 | 20.0 |
| | GN-13 | 25-Jul-17 | 14 | 517986 | 6213183 | 20.6 | 3.4 | 7.6 | 20.0 |
| | SN-05 | 26-Jul-17 | 14 | 523296 | 6217424 | 19.3 | 3.0 | 1.5 | 20.0 |
| | SN-08 | 25-Jul-17 | 14 | 523787 | 6213794 | 23.2 | 4.7 | 4.6 | 20.0 |
| | SN-12 | 24-Jul-17 | 14 | 517108 | 6213473 | 21.6 | 3.2 | 2.4 | 20.0 |
| | GN-01 | 27-Jul-18 | 14 | 525931 | 6216987 | 23.1 | 6.7 | 6.6 | 18.0 |
| | GN-02 | 27-Jul-18 | 14 | 525437 | 6216146 | 24.0 | 1.8 | 4.7 | 18.0 |
| | GN-05 | 27-Jul-18 | 14 | 523423 | 6217408 | 21.7 | 3.6 | 3.6 | 18.0 |
| | GN-08 | 24-Jul-18 | 14 | 523790 | 6213768 | 23.7 | 4.9 | 6.2 | 19.0 |
| | GN-09 | 24-Jul-18 | 14 | 521695 | 6213672 | 23.0 | 3.0 | 3.4 | 18.0 |
| | GN-10 | 26-Jul-18 | 14 | 518466 | 6209945 | 23.3 | 5.9 | 6.9 | 18.0 |
| | GN-11 | 25-Jul-18 | 14 | 519764 | 6209115 | 24.0 | 11.3 | 11.3 | 18.0 |
| | GN-12 | 25-Jul-18 | 14 | 517118 | 6213434 | 25.2 | 4.8 | 4.4 | 18.0 |
| | GN-13 | 26-Jul-18 | 14 | 517898 | 6213124 | 22.6 | 7.8 | 8.9 | 18.0 |
| | SN-05 | 27-Jul-18 | 14 | 523394 | 6217410 | 21.7 | 3.5 | 3.6 | 18.0 |



Table A5-1-1. continued.

| Waterbody | Site | Set Date | | UTM Coordin | | Duration (dec. hrs) | Wa Dep (m | oth | Set Water Temperature (°C) |
|-----------|-------|-------------|------|----------------|----------|------------------------|-----------------|------|----------------------------------|
| | | | Zone | Easting | Northing | | Start | End | () |
| LEFT | SN-08 | 24-Jul-18 | 14 | 523802 | 6213795 | 23.7 | 4.5 | 4.9 | 19.0 |
| | SN-12 | 25-Jul-18 | 14 | 517110 | 6213461 | 25.2 | 1.6 | 4.8 | 18.0 |
| | GN-01 | 24-Jul-19 | 14 | 525806 | 6216940 | 23.6 | 5.9 | 3.7 | 22.0 |
| | GN-02 | 24-Jul-19 | 14 | 525498 | 6216030 | 26.2 | 2.3 | 4.0 | 21.0 |
| | GN-05 | 24-Jul-19 | 14 | 523389 | 6217416 | 21.9 | 3.3 | 3.3 | 21.0 |
| | GN-08 | 23-Jul-19 | 14 | 523763 | 6213789 | 23.4 | 4.4 | 5.0 | 21.0 |
| | GN-09 | 23-Jul-19 | 14 | 521716 | 6213649 | 22.5 | 2.9 | 3.0 | 20.0 |
| | GN-10 | 22-Jul-19 | 14 | 518463 | 6209944 | 23.6 | 5.7 | 6.3 | 19.0 |
| | GN-11 | 21-Jul-19 | 14 | 519775 | 6209123 | 22.4 | 11.2 | 11.7 | 19.0 |
| | GN-12 | 21-Jul-19 | 14 | 517140 | 6213443 | 23.3 | 3.7 | 5.5 | 19.0 |
| | GN-13 | 22-Jul-19 | 14 | 517906 | 6213112 | 22.9 | 7.4 | 6.6 | 20.0 |
| | SN-05 | 24-Jul-19 | 14 | 523359 | 6217420 | 21.9 | 3.6 | 3.6 | 21.0 |
| | SN-08 | 23-Jul-19 | 14 | 523795 | 6213790 | 23.4 | 4.5 | 4.4 | 21.0 |
| | SN-12 | 21-Jul-19 | 14 | 517120 | 6213470 | 23.3 | 2.3 | 3.7 | 19.0 |
| | GN-01 | 23-Jul-20 | 14 | 525901 | 6216989 | 24.3 | 6.2 | 6.1 | 21.0 |
| | GN-02 | 22-Jul-20 | 14 | 525462 | 6216067 | 21.5 | 1.8 | 5.4 | 21.5 |
| | GN-05 | 22-Jul-20 | 14 | 523305 | 6217424 | 23.3 | 3.8 | 3.9 | 21.5 |
| | GN-08 | 21-Jul-20 | 14 | 523846 | 6213770 | 22.0 | 5.5 | 6.2 | 22.5 |
| | GN-09 | 21-Jul-20 | 14 | 521623 | 6213762 | 20.5 | 3.5 | 3.6 | 22.5 |
| | GN-10 | 24-Jul-20 | 14 | 518385 | 6210042 | 21.7 | 6.0 | 6.1 | 22.0 |
| | GN-11 | 24-Jul-20 | 14 | 519675 | 6209068 | 20.8 | 4.6 | 11.7 | 22.0 |
| | GN-12 | 23-Jul-20 | 14 | 517135 | 6213432 | 24.0 | - | - | 21.0 |
| | GN-13 | 24-Jul-20 | 14 | 517981 | 6213159 | 22.0 | 6.4 | 9.0 | 22.0 |
| | SN-05 | 22-Jul-20 | 14 | 523294 | 6217430 | 23.3 | 2.1 | 3.8 | 21.5 |
| | SN-08 | 21-Jul-20 | 14 | 523831 | 6213770 | 22.0 | 5.1 | 5.5 | 22.5 |
| | SN-12 | 23-Jul-20 | 14 | 517126 | 6213459 | 24.0 | 3.0 | 4.8 | 21.0 |
| | GN-01 | 24-Jul-21 | 14 | 523913 | 6213682 | 21.8 | 5.8 | 7.6 | 20.0 |
| | GN-02 | 24-Jul-21 | 14 | 523332 | 6217420 | 22.6 | 1.8 | 4.3 | 20.0 |
| | GN-05 | 24-Jul-21 | 14 | 525866 | 6217041 | 25.1 | 3.7 | 3.4 | 20.0 |
| | GN-08 | 25-Jul-21 | 14 | 518405 | 6210004 | 22.2 | 5.9 | 4.6 | 20.0 |
| | GN-09 | 25-Jul-21 | 14 | 521615 | 6213734 | 23.4 | 3.0 | 3.1 | 20.0 |
| | GN-10 | 27-Jul-21 | 14 | 517111 | 6213458 | 23.0 | 5.8 | 6.3 | 20.0 |
| | GN-11 | 26-Jul-21 | 14 | 523295 | 6217434 | 23.9 | 8.8 | 11.4 | 20.0 |
| | GN-12 | 26-Jul-21 | 14 | 517121 | 6213414 | 23.6 | 5.1 | 4.6 | 20.0 |
| | GN-13 | 27-Jul-21 | 14 | 523986 | 6213798 | 22.7 | 4.1 | 7.3 | 20.0 |



Table A5-1-1. continued.

| Waterbody | Site | Set Date | | UTM Coordina | | Duration (dec. hrs) | Wa Dep (m | oth | Set Water Temperature (°C) |
|-----------|-------|-------------|-----------|-----------------|----------|------------------------|-----------------|-----|----------------------------------|
| | | | Zone | Easting | Northing | | Start | End | () |
| LEFT | SN-05 | 24-Jul-21 | 14 | 525405 | 6216123 | 25.1 | 1.4 | 3.7 | 20.0 |
| | SN-08 | 25-Jul-21 | 14 519703 | | 6209065 | 22.2 | 4.6 | 4.4 | 20.0 |
| | SN-12 | 26-Jul-21 | 14 | 517979 | 6213168 | 23.6 | 2.5 | 5.1 | 20.0 |



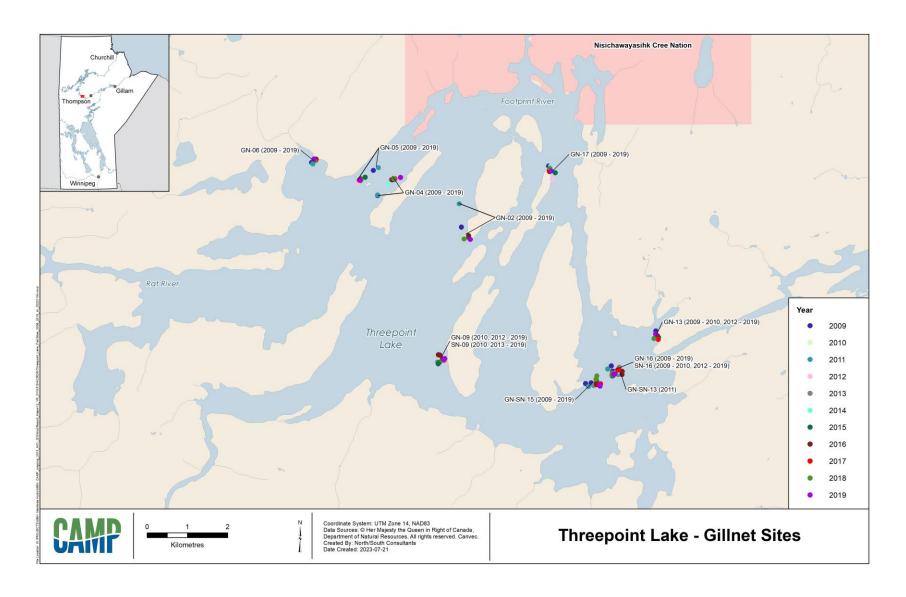


Figure A5-1-1. 2008-2019 Gillnetting sites in Threepoint Lake.



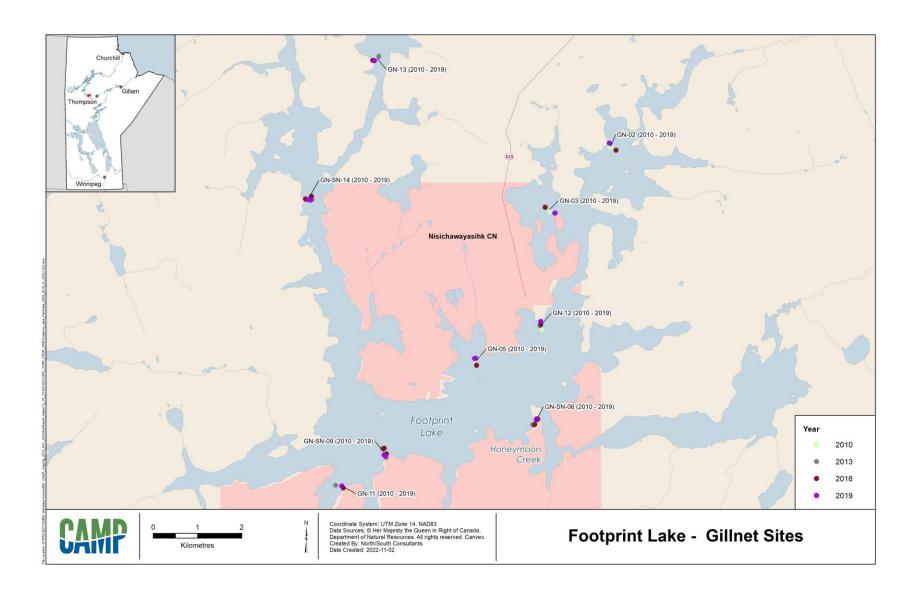


Figure A5-1-2. 2008-2019 Gillnetting sites in Footprint Lake.



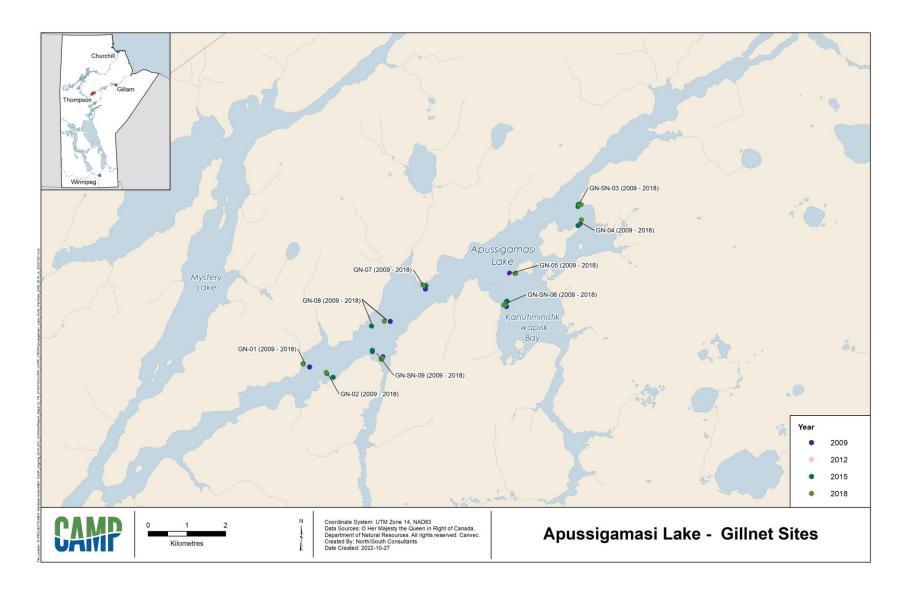


Figure A5-1-3. 2008-2019 Gillnetting sites in Apussigamasi Lake.



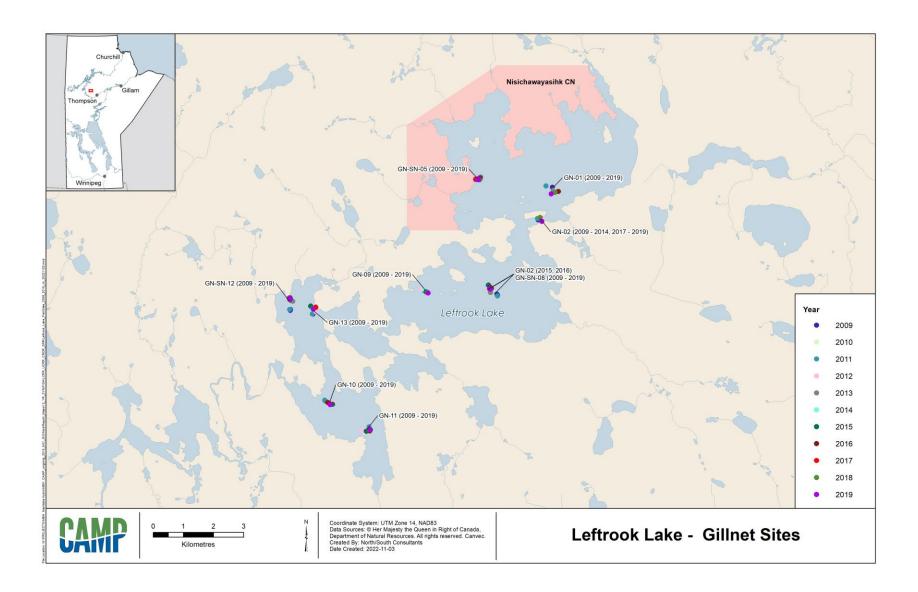


Figure A5-1-4. 2008-2019 Gillnetting sites in Leftrook Lake.



6.0 MERCURY IN FISH

6.1 INTRODUCTION

The following presents the results of fish mercury monitoring conducted from 2008-2019 in the Churchill River Diversion Region. Fish mercury sampling was conducted annually beginning in 2010 at one on-system waterbody, Threepoint Lake, and one off-system waterbody, Leftrook Lake (Table 6-1.1; Figure 6.1-1). Threepoint and Leftrook lakes are monitored annually to ensure short-term changes in mercury concentrations that may be indicative of regional (i.e., northern Manitoba) effects on the rates of mercury methylation and biomagnification are not being missed by rotational sampling. Leftrook Lake is the only reference water body sampled under CAMP with a strong historic record and Threepoint Lake is one of the waterbodies that has experienced the longest recovery time since construction of the CRD.

Mercury concentrations are measured in muscle tissue of commercially important fish species – Northern Pike, Walleye, and Lake Whitefish. Monitoring of mercury in 1-year-old Yellow Perch (*Perca flavescens*) is also conducted as a potential early indicator of changes in mercury in the food web. Samples of fish muscle are collected during the conduct of fish community monitoring. Mercury is analysed in the trunk muscle of Northern Pike, Lake Whitefish, and Walleye selected over a range of fork lengths. Yearling Yellow Perch are analyzed for mercury as carcass with the head, pelvic and pectoral girdles, caudal fin, and digestive tract removed.

There were no departures from the planned field sampling schedule during the 12-year period.

Two metrics were selected for detailed reporting: arithmetic mean mercury concentrations; and length-standardized mean mercury concentrations (also referred to as "standard mean(s)"; Table 6.1-2). Standard lengths varied by species as follows: Lake Whitefish (350 mm); Northern Pike (550 mm); and Walleye (400 mm). As CAMP targets a specific age class of Yellow Perch, fish captured for this component are inherently of a limited size range; therefore, length-standardization for this species was not undertaken.

A detailed description of the program design and sampling methods is provided in Technical Document 1, Section 2.6.



Table 6.1-1. 2008-2019 Inventory of fish mercury sampling.

| Waterbody/Area | Sampling Year | | | | | | | | | | | | |
|----------------|---------------|------|------|------|------|------|------|------|------|------|------|------|--|
| | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | |
| 3РТ | | | • | • | • | • | • | • | • | • | • | • | |
| LEFT | | | • | • | • | • | • | • | • | • | • | • | |

Table 6.1-2. Mercury in fish indicators and metrics.

| Indicator | Metric | Units |
|-----------------|---|-------------------------|
| | Arithmetic mean mercury concentration | Parts per million (ppm) |
| Mercury in Fish | Length-standardized mean mercury concentration of large-bodied species | ppm |



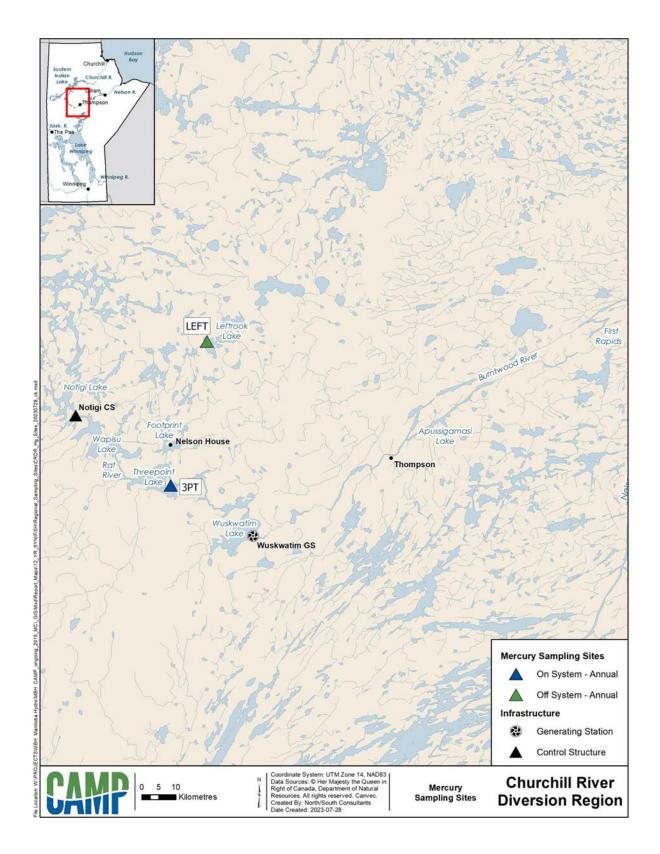


Figure 6.1-1. 2008-2019 Fish mercury sampling sites.



6.2 MERCURY IN FISH

6.2.1 ARITHMETIC MEAN MERCURY CONCENTRATION

6.2.1.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Lake Whitefish

The arithmetic mean mercury concentration of Lake Whitefish over the 10 years of monitoring ranged from a low of 0.053 ppm in 2014 to a high of 0.148 ppm in 2017 (Table 6.2-1). The mercury concentration typically increased with fork length, although there was variation in the mercury concentration of Lake Whitefish of the same length (Figure 6.2-1).

Northern Pike

The arithmetic mean mercury concentration of Northern Pike over the 10 years of monitoring ranged from a low of 0.323 ppm in 2011 and 2015 to a high of 0.507 ppm in 2018 (Table 6.2-1). The mercury concentration typically increased with fork length, although there was variation in the mercury concentration of Northern Pike of the same length (Figure 6.2-2).

Walleye

The arithmetic mean mercury concentration of Walleye over the 10 years of monitoring ranged from a low of 0.361 ppm in 2012 to a high of 0.541 ppm in 2019 (Table 6.2-1). The mercury concentration typically increased with fork length, although there was variation in the mercury concentration of Walleye of the same length (Figure 6.2-3).

Yellow Perch

Over 10 years of monitoring, 1-year-old Yellow Perch were only submitted for mercury analysis in 2018 (Table 6.2-2). In this year the arithmetic mean mercury concentration was 0.033 ppm (Figure 6.2-4).

ROTATIONAL SITES

There are no on-system waterbodies in the Churchill River Diversion Region that are monitored for fish mercury on a rotational basis.



6.2.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

Lake Whitefish

The arithmetic mean mercury concentration of Lake Whitefish over the 10 years of monitoring ranged from a low of 0.029 ppm in 2014 to a high of 0.064 ppm in 2017 (Table 6.2-1). The mercury concentration typically increased with fork length, although there was variation in the mercury concentration of Lake Whitefish of the same length (Figure 6.2-1).

Northern Pike

The arithmetic mean mercury concentration of Northern Pike over the 10 years of monitoring ranged from a low of 0.140 ppm in 2014 to a high of 0.247 ppm in 2010 (Table 6.2-1). The mercury concentration typically increased with fork length, although there was variation in the mercury concentration of Northern Pike of the same length (Figure 6.2-2).

Walleye

The arithmetic mean mercury concentration of Walleye over the 10 years of monitoring ranged from a low of 0.184 ppm in 2011 to a high of 0.243 ppm in 2015 (Table 6.2-1). The mercury concentration typically increased with fork length, although there was variation in the mercury concentration of Walleye of the same length (Figure 6.2-3).

Yellow Perch

The arithmetic mean mercury concentration of 1-year-old Yellow Perch over the 10 years of monitoring ranged from a low of <0.010 ppm in 2014 and 2016 to a high of 0.038 ppm in 2015 (Figure 6.2-4). No Yellow Perch were submitted for mercury analysis in 2011, 2012, 2013, or 2019.

ROTATIONAL SITES

There are no off-system waterbodies in the Churchill River Diversion Region that are monitored for fish mercury on a rotational basis.



Table 6.2-1. 2010-2019 Fork length, age, and mercury concentrations of Lake Whitefish, Northern Pike, and Walleye.

| 6 | M | V | | Fork | Length | (mm) | | | Age | e (year | s) | | | | | | Mercu | ıry (ppm) | |
|---------|-----------|------|----|------|------------------|------------------|-----------------|----|------|---------|-----|----|----|-------|--------|-------|-------|----------------------------|-------------|
| Species | Waterbody | Year | n¹ | Mean | Min ² | Max ² | SE ² | n | Mean | Min | Max | SE | n | Mean | Min | Max | SE | Standard Mean ³ | 95% CL⁴ |
| LKWH | 3PT | 2010 | 2 | 389 | 335 | 442 | 54 | 2 | 8 | 5 | 11 | 3 | 2 | 0.082 | 0.042 | 0.121 | 0.040 | - | |
| | | 2011 | 18 | 402 | 265 | 481 | 13 | 18 | 10 | 3 | 26 | 1 | 18 | 0.087 | 0.036 | 0.169 | 0.011 | 0.055 | 0.043-0.070 |
| | | 2012 | 7 | 399 | 277 | 480 | 25 | 6 | 12 | 5 | 28 | 4 | 7 | 0.107 | 0.041 | 0.251 | 0.029 | 0.062 | 0.034-0.111 |
| | | 2013 | 15 | 392 | 273 | 483 | 14 | 15 | 10 | 4 | 24 | 1 | 15 | 0.092 | 0.030 | 0.238 | 0.014 | 0.055 | 0.045-0.066 |
| | | 2014 | 10 | 359 | 196 | 442 | 23 | 10 | 6 | 2 | 10 | 1 | 10 | 0.053 | 0.027 | 0.087 | 0.007 | 0.049 | 0.040-0.059 |
| | | 2015 | 13 | 423 | 333 | 493 | 13 | 13 | 9 | 4 | 19 | 1 | 13 | 0.095 | 0.039 | 0.190 | 0.015 | 0.035 | 0.025-0.050 |
| | | 2016 | 6 | 332 | 147 | 455 | 54 | 6 | 9 | 1 | 20 | 3 | 6 | 0.067 | 0.020 | 0.135 | 0.022 | not significant | |
| | | 2017 | 15 | 429 | 200 | 548 | 20 | 15 | 14 | 4 | 24 | 2 | 15 | 0.148 | 0.0331 | 0.269 | 0.020 | 0.081 | 0.059-0.112 |
| | | 2018 | 7 | 360 | 191 | 461 | 33 | 7 | 7 | 3 | 13 | 1 | 7 | 0.074 | 0.023 | 0.131 | 0.015 | 0.065 | 0.049-0.085 |
| | | 2019 | 11 | 365 | 181 | 454 | 28 | 11 | 8 | 2 | 16 | 1 | 11 | 0.077 | 0.032 | 0.123 | 0.011 | 0.069 | 0.057-0.083 |
| | LEFT | 2010 | 36 | 418 | 282 | 520 | 10 | 30 | 14 | 5 | 28 | 1 | 36 | 0.044 | 0.017 | 0.093 | 0.004 | 0.026 | 0.022-0.031 |
| | | 2011 | 36 | 397 | 270 | 485 | 10 | 36 | 10 | 4 | 24 | 1 | 36 | 0.042 | 0.017 | 0.113 | 0.004 | 0.026 | 0.023-0.029 |
| | | 2012 | 36 | 424 | 320 | 530 | 9 | 36 | 12 | 4 | 27 | 1 | 36 | 0.049 | 0.013 | 0.137 | 0.005 | 0.021 | 0.017-0.025 |
| | | 2013 | 36 | 408 | 354 | 485 | 5 | 36 | 9 | 4 | 26 | 1 | 36 | 0.048 | 0.025 | 0.176 | 0.004 | 0.027 | 0.021-0.034 |
| | | 2014 | 36 | 402 | 297 | 491 | 7 | 34 | 8 | 4 | 22 | 1 | 36 | 0.029 | <0.010 | 0.116 | 0.003 | 0.014 | 0.012-0.016 |
| | | 2015 | 35 | 429 | 320 | 495 | 6 | 35 | 9 | 4 | 24 | 1 | 35 | 0.039 | 0.016 | 0.093 | 0.003 | 0.019 | 0.015-0.025 |
| | | 2016 | 36 | 414 | 213 | 492 | 8 | 36 | 10 | 2 | 28 | 1 | 36 | 0.049 | 0.019 | 0.090 | 0.003 | 0.032 | 0.027-0.039 |
| | | 2017 | 36 | 435 | 357 | 501 | 6 | 36 | 12 | 4 | 26 | 1 | 36 | 0.064 | 0.025 | 0.135 | 0.004 | 0.029 | 0.023-0.037 |
| | | 2018 | 35 | 440 | 367 | 535 | 6 | 35 | 10 | 4 | 23 | 1 | 35 | 0.057 | 0.027 | 0.086 | 0.003 | 0.029 | 0.022-0.039 |
| | | 2019 | 36 | 432 | 233 | 486 | 7 | 36 | 13 | 3 | 26 | 1 | 36 | 0.059 | 0.016 | 0.153 | 0.004 | 0.033 | 0.028-0.041 |
| NRPK | 3PT | 2010 | 32 | 484 | 325 | 760 | 22 | 32 | 5 | 2 | 10 | 0 | 32 | 0.502 | 0.129 | 0.917 | 0.039 | 0.591 | 0.527-0.663 |
| | | 2011 | 31 | 423 | 303 | 742 | 19 | 30 | 5 | 3 | 10 | 0 | 31 | 0.323 | 0.096 | 1.07 | 0.036 | 0.485 | 0.396-0.592 |
| | | 2012 | 20 | 444 | 236 | 871 | 31 | 20 | 5 | 2 | 11 | 0 | 20 | 0.344 | 0.083 | 1.00 | 0.045 | 0.460 | 0.373-0.568 |
| | | 2013 | 26 | 413 | 190 | 790 | 28 | 25 | 4 | 1 | 10 | 0 | 26 | 0.370 | 0.093 | 1.28 | 0.065 | 0.516 | 0.407-0.653 |
| | | 2014 | 33 | 431 | 228 | 720 | 23 | 33 | 4 | 1 | 8 | 0 | 33 | 0.338 | 0.057 | 1.25 | 0.048 | 0.485 | 0.416-0.565 |
| | | 2015 | 32 | 445 | 275 | 875 | 26 | 32 | 5 | 2 | 11 | 0 | 32 | 0.323 | 0.110 | 0.793 | 0.034 | 0.422 | 0.373-0.478 |
| | | 2016 | 29 | 440 | 224 | 646 | 22 | 29 | 4 | 1 | 6 | 0 | 29 | 0.350 | 0.069 | 0.683 | 0.031 | 0.484 | 0.419-0.559 |
| | | 2017 | 26 | 496 | 305 | 728 | 26 | 26 | 5 | 2 | 9 | 0 | 26 | 0.407 | 0.100 | 0.879 | 0.048 | 0.456 | 0.402-0.518 |
| | | 2018 | 21 | 513 | 296 | 640 | 20 | 21 | 5 | 2 | 7 | 0 | 21 | 0.507 | 0.138 | 1.02 | 0.056 | 0.545 | 0.469-0.633 |
| | | 2019 | 13 | 467 | 290 | 594 | 25 | 12 | 4 | 2 | 8 | 1 | 13 | 0.434 | 0.134 | 0.955 | 0.079 | 0.579 | 0.410-0.817 |
| | LEFT | 2010 | 36 | 470 | 345 | 548 | 7 | 36 | 5 | 3 | 7 | 0 | 36 | 0.247 | 0.082 | 0.432 | 0.017 | 0.392 | 0.317-0.484 |
| | | 2011 | 36 | 494 | 355 | 871 | 17 | 35 | 6 | 2 | 15 | 0 | 36 | 0.205 | 0.057 | 0.778 | 0.023 | 0.241 | 0.213-0.272 |
| | | 2012 | 36 | 449 | 275 | 675 | 13 | 36 | 5 | 2 | 8 | 0 | 36 | 0.171 | 0.039 | 0.453 | 0.017 | 0.245 | 0.195-0.308 |
| | | 2013 | 35 | 445 | 236 | 687 | 15 | 33 | 4 | 1 | 8 | 0 | 35 | 0.166 | 0.039 | 0.492 | 0.018 | 0.230 | 0.193-0.274 |
| | | 2014 | 36 | 457 | 284 | 700 | 17 | 36 | 4 | 2 | 10 | 0 | 36 | 0.140 | 0.042 | 0.327 | 0.015 | 0.187 | 0.156-0.223 |
| | | 2015 | 35 | 481 | 320 | 636 | 12 | 35 | 6 | 3 | 12 | 0 | 35 | 0.220 | 0.074 | 0.454 | 0.017 | 0.283 | 0.246-0.325 |
| | | 2016 | 36 | 494 | 335 | 672 | 12 | 36 | 5 | 2 | 7 | 0 | 36 | 0.221 | 0.074 | 0.417 | 0.014 | 0.247 | 0.212-0.288 |
| | | 2017 | 36 | 457 | 283 | 603 | 11 | 36 | 4 | 2 | 7 | 0 | 36 | 0.186 | 0.066 | 0.467 | 0.016 | 0.283 | 0.242-0.330 |
| | | 2018 | 32 | 463 | 290 | 653 | 14 | 31 | 5 | 2 | 12 | 0 | 32 | 0.171 | 0.0603 | 0.365 | 0.016 | 0.237 | 0.204-0.275 |



Table 6.2-1. continued.

| Caraina | NA/ataulaada | V | | Fork | Length | (mm) | | | Age | e (year | s) | | Mercury (ppm) | | | | | | |
|---------|--------------|------|----|------|------------------|------------------|-----------------|----|------|---------|-----|----|---------------|-------|-------|-------|-------|----------------------------|-------------|
| Species | Waterbody | Year | n¹ | Mean | Min ² | Max ² | SE ³ | n | Mean | Min | Max | SE | n | Mean | Min | Max | SE | Standard Mean ⁴ | 95% CL⁵ |
| NRPK | LEFT | 2019 | 38 | 478 | 337 | 780 | 13 | 38 | 4 | 2 | 9 | 0 | 38 | 0.193 | 0.042 | 0.760 | 0.021 | 0.251 | 0.216-0.291 |
| WALL | 3PT | 2010 | 36 | 358 | 228 | 444 | 9 | 36 | 11 | 4 | 20 | 1 | 36 | 0.510 | 0.170 | 0.984 | 0.036 | 0.577 | 0.495-0.673 |
| | | 2011 | 36 | 361 | 237 | 455 | 10 | 35 | 11 | 5 | 22 | 1 | 36 | 0.386 | 0.134 | 0.698 | 0.028 | 0.431 | 0.370-0.502 |
| | | 2012 | 36 | 287 | 127 | 444 | 17 | 35 | 7 | 1 | 17 | 1 | 36 | 0.361 | 0.057 | 0.910 | 0.038 | 0.567 | 0.471-0.682 |
| | | 2013 | 35 | 360 | 150 | 486 | 13 | 35 | 10 | 1 | 19 | 1 | 35 | 0.468 | 0.068 | 1.24 | 0.045 | 0.520 | 0.455-0.595 |
| | | 2014 | 34 | 371 | 230 | 485 | 10 | 33 | 10 | 3 | 18 | 1 | 34 | 0.477 | 0.220 | 0.950 | 0.034 | 0.502 | 0.437-0.576 |
| | | 2015 | 36 | 352 | 205 | 465 | 12 | 36 | 9 | 3 | 22 | 1 | 36 | 0.466 | 0.142 | 1.10 | 0.043 | 0.558 | 0.492-0.632 |
| | | 2016 | 36 | 331 | 163 | 488 | 15 | 36 | 8 | 1 | 18 | 1 | 36 | 0.378 | 0.059 | 1.03 | 0.046 | 0.476 | 0.394-0.576 |
| | | 2017 | 35 | 354 | 189 | 541 | 13 | 33 | 9 | 3 | 19 | 1 | 35 | 0.404 | 0.118 | 1.11 | 0.044 | 0.456 | 0.393-0.530 |
| | | 2018 | 35 | 332 | 132 | 482 | 16 | 35 | 9 | 1 | 17 | 1 | 35 | 0.410 | 0.051 | 0.971 | 0.043 | 0.530 | 0.454-0.620 |
| | | 2019 | 32 | 355 | 169 | 501 | 14 | 32 | 10 | 2 | 24 | 1 | 32 | 0.541 | 0.092 | 1.10 | 0.049 | 0.618 | 0.539-0.708 |
| | LEFT | 2010 | 36 | 353 | 210 | 393 | 6 | 36 | 11 | 4 | 21 | 1 | 36 | 0.220 | 0.087 | 0.498 | 0.015 | 0.255 | 0.216-0.301 |
| | | 2011 | 36 | 338 | 212 | 422 | 11 | 36 | 9 | 3 | 17 | 1 | 36 | 0.184 | 0.061 | 0.445 | 0.016 | 0.244 | 0.215-0.276 |
| | | 2012 | 38 | 342 | 211 | 455 | 11 | 38 | 10 | 3 | 21 | 1 | 38 | 0.226 | 0.082 | 0.515 | 0.021 | 0.288 | 0.252-0.329 |
| | | 2013 | 36 | 352 | 220 | 480 | 12 | 36 | 9 | 2 | 18 | 1 | 36 | 0.201 | 0.064 | 0.547 | 0.017 | 0.235 | 0.207-0.266 |
| | | 2014 | 35 | 340 | 211 | 477 | 13 | 35 | 9 | 2 | 16 | 1 | 35 | 0.231 | 0.043 | 0.522 | 0.022 | 0.293 | 0.250-0.344 |
| | | 2015 | 36 | 389 | 225 | 487 | 10 | 36 | 10 | 3 | 19 | 1 | 36 | 0.243 | 0.115 | 0.511 | 0.018 | 0.239 | 0.214-0.266 |
| | | 2016 | 35 | 345 | 125 | 475 | 13 | 34 | 9 | 2 | 17 | 1 | 35 | 0.225 | 0.032 | 0.550 | 0.023 | 0.277 | 0.248-0.309 |
| | | 2017 | 35 | 337 | 213 | 490 | 12 | 35 | 7 | 3 | 17 | 1 | 35 | 0.220 | 0.085 | 0.539 | 0.016 | 0.278 | 0.257-0.301 |
| | | 2018 | 35 | 355 | 160 | 523 | 16 | 35 | 9 | 1 | 20 | 1 | 35 | 0.210 | 0.045 | 0.652 | 0.024 | 0.237 | 0.214-0.263 |
| | | 2019 | 30 | 353 | 170 | 581 | 19 | 30 | 8 | 2 | 21 | 1 | 30 | 0.205 | 0.048 | 0.555 | 0.026 | 0.223 | 0.195-0.255 |

Notes:

- 1. n = Sample size.
- 2. Min = minimum; Max = maximum.
- 3. SE = standard error.
- 4. For standard lengths of 350 mm for Lake Whitefish (LKWH), 550 mm for Northern Pike (NRPK), and 400 mm for Walleye (WALL.).
- 5. CL = confidence limits.



Table 6.2-2. 2010-2019 Fork length and mercury concentrations of 1-year-old Yellow Perch.

| 6 | Maria di La | | .1 | For | rk Leng | th (mm) |) | | Mercury | y (ppm) | |
|---------|-------------|------|----|------|------------------|------------------|-----------------|--------|---------|---------|-------|
| Species | Waterbody | Year | n¹ | Mean | Min ² | Max ² | SE ³ | Mean | Min | Max | SE |
| YLPR | 3PT | 2010 | 0 | - | - | - | - | - | - | - | - |
| | | 2011 | 0 | - | - | - | - | - | - | - | - |
| | | 2012 | 0 | - | - | - | 1 | ı | - | - | - |
| | | 2013 | 0 | - | - | - | 1 | ı | - | - | - |
| | | 2014 | 0 | - | - | - | - | • | - | - | - |
| | | 2015 | 0 | - | - | - | - | • | - | - | - |
| | | 2016 | 0 | - | - | - | - | - | - | - | - |
| | | 2017 | 0 | - | - | - | 1 | ı | - | - | - |
| | | 2018 | 3 | 80 | 73 | 93 | 5 | 0.033 | 0.0294 | 0.040 | 0.003 |
| | | 2019 | 0 | - | - | - | - | • | - | - | - |
| | LEFT | 2010 | 3 | 78 | 70 | 85 | 4 | 0.028 | 0.017 | 0.044 | 0.007 |
| | | 2011 | 0 | - | - | - | - | - | - | - | - |
| | | 2012 | 0 | - | - | - | - | - | - | - | - |
| | | 2013 | 0 | - | - | - | - | • | - | - | - |
| | | 2014 | 10 | 79 | 73 | 86 | 1 | <0.010 | <0.010 | 0.015 | - |
| | | 2015 | 2 | 72 | 70 | 74 | 1 | 0.038 | 0.032 | 0.043 | 0.004 |
| | | 2016 | 1 | 77 | - | - | - | <0.010 | - | - | - |
| | | 2017 | 16 | 69 | 62 | 76 | 1 | 0.018 | 0.012 | 0.028 | 0.001 |
| | | 2018 | 12 | 82 | 67 | 96 | 3 | 0.010 | 0.007 | 0.013 | 0.001 |
| | | 2019 | 0 | - | - | - | - | - | - | - | - |

Notes:



^{1.} n = sample size.

^{2.} Min = minimum; Max = maximum.

^{3.} SE = standard error.

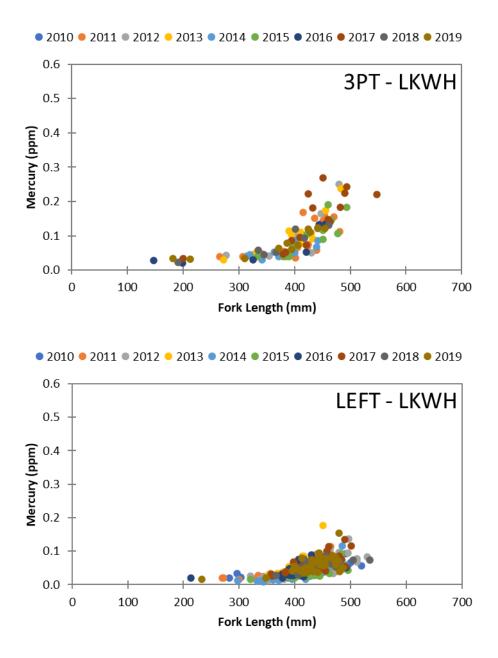


Figure 6.2-1. 2010-2019 Mercury concentration versus fork length of Lake Whitefish.



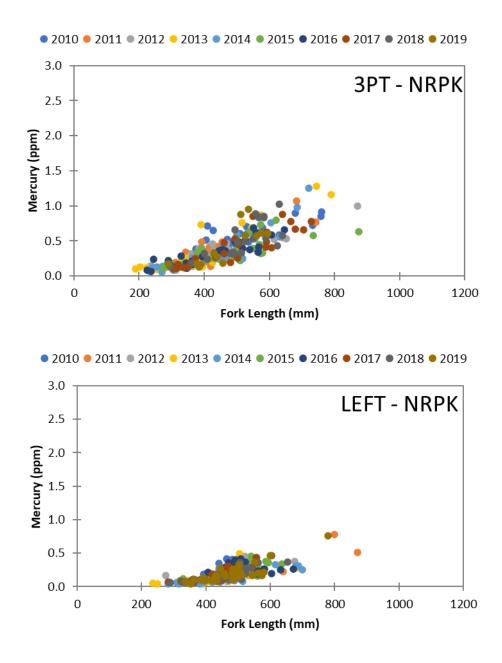


Figure 6.2-2. 2010-2019 Mercury concentration versus fork length of Northern Pike.



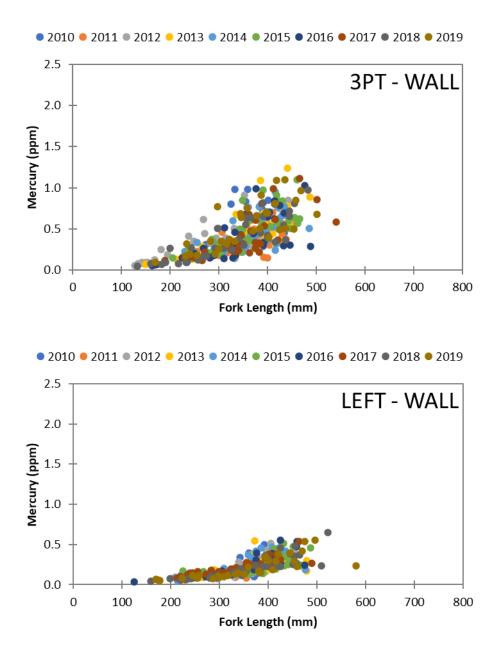
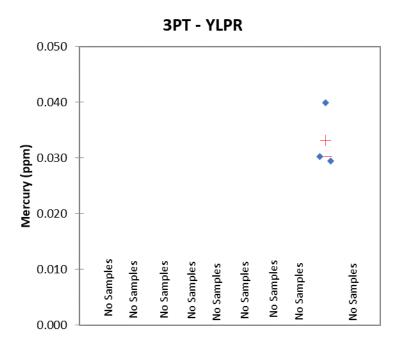


Figure 6.2-3. 2010-2019 Mercury concentration versus fork length of Walleye.







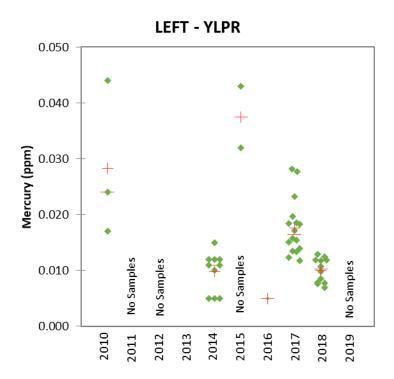


Figure 6.2-4. 2010-2019 Mercury concentrations of 1-year-old Yellow Perch.



6.2.2 LENGTH-STANDARDIZED MEAN CONCENTRATION

6.2.2.1 ON-SYSTEM SITES

ANNUAL SITES

Threepoint Lake

Lake Whitefish

The length-standardized mean mercury concentration of a 350 mm Lake Whitefish over the 10 years of monitoring ranged from 0.035 in 2015 to a high of 0.081 ppm in 2017 (Table 6.2-1). A standard mean could not be calculated for 2010 because only two fish were submitted for mercury analysis or in 2016 because there was not a significant relationship between mercury concentration and fork length.

The overall mean concentration was 0.059 ppm, the median concentration was 0.058 ppm, and the IQR was 0.053–0.066 ppm (Figure 6.2-5). The annual mean mercury concentration fell within the IQR except in 2014 and 2015 when it was below the IQR and in 2017 and 2019 when it was above the IQR.

Northern Pike

The length-standardized mean mercury concentration of a 550 mm Northern Pike over the 10 years of monitoring ranged from a low of 0.422 ppm in 2015 to a high of 0.591 ppm in 2010 (Table 6.2-1).

The overall mean concentration was 0.502 ppm, the median concentration was 0.485 ppm, and the IQR was 0.466–0.538 ppm (Figure 6.2-6). The annual mean mercury concentration fell within the IQR except in 2012, 2015, and 2017 when it was below the IQR and in 2010, 2018, and 2019 when it was above the IQR.

Walleye

The length-standardized mean mercury concentration of a 400 mm Walleye over the 10 years of monitoring ranged from a low of 0.431 ppm in 2011 to a high of 0.618 ppm in 2019 (Table 6.2-1).

The overall mean concentration was 0.524 ppm, the median concentration was 0.525 ppm, and the IQR was 0.483–0.565 ppm (Figure 6.2-7). The annual mean mercury concentration fell within the IQR except in 2011, 2016, and 2017 when it was below the IQR and in 2010, 2012, and 2019 when it was above the IQR.



ROTATIONAL SITES

There are no on-system waterbodies in the Churchill River Diversion Region that are monitored for fish mercury on a rotational basis.

6.2.2.2 OFF-SYSTEM SITES

ANNUAL SITES

Leftrook Lake

Lake 2333

The length-standardized mean mercury concentration of a 350 mm Lake Whitefish over the 10 years of monitoring ranged from a low of 0.014 ppm in 2014 to a high of 0.033 ppm in 2019 (Table 6.2-1).

The overall mean concentration was 0.026 ppm, the median concentration was 0.027 ppm, and the IQR was 0.022–0.029 ppm (Figure 6.2-5). The annual mean mercury concentration fell within the IQR except in 2012, 2014, and 2015 when it was below the IQR and in 2016 and 2019 when it was above the IQR.

Northern Pike

The length-standardized mean mercury concentration of a 550 mm Northern Pike over the 10 years of monitoring ranged from a low of 0.187 ppm in 2014 to a high of 0.392 ppm in 2010 (Table 6.2-1).

The overall mean concentration was 0.260 ppm, the median concentration was 0.246 ppm, and the IQR was 0.238–0.275 ppm (Figure 6.2-6). The annual mean mercury concentration fell within the IQR except in 2013, 2014, 2018 when it was below the IQR and in 2010, 2015, and 2017 when it was above the IQR.

Walleye

The length-standardized mean mercury concentration of a 400 mm Walleye over the 10 years of monitoring ranged from a low of 0.223 ppm in 2019 to a high of 0.293 ppm in 2014 (Table 6.2-1).

The overall mean concentration was 0.257 ppm, the median concentration was 0.249 ppm, and the IQR was 0.238–0.278 ppm (Figure 6.2-7). The annual mean mercury concentration fell within the IQR except in 2013, 2018, and 2019 when it was below the IQR and in 2012 and 2014 when it was above the IQR.



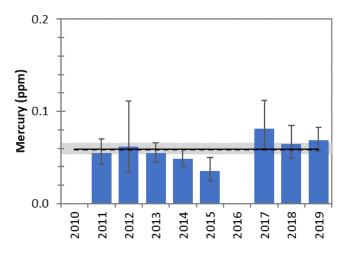
ROTATIONAL SITES

There are no off-system waterbodies in the Churchill River Diversion Region that are monitored for fish mercury on a rotational basis.





3PT - 350 mm LKWH



LEFT - 350 mm LKWH

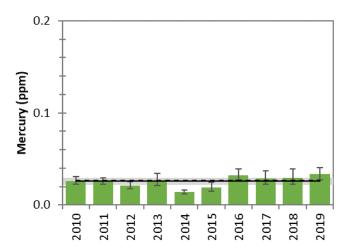
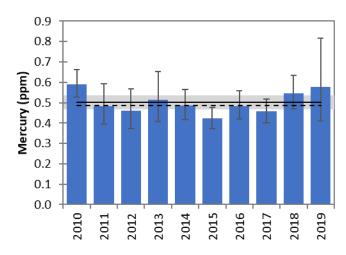


Figure 6.2-5. 2010-2019 Length-standardized mean mercury concentrations (±95% confidence intervals) of Lake Whitefish.





3PT - 550 mm NRPK



LEFT - 550 mm NRPK

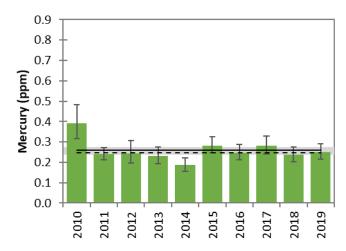
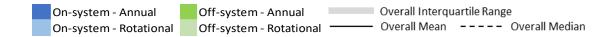
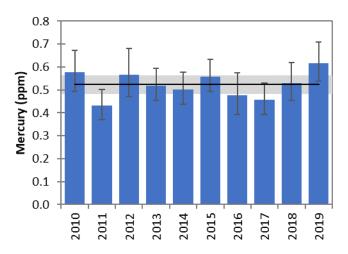


Figure 6.2-6. 2010-2019 Length-standardized mean mercury concentrations (±95% confidence intervals) of Northern Pike.





3PT - 400 mm WALL



LEFT - 400 mm WALL

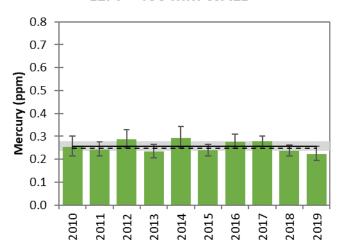


Figure 6.2-7. 2010-2019 Length-standardized mean mercury concentrations (±95% confidence intervals) of Walleye.



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