



Coordinated Aquatic Monitoring Program

CAMP Twelve Year Data Report (2008-2019)

Technical Document 2:

Winnipeg River Region

Prepared by

Manitoba Hydro

And

North/South Consultants Inc.

2024

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CAMP TWELVE YEAR DATA REPORT (2008-2019)

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EXECUTIVE SUMMARY

This report presents the results of monitoring conducted under the Coordinated Aquatic Monitoring Program (CAMP) for years 1 through 12 (i.e., 2008/2009 through 2019/2020) in the Winnipeg River Region. The Winnipeg River Region is composed of the Winnipeg River from the Manitoba-Ontario border downstream to the mouth of the river at Traverse Bay on Lake Winnipeg. Waterbodies and sites monitored in this region over this period included three on-system and two off-system waterbodies or river reaches as follows:

- the Pointe du Bois Forebay;
- Lac du Bonnet;
- the Pine Falls Forebay;
- Manigotagan Lake (off-system); and
- Eaglenest Lake (off-system).

Monitoring on-system waterbodies addresses the primary objective of CAMP – to monitor aquatic ecosystem health along Manitoba Hydro’s hydraulic operating system. The off-system waterbodies were included in CAMP to provide regional information collected in a manner consistent with monitoring of on-system waterbodies that will assist in interpreting any observed environmental changes over time. Such comparisons are intended to help distinguish between hydroelectric-related effects and other external factors (e.g., climate change) in each CAMP region.

Monitoring was conducted annually at some waterbodies and river reaches and on a three-year rotation at other sites. Components monitored under CAMP in the Winnipeg River Region presented in this report include the physical environment (water regime and sedimentation), water quality, benthic invertebrates, fish community, and mercury in fish. Climatological data for the region are also included to provide supporting information to assist with interpretation of CAMP monitoring results.

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ABBREVIATIONS, ACRONYMS, AND UNITS

| | |
|----------------|--|
| ANN | Annual |
| CAMP | Coordinated Aquatic Monitoring Program |
| CCME | Canadian Council of Ministers of the Environment |
| CL(s) | Confidence limit(s) |
| cms | Cubic metres per second |
| CONT | Continuous |
| COSEWIC | Committee on the Status of Endangered Wildlife in Canada |
| CPUE | Catch-per-unit-effort |
| DELTs | Deformities, Erosion, Lesions, and Tumours |
| DL(s) | Detection limit(s) |
| DO | Dissolved oxygen |
| ECCC | Environment and Climate Change Canada |
| EPT | Ephemeroptera (mayflies), Plecoptera (stoneflies), and Trichoptera (caddisflies) |
| FA | Fall |
| FLA | Fork length-at-age |
| FNU | Formazin nephelometric unit |
| GN | Standard gang index gill net |
| GS(s) | Generating station(s) |
| h | hour |
| IC | Ice-cover season |
| IQR | Interquartile range |
| KF | Fulton's Condition Factor |
| LWCB | Lake of the Woods Control Board |
| m | Metre |
| m ² | Metre squared |
| Max | Maximum |
| µg/L | Micrograms per litre |
| mg/L | Milligrams per litre |
| Min | Minimum |
| mm | Millimetre |
| MWQSOGs | Manitoba Water Quality Standards, Objectives, and Guidelines |
| MWS | Manitoba Water Stewardship |
| n | Sample size or number of samples |
| n _F | Number of fish |
| n _S | Number of sites |

| | |
|------------------|---|
| ND | No data |
| no. | Number |
| NS | Nearshore |
| n _{spp} | Number of species caught in standard and small mesh gill nets |
| NTU | Nephelometric turbidity units |
| O+C | Oligochaeta and Chironomidae |
| OECD | Organization for Economic Cooperation and Development |
| OS | Offshore |
| OW | Open-water season |
| PAL | Protection of aquatic life |
| ppm | Parts per million |
| ROT | Rotational |
| RSA | Relative species abundance |
| RYCS | Relative year-class strength |
| SD | Standard deviation |
| SE | Standard error |
| SN | Small mesh index gillnet gang |
| SP | Spring |
| SU | Summer |
| T/day | Tonnes per day |
| TN | Total nitrogen |
| TOC | Total organic carbon |
| TP | Total phosphorus |
| TSS | Total suspended solids |
| unid. | unidentified |
| WI | Winter |
| Wr | Relative weight |
| °C | Degrees Celsius |

WATERBODY ABBREVIATIONS

| Abbreviation | Waterbody |
|--------------|------------------------|
| EAGLE | Eaglenest Lake |
| LDB | Lac du Bonnet |
| MANIG | Manigotagan Lake |
| PDB | Pointe du Bois Forebay |
| PFF | Pine Falls Forebay |

FISH SPECIES LIST

| Abbreviation | Common Species Name | Species Name |
|--------------|---------------------|---------------------------------|
| BLBL | Black Bullhead | <i>Ameiurus melas</i> |
| BLCR | Black Crappie | <i>Pomoxis nigromaculatus</i> |
| BRBL | Brown Bullhead | <i>Ameiurus nebulosus</i> |
| BURB | Burbot | <i>Lota lota</i> |
| CHCT | Channel Catfish | <i>Ictalurus punctatus</i> |
| CISC | Cisco | <i>Coregonus artedi</i> |
| EMSH | Emerald Shiner | <i>Notropis atherinoides</i> |
| GOLD | Goldeye | <i>Hiodon alosoides</i> |
| LGPR | Logperch | <i>Percina caprodes</i> |
| LKST | Lake Sturgeon | <i>Acipenser fulvescens</i> |
| LKWH | Lake Whitefish | <i>Coregonus clupeaformis</i> |
| LNSC | Longnose Sucker | <i>Catostomus catostomus</i> |
| MOON | Mooneye | <i>Hiodon tergisus</i> |
| MTSC | Mottled Sculpin | <i>Cottus bairdii</i> |
| NRPK | Northern Pike | <i>Esox lucius</i> |
| RCBS | Rock Bass | <i>Ambloplites rupestris</i> |
| RNSM | Rainbow Smelt | <i>Osmerus mordax</i> |
| SAUG | Sauger | <i>Sander canadensis</i> |
| SHRD | Shorthead Redhorse | <i>Moxostoma macrolepidotum</i> |
| SLLM | Silver Lamprey | <i>Ichthyomyzon unicuspis</i> |
| SLRD | Silver Redhorse | <i>Moxostoma anisurum</i> |
| SLSC | Slimy Sculpin | <i>Cottus cognatus</i> |
| SMBS | Smallmouth Bass | <i>Micropterus dolomieu</i> |
| SPSH | Spottail Shiner | <i>Notropis hudsonius</i> |
| TRPR | Trout-perch | <i>Percopsis omiscomaycus</i> |
| WALL | Walleye | <i>Sander vitreus</i> |
| WHSC | White Sucker | <i>Catostomus commersonii</i> |
| YLPR | Yellow Perch | <i>Perca flavescens</i> |

1.0 INTRODUCTION

This report presents the results of monitoring conducted under the Coordinated Aquatic Monitoring Program (CAMP) for years 1 through 12 (i.e., 2008/2009 through 2019/2020) in the Winnipeg River Region. The Winnipeg River Region is composed of the Winnipeg River from the Manitoba-Ontario border downstream to the mouth of the river at Traverse Bay on Lake Winnipeg. Waterbodies and sites monitored in this region over this period included three on-system and two off-system waterbodies or river reaches as follows:

- the Pointe du Bois Forebay;
- Lac du Bonnet;
- the Pine Falls Forebay;
- Manigotagan Lake (off-system); and
- Eaglenest Lake (off-system).

Monitoring on-system waterbodies addresses the primary objective of CAMP – to monitor aquatic ecosystem health along Manitoba Hydro’s hydraulic operating system. The off-system waterbodies were included in CAMP to provide regional information collected in a manner consistent with monitoring of on-system waterbodies that will assist in interpreting any observed environmental changes over time. Such comparisons are intended to help distinguish between hydroelectric-related effects and other external factors (e.g., climate change) in each CAMP region.

A summary of monitoring conducted by waterbody or river reach presented in this data report is provided in Table 1-1 and monitoring areas are shown in Figure 1-1. As noted in Table 1-1, monitoring was conducted annually at some waterbodies and river reaches and on a three-year rotation at other sites. Components monitored under CAMP in the Winnipeg River Region presented in this report include the physical environment (water regime and sedimentation), water quality, benthic invertebrates, fish community, and mercury in fish. Climatological data for the region are also included to provide supporting information to assist with interpretation of CAMP monitoring results.

Table 1-1. Winnipeg River Region CAMP monitoring summary.

| Waterbody/ Area | Abbreviation | On/Off-System | | Component | | | | | |
|------------------------|--------------|---------------|------------|--------------|---------------|---------------|-----------------------|----------------|--------------|
| | | On-System | Off-System | Water Regime | Sedimentation | Water Quality | Benthic Invertebrates | Fish Community | Fish Mercury |
| Pointe du Bois Forebay | PDB | ● | | CONT | | ANN | ANN | ANN | ROT |
| Slave Falls GS | SFGS | ● | | CONT | | | | | |
| Lac du Bonnet | LDB | ● | | CONT | | ANN | ANN | ANN | |
| Pine Falls Forebay | PFF | ● | | CONT | | ROT | ROT | ROT | ROT |
| Manigotagan Lake | MANIG | | ● | CONT | | ANN | ANN | ANN | |
| Eaglenest Lake | EAGLE | | ● | | | ROT | ROT | ROT | |

Notes:

1. CONT = site monitored continuously; ANN = site sampled each year; ROT = site sampled every 3 years.

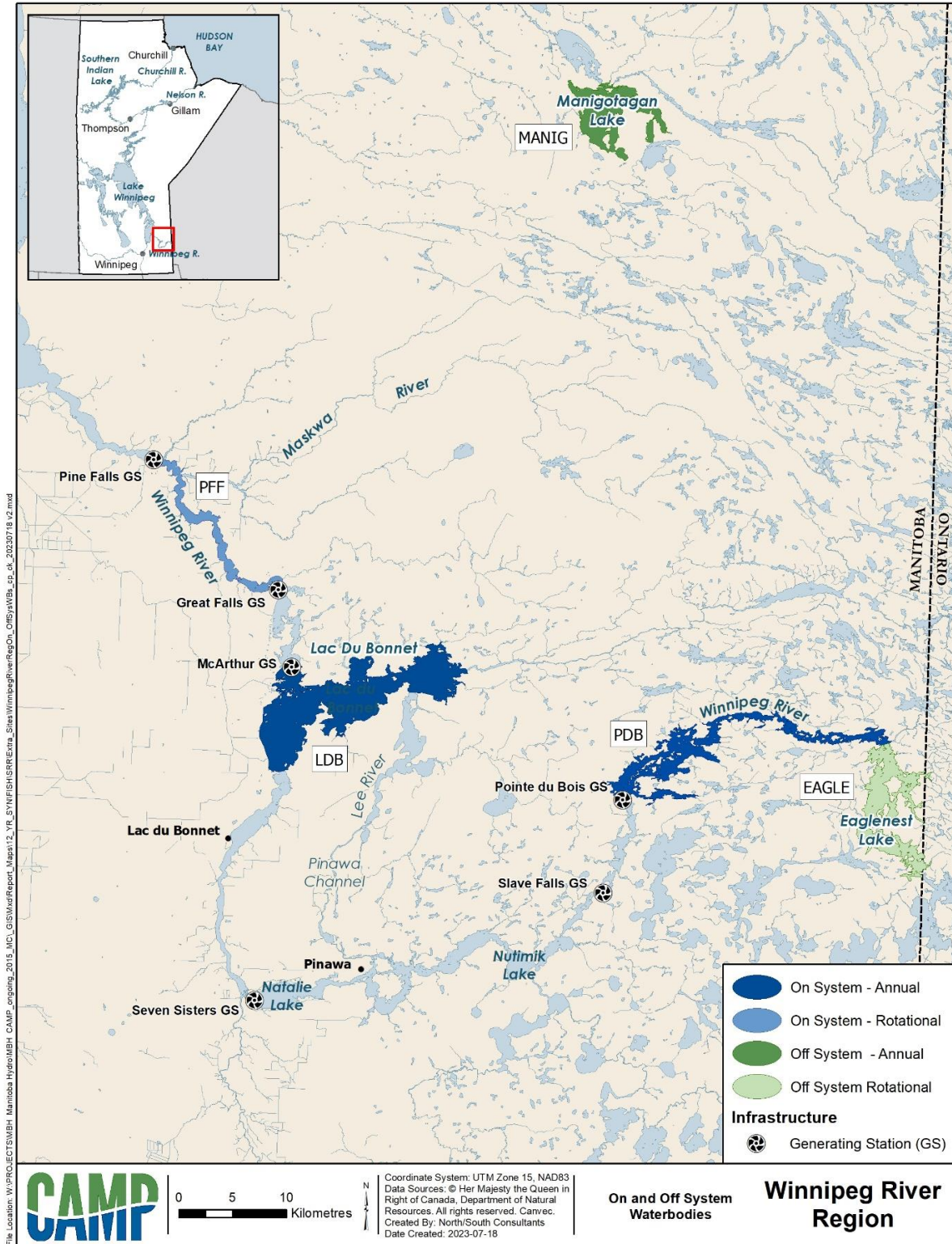


Figure 1-1. On-system and off-system waterbodies and river reaches sampled under CAMP in the Winnipeg River Region: 2008-2019.



Photograph 1. The Pointe du Bois Forebay.



Photograph 2. Lac du Bonnet.



Photograph 3. The Pine Falls Forebay.



Photograph 4. Manigotagan Lake.



Photograph 5. Eaglenest Lake.

2.0 PHYSICAL ENVIRONMENT

2.1 INTRODUCTION

The following presents the results of the physical environment monitoring conducted from 2008 to 2019 in the Winnipeg River Region. Five waterbodies were monitored in the Winnipeg River Region: Three on-system sites (Pointe du Bois Generating Station (GS) forebay, Lac du Bonnet (McArthur Falls GS forebay), and the Pine Falls GS forebay) and two off system sites (Manigotagan Lake; Eaglenest Lake). Although Slave Falls GS is not part of CAMP monitoring, Winnipeg River flow is reported from Slave Falls GS because it has the longest and most reliable flow record and flows do not change significantly along the Winnipeg River within Manitoba. A continuous water quality monitoring station was installed at the Pointe du Bois GS in 2020, after the monitoring period for this report and this data will be presented in the next reporting cycle. Though CAMP does not directly monitor climate, data from Environment and Climate Change Canada (ECCC) is included in reporting to contextualize the data collected under each CAMP component. For the Winnipeg River Region, meteorological conditions from ECCC's Pinawa station are reported.

Three indicators (climate; water regime; and sedimentation) were selected for detailed reporting (Table 2.1-1). Metrics for these indicators include temperature, precipitation, water flow, level and variability, water temperature, continuous turbidity, and suspended sediment load (Table 2.1-1). A detailed description of these indicators is provided in CAMP (2024).

A detailed description of the program design and sampling methods is provided in Technical Document 1, Sections 2.1 and 2.2.

Table 2.1-1. Physical Environment indicators and metrics.

| Indicator | Metric | Units |
|----------------------|-------------------------------|---|
| Climate ¹ | • Temperature | Degrees Celsius (°C) |
| | • Precipitation | Millimetres (mm) |
| Water Regime | • Flow | Cubic meters per second (cms) |
| | • Water Level and Variability | Metres (m) |
| | • Water Temperature | Duration of temperature in 5-degree Celsius increments (#days/5 °C) |
| Sedimentation | • Continuous Turbidity | Formazin nephelometric unit (FNU) |
| | • Suspended Sediment Load | Tonnes/day (T/day) |

Notes:

1. Climate is not monitored through CAMP; data are included for reporting purposes only.

2.2 CLIMATE

In this section, mean monthly air temperatures and total monthly precipitation for each year in the monitoring program (2008-2020) are compared to ECCC climate normals to provide a summary of the Pinawa station meteorological conditions. Climate normals are used to summarize the average climatic conditions of a particular location. As recommended by the World Meteorological Organization, ECCC calculates climate normals using a 30-year period (e.g., 1981-2010). The Pinawa station is used herein to illustrate climate conditions in the Winnipeg River Region.

Historical monthly average air temperature and total monthly precipitation during the monitoring period were calculated based on daily data from ECCC at multiple stations. It is important to note that the use of multiple stations could introduce inhomogeneities in observations between various stations and the station used for climate normals (Climate ID: 5032162). For instances where datasets were missing more than 10% of the daily data in a month, monthly values were gap-filled using ERA5-Land data (Muñoz Sabater 2019). Seasonal and annual maps derived from ERA5-Land data are also provided in Appendices 2-1 and 2-2 to complement the station data and offer a broader spatial representation of temperature and precipitation conditions across Manitoba. Although the ERA5-Land data correlated well with the actual observed ECCC data for the Pinawa station, it should be noted that ERA5-Land is a gridded reanalysis product, meaning the dataset combines modelled data with observations, and therefore may not provide an entirely accurate representation of observed climate.

2.2.1 TEMPERATURE

Figure 2.2-1 illustrates the mean monthly air temperatures (in °C) for each year during the monitoring period compared to the 1981-2010 normal mean temperature. As shown, air temperatures at this location follow a distinct seasonal pattern; warmer in the summer (warmest in July) and cooler in the winter (coldest in January). In general, recorded air temperatures for the monitoring period were consistent with the climate normal pattern. Some deviations can be seen, for example, 2014 recorded considerably cooler temperatures from January to April.

Table 2.2-1 summarizes the mean monthly air temperature data and categorizes each month in the monitoring period as “below normal”, “near normal” or “above normal” conditions. It should be noted that the “near normal” category was subjectively defined as +/- 1°C of the ECCC climate normal. Months “below normal” are highlighted in blue, “near normal” are highlighted in grey, and “above normal” are highlighted in orange. Over the monitoring period, October and November generally experienced warmer than normal conditions (≥ 7 out of 13 months above normal), while April generally experienced below normal conditions (≥ 7 out of 13 months below normal). On an annual basis, no distinct patterns in the data were identified as there was nearly equal number of years below, near, and above normal conditions; 2012 had the warmest annual average temperature at 4.0°C, while 2014 had the coolest annual average temperature at 0.9°C. The maximum and minimum monthly average air temperatures during the monitoring period were 21.3°C (July 2012) and -20.6°C (January 2014), respectively.

Table 2.2-1. Pinawa mean monthly and annual air temperature (in °C) compared to 1981-2010 normal.

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec | Annual |
|------------------|-------|-------|-------|------|------|------|------|------|------|-----|------|-------|--------|
| 2008 | -16.2 | -17.8 | -9.0 | 2.1 | 8.0 | 15.1 | 17.8 | 18.4 | 12.5 | 6.6 | -3.4 | -20.0 | 1.2 |
| 2009 | -19.4 | -14.2 | -7.6 | 2.8 | 7.2 | 14.4 | 15.7 | 16.7 | 17.2 | 3.9 | 1.4 | -14.5 | 2.0 |
| 2010 | -13.7 | -13.2 | 0.2 | 7.5 | 11.1 | 15.4 | 19.5 | 18.5 | 10.9 | 7.7 | -2.8 | -14.0 | 3.9 |
| 2011 | -19.5 | -14.4 | -8.8 | 3.7 | 10.4 | 16.2 | 20.3 | 19.2 | 13.4 | 8.3 | -1.6 | -8.6 | 3.2 |
| 2012 | -11.0 | -9.8 | 1.2 | 4.9 | 11.5 | 17.4 | 21.3 | 17.9 | 11.8 | 3.4 | -5.7 | -14.6 | 4.0 |
| 2013 | -17.8 | -14.1 | -9.8 | -1.3 | 10.5 | 16.4 | 18.4 | 18.6 | 14.9 | 5.4 | -5.3 | -20.5 | 1.3 |
| 2014 | -20.6 | -19.9 | -13.4 | -0.2 | 10.6 | 16.3 | 17.9 | 18.4 | 12.8 | 6.3 | -8.3 | -9.5 | 0.9 |
| 2015 | -14.7 | -20.1 | -3.8 | 4.2 | 10.1 | 17.4 | 19.7 | 17.7 | 15.5 | 7.2 | -0.3 | -7.1 | 3.8 |
| 2016 | -15.0 | -12.5 | -2.6 | 1.2 | 12.6 | 16.2 | 19.0 | 18.3 | 13.9 | 6.7 | 2.9 | -13.3 | 3.9 |
| 2017 | -12.6 | -10.1 | -6.1 | 4.3 | 10.8 | 16.2 | 18.7 | 16.7 | 13.6 | 6.3 | -6.4 | -16.2 | 2.9 |
| 2018 | -15.9 | -17.6 | -4.5 | -0.4 | 13.0 | 18.2 | 19.8 | 18.3 | 10.5 | 2.5 | -8.0 | -10.0 | 2.2 |
| 2019 | -18.6 | -18.9 | -7.9 | 3.1 | 8.6 | 16.1 | 18.9 | 17.4 | 12.5 | 3.5 | -5.4 | -12.8 | 1.4 |
| 2020 | -12.8 | -13.2 | -5.3 | 0.6 | 10.0 | 17.7 | 20.2 | 18.7 | 11.2 | 1.8 | -2.6 | -9.2 | 3.1 |
| 1981-2010 Normal | -16.6 | -13.2 | -5.7 | 3.9 | 11.2 | 16.4 | 19.3 | 18.2 | 12.3 | 5.1 | -4.5 | -13.1 | 2.8 |

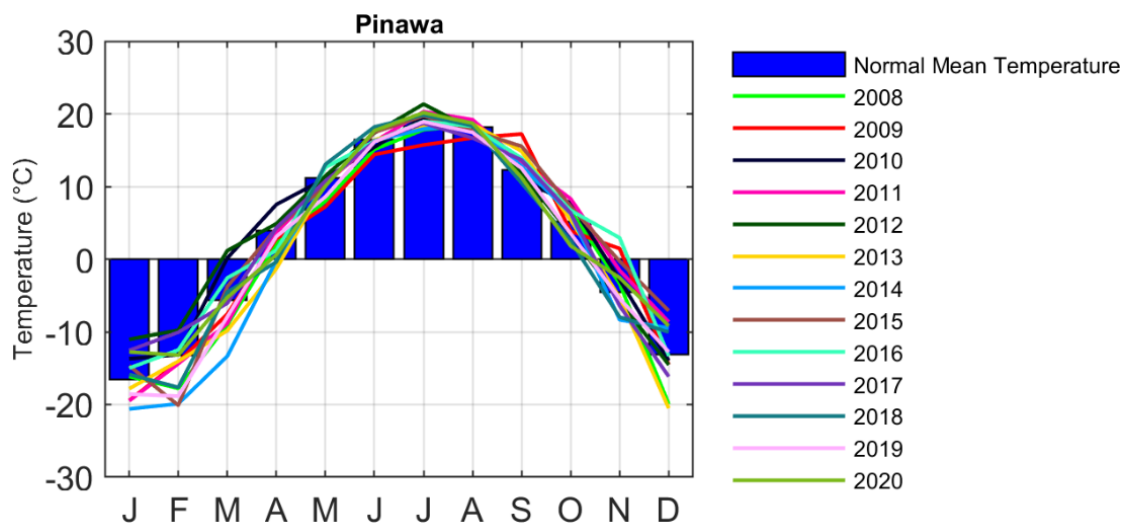
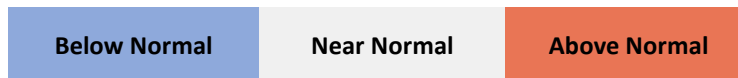


Figure 2.2-1. Pinawa mean monthly air temperature (in °C) compared to 1981-2010 normal.

2.2.2 PRECIPITATION

Figure 2.2-2 illustrates the monthly total precipitation (in mm) for each year during the monitoring period compared to the 1981-2010 normal total precipitation. Total precipitation refers to the water equivalent of all types of precipitation. The total precipitation at Pinawa follows a noticeable seasonal pattern, where generally the highest amounts of precipitation fall during the summer months (June and July) and the lowest amounts fall during the winter months (January and February). Overall, recorded precipitation for the monitoring period followed similar patterns to the climate normal pattern. Some deviations can be seen, such as 2010, where the recorded total precipitation for May and August was much higher than normal and for 2011 (July), which recorded total precipitation well below the normal condition.

Table 2.2-2 summarizes the total monthly precipitation data and categorizes each month in the monitoring period as "below normal", "near normal" or "above normal" conditions. It should be noted that the "near normal" was subjectively defined as +/- 10% of the ECCC climate normal. Months "below normal" are highlighted in light brown, "near normal" are highlighted in grey and "above normal" are highlighted in green. Over the monitoring period, February, April, July, and September generally experienced more than normal precipitation (≥ 7 out of 13 months above normal). On an annual basis, no distinct patterns in the data were identified as there was nearly equal number of years below, near, and above normal conditions; 2010 had the highest annual total precipitation (870.8 mm), while 2011 had the lowest annual total precipitation (389.0 mm). The maximum and minimum monthly total precipitation recorded during the monitoring period were 184.3 mm (September 2019) and 0.5 mm (April 2018), respectively.

Table 2.2-2. Pinawa total monthly and annual precipitation (in mm) compared to 1981-2010 normal.

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec | Annual |
|------------------|------|------|------|------|-------|-------|-------|-------|-------|------|------|------|--------|
| 2008 | 2.6 | 10.8 | 26.5 | 15.9 | 39.8 | 100.2 | 139.6 | 47.0 | 83.0 | 63.0 | 51.2 | 24.4 | 604.0 |
| 2009 | 14.1 | 31.2 | 46.7 | 18.5 | 74.4 | 143.1 | 133.8 | 100.5 | 111.3 | 26.3 | 14.2 | 18.8 | 732.9 |
| 2010 | 29.3 | 13.7 | 9.8 | 34.7 | 162.4 | 129.9 | 115.6 | 175.2 | 78.1 | 69.3 | 43.4 | 9.4 | 870.8 |
| 2011 | 30.5 | 7.6 | 24.6 | 44.0 | 51.1 | 64.6 | 15.7 | 23.9 | 68.7 | 21.8 | 25.5 | 11.0 | 389.0 |
| 2012 | 22.8 | 21.8 | 40.8 | 27.3 | 119.1 | 89.8 | 45.4 | 85.7 | 21.6 | 90.2 | 40.3 | 12.5 | 617.3 |
| 2013 | 32.3 | 11.1 | 20.1 | 41.9 | 60.2 | 55.5 | 71.6 | 35.9 | 59.2 | 24.6 | 41.4 | 32.4 | 486.2 |
| 2014 | 41.8 | 26.0 | 17.4 | 41.0 | 71.8 | 143.1 | 30.5 | 59.2 | 32.2 | 47.7 | 32.1 | 14.7 | 557.5 |
| 2015 | 22.2 | 37.2 | 25.3 | 24.4 | 85.2 | 80.2 | 121.5 | 100.8 | 34.6 | 50.9 | 53.1 | 48.6 | 683.9 |
| 2016 | 19.0 | 22.9 | 64.2 | 55.0 | 61.1 | 113.5 | 98.1 | 115.1 | 71.4 | 50.7 | 37.3 | 80.3 | 788.6 |
| 2017 | 47.8 | 22.5 | 29.0 | 42.0 | 28.1 | 64.1 | 112.6 | 23.4 | 84.7 | 49.1 | 31.1 | 27.9 | 562.3 |
| 2018 | 21.0 | 7.3 | 5.1 | 0.5 | 64.9 | 124.4 | 30.9 | 35.5 | 74.5 | 52.8 | 16.8 | 33.7 | 467.4 |
| 2019 | 68.0 | 33.6 | 14.5 | 25.0 | 45.8 | 57.4 | 139.5 | 48.6 | 184.3 | 97.4 | 13.8 | 11.8 | 739.8 |
| 2020 | 19.4 | 6.6 | 8.4 | 32.3 | 26.2 | 72.3 | 90.0 | 111.3 | 20.8 | 22.4 | 7.9 | 23.8 | 441.4 |
| 1981-2010 Normal | 21.7 | 16.7 | 25.8 | 29.1 | 66.6 | 98.8 | 89.1 | 65.3 | 61.9 | 48.2 | 29.5 | 25.6 | 578.3 |

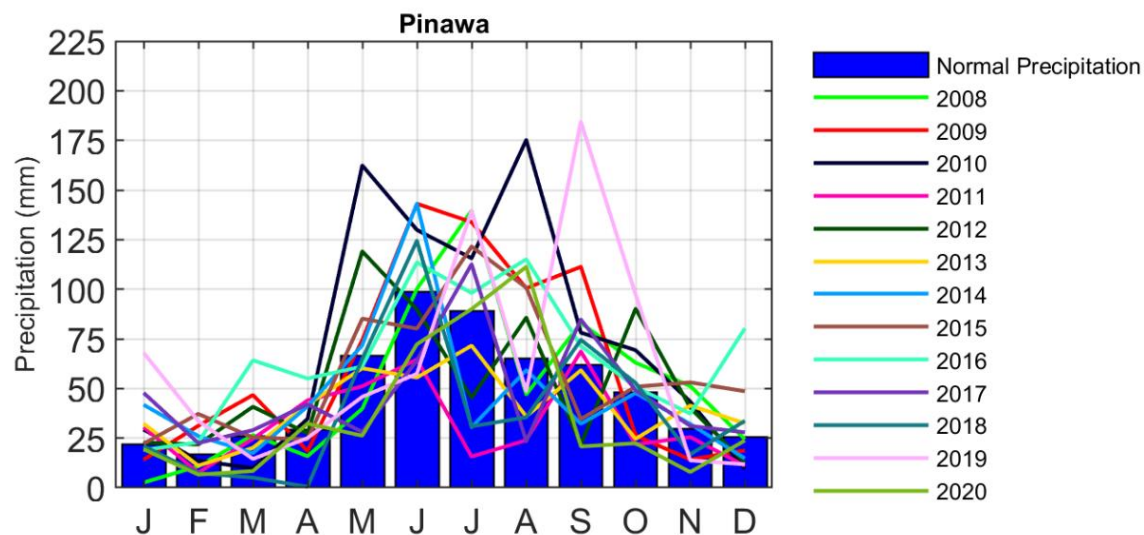


Figure 2.2-2. Pinawa total monthly precipitation (in mm) compared to 1981-2010 normal.

2.3 WATER REGIME

Winnipeg River flow is primarily determined by precipitation within the river's drainage basin, however a major influence on flow is releases from upstream storage reservoirs in Ontario, which are regulated by the Lake of the Woods Control Board (LWCB). The LWCB regulates the water levels of Lake of the Woods and Lac Seul, and the flows in the Winnipeg and English Rivers downstream of these lakes for the long-term benefit of all users and interests. More information is available online at [Lake of the Woods Control Board \(lwcb.ca\)](http://lwcb.ca). Outflows from Lake of the Woods on the Winnipeg River and Lac Seul on the English River combine at Boundary Falls just east of the Manitoba-Ontario border. Within Manitoba, six Manitoba Hydro generating stations along the Winnipeg River (Figure 2.3-1) create stable water levels in the forebays upstream of the generating stations under most flow conditions. Water levels downstream from generating stations continue to vary with flow in the Winnipeg River.

On-System Sites

On-system CAMP monitoring along the Winnipeg River occurred on the Pointe du Bois GS forebay, Lac du Bonnet (McArthur Falls GS forebay), and the Pine Falls GS forebay (Figure 2.3-1). Although Slave Falls GS is not part of CAMP monitoring, Winnipeg River flow is reported from Slave Falls GS because it has the longest and most reliable flow record and flows do not change significantly along the Winnipeg River within Manitoba. CAMP water level measurements occur on the upstream side of the GS for both Pointe du Bois and Pine Falls forebays and at Water Survey of Canada gauge 05PF062 at Lac du Bonnet. Continuous water temperature is measured at the Pointe du Bois GS at the continuous water quality monitoring site because it is the most upstream CAMP location. Monitoring at this station was initiated in 2020, therefore no data is available for the time period presented in this report.

Off-System Sites

CAMP monitors Manigotagan and Eaglenest lakes as the off-system waterbodies for this region. Manigotagan River is located north of the Winnipeg River and is a smaller river that flows into the east side of Lake Winnipeg. Water levels are not measured on Manigotagan Lake, but the level varies up and down with the measured flow in the Manigotagan River. Manigotagan River flow is reported at Water Survey of Canada Gauge 05RA001 located about 50 km downstream from Manigotagan Lake (Figure 2.3-1). The Manigotagan River flow record at Gauge 05RA001 was discontinued by Water Survey of Canada in 1997 and was re-established as part of CAMP in 2010. As a result, Manigotagan Flow data is not available for the first two years of CAMP (2008 and

2009) and the historical 30-year reference period is selected as 1965-1996 rather than 1981 to 2010 as it is at other locations.

Eaglenest Lake is located on the Winnipeg River upstream from Pointe du Bois GS near the Manitoba Ontario border. Although it is along the Winnipeg River, water levels on Eaglenest Lake are not affected by Manitoba Hydro's operations and Eaglenest Lake is therefore considered an off-system site despite being affected by upstream regulation by LWCB. Water levels are not measured on Eaglenest Lake but varies up and down with the measured flow in the Winnipeg River (reported at Slave Falls as described above). Continuous water temperature is not monitored at the off-system waterbodies.

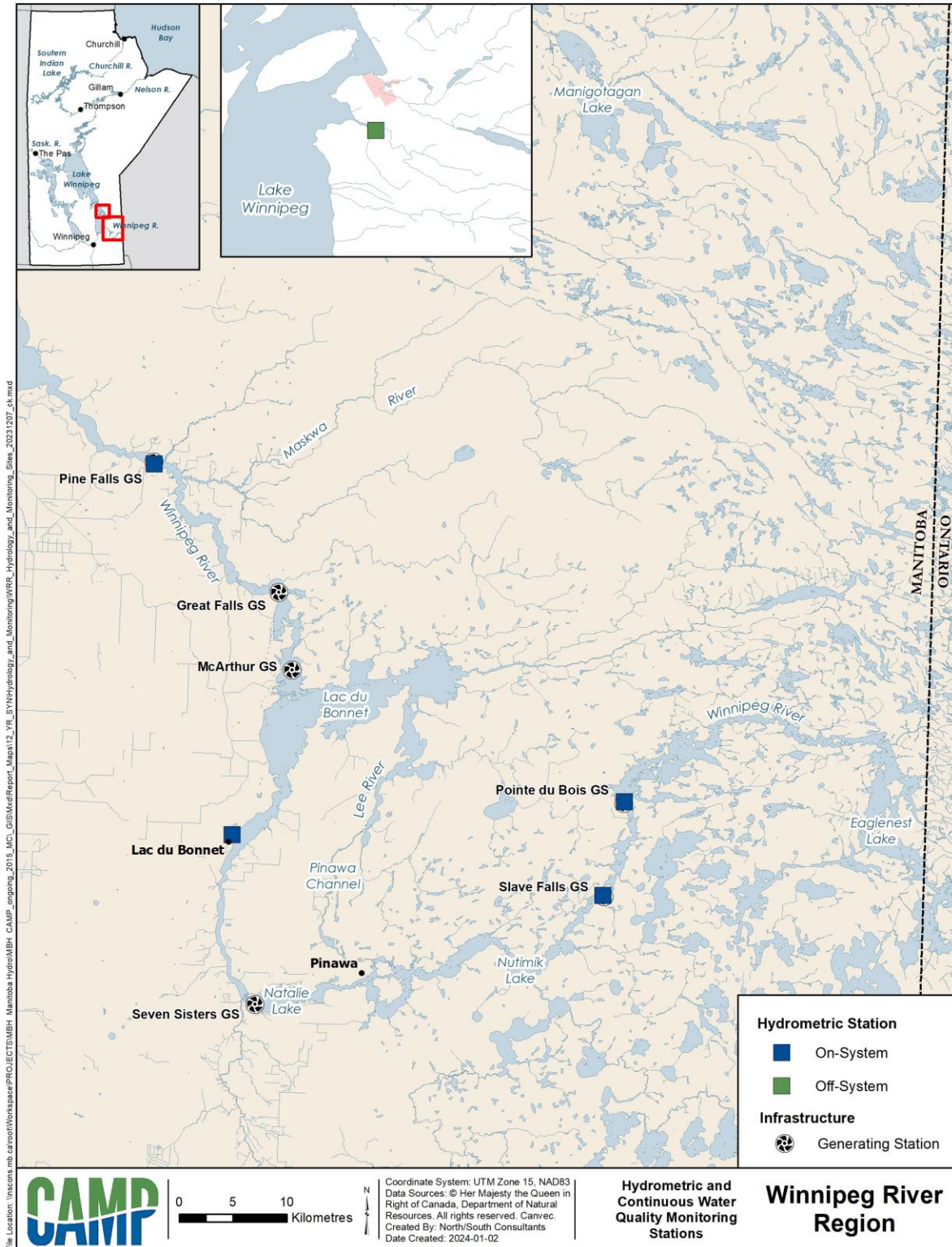


Figure 2.3-1. Hydrometric and continuous water quality monitoring stations in the Winnipeg River Region.

2.3.1 FLOW

2.3.1.1 ON-SYSTEM SITES

Winnipeg River

From 2008 to 2020, flow conditions on the Winnipeg River ranged from very dry to very wet and were more frequently above average than below average compared to the reference period from 1981 to 2010 (Figure 2.3-2 and Table 2.3-1). Monthly mean flow ranged from 278 to 2,503 cms with the overall mean from 2008 to 2020 at 1,025 cms. Flow conditions were very dry, defined as lower than 10th percentile, in parts of three years during the 2008 to 2020 CAMP monitoring period in August-November 2011, August and September 2018, and August-November 2020 (Table 2.3-1). Flow conditions were very wet, defined as above the 90th percentile, in seven years during CAMP, during the following months; June-August 2008, May-July, and September 2009, August 2010, May and June 2011, June and July 2013, May-September 2014, and October-December 2019 (Table 2.3-1).

2.3.1.2 OFF-SYSTEM SITES

Manigotagan River

From October 2010 to 2020, flow conditions on the Manigotagan River ranged from very dry to very wet and were more frequently above average than below average as compared to the reference period from 1965 to 1996 (Figure 2.3-3 and Table 2.3-2). There is no flow data available for the Manigotagan River from 1997 to September 2010. The monthly mean values ranged from 1.1 to 52.4 cms with the overall average from 2010 to 2020 at 10 cms. Flow conditions were very dry (below 10th percentile) with coincident very low Manigotagan Lake levels during the following months; January, February, November, and December 2012 (Table 2.3-2). Flow conditions were very wet (above 90th percentile) with coincident very high Manigotagan Lake levels during the following months; October 2010, May and June 2011, June, July and November 2012, May 2013, May-July 2014, September 2015, April and May 2016, and October and November 2019 (Table 2.3-2).

Table 2.3-1. Winnipeg River monthly average flow (cms).

| Year | Annual | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|--------|------|------|------|------|------|------|------|------|------|------|------|------|
| 2008 | 1216 | 1230 | 1182 | 1053 | 882 | 996 | 1628 | 2062 | 1939 | 968 | 719 | 885 | 1029 |
| 2009 | 1340 | 1000 | 1055 | 1176 | 1262 | 1873 | 2140 | 1754 | 1423 | 1466 | 1115 | 890 | 908 |
| 2010 | 1004 | 952 | 979 | 998 | 821 | 528 | 840 | 1078 | 1459 | 1193 | 1177 | 1041 | 972 |
| 2011 | 867 | 987 | 1052 | 1073 | 1222 | 1631 | 1533 | 1036 | 354 | 278 | 338 | 372 | 545 |
| 2012 | 725 | 607 | 649 | 600 | 587 | 606 | 878 | 1192 | 891 | 690 | 526 | 702 | 767 |
| 2013 | 997 | 776 | 868 | 934 | 674 | 814 | 1616 | 1526 | 1062 | 1004 | 1008 | 851 | 823 |
| 2014 | 1343 | 851 | 985 | 1003 | 945 | 1620 | 2267 | 2495 | 2124 | 1565 | 757 | 691 | 780 |
| 2015 | 931 | 849 | 897 | 894 | 692 | 528 | 1009 | 1162 | 1064 | 1025 | 855 | 866 | 1321 |
| 2016 | 1216 | 1358 | 1265 | 1168 | 1333 | 1406 | 1055 | 1289 | 1085 | 946 | 1261 | 1223 | 1195 |
| 2017 | 1020 | 1235 | 1260 | 1281 | 1396 | 1135 | 840 | 948 | 787 | 535 | 864 | 1067 | 904 |
| 2018 | 686 | 863 | 971 | 927 | 612 | 488 | 555 | 507 | 391 | 323 | 537 | 1080 | 998 |
| 2019 | 1198 | 979 | 945 | 1025 | 1160 | 1263 | 923 | 759 | 668 | 764 | 2161 | 2199 | 1518 |
| 2020 | 778 | 1361 | 1267 | 1205 | 1193 | 698 | 656 | 690 | 417 | 436 | 402 | 395 | 637 |

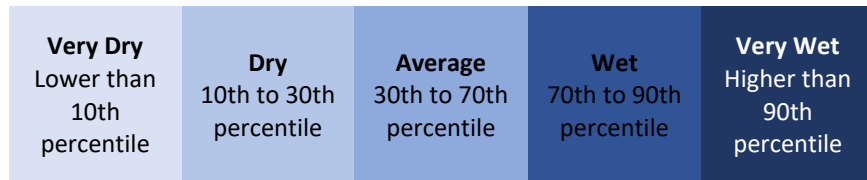


Table 2.3-2. Manigotagan River monthly average flow (cms).

| Year | Annual | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|--------|-----|-----|-----|-----|-----|-----|-----|-----|-----|-----|-----|-----|
| 2008 | | | | | | | | | | | | | |
| 2009 | | | | | | | | | | | | | |
| 2010 | 15 | | | | | | | | | | 21 | 19 | 10 |
| 2011 | 11 | 6 | 5 | 4 | 16 | 39 | 29 | 12 | 5 | 4 | 3 | 2 | 1 |
| 2012 | 12 | 1 | 1 | 3 | 4 | 5 | 27 | 34 | 19 | 10 | 10 | 21 | 12 |
| 2013 | 8 | 7 | 5 | 3 | 4 | 23 | 18 | 8 | 6 | | 3 | 3 | 3 |
| 2014 | 13 | 3 | 3 | 3 | 8 | 52 | 34 | 22 | 10 | 7 | 8 | 5 | 4 |
| 2015 | 12 | 3 | 3 | 3 | 6 | 14 | 18 | 11 | 19 | 30 | 14 | 14 | 14 |
| 2016 | 13 | 10 | 6 | 8 | 26 | 29 | 18 | 12 | 7 | 10 | 10 | 13 | 12 |
| 2017 | 7 | 8 | 6 | 5 | 14 | 15 | 8 | 5 | 4 | 4 | 5 | 7 | 6 |
| 2018 | 4 | 4 | 4 | 3 | 3 | 3 | 3 | 3 | 2 | 2 | 4 | 8 | 7 |
| 2019 | 10 | 5 | 4 | 3 | 7 | 17 | 10 | 9 | 5 | 5 | 23 | 25 | 10 |
| 2020 | 8 | 6 | 5 | 3 | 8 | 18 | 18 | 17 | 6 | 4 | 4 | 4 | 3 |

| | | | | |
|---|--|--|--|--|
| Very Dry Lower than 10th percentile | Dry 10th to 30th percentile | Average 30th to 70th percentile | Wet 70th to 90th percentile | Very Wet Higher than 90th percentile |
|---|--|--|--|--|

Notes:

- Blank cell indicates no data.

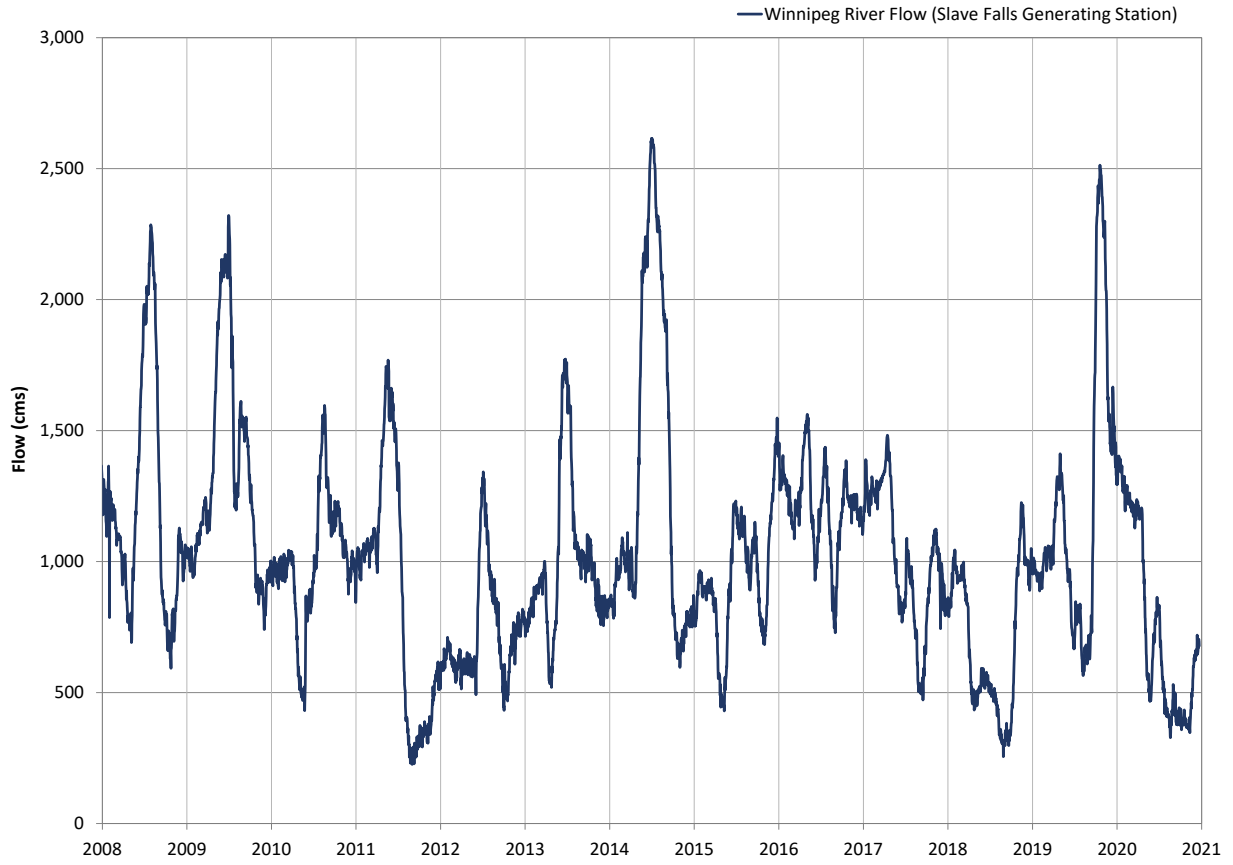


Figure 2.3-2. 2008-2020 Winnipeg River daily mean flow.

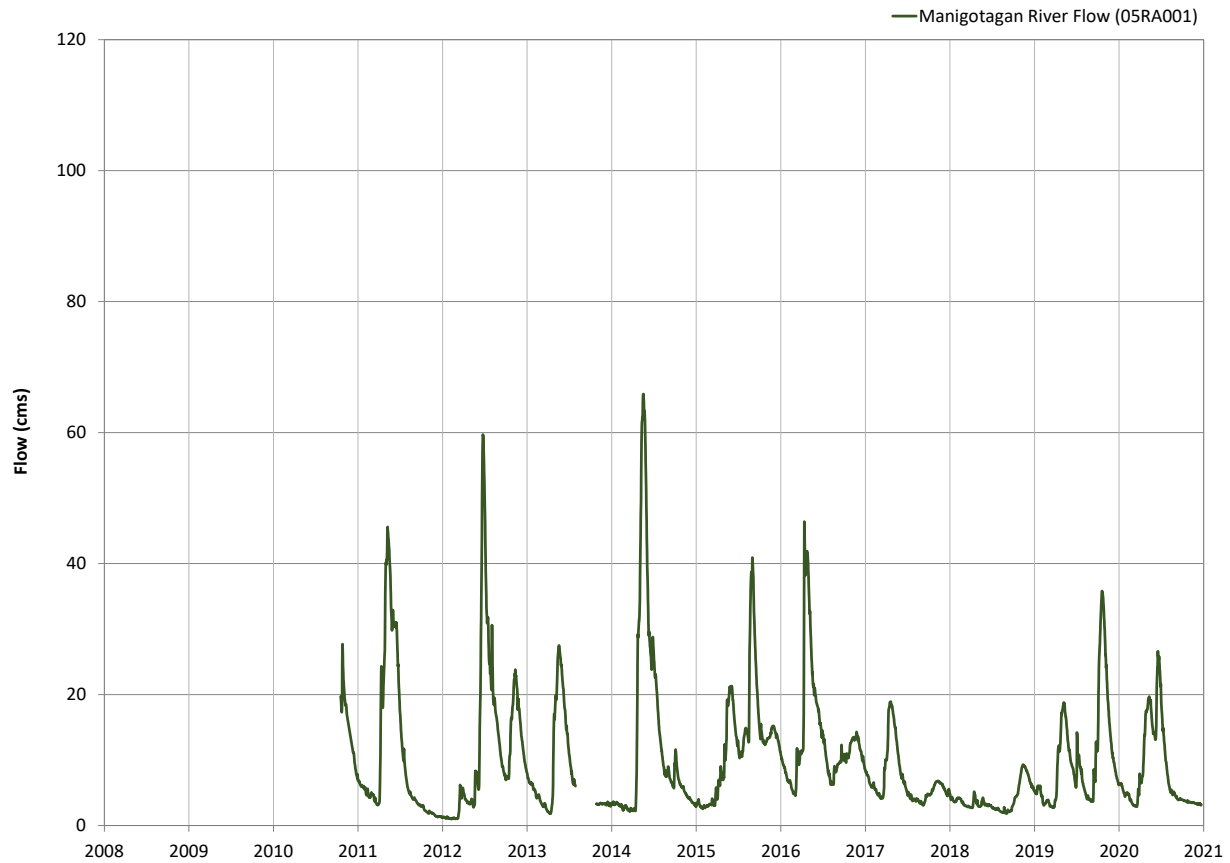


Figure 2.3-3. 2010-2020 Manigotagan River daily mean flow.

2.3.2 WATER LEVEL AND VARIABILITY

2.3.2.1 ON-SYSTEM SITES

Despite the changing flow conditions on the Winnipeg River, the water level at the three on-system CAMP locations (the Pointe du Bois Forebay, Lac du Bonnet (McArthur Forebay), and the Pine Falls Forebay) remained very stable as the flow through the generating stations are regulated to maintain stable upstream water levels.

Pointe du Bois Forebay

The Pointe du Bois forebay water level remained just below 299.1 m and close to average (within 0.5 m above or below) from 2008 to 2020 (Figure 2.3-4 and Table 2.3-3). Monthly average water levels ranged from 299.02 to 299.08 m. Water level variability remained low (less than 0.25 m) for the entire period from 2008 to 2020 (Table 2.3-4).

Lac du Bonnet

The water level on Lac du Bonnet remained steady near 254.9 m from 2008 to 2020 (Figure 2.3-5 and Table 2.3-5). Monthly average water levels ranged from 254.82 to 255.01 m. Water level variability also remained low (less than 0.25 m) for most of the period from 2008 to 2020, with the exception of six one-month periods in the record showing moderate variability (0.25 to 0.75 m; Table 2.3-6).

ROTATIONAL SITES

Pine Falls Forebay

The Pine Falls forebay remained near 229.1 m from 2008 to 2020 (Figure 2.3-6 and Table 2.3-7). Monthly average water levels ranged from 229.04 to 229.16 m. Water level variability remained low (less than 0.25 m) during this period, with the exception of three one-month periods in the record with moderate variability (0.25 to 0.75 m; Table 2.3-8).

2.3.2.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

Water levels are not recorded on off-system Manigotagan Lake. Water levels vary with Manigotagan River flow, which is measured downstream from Manigotagan Lake (Figure 2.3-1). As a result, Manigotagan specific water levels are not provided but rather may be inferred/visualized from Manigotagan flow presented in Figure 2.3-3 and Table 2.3-2 above).

ROTATIONAL SITES

Eaglenest Lake

Water levels are also not recorded on off-system Eaglenest Lake. Water levels vary with Winnipeg River flow, which is measured downstream from Eaglenest Lake. As a result, Winnipeg River Flow summary information in Figure 2.3-2 and Table 2.3-1 may be used to visualize water level changes on Eaglenest Lake.

Table 2.2-3. Pointe du Bois Forebay monthly average water level (m).

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|
| 2008 | 299.07 | 299.06 | 299.07 | 299.08 | 299.08 | 299.08 | 299.08 | 299.07 | 299.07 | 299.07 | 299.08 | 299.06 |
| 2009 | 299.06 | 299.07 | 299.07 | 299.07 | 299.08 | 299.07 | 299.06 | 299.06 | 299.06 | 299.07 | 299.07 | 299.07 |
| 2010 | 299.08 | 299.07 | 299.08 | 299.06 | 299.06 | 299.07 | 299.07 | 299.08 | 299.07 | 299.08 | 299.07 | 299.07 |
| 2011 | 299.07 | 299.07 | 299.07 | 299.07 | 299.07 | 299.07 | 299.07 | 299.05 | 299.06 | 299.06 | 299.05 | 299.02 |
| 2012 | 299.04 | 299.04 | 299.05 | 299.05 | 299.07 | 299.07 | 299.06 | 299.06 | 299.06 | 299.06 | 299.04 | 299.05 |
| 2013 | 299.05 | 299.06 | 299.06 | 299.05 | 299.06 | 299.06 | 299.05 | 299.06 | 299.07 | 299.06 | 299.06 | 299.06 |
| 2014 | 299.07 | 299.06 | 299.06 | 299.06 | 299.07 | 299.06 | 299.05 | 299.06 | 299.06 | 299.08 | 299.07 | 299.08 |
| 2015 | 299.08 | 299.08 | 299.08 | 299.07 | 299.08 | 299.08 | 299.07 | 299.07 | 299.08 | 299.06 | 299.08 | 299.08 |
| 2016 | 299.07 | 299.07 | 299.08 | 299.08 | 299.07 | 299.07 | 299.08 | 299.07 | 299.07 | 299.07 | 299.08 | 299.08 |
| 2017 | 299.08 | 299.08 | 299.08 | 299.08 | 299.06 | 299.08 | 299.08 | 299.08 | 299.08 | 299.08 | 299.08 | 299.08 |
| 2018 | 299.08 | 299.08 | 299.08 | 299.07 | 299.08 | 299.07 | 299.07 | 299.07 | 299.07 | 299.08 | 299.08 | 299.08 |
| 2019 | 299.08 | 299.07 | 299.06 | 299.04 | 299.07 | 299.07 | 299.07 | 299.07 | 299.08 | 299.08 | 299.07 | 299.06 |
| 2020 | 299.07 | 299.06 | 299.08 | 299.07 | 299.06 | 299.07 | 299.07 | 299.07 | 299.07 | 299.07 | 299.08 | 299.08 |

| | | |
|--|--|--|
| Lower Lower than 0.5 m below average | Average Within 0.5 m below and above average | Higher More than 0.5 m above average |
|--|--|--|

Table 2.2-4. Pointe du Bois Forebay monthly water level range (m).

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|------|------|------|------|------|------|------|------|------|------|------|------|
| 2008 | 0.08 | 0.06 | 0.06 | 0.06 | 0.05 | 0.08 | 0.05 | 0.07 | 0.09 | 0.10 | 0.06 | 0.05 |
| 2009 | 0.05 | 0.06 | 0.04 | 0.05 | 0.02 | 0.03 | 0.05 | 0.04 | 0.04 | 0.06 | 0.04 | 0.04 |
| 2010 | 0.03 | 0.03 | 0.05 | 0.07 | 0.08 | 0.05 | 0.06 | 0.07 | 0.04 | 0.03 | 0.05 | 0.06 |
| 2011 | 0.05 | 0.03 | 0.03 | 0.03 | 0.02 | 0.03 | 0.04 | 0.09 | 0.06 | 0.07 | 0.09 | 0.07 |
| 2012 | 0.07 | 0.06 | 0.05 | 0.06 | 0.06 | 0.05 | 0.09 | 0.04 | 0.05 | 0.04 | 0.07 | 0.05 |
| 2013 | 0.04 | 0.03 | 0.04 | 0.04 | 0.07 | 0.05 | 0.06 | 0.05 | 0.05 | 0.05 | 0.05 | 0.04 |
| 2014 | 0.04 | 0.04 | 0.04 | 0.05 | 0.05 | 0.05 | 0.05 | 0.06 | 0.06 | 0.06 | 0.05 | 0.04 |
| 2015 | 0.05 | 0.06 | 0.03 | 0.05 | 0.04 | 0.03 | 0.06 | 0.06 | 0.05 | 0.06 | 0.04 | 0.05 |
| 2016 | 0.04 | 0.04 | 0.04 | 0.05 | 0.04 | 0.05 | 0.03 | 0.05 | 0.04 | 0.04 | 0.03 | 0.03 |
| 2017 | 0.03 | 0.04 | 0.02 | 0.03 | 0.06 | 0.04 | 0.02 | 0.04 | 0.04 | 0.03 | 0.04 | 0.04 |
| 2018 | 0.04 | 0.11 | 0.03 | 0.10 | 0.03 | 0.04 | 0.04 | 0.04 | 0.05 | 0.04 | 0.03 | 0.03 |
| 2019 | 0.03 | 0.06 | 0.09 | 0.18 | 0.06 | 0.06 | 0.12 | 0.06 | 0.04 | 0.04 | 0.05 | 0.07 |
| 2020 | 0.05 | 0.06 | 0.05 | 0.11 | 0.06 | 0.05 | 0.05 | 0.05 | 0.05 | 0.04 | 0.04 | 0.05 |

| | | |
|--|---|---|
| Lower Variability Below 0.25 m | Moderate Variability 0.25 to 0.75 m | Higher Variability Above 0.75 m |
|--|---|---|

Table 2.2-5. Lac du Bonnet Monthly average water level (m).

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|
| 2008 | 254.90 | 254.93 | 254.92 | 254.90 | 254.90 | 254.91 | 254.94 | 254.91 | 254.82 | 254.86 | 254.88 | 254.90 |
| 2009 | 254.90 | 254.92 | 254.93 | 254.92 | 254.95 | 254.95 | 254.92 | 254.92 | 254.90 | 254.89 | 254.88 | 254.89 |
| 2010 | 254.89 | 254.91 | 254.90 | 254.86 | 254.88 | 254.89 | 254.90 | 254.91 | 254.88 | 254.88 | 254.87 | 254.90 |
| 2011 | 254.89 | 254.92 | 254.92 | 254.92 | 254.92 | 254.90 | 254.86 | 254.88 | 254.89 | 254.87 | 254.87 | 254.87 |
| 2012 | 254.87 | 254.88 | 254.85 | 254.87 | 254.84 | 254.88 | 254.89 | 254.87 | 254.87 | 254.84 | 254.90 | 254.89 |
| 2013 | 254.92 | 254.92 | 254.88 | 254.86 | 254.92 | 254.92 | 254.92 | 254.90 | 254.88 | 254.88 | 254.88 | 254.93 |
| 2014 | 254.95 | 254.91 | 254.91 | 254.89 | 254.94 | 254.99 | 255.01 | | 254.91 | 254.89 | 254.93 | 254.95 |
| 2015 | 254.93 | 254.89 | 254.88 | 254.87 | 254.88 | 254.91 | 254.90 | 254.89 | 254.89 | 254.88 | 254.91 | 254.90 |
| 2016 | 254.93 | 254.92 | 254.92 | 254.92 | 254.89 | 254.88 | 254.90 | 254.87 | 254.88 | 254.88 | 254.87 | 254.89 |
| 2017 | 254.93 | 254.92 | 254.91 | 254.90 | 254.88 | 254.85 | 254.86 | 254.84 | 254.85 | 254.89 | 254.86 | 254.87 |
| 2018 | 254.91 | 254.88 | 254.90 | 254.88 | 254.91 | 254.89 | 254.84 | 254.86 | 254.89 | 254.90 | 254.90 | 254.90 |
| 2019 | 254.90 | 254.91 | 254.90 | 254.91 | 254.91 | 254.86 | 254.90 | 254.87 | 254.88 | 254.97 | 254.91 | 254.91 |
| 2020 | 254.93 | 254.93 | 254.92 | 254.92 | 254.88 | 254.90 | 254.87 | 254.89 | 254.89 | 254.88 | 254.89 | 254.90 |

| | | |
|--|--|--|
| Lower Lower than 0.5 m below average | Average Within 0.5 m below and above average | Higher More than 0.5 m above average |
|--|--|--|

Notes:

- Blank cell indicates no data.

Table 2.2-6. Lac du Bonnet monthly water level range (m).

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|------|------|------|------|------|------|------|------|------|------|------|------|
| 2008 | 0.10 | 0.07 | 0.08 | 0.22 | 0.09 | 0.15 | 0.12 | 0.16 | 0.16 | 0.11 | 0.09 | 0.08 |
| 2009 | 0.06 | 0.05 | 0.10 | 0.08 | 0.10 | 0.12 | 0.17 | 0.14 | 0.05 | 0.11 | 0.12 | 0.07 |
| 2010 | 0.08 | 0.06 | 0.06 | 0.19 | 0.17 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.08 | 0.08 |
| 2011 | 0.11 | 0.06 | 0.07 | 0.08 | 0.09 | 0.11 | 0.13 | 0.16 | 0.09 | 0.06 | 0.04 | 0.11 |
| 2012 | 0.17 | 0.08 | 0.14 | 0.08 | 0.25 | 0.14 | 0.06 | 0.10 | 0.12 | 0.25 | 0.14 | 0.13 |
| 2013 | 0.08 | 0.10 | 0.07 | 0.25 | 0.12 | 0.08 | 0.08 | 0.05 | 0.10 | 0.10 | 0.16 | 0.12 |
| 2014 | 0.05 | 0.11 | 0.12 | 0.13 | 0.12 | 0.08 | 0.07 | | 0.09 | 0.07 | 0.14 | 0.15 |
| 2015 | 0.11 | 0.10 | 0.13 | 0.25 | 0.30 | 0.12 | 0.09 | 0.07 | 0.06 | 0.20 | 0.13 | 0.10 |
| 2016 | 0.08 | 0.07 | 0.06 | 0.09 | 0.12 | 0.11 | 0.10 | 0.11 | 0.14 | 0.10 | 0.08 | 0.06 |
| 2017 | 0.08 | 0.06 | 0.05 | 0.04 | 0.06 | 0.11 | 0.15 | 0.18 | 0.27 | 0.14 | 0.04 | 0.15 |
| 2018 | 0.14 | 0.07 | 0.10 | 0.16 | 0.11 | 0.09 | 0.21 | 0.18 | 0.08 | 0.11 | 0.07 | 0.07 |
| 2019 | 0.07 | 0.06 | 0.11 | 0.07 | 0.05 | 0.14 | 0.15 | 0.11 | 0.13 | 0.16 | 0.10 | 0.07 |
| 2020 | 0.08 | 0.04 | 0.05 | 0.08 | 0.13 | 0.14 | 0.14 | 0.07 | 0.08 | 0.10 | 0.13 | 0.06 |

| | | |
|--|---|---|
| Lower Variability Below 0.25 m | Moderate Variability 0.25 to 0.75 m | Higher Variability Above 0.75 m |
|--|---|---|

Notes:

- Blank cell indicates no data.

Table 2.2-7. Pine Falls Forebay monthly average water level (m).

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|
| 2008 | 229.13 | 229.15 | 229.13 | 229.11 | 229.12 | 229.12 | 229.12 | 229.11 | 229.14 | 229.09 | 229.10 | 229.12 |
| 2009 | 229.14 | 229.15 | 229.16 | 229.15 | 229.14 | 229.13 | 229.13 | 229.14 | 229.14 | 229.15 | 229.13 | 229.14 |
| 2010 | 229.13 | 229.15 | 229.14 | 229.13 | 229.11 | 229.15 | 229.14 | 229.14 | 229.14 | 229.14 | 229.15 | 229.14 |
| 2011 | 229.15 | 229.16 | 229.15 | 229.16 | 229.15 | 229.16 | 229.14 | 229.13 | 229.15 | 229.10 | 229.13 | 229.11 |
| 2012 | 229.11 | 229.12 | 229.10 | 229.11 | 229.09 | 229.13 | 229.15 | 229.09 | 229.11 | 229.10 | 229.08 | 229.07 |
| 2013 | 229.08 | 229.08 | 229.11 | 229.09 | 229.15 | 229.15 | 229.15 | 229.16 | 229.15 | 229.16 | 229.11 | 229.10 |
| 2014 | 229.11 | 229.15 | 229.16 | 229.16 | 229.15 | 229.13 | 229.11 | 229.11 | 229.14 | 229.11 | 229.11 | 229.09 |
| 2015 | 229.10 | 229.10 | 229.11 | 229.10 | 229.10 | 229.14 | 229.15 | 229.15 | 229.14 | 229.12 | 229.12 | 229.14 |
| 2016 | 229.14 | 229.15 | 229.15 | 229.15 | 229.15 | 229.14 | 229.15 | 229.15 | 229.14 | 229.15 | 229.15 | 229.15 |
| 2017 | 229.15 | 229.15 | 229.15 | 229.15 | 229.15 | 229.13 | 229.15 | 229.11 | 229.10 | 229.11 | 229.15 | 229.12 |
| 2018 | 229.11 | 229.14 | 229.11 | 229.08 | 229.09 | 229.07 | 229.09 | 229.10 | 229.09 | 229.06 | 229.15 | 229.12 |
| 2019 | 229.10 | 229.10 | 229.14 | 229.15 | 229.15 | 229.11 | 229.08 | 229.07 | 229.04 | 229.10 | 229.11 | 229.11 |
| 2020 | 229.14 | 229.14 | 229.15 | 229.14 | 229.09 | 229.10 | 229.07 | 229.08 | 229.06 | 229.08 | 229.06 | 229.08 |

| | | |
|--|--|--|
| Lower Lower than 0.5 m below average | Average Within 0.5 m below and above average | Higher More than 0.5 m above average |
|--|--|--|

Table 2.2-8. Pine Falls Forebay monthly water level range (m).

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
|------|------|------|------|------|------|------|------|------|------|------|------|------|
| 2008 | 0.20 | 0.07 | 0.11 | 0.16 | 0.16 | 0.11 | 0.12 | 0.13 | 0.09 | 0.24 | 0.14 | 0.26 |
| 2009 | 0.11 | 0.08 | 0.04 | 0.06 | 0.08 | 0.08 | 0.09 | 0.08 | 0.09 | 0.08 | 0.09 | 0.11 |
| 2010 | 0.10 | 0.10 | 0.12 | 0.10 | 0.11 | 0.08 | 0.06 | 0.10 | 0.07 | 0.10 | 0.10 | 0.07 |
| 2011 | 0.08 | 0.07 | 0.04 | 0.06 | 0.05 | 0.06 | 0.08 | 0.12 | 0.09 | 0.21 | 0.09 | 0.12 |
| 2012 | 0.10 | 0.09 | 0.08 | 0.09 | 0.14 | 0.09 | 0.07 | 0.13 | 0.10 | 0.12 | 0.13 | 0.15 |
| 2013 | 0.09 | 0.16 | 0.11 | 0.15 | 0.08 | 0.06 | 0.04 | 0.02 | 0.08 | 0.06 | 0.16 | 0.19 |
| 2014 | 0.12 | 0.10 | 0.02 | 0.04 | 0.05 | 0.07 | 0.07 | 0.06 | 0.05 | 0.12 | 0.11 | 0.17 |
| 2015 | 0.10 | 0.16 | 0.17 | 0.11 | 0.15 | 0.11 | 0.06 | 0.06 | 0.06 | 0.09 | 0.12 | 0.06 |
| 2016 | 0.09 | 0.03 | 0.04 | 0.02 | 0.06 | 0.12 | 0.05 | 0.06 | 0.13 | 0.06 | 0.06 | 0.04 |
| 2017 | 0.05 | 0.03 | 0.04 | 0.03 | 0.05 | 0.18 | 0.14 | 0.16 | 0.16 | 0.22 | 0.06 | 0.15 |
| 2018 | 0.14 | 0.08 | 0.20 | 0.11 | 0.14 | 0.11 | 0.13 | 0.09 | 0.19 | 0.25 | 0.06 | 0.11 |
| 2019 | 0.18 | 0.12 | 0.15 | 0.04 | 0.04 | 0.14 | 0.13 | 0.11 | 0.42 | 0.14 | 0.08 | 0.11 |
| 2020 | 0.07 | 0.06 | 0.05 | 0.06 | 0.16 | 0.14 | 0.15 | 0.13 | 0.16 | 0.17 | 0.20 | 0.11 |

| | | |
|--|---|---|
| Lower Variability Below 0.25 m | Moderate Variability 0.25 to 0.75 m | Higher Variability Above 0.75 m |
|--|---|---|

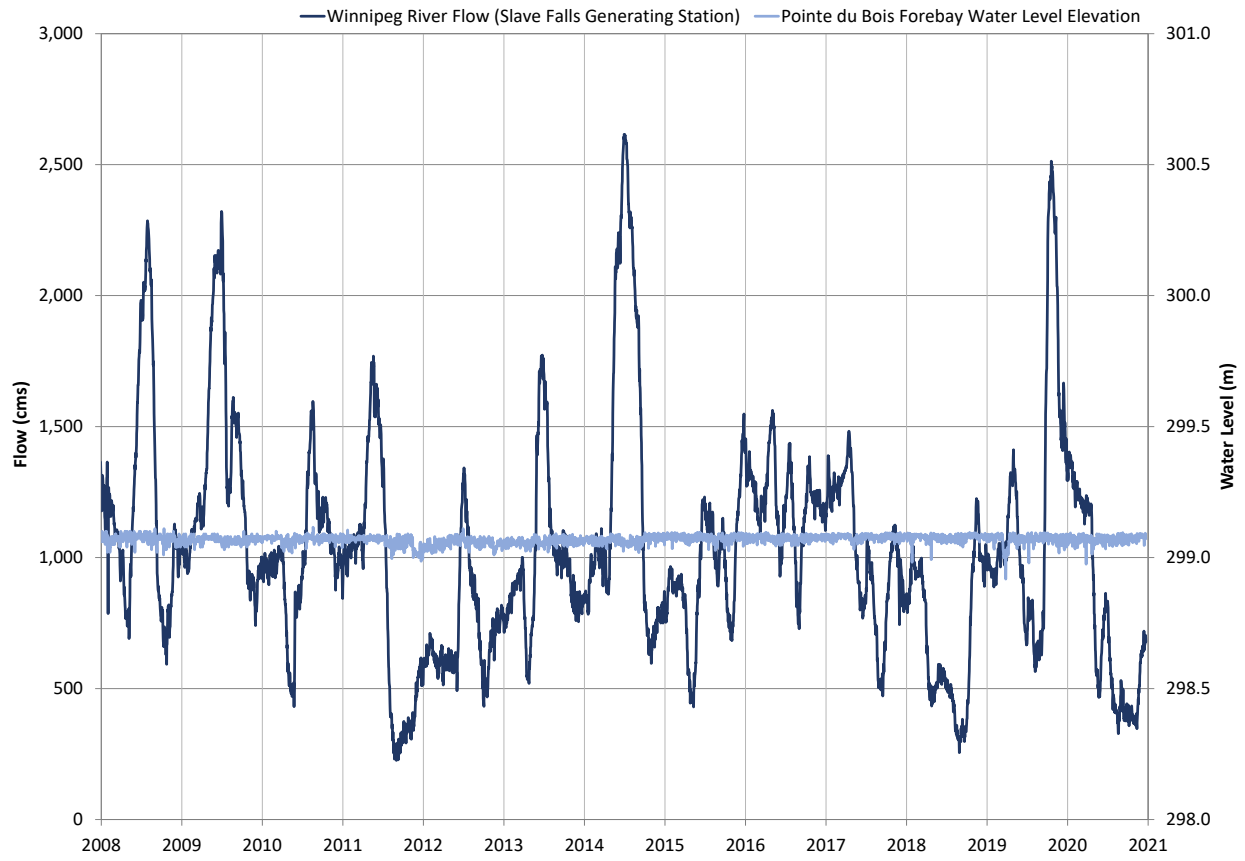


Figure 2.3-4. 2008-2020 Winnipeg River daily mean flow and Pointe du Bois Forebay daily average water level.

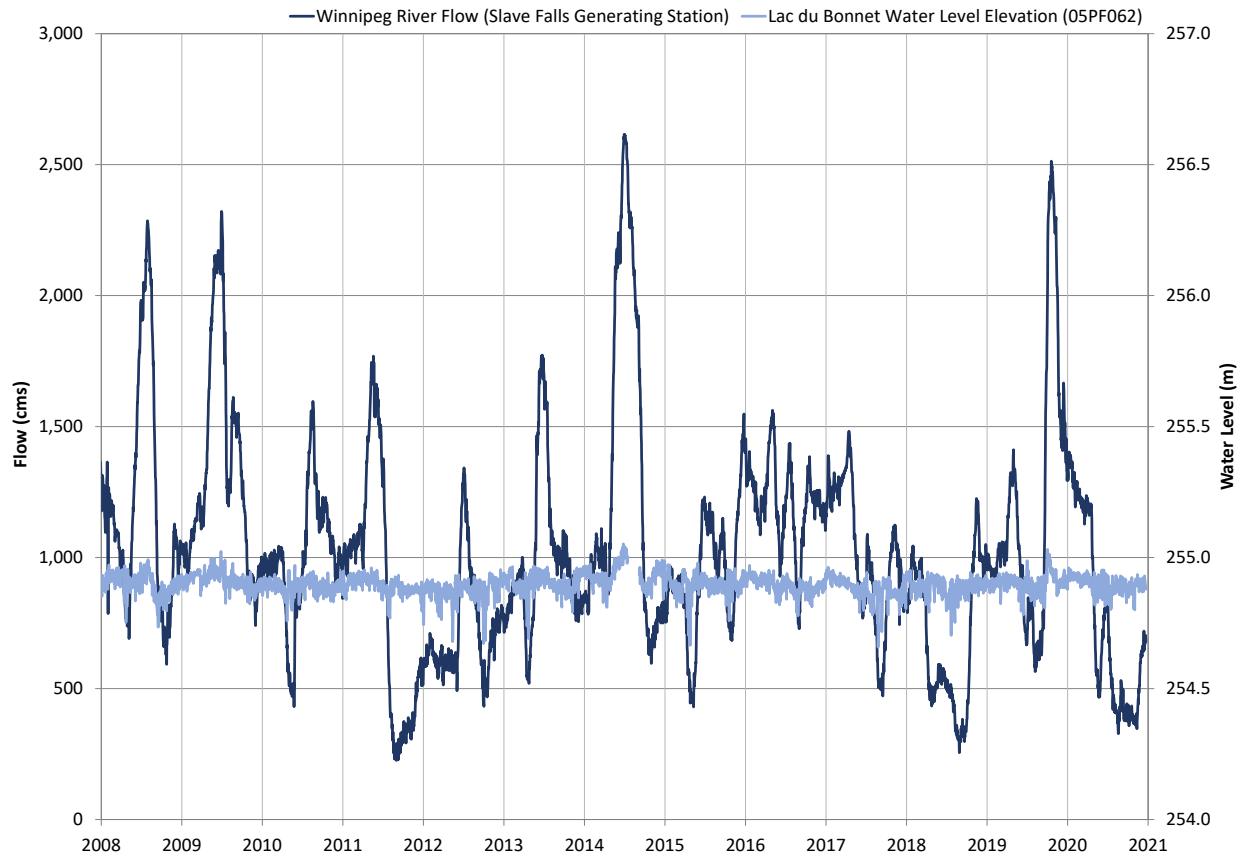


Figure 2.3-5. 2008-2020 Winnipeg River daily mean flow and Lac du Bonnet daily average water level.

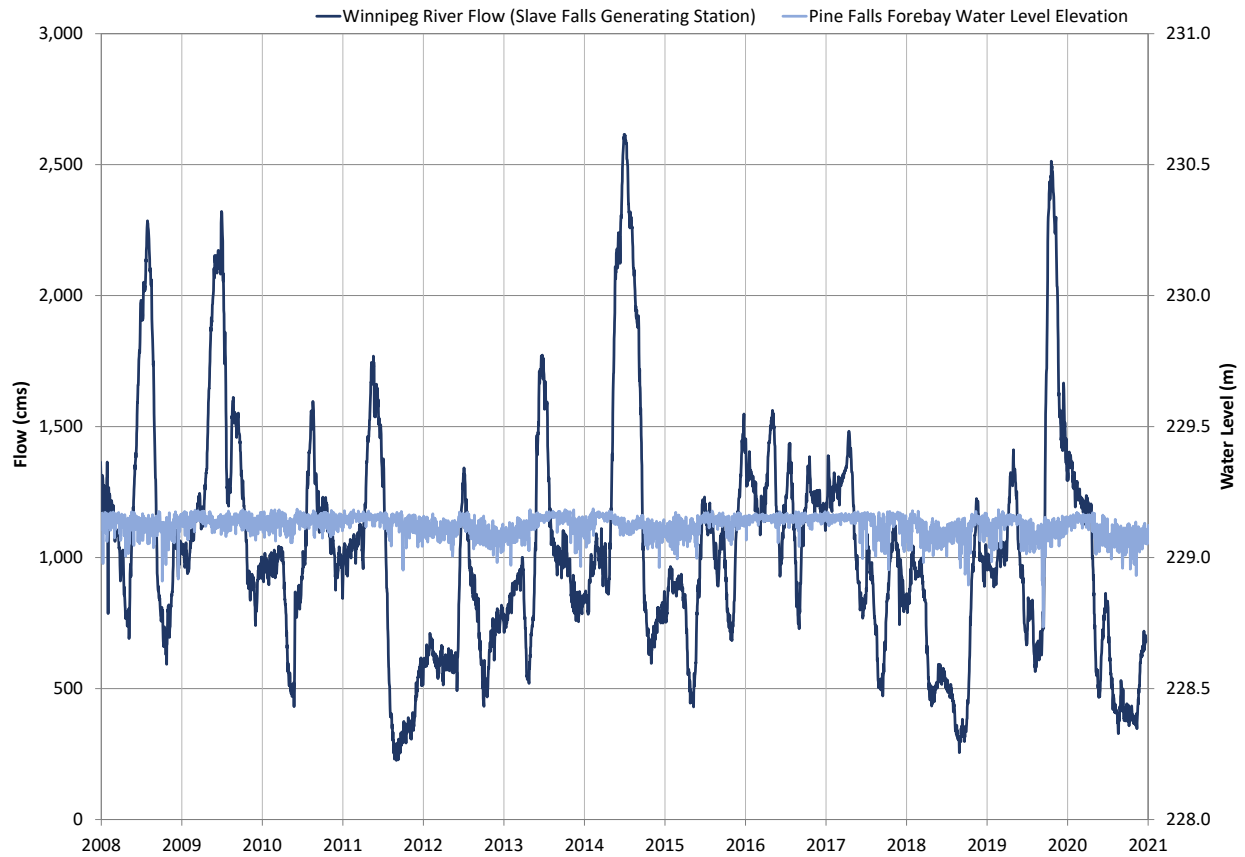


Figure 2.3-6. 2008-2020 Winnipeg River daily mean flow and the Pine Falls GS Forebay daily average water level.

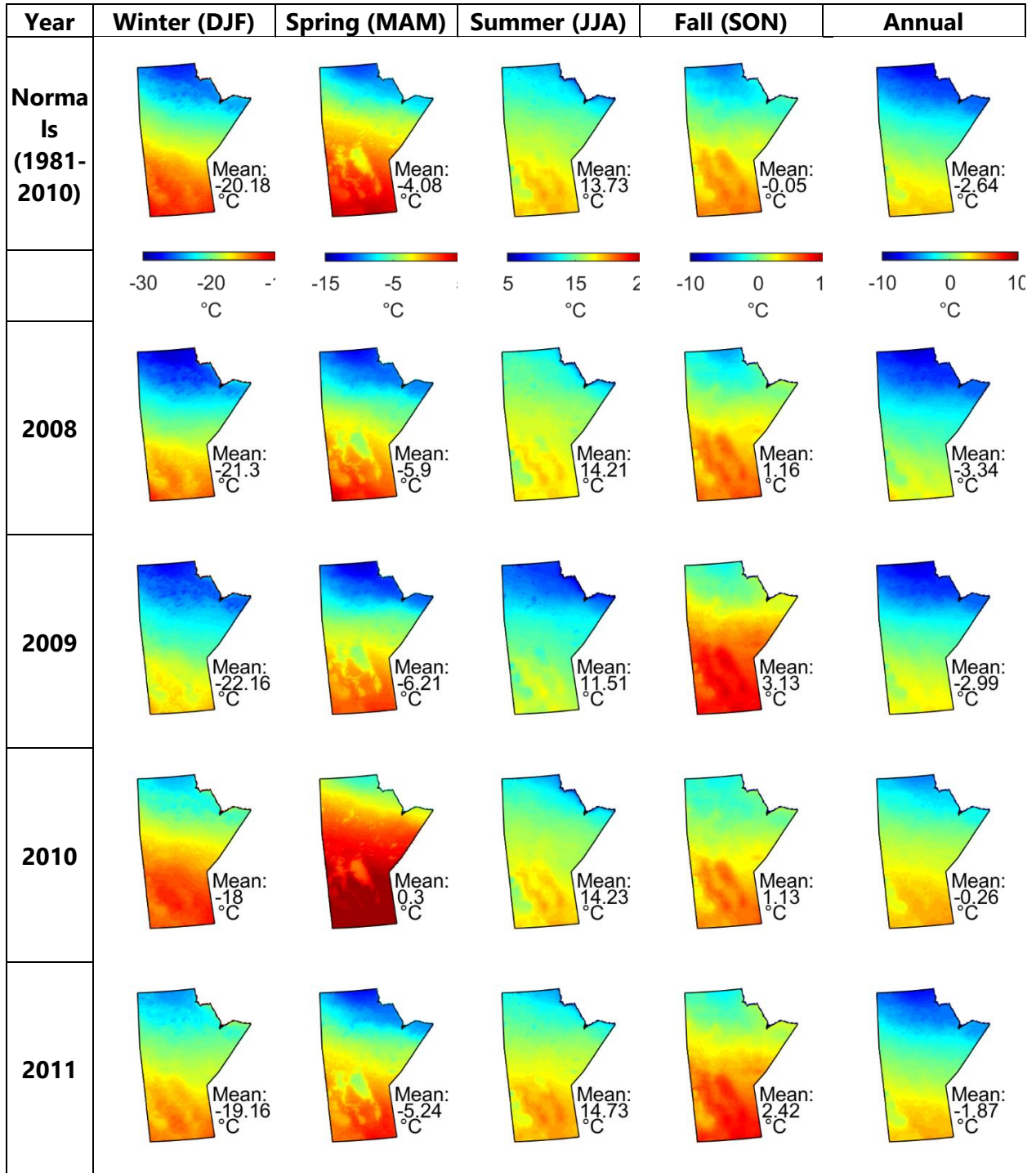
2.3.3 WATER TEMPERATURE

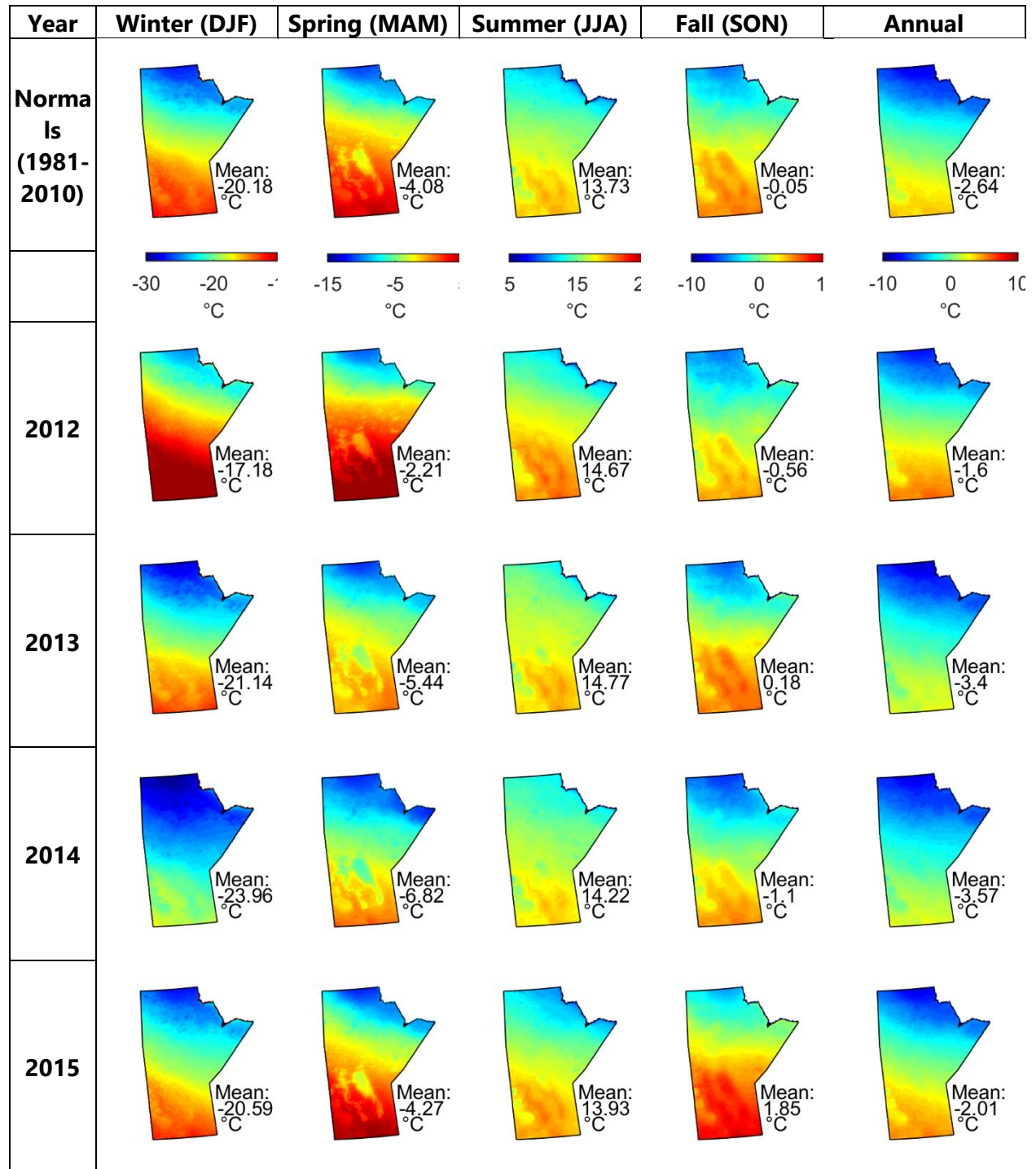
Continuous water temperature data are not available for this region during the period from 2008 to 2020. Continuous water temperature monitoring was initiated at the Pointe du Bois GS in 2020, after the monitoring period for this report and this data will be presented in the next reporting cycle.

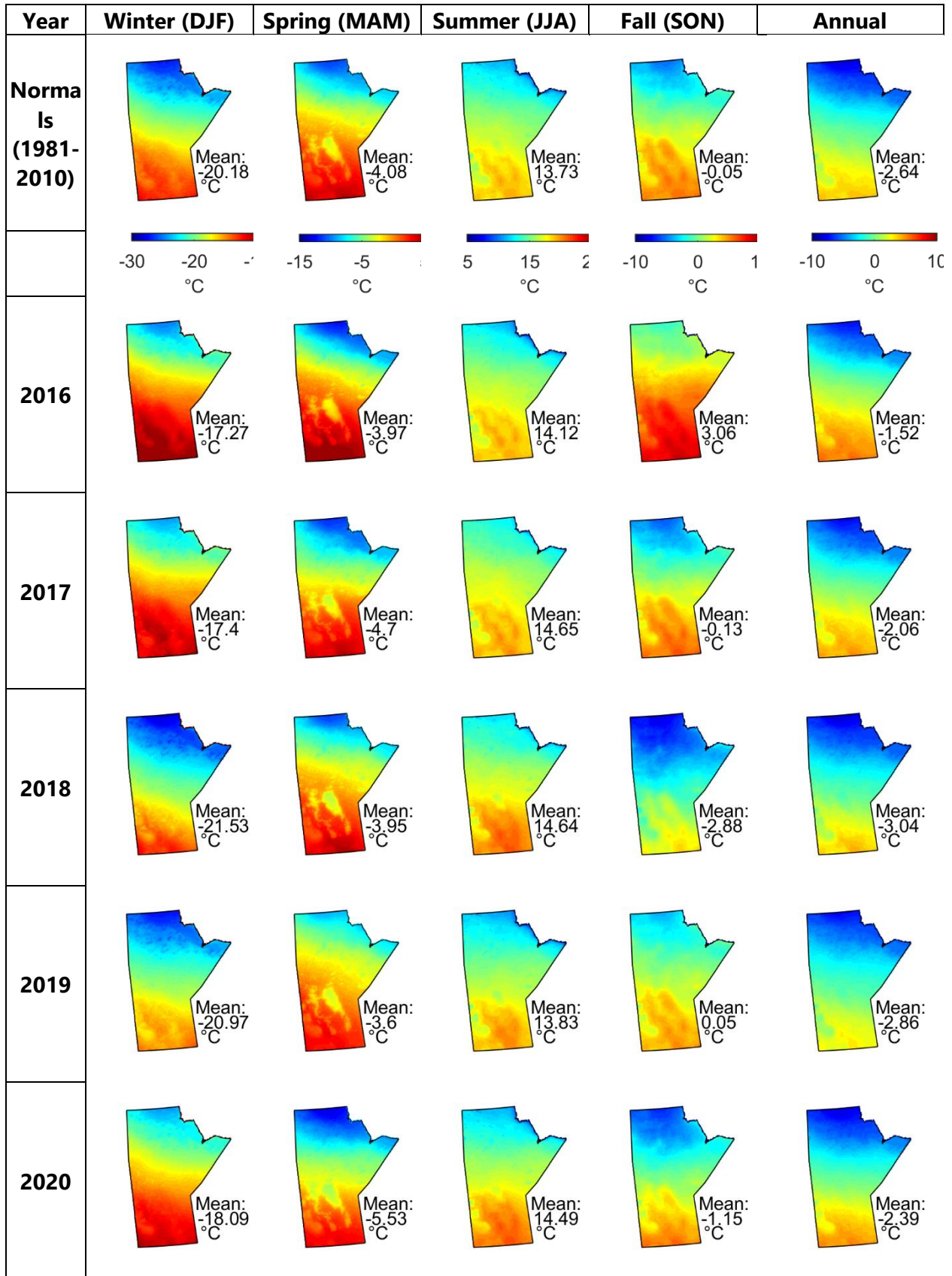
2.4 SEDIMENTATION

Sedimentation data are not available for this region during the period from 2008 to 2020. Monitoring was initiated at the Pointe du Bois GS in 2020, after the monitoring period for this report and this data will be presented in the next reporting cycle.

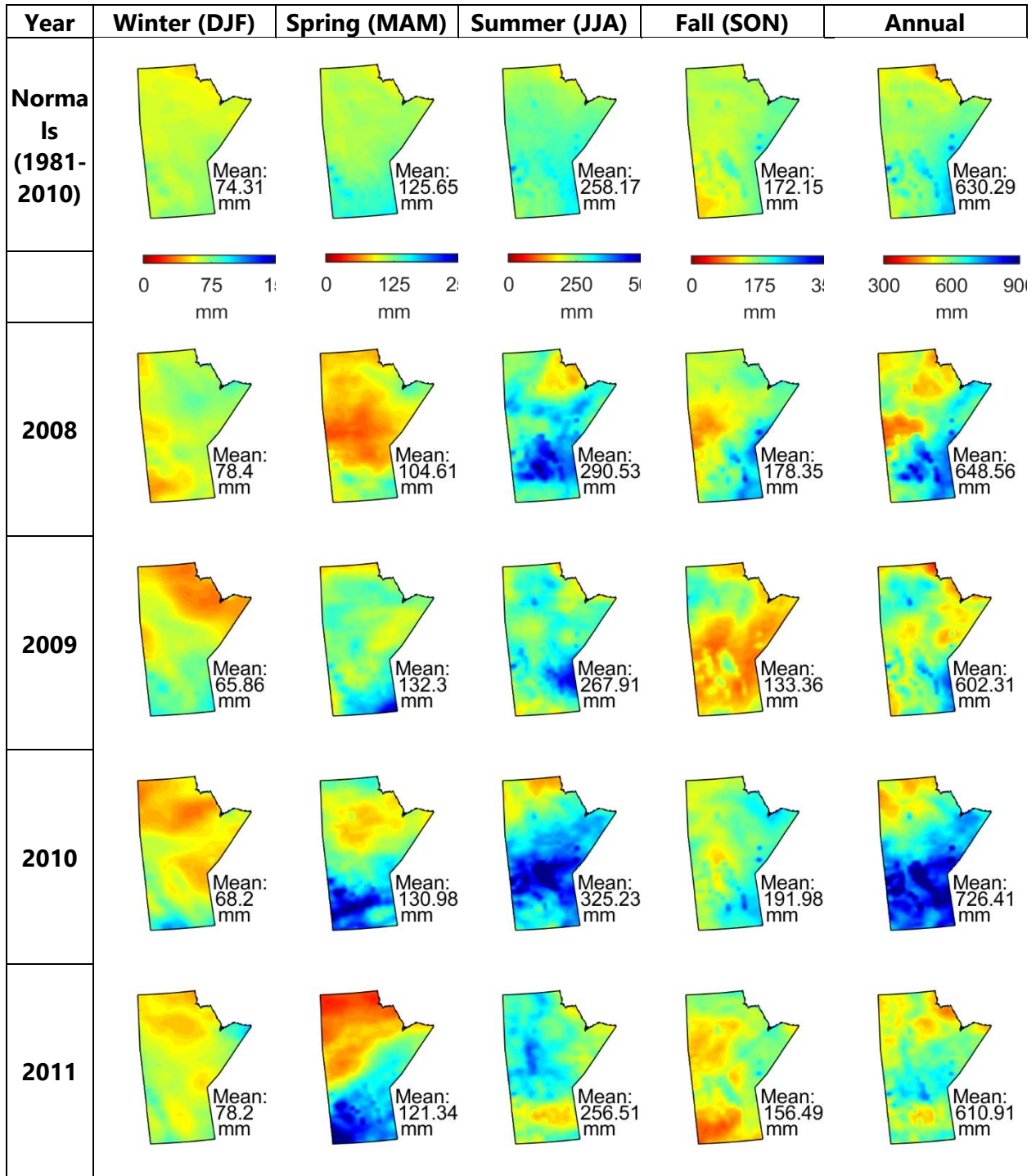
**APPENDIX 2-1. SEASONAL AND ANNUAL TEMPERATURE
NORMALS DERIVED FROM ERA5-LAND
DATA**

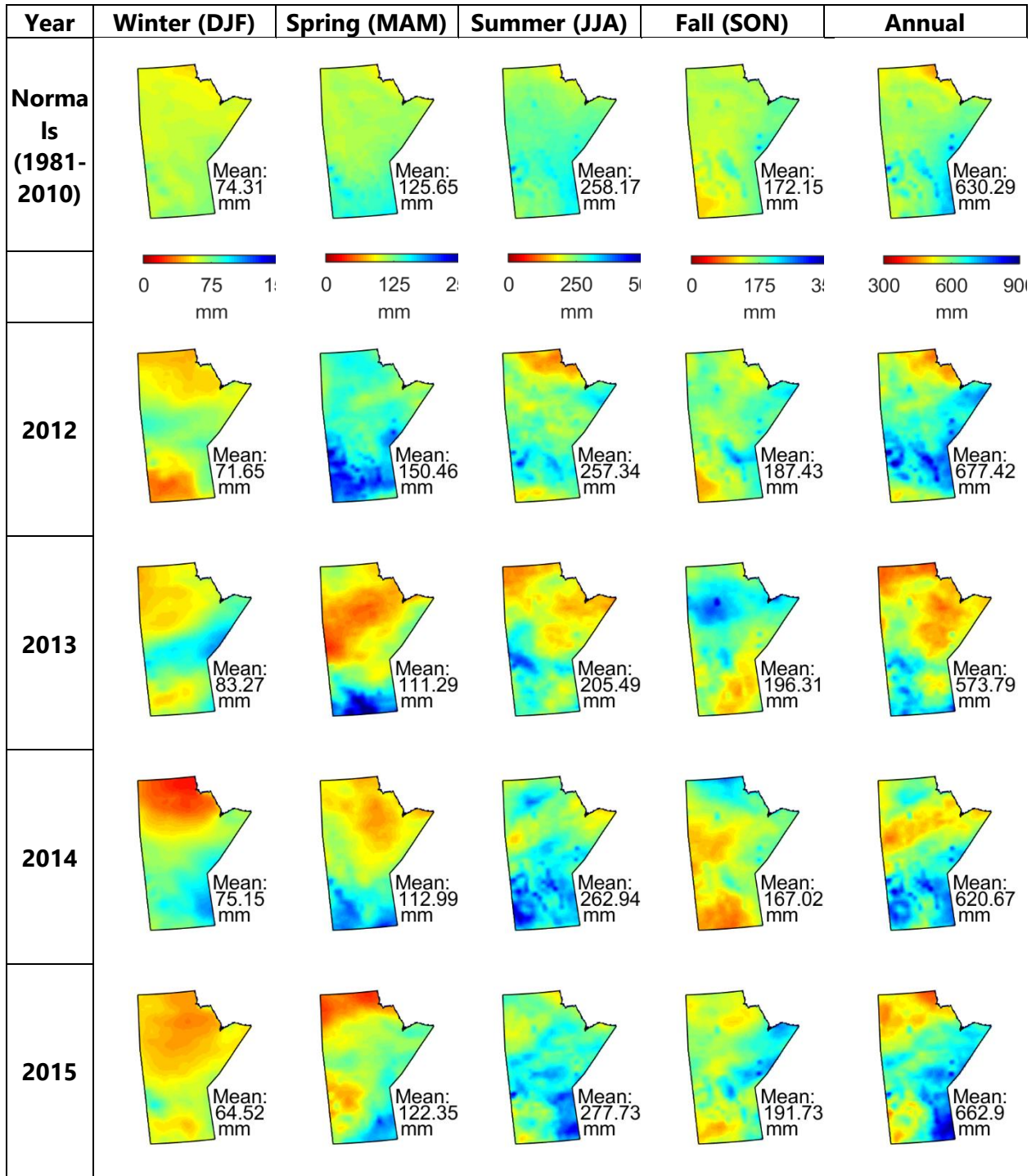


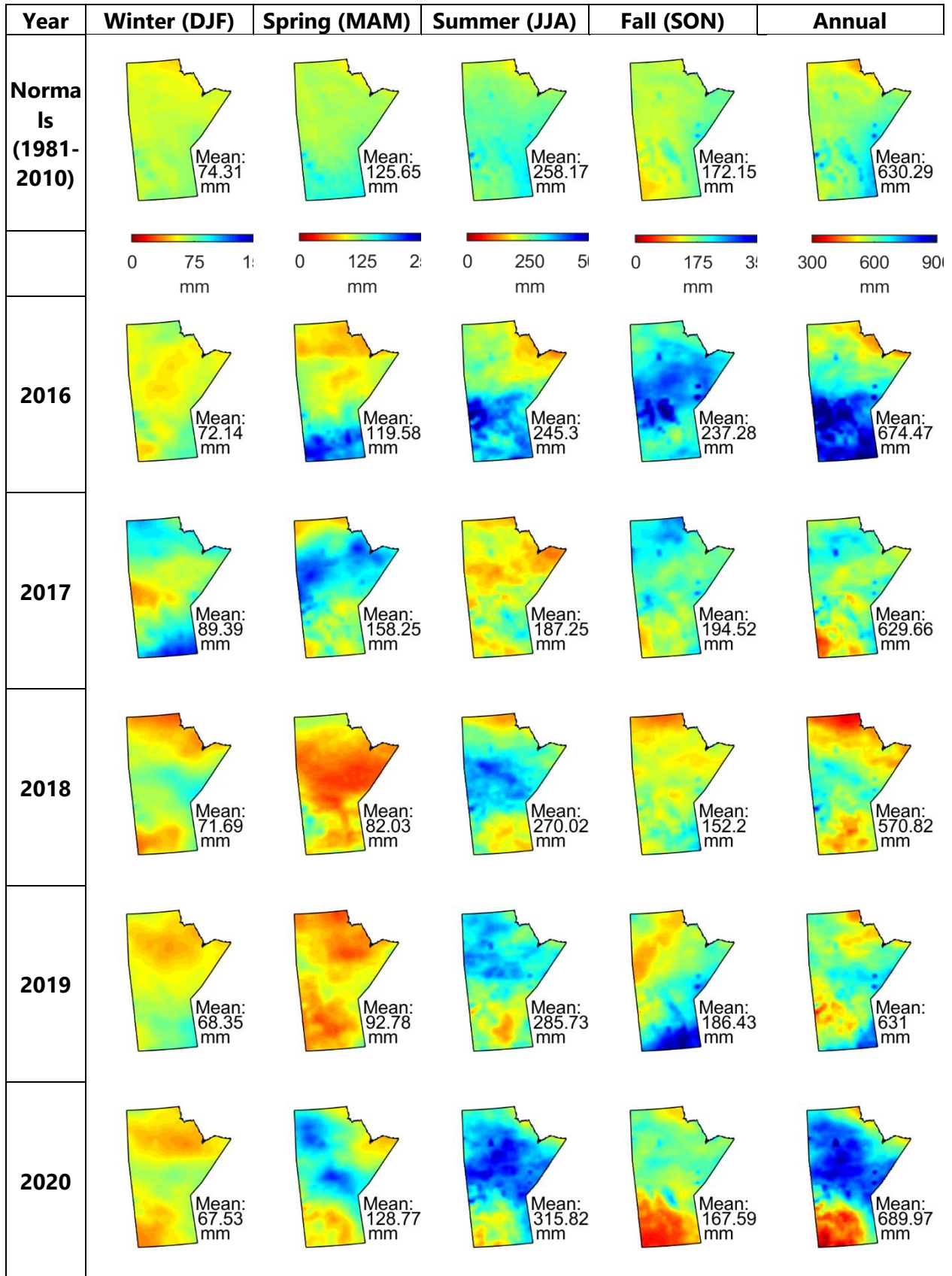




**APPENDIX 2-2. SEASONAL AND PRECIPITATION NORMALS
DERIVED FROM ERA5-LAND DATA**







3.0 WATER QUALITY

3.1 INTRODUCTION

The following presents the results of water quality monitoring conducted from 2008 to 2019 in the Winnipeg River Region. Five waterbodies were monitored in the Winnipeg River Region: two on-system annual sites (the Pointe du Bois Forebay and Lac du Bonnet) and one on-system rotational site (the Pine Falls Forebay); and one off-system annual site (Manigotagan Lake) and one off-system rotational site (Eaglenest Lake; Table 3.1-1 and Figure 3.1-1). Annual sites are sampled each year, whereas rotational sites are sampled once every three years on a rotational basis and are therefore limited to three or four years of data for the 12-year period. Eaglenest Lake is located on the Winnipeg River upstream of the Pointe du Bois Forebay and is not affected by Manitoba Hydro's hydraulic operating system and is therefore identified as an "off-system" site.

The CAMP water quality program includes four sampling periods (referred to as spring, summer, fall, and winter) per monitoring year (i.e., April-March) at a single location within each waterbody or area of a waterbody/river reach. Over the 12-year period, water quality sampling was conducted at each sampling location during each sampling period (i.e., n=48 for annual sites) with two exceptions (Table 3.1-1; Appendix 3-1):

- sampling could not be completed in the Pointe du Bois Forebay due to thin ice in the winter of 2009 therefore only 11 winter samples were collected over the 12-year period (i.e., n=47); and
- sampling could not be completed in Eaglenest Lake due to thin ice in the winter of 2019, therefore only three winter samples were collected over the 12-year period (i.e., n=15).

A detailed description of the program design and sampling methods is provided in Technical Document 1, Section 2.3.

Three indicators (dissolved oxygen (DO); water clarity; and nutrients/trophic status) were selected for detailed reporting (Table 3.1-2). Metrics for these indicators include DO and its supporting metric temperature/stratification, Secchi disk depth, turbidity, total suspended solids (TSS), total phosphorus (TP), total nitrogen (TN), and chlorophyll *a* (Table 3.1-2). A detailed description of these indicators is provided in CAMP (2024).

Table 3.1-1. 2008-2019 Water quality sampling inventory.

| Waterbody/ Area | Sampling Year ¹ | | | | | | | | | | | |
|--------------------|----------------------------|----------------|------|------|------|------|------|------|------|------|------|----------------|
| | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
| PDB | • | • ² | • | • | • | • | • | • | • | • | • | • |
| LDB | • | • | • | • | • | • | • | • | • | • | • | • |
| PFF | | | | • | | | • | | | • | | |
| MANIG | • | • | • | • | • | • | • | • | • | • | • | • |
| EAGLE | | | • | | | • | | | • | | | • ² |

Notes:

1. Sampling year is from April-March.
2. No winter sample collected due to unsafe ice conditions.

Table 3.1-2. Water quality indicators and metrics.

| Indicator | Metric | Units |
|------------------------------|---|--|
| Dissolved Oxygen | • Dissolved oxygen (DO) | milligrams per litre (mg/L) and percent (%) saturation |
| | • Temperature/stratification ¹ | °C |
| Water Clarity | • Secchi disk depth | m |
| | • Turbidity | Nephelometric turbidity units (NTU) |
| | • Total suspended solids (TSS) | mg/L |
| Nutrients and Trophic Status | • Total phosphorus (TP) | mg/L |
| | • Total nitrogen (TN) | mg/L |
| | • Chlorophyll <i>a</i> | micrograms per litre (µg/L) |

Notes:

1. Supporting metric.

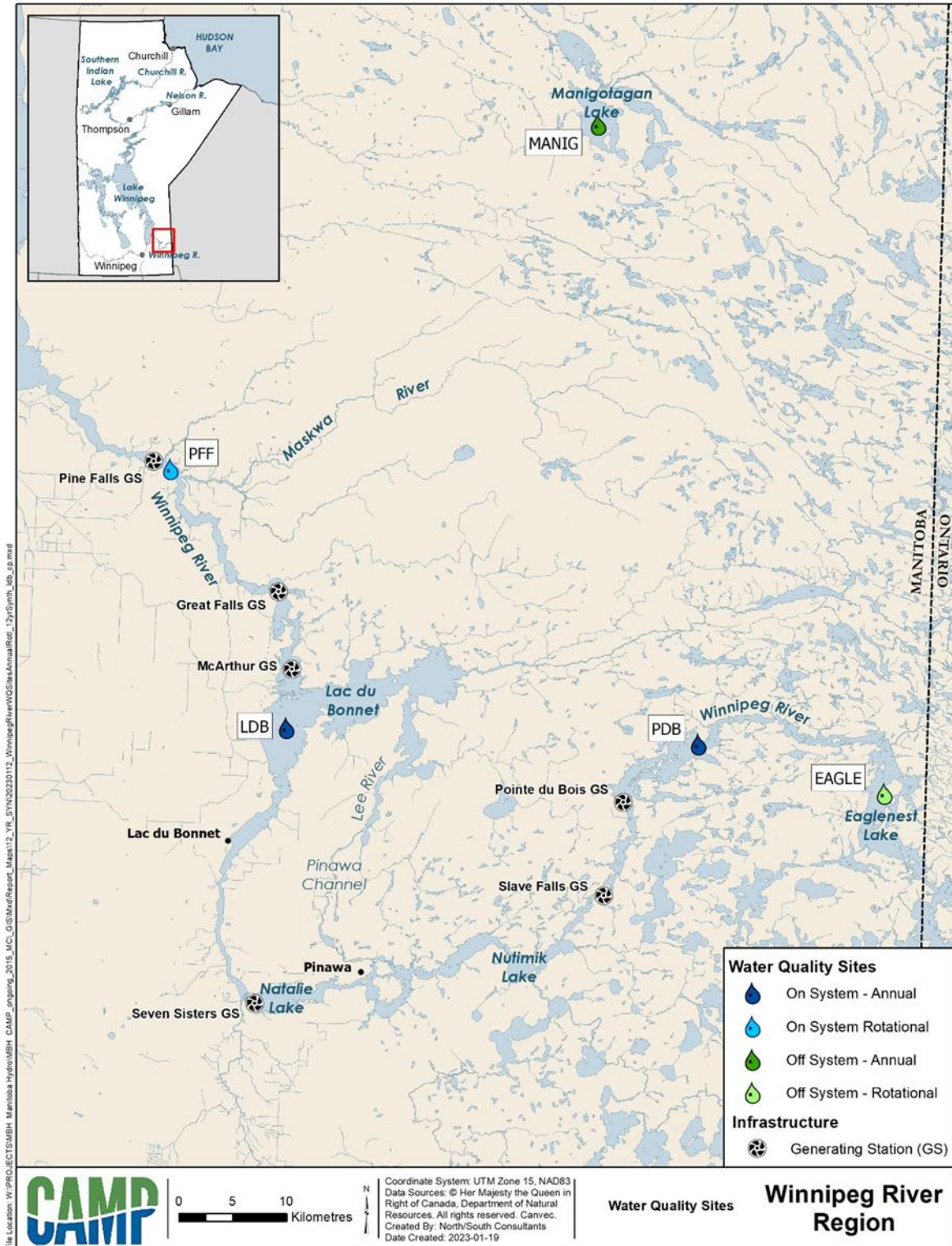


Figure 3.1-1. 2008-2019 Winnipeg River Region water quality sites.

3.2 DISSOLVED OXYGEN

3.2.1 DISSOLVED OXYGEN

3.2.1.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

The Pointe du Bois Forebay was well-oxygenated year-round and DO concentrations throughout the water column consistently met the Manitoba Water Quality Standards, Objectives, and Guidelines (MWQSOGs) instantaneous minimum objectives for cool- and cold-water aquatic life during the open-water and ice-cover seasons (Manitoba Water Stewardship (MWS) 2011; Table 3.2-1).

The Pointe du Bois Forebay was isothermal (i.e., thermal stratification was not observed) and DO concentrations were similar throughout the water column during each sampling period (Figures 3.2-1 and 3.2-2). During the open-water season, DO concentrations ranged from 7.22 to 14.04 mg/L at the surface and 7.19 to 14.07 mg/L near the bottom (maximum site water depth = 22.0 m). During the ice-cover season, DO concentrations ranged from 11.86 to 15.48 mg/L at the surface and 11.74 to 15.56 mg/L near the bottom (Table 3.2-2 and Figure 3.2-3).

DO concentrations varied between seasons with seasonal mean DO concentrations being higher in winter and spring when the water was cooler, and lower in the summer and fall when the water was warmer (Figure 3.2-4).

DO saturation was near 100% at both the surface and near the bottom during each season sampled (Figure 3.2-5). During the open-water season, surface DO saturation ranged from 86.1 to 121.9% with a mean of 101.4% and a median of 100.9% over the 12 years of monitoring. Mean surface DO saturation levels in the open-water season were similar from year to year ranging from 91.4 to 108.6% and were within or near the interquartile range (IQR) of 93.3 to 107.4%. Bottom DO saturation during the open-water season ranged from 85.6 to 122.1% with a mean of 100.3% and a median of 98.4% over the 12 years of monitoring. Mean bottom DO saturation levels in the open-water season were similar from year to year ranging from 90.9 to 107.2% and were within or near the IQR of 92.4 to 106.5% (Table 3.2-2 and Figure 3.2-6).

During the ice-cover season, DO saturation at the surface ranged from 81.1 to 110.3% with a mean of 96.4% and a median of 97.1%. The IQR was 94.0 to 98.7%. Bottom DO saturation during the ice-cover season ranged from 80.3 to 111.0% with a mean of 95.5% and a median of 94.9%. The IQR was 92.4 to 98.6% (Table 3.2-2 and Figure 3.2-7).

Lac du Bonnet

Lac du Bonnet was well-oxygenated year-round and DO concentrations throughout the water column consistently met the MWQSOGs instantaneous minimum objectives for cool- and cold-water aquatic life during the open-water and ice-cover seasons (Table 3.2-1).

Lac du Bonnet was typically isothermal; however, there were two occurrences of thermal stratification near the surface (thermocline at 0-1 m) in spring (2015 and 2018; Table 3.2-1 and Figure 3.2-1).

DO concentrations were similar throughout the water column during each sampling period (Figure 3.2-2). During the open-water season, DO concentrations ranged from 7.38 to 14.11 mg/L at the surface and 6.92 to 13.63 mg/L near the bottom (maximum site water depth = 23.8 m). During the ice-cover season, DO concentrations ranged from 13.60 to 15.62 mg/L at the surface and 13.56 to 15.36 mg/L near the bottom (Table 3.2-2 and Figure 3.2-8).

DO concentrations varied between seasons, with seasonal mean DO concentrations being higher in winter and spring when the water was cooler, and lower in the summer and fall when the water was warmer (Figure 3.2-4).

DO saturation was near 100% at the surface and near the bottom during each season sampled (Figure 3.2-5). During the open-water season, surface DO saturation ranged from 88.1 to 124.2% over the 12 years of monitoring, with a mean of 104.0% and a median of 102.2%. Mean surface DO saturation levels in the open-water season were similar from year to year ranging from 93.9 to 112.7 % and were within or near the IQR of 95.1 to 113.1%. Bottom DO saturation during the open-water season ranged from 80.1 to 124.0% with a mean of 100.5% and a median of 99.0% over the 12 years of monitoring. Mean bottom DO saturation levels in the open-water season were similar from year to year ranging from 90.2 to 109.0% and were within or near the IQR of 92.3 to 108.4% (Table 3.2-2 and Figure 3.2-6).

During the ice-cover season, DO saturation at the surface ranged from 94.1 to 110.0% with a mean of 104.8% and a median of 105.3%. The IQR was 102.7 to 108.7%. Bottom DO saturation during

the ice-cover season ranged from 93.0 to 107.7% with a mean of 103.8% and a median of 105.1%. The IQR was 103.4 to 106.2% (Table 3.2-2 and Figure 3.2-7).

ROTATIONAL SITES

Pine Falls Forebay

The Pine Falls Forebay was well-oxygenated year-round and DO concentrations throughout the water column consistently met the MWQSOGs instantaneous minimum objectives for cool- and cold-water aquatic life during the open-water and ice-cover seasons (Table 3.2-1).

The Pine Falls Forebay was isothermal and DO concentrations were similar throughout the water column during each sampling period (Table 3.2-1, and Figures 3.2-1 and 3.2-2). During the open-water season, DO concentrations ranged from 8.09 to 13.83 mg/L at the surface and 8.06 to 13.68 mg/L near the bottom (maximum site water depth = 39.0 m). During the ice-cover season, the surface DO concentration was 13.66 mg/L in 2011 and 15.31 mg/L in 2017 and near the bottom it was 13.70 mg/L in 2011 and 14.87 mg/L in 2017 (Table 3.2-2 and Figure 3.2-9).

DO saturation in the Pine Falls Forebay was near 100% at both the surface and near the bottom of the water column during each season sampled. During the open-water season, surface DO saturation ranged from 92.4 to 122.5% with a mean of 106.7% and a median of 109.6% over the three years of monitoring. Mean surface DO saturation levels in the open-water season were similar from year to year ranging from 100.8 to 116.3% and were within or near the IQR of 93.7 to 118.2%. Bottom DO saturation during the open-water season ranged from 91.2 to 121.2% with a mean of 105.4 and median of 108.5% over the three years of monitoring. Mean bottom DO saturation levels in the open-water season were similar from year to year ranging from 99.2 to 115.3% and were within or near the IQR of 92.2 to 116.7% (Table 3.2-2 and Figure 3.2-6).

During the ice-cover season, DO saturation at the surface was 100.9% in 2011 and 108.6% in 2017 with a mean of 104.7%. Bottom DO saturation during the ice-cover season was 101.2% in 2011 and 105.5% in 2017 with a mean of 103.3% (Table 3.2-2 and Figure 3.2-7).

3.2.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

Manigotagan Lake was well-oxygenated near the surface and DO concentrations met the MWQSOGs during all sampling periods. DO concentrations decreased with water depth and fell below the MWQSOGs instantaneous minimum objectives for cool- and cold-water aquatic life in the open-water season.

Manigotagan Lake was thermally stratified during most open-water sampling events conducted over the 12 years of monitoring (Figure 3.2-1). Specifically, stratification was observed in five of the 12 spring monitoring events (2008, 2010, 2014, 2015, and 2018), during each summer sampling event, and all but one (2012) fall sampling event (Table 3.2-3).

During late summer or fall of each year, DO concentrations decreased down the water column in Manigotagan Lake to levels below one or more of the MWQSOGs instantaneous minimum objectives for cool- and cold-water aquatic life (5.0 and 4.0 mg/L, respectively) at approximately 14-20 m from the surface (Table 3.2-3 and Figure 3.2-2). During the open-water season, DO concentrations ranged from 7.29 to 12.93 mg/L at the surface and from 0.12 to 13.39 mg/L near the bottom (maximum site water depth = 24.0 m; Table 3.2-4 and Figure 3.2-10).

Similarly, during the ice-cover season Manigotagan Lake was well-oxygenated near the surface while DO concentrations decreased with depth and fell below the MWQSOGs instantaneous minimum objectives for cold-water aquatic life near the bottom of the water column in some winters (Figure 3.2-2). Specifically, DO concentrations near the bottom were below the MWQSOGs instantaneous minimum objectives for cold-water aquatic life (8.0 mg/L) in the winters of 2009, 2018, and 2019 (Table 3.2-3). In the ice-cover season, DO concentrations ranged from 11.87 to 15.16 mg/L at the surface and from 4.67 to 10.82 mg/L near the bottom (Table 3.2-4 and Figure 3.2-10). The decrease in DO concentrations with depth occurred despite the lake being isothermal in winter (Table 3.2-3 and Figure 3.2-1).

DO concentrations varied between seasons with seasonal mean DO concentrations being higher in winter and spring when the water was colder, and lower in the summer and fall when the water was warmer (Figure 3.2-11).

DO saturation was near 100% at the surface during each season sampled (Figure 3.2-12). In the open-water season, surface DO saturation ranged from 83.6 to 117.4% with a mean of 100.2% and a median of 100.0% over the 12 years of monitoring. Mean surface DO saturation levels in the open-water season were similar from year to year ranging from 92.1 to 108.7% and were within or near the IQR of 94.6 to 106.0% (Table 3.2-4 and Figure 3.2-13). During the ice-cover season, surface DO saturation ranged from 85.4 to 112.3% with a mean of 100.3% and a median of 102.2%. The IQR for the ice-cover season was 96.9 to 106.5% (Table 3.2-4 and Figure 3.2-14).

Seasonal differences in both DO concentration and percent saturation occurred near the bottom of the water column where DO saturation decreased throughout the open-water season and was somewhat reduced in winter (Figures 3.2-11 and 3.2-12). During the open-water season, bottom DO saturation ranged from 0.8 to 111.6% with a mean of 57.7% and a median of 56.2% over the 12 years of monitoring (Table 3.2-4). Bottom DO saturation tended to be above the IQR for the open-water season (31.3 to 93.4%) in spring (mean = 99.0%) and below the IQR in fall (mean = 26.4%; Figure 3.2-13). During the ice-cover season, bottom DO saturation ranged from 35.3 to 85.0% with a mean of 62.2% and a median of 66.3%. The IQR for the ice-cover season was 44.3 to 74.6% (Table 3.2-4 and Figure 3.2-14).

ROTATIONAL SITES

Eaglenest Lake

Eaglenest Lake was well-oxygenated year-round and DO concentrations throughout the water column consistently met the MWQSOGs instantaneous minimum objectives for cool- and cold-water aquatic life during the open-water and ice-cover seasons (Table 3.2-3).

Eaglenest Lake was isothermal with the exception of spring 2010 when thermal stratification occurred near the surface (thermocline at 0-1 m; Table 3.2-3 and Figure 3.2-1).

DO concentrations were similar throughout the water column during each sampling period (Figure 3.2-2). During the open-water season, DO concentrations ranged from 8.23 to 13.24 mg/L at the surface and from 8.06 to 13.21 mg/L near the bottom (maximum site water depth = 21.6 m). During the ice-cover season, the DO concentration at the surface was 13.65 mg/L in 2013 and 14.19 mg/L in 2016, and near the bottom it was 13.32 mg/L in 2013 and 14.20 mg/L in 2016 (Table 3.2-4 and Figure 3.2-15).

DO saturation in Eaglenest Lake was near 100% at both the surface and near the bottom of the water column during each season sampled. During the open-water season, surface DO saturation ranged from 90.5 to 118.3% with a mean of 102.7% and a median of 103.7% over the four years of monitoring. Mean surface DO saturation levels in the open-water season were similar from year to year ranging from 97.2 to 105.0% and were within the IQR of 95.9 to 105.9%. Bottom DO saturation during the open-water season ranged from 90.0 to 114.7% with a mean of 100.3 and median of 97.9% over the four years of monitoring. Mean bottom DO saturation levels in the open-water season were similar from year to year ranging from 91.8 to 103.8% and were within or near the IQR of 94.3 to 104.8% (Table 3.2-4 and Figure 3.2 13).

During the ice-cover season, DO saturation at the surface was 101.5% in 2013 and 106.7% in 2016 with a mean of 104.1%, and bottom DO saturation during the ice-cover season was 99.1% in 2013 and 100.9% in 2016 with a mean of 100.0% (Table 3.2-4 and Figure 3.2-14).

Table 3.2-1. 2008-2019 On-system sites summary of thermal stratification and DO concentrations.

| Metric | Sampling Year | Surface or Bottom | PDB | | | | LDB | | | | PFF | | | |
|-------------------------------|---------------|-------------------|------------|-----|-----|-----------|------------|-----|-----|-----------|------------|-----|-----|-----------|
| | | | Open-Water | | | Ice-Cover | Open-Water | | | Ice-Cover | Open-Water | | | Ice-Cover |
| | | | SP | SU | FA | WI | SP | SU | FA | WI | SP | SU | FA | WI |
| Thermal Stratification | 2008 | | No | No | No | No | No | No | No | No | | | | |
| | 2009 | | No | No | No | ND | No | No | No | No | | | | |
| | 2010 | | No | No | No | No | No | No | No | No | | | | |
| | 2011 | | No | No | No | No | No | No | No | No | No | No | No | No |
| | 2012 | | No | No | No | No | No | No | No | No | | | | |
| | 2013 | | No | No | No | No | No | No | No | No | | | | |
| | 2014 | | No | No | No | No | No | No | No | No | No | No | No | No |
| | 2015 | | No | No | No | No | 2015 | No | No | No | | | | |
| | 2016 | | No | No | No | No | No | No | No | No | | | | |
| | 2017 | | No | No | No | No | No | No | No | No | No | No | No | No |
| | 2018 | | No | No | No | No | 2018 | No | No | No | | | | |
| 2019 | | No | No | No | No | No | No | No | No | | | | | |
| DO met MWQSOGs PAL objectives | 2008 | Surface | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | | | | |
| | | Bottom | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | | | | |
| | 2009 | Surface | Yes | ND | Yes | ND | Yes | ND | Yes | Yes | | | | |
| | | Bottom | Yes | ND | Yes | ND | Yes | ND | Yes | Yes | | | | |
| | 2010 | Surface | ND | Yes | Yes | ND | ND | Yes | Yes | ND | | | | |
| | | Bottom | ND | Yes | Yes | ND | ND | Yes | Yes | ND | | | | |
| | 2011 | Surface | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes |
| | | Bottom | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes |
| | 2012 | Surface | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | | | | |
| | | Bottom | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | | | | |
| | 2013 | Surface | ND | Yes | Yes | Yes | ND | Yes | Yes | Yes | | | | |
| | | Bottom | ND | Yes | Yes | Yes | ND | Yes | Yes | Yes | | | | |
| | 2014 | Surface | Yes | Yes | Yes | ND | Yes | Yes | Yes | ND | Yes | Yes | Yes | ND |
| | | Bottom | Yes | Yes | Yes | ND | Yes | Yes | Yes | ND | Yes | Yes | Yes | ND |
| | 2015 | Surface | ND | Yes | Yes | Yes | Yes | Yes | Yes | Yes | | | | |
| | | Bottom | ND | Yes | Yes | Yes | Yes | Yes | Yes | Yes | | | | |
| | 2016 | Surface | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | | | | |
| | | Bottom | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | | | | |
| | 2017 | Surface | ND | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes |
| | | Bottom | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes |
| 2018 | Surface | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | | | | | |
| | Bottom | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | | | | | |
| 2019 | Surface | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | | | | | |
| | Bottom | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | | | | | |

Notes:

1. SP = spring; SU = summer; FA = fall; WI = winter.
2. ND = No data.
3. MWQSOGs = Manitoba Water Quality Standards, Objectives, and Guidelines; PAL = Protection of aquatic life.
4. DO concentrations were compared to the most stringent MWQSOGs instantaneous minimum PAL objectives for each season; i.e., 5 mg/L for cool-water early life for the open-water season and 8 mg/L for cold-water early life the ice-cover season.
5. Cells with a year indicated denote instances of stratification or non-compliance with MWQSOGs instantaneous minimum objectives.
6. = Sampling did not occur.

Table 3.2-2. 2008-2019 On-system sites DO, water depth, and ice thickness summary statistics.

| Site | Statistic | Dissolved Oxygen | | | | | | | | Water Depth at Site (m) | | Ice Thickness at Site (m) |
|--------------|----------------|---------------------|-------|--------------------|-------|-----------------------------|-------|----------------------------|-------|-------------------------|------|---------------------------|
| | | DO - Surface (mg/L) | | DO - Bottom (mg/L) | | DO Saturation - Surface (%) | | DO Saturation - Bottom (%) | | OW | IC | IC |
| | | OW | IC | OW | IC | OW | IC | OW | IC | | | |
| PDB | Mean | 9.92 | 13.51 | 9.83 | 13.37 | 101.4 | 96.4 | 100.3 | 95.5 | 14.0 | 8.3 | 0.57 |
| | Median | 9.21 | 13.28 | 9.19 | 13.24 | 100.9 | 97.1 | 98.4 | 94.9 | 14.0 | 8.2 | 0.52 |
| | Minimum | 7.22 | 11.86 | 7.19 | 11.74 | 86.1 | 81.1 | 85.6 | 80.3 | 6.0 | 7.5 | 0.38 |
| | Maximum | 14.04 | 15.48 | 14.07 | 15.56 | 121.9 | 110.3 | 122.1 | 111.0 | 22.0 | 8.9 | 0.77 |
| | SD | 2.01 | 1.04 | 1.99 | 1.09 | 10.4 | 7.63 | 10.42 | 8.33 | 3.75 | 0.45 | 0.12 |
| | SE | 0.361 | 0.348 | 0.358 | 0.365 | 1.87 | 2.54 | 1.87 | 2.78 | 0.62 | 0.10 | 0.00 |
| | Lower Quartile | 8.57 | 13.07 | 8.40 | 12.73 | 93.3 | 94.0 | 92.4 | 92.4 | 20.8 | 8.8 | 0.80 |
| | Upper Quartile | 10.27 | 14.04 | 10.13 | 13.87 | 107.4 | 98.7 | 106.5 | 98.6 | 7.1 | 7.7 | 0.40 |
| | n | 31 | 9 | 31 | 9 | 31 | 9 | 31 | 9 | 36 | 11 | 11 |
| % Detections | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | - | - | - | |
| LDB | Mean | 10.14 | 14.65 | 9.94 | 14.51 | 104.0 | 104.8 | 100.5 | 103.8 | 13.5 | 13.3 | 0.71 |
| | Median | 9.50 | 14.77 | 9.47 | 14.66 | 102.2 | 105.3 | 99.0 | 105.1 | 13.0 | 12.7 | 0.68 |
| | Minimum | 7.38 | 13.60 | 6.92 | 13.56 | 88.1 | 94.1 | 80.1 | 93.0 | 8.6 | 7.6 | 0.5 |
| | Maximum | 14.11 | 15.62 | 13.63 | 15.36 | 124.2 | 110.0 | 124.0 | 107.7 | 18.3 | 23.8 | 1.15 |
| | SD | 1.87 | 0.638 | 1.92 | 0.645 | 10.3 | 4.91 | 10.6 | 4.37 | 2.77 | 4.16 | 0.19 |
| | SE | 0.325 | 0.202 | 0.335 | 0.204 | 1.80 | 1.55 | 1.84 | 1.38 | 0.46 | 1.20 | 0.10 |
| | Lower Quartile | 8.74 | 14.30 | 8.66 | 14.04 | 95.1 | 102.7 | 92.3 | 103.4 | 17.7 | 19.1 | 1.1 |
| | Upper Quartile | 11.51 | 14.92 | 11.45 | 14.91 | 113.1 | 108.7 | 108.4 | 106.2 | 9.8 | 8.2 | 0.5 |
| | n | 33 | 10 | 33 | 10 | 33 | 10 | 33 | 10 | 36 | 12 | 12 |
| % Detections | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | - | - | - | |
| PFF | Mean | 10.60 | 14.49 | 10.53 | 14.29 | 106.7 | 104.7 | 105.4 | 103.3 | 27.4 | 21.0 | 0.84 |
| | Median | 10.53 | - | 10.51 | - | 109.6 | - | 108.5 | - | 26.6 | 19.6 | 0.84 |
| | Minimum | 8.09 | 13.66 | 8.06 | 13.70 | 92.4 | 100.9 | 91.2 | 101.2 | 13.0 | 15.8 | 0.81 |
| | Maximum | 13.83 | 15.31 | 13.68 | 14.87 | 122.5 | 108.6 | 121.2 | 105.5 | 39.0 | 27.5 | 0.87 |
| | SD | 2.19 | - | 2.21 | - | 12.4 | - | 12.9 | - | 8.32 | 5.97 | 0.04 |
| | SE | 0.732 | - | 0.736 | - | 4.15 | - | 4.30 | - | 2.77 | 3.40 | 0.00 |
| | Lower Quartile | 8.98 | - | 8.99 | - | 93.7 | - | 92.2 | - | 37.7 | 26.7 | 0.9 |
| | Upper Quartile | 12.03 | - | 12.10 | - | 118.2 | - | 116.7 | - | 16.2 | 16.2 | 0.8 |
| | n | 9 | 2 | 9 | 2 | 9 | 2 | 9 | 2 | 9 | 3 | 2 |
| % Detections | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | - | - | - | |

Notes:
 1. OW = Open-water season; IC = Ice-cover season.
 2. SD = standard deviation; SE = standard error; n = number of samples.

Table 3.2-3. 2008-2019 Off-system sites summary of thermal stratification and DO concentrations.

| Metric | Sampling Year | Surface or Bottom | MANIG | | | | EAGLE | | | |
|-------------------------------|---------------|-------------------|------------|------|------|-----------|------------|-----|-----|-----------|
| | | | Open-Water | | | Ice-Cover | Open-Water | | | Ice-Cover |
| | | | SP | SU | FA | WI | SP | SU | FA | WI |
| Thermal Stratification | 2008 | | 2008 | 2008 | 2008 | No | | | | |
| | 2009 | | No | 2009 | 2009 | No | | | | |
| | 2010 | | 2010 | 2010 | 2010 | No | 2010 | No | No | No |
| | 2011 | | No | 2011 | 2011 | No | | | | |
| | 2012 | | No | 2012 | No | No | | | | |
| | 2013 | | No | 2013 | 2013 | No | No | No | No | No |
| | 2014 | | 2014 | 2014 | 2014 | No | | | | |
| | 2015 | | 2015 | 2015 | 2015 | No | | | | |
| | 2016 | | No | 2016 | 2016 | No | No | No | No | No |
| | 2017 | | No | 2017 | 2017 | No | | | | |
| 2018 | | 2018 | 2018 | 2018 | No | | | | | |
| 2019 | | No | 2019 | 2019 | No | No | No | No | ND | |
| DO met MWQSOGs PAL objectives | 2008 | Surface | Yes | Yes | Yes | Yes | | | | |
| | | Bottom | ND | Yes | 2008 | Yes | | | | |
| | 2009 | Surface | Yes | ND | Yes | Yes | | | | |
| | | Bottom | ND | ND | Yes | 2009 | | | | |
| | 2010 | Surface | Yes | Yes | Yes | ND | Yes | Yes | Yes | ND |
| | | Bottom | ND | Yes | 2010 | ND | ND | Yes | Yes | ND |
| | 2011 | Surface | Yes | Yes | Yes | Yes | | | | |
| | | Bottom | Yes | Yes | 2011 | Yes | | | | |
| | 2012 | Surface | Yes | Yes | Yes | Yes | | | | |
| | | Bottom | Yes | 2012 | Yes | Yes | | | | |
| | 2013 | Surface | ND | Yes | Yes | Yes | ND | Yes | Yes | Yes |
| | | Bottom | ND | ND | ND | Yes | ND | Yes | Yes | Yes |
| | 2014 | Surface | Yes | Yes | Yes | ND | | | | |
| | | Bottom | Yes | ND | ND | ND | | | | |
| | 2015 | Surface | Yes | Yes | Yes | Yes | | | | |
| | | Bottom | Yes | Yes | 2015 | Yes | | | | |
| | 2016 | Surface | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes |
| | | Bottom | Yes | Yes | 2016 | Yes | Yes | Yes | Yes | Yes |
| 2017 | Surface | Yes | Yes | Yes | Yes | | | | | |
| | Bottom | Yes | 2017 | 2017 | Yes | | | | | |
| 2018 | Surface | Yes | Yes | Yes | Yes | | | | | |
| | Bottom | Yes | Yes | 2018 | 2018 | | | | | |
| 2019 | Surface | Yes | Yes | Yes | Yes | Yes | Yes | Yes | ND | |
| | Bottom | Yes | Yes | 2019 | 2019 | Yes | Yes | Yes | ND | |

Notes:

1. SP = spring; SU = summer; FA = fall; WI = winter; DO = dissolved oxygen; MWQSOG = Manitoba Water Quality Standards, Objectives, and Guidelines; PAL = Protection of Aquatic Life.
2. ND = No data.
4. DO concentrations were compared to the most stringent MWQSOGs instantaneous minimum PAL objectives for each season; i.e., 5 mg/L for cool-water early life for the open-water season and 8 mg/L for cold-water early life the ice-cover season.
5. Cells with a year indicated denote instances of stratification or non-compliance with MWQSOGs instantaneous minimum objectives.
6. = Sampling did not occur.

Table 3.2-4. 2008-2019 Off-system sites DO, water depth, and ice thickness summary statistics.

| Site | Statistic | Dissolved Oxygen | | | | | | | | Water Depth at Site (m) | | Ice Thickness at Site (m) |
|--------------|----------------|---------------------|-------|--------------------|-------|-----------------------------|-------|----------------------------|-------|-------------------------|------|---------------------------|
| | | DO - Surface (mg/L) | | DO - Bottom (mg/L) | | DO Saturation - Surface (%) | | DO Saturation - Bottom (%) | | OW | IC | IC |
| | | OW | IC | OW | IC | OW | IC | OW | IC | OW | IC | IC |
| MANIG | Mean | 9.86 | 13.71 | 6.71 | 7.84 | 100.2 | 100.3 | 57.7 | 62.2 | 21.2 | 20.1 | 0.65 |
| | Median | 9.22 | 13.87 | 6.45 | 8.48 | 100.0 | 102.2 | 56.2 | 66.3 | 21.0 | 20.6 | 0.62 |
| | Minimum | 7.29 | 11.87 | 0.12 | 4.67 | 83.6 | 85.4 | 0.8 | 35.3 | 19.1 | 13.8 | 0.49 |
| | Maximum | 12.93 | 15.16 | 13.39 | 10.82 | 117.4 | 112.3 | 111.6 | 85.0 | 23.8 | 24.0 | 0.9 |
| | SD | 1.55 | 1.06 | 4.13 | 2.20 | 9.20 | 8.65 | 34.4 | 18.8 | 1.00 | 2.45 | 0.13 |
| | SE | 0.266 | 0.334 | 0.780 | 0.696 | 1.58 | 2.74 | 6.50 | 5.93 | 0.17 | 0.70 | 0.00 |
| | Lower Quartile | 8.73 | 13.02 | 3.47 | 5.86 | 94.6 | 96.9 | 31.3 | 44.3 | 23.2 | 22.6 | 0.9 |
| | Upper Quartile | 11.47 | 14.21 | 11.01 | 9.22 | 106.0 | 106.5 | 93.4 | 74.6 | 19.9 | 16.1 | 0.5 |
| | n | 34 | 10 | 28 | 10 | 34 | 10 | 28 | 10 | 36 | 12 | 12 |
| % Detections | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | - | - | - | |
| EAGLE | Mean | 10.04 | 13.92 | 9.79 | 13.76 | 102.7 | 104.1 | 100.3 | 100.0 | 19.8 | 15.7 | 0.47 |
| | Median | 9.45 | - | 9.43 | - | 103.7 | - | 97.9 | - | 20.4 | 16.0 | 0.47 |
| | Minimum | 8.23 | 13.65 | 8.06 | 13.32 | 90.5 | 101.5 | 90.0 | 99.1 | 14.9 | 10.0 | 0.42 |
| | Maximum | 13.24 | 14.19 | 13.21 | 14.20 | 118.3 | 106.7 | 114.7 | 100.9 | 21.6 | 21.2 | 0.51 |
| | SD | 1.74 | - | 1.90 | - | 8.46 | - | 8.99 | - | 1.89 | 5.60 | 0.05 |
| | SE | 0.524 | - | 0.600 | - | 2.55 | - | 2.84 | - | 0.54 | 3.20 | 0.00 |
| | Lower Quartile | 8.85 | - | 8.41 | - | 95.9 | - | 94.3 | - | 21.6 | 20.7 | 0.5 |
| | Upper Quartile | 10.77 | - | 9.75 | - | 105.9 | - | 104.8 | - | 16.8 | 10.6 | 0.4 |
| | n | 11 | 2 | 10 | 2 | 11 | 2 | 10 | 2 | 12 | 3 | 3 |
| % Detections | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | - | - | - | |

- Notes:**
1. OW = Open-water season; IC = Ice-cover season.
 2. SD = standard deviation; SE = standard error; n = number of samples.

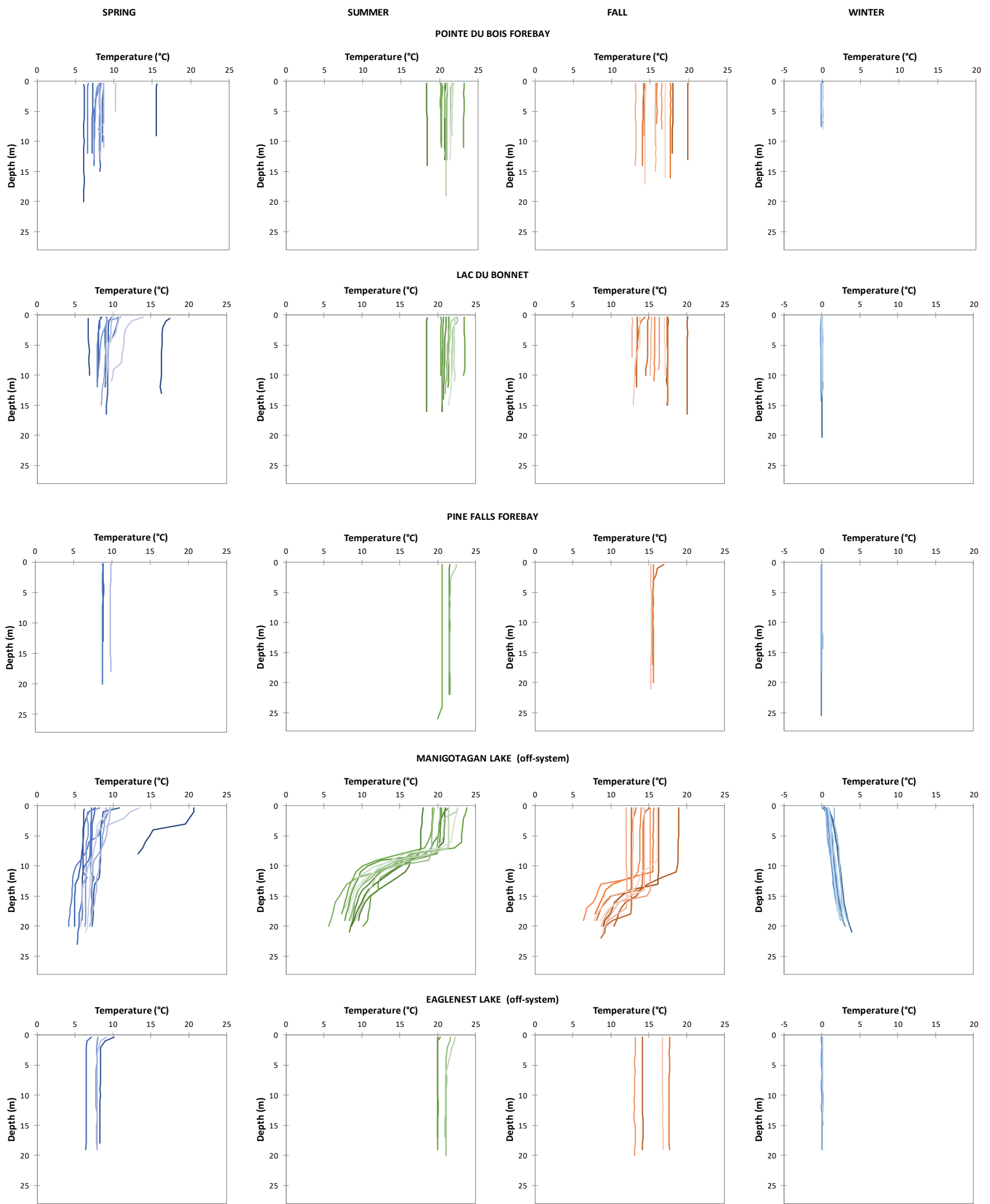


Figure 3.2-1. 2008-2019 On-system and off-system water temperature depth profiles.

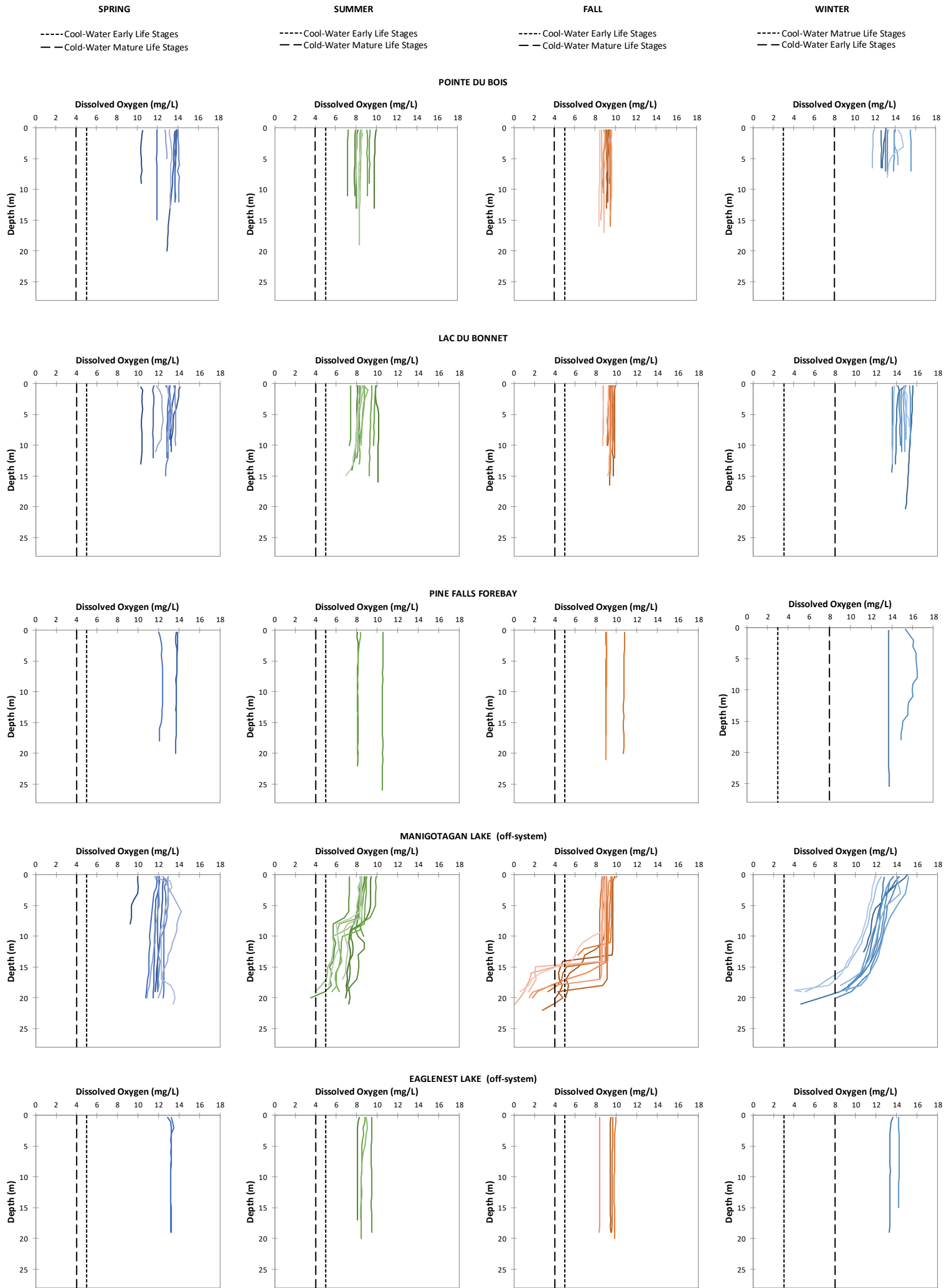


Figure 3.2-2. 2008-2019 On-system and off-system dissolved oxygen depth profiles and comparison to instantaneous minimum objectives for the protection of aquatic life.

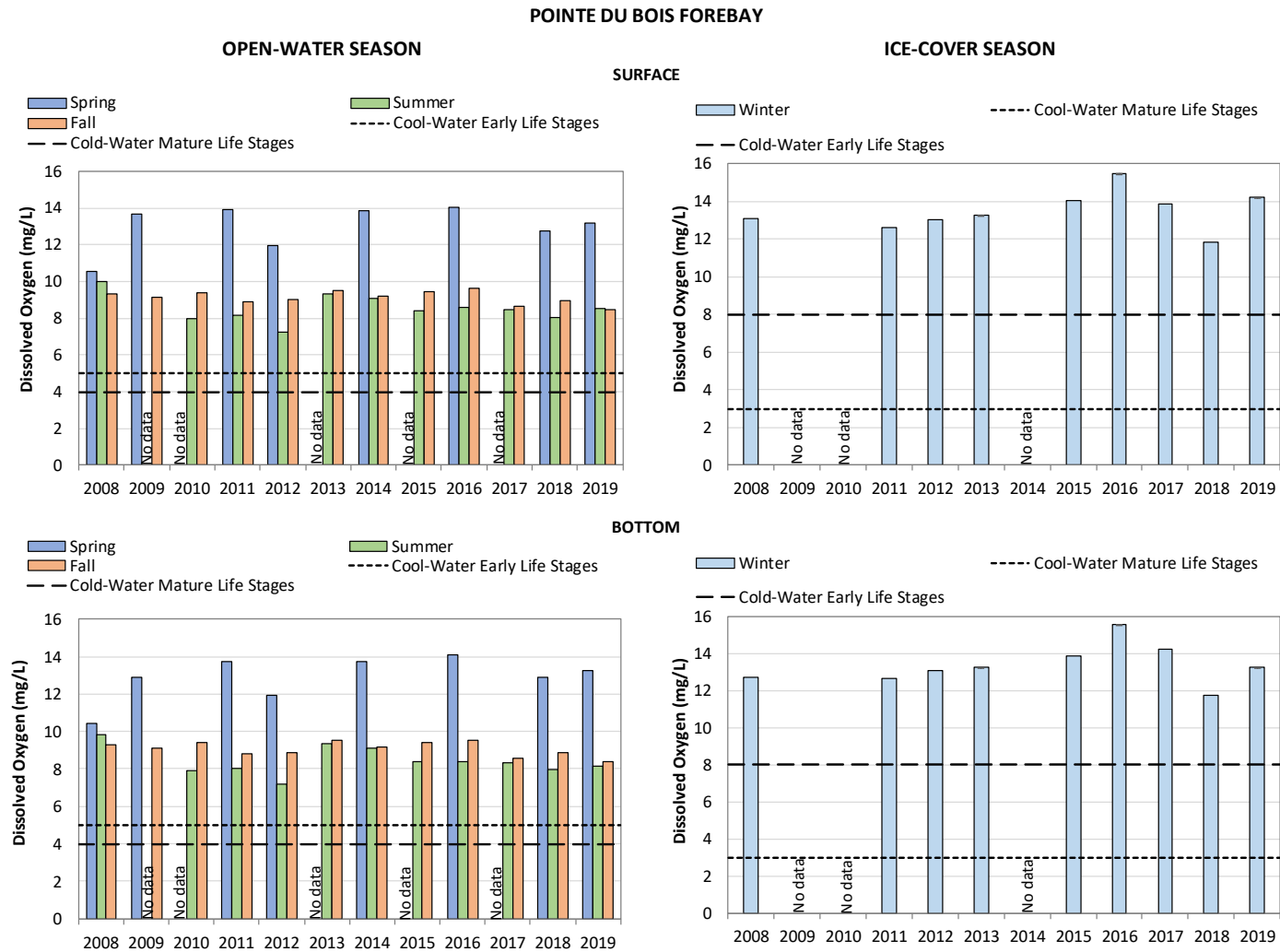


Figure 3.2-3. 2008-2019 Pointe du Bois Forebay surface and bottom dissolved oxygen concentrations with comparison to instantaneous minimum objectives for the protection of aquatic life.

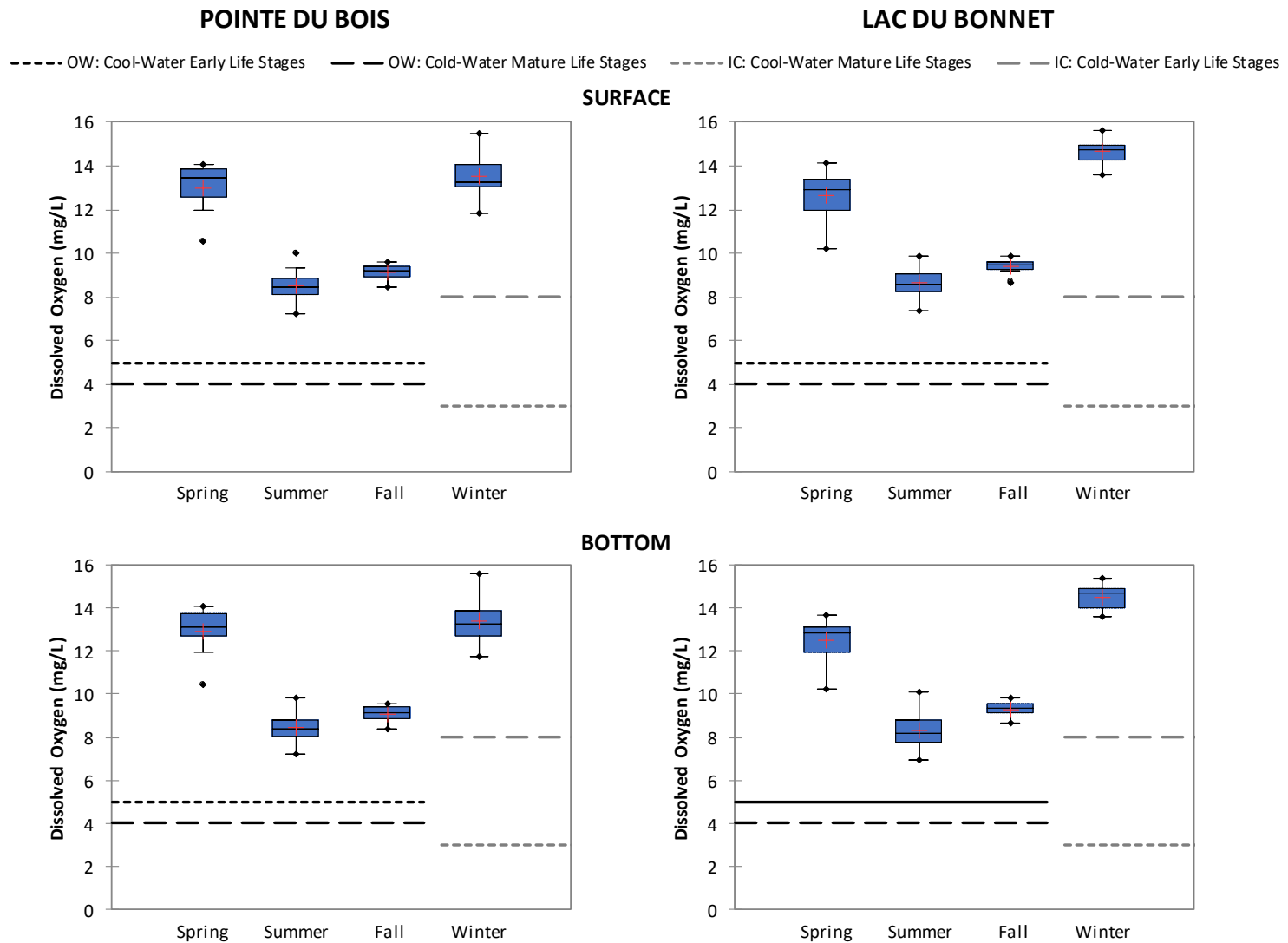


Figure 3.2-4. 2008-2019 On-system seasonal surface and bottom dissolved oxygen concentrations with comparison to instantaneous minimum objectives for the protection of aquatic life.

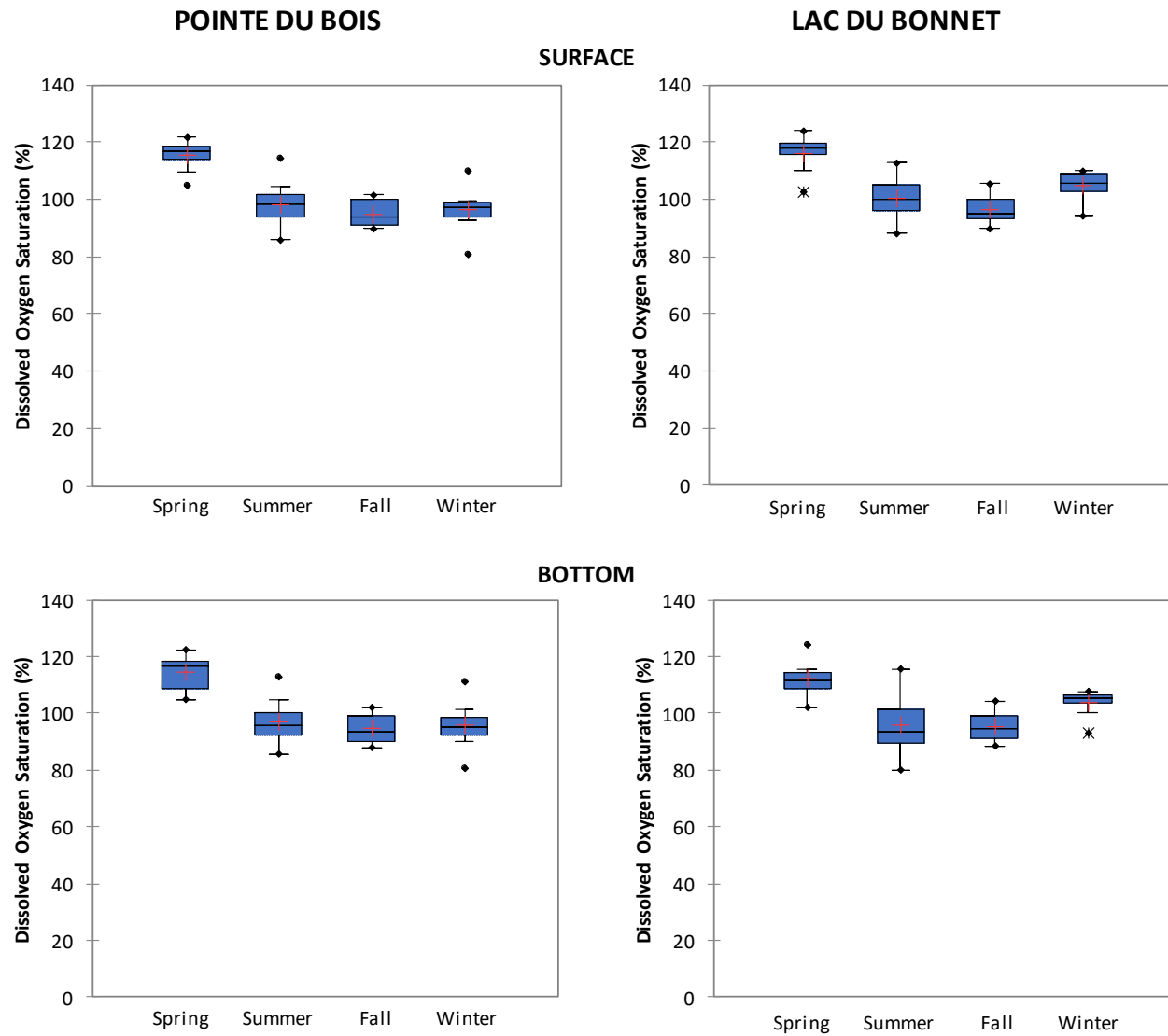


Figure 3.2-5. 2008-2019 On-system seasonal surface and bottom dissolved oxygen saturation.

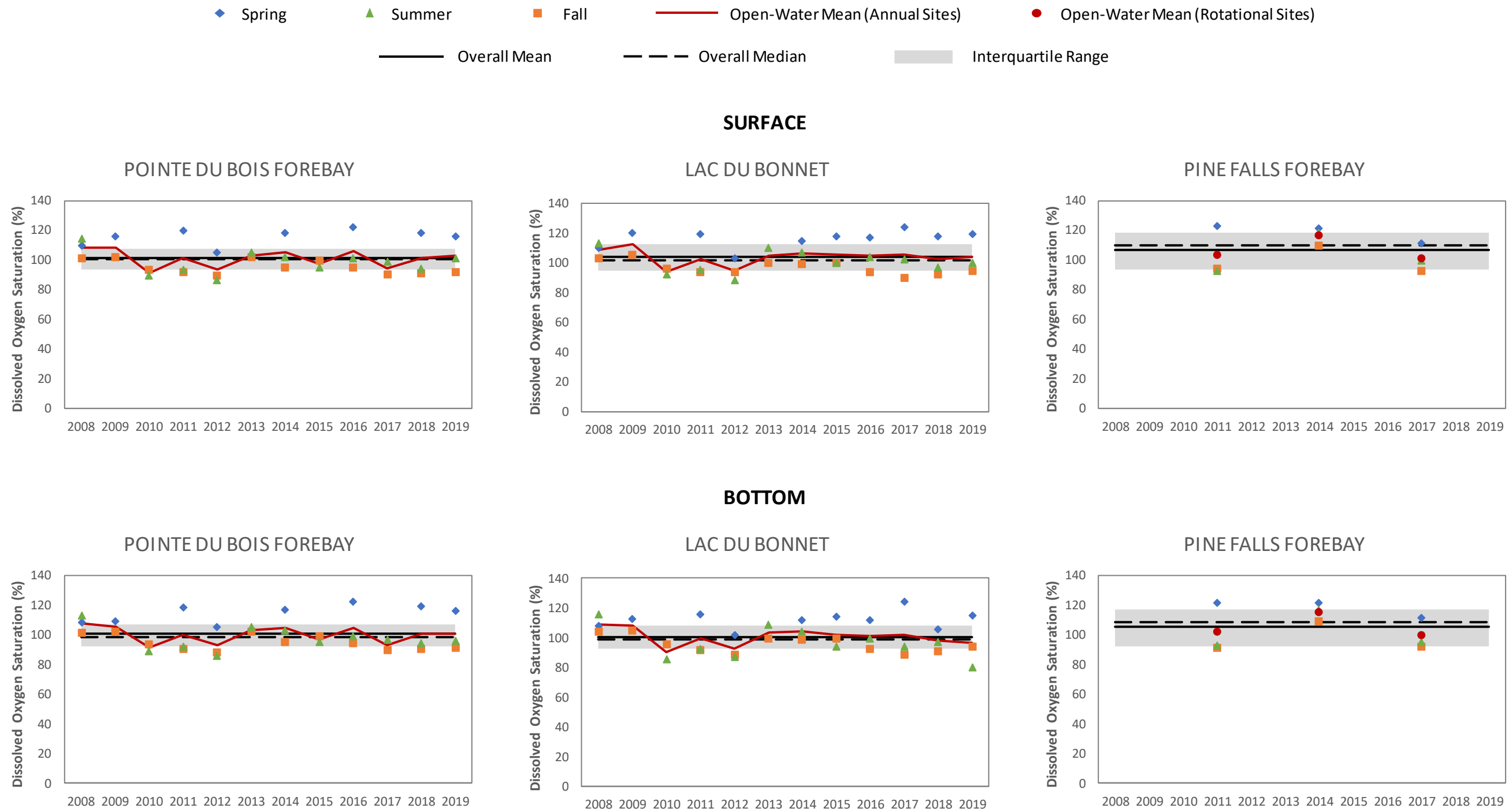


Figure 3.2-6. 2008-2019 On-system open-water season surface and bottom dissolved oxygen saturation.

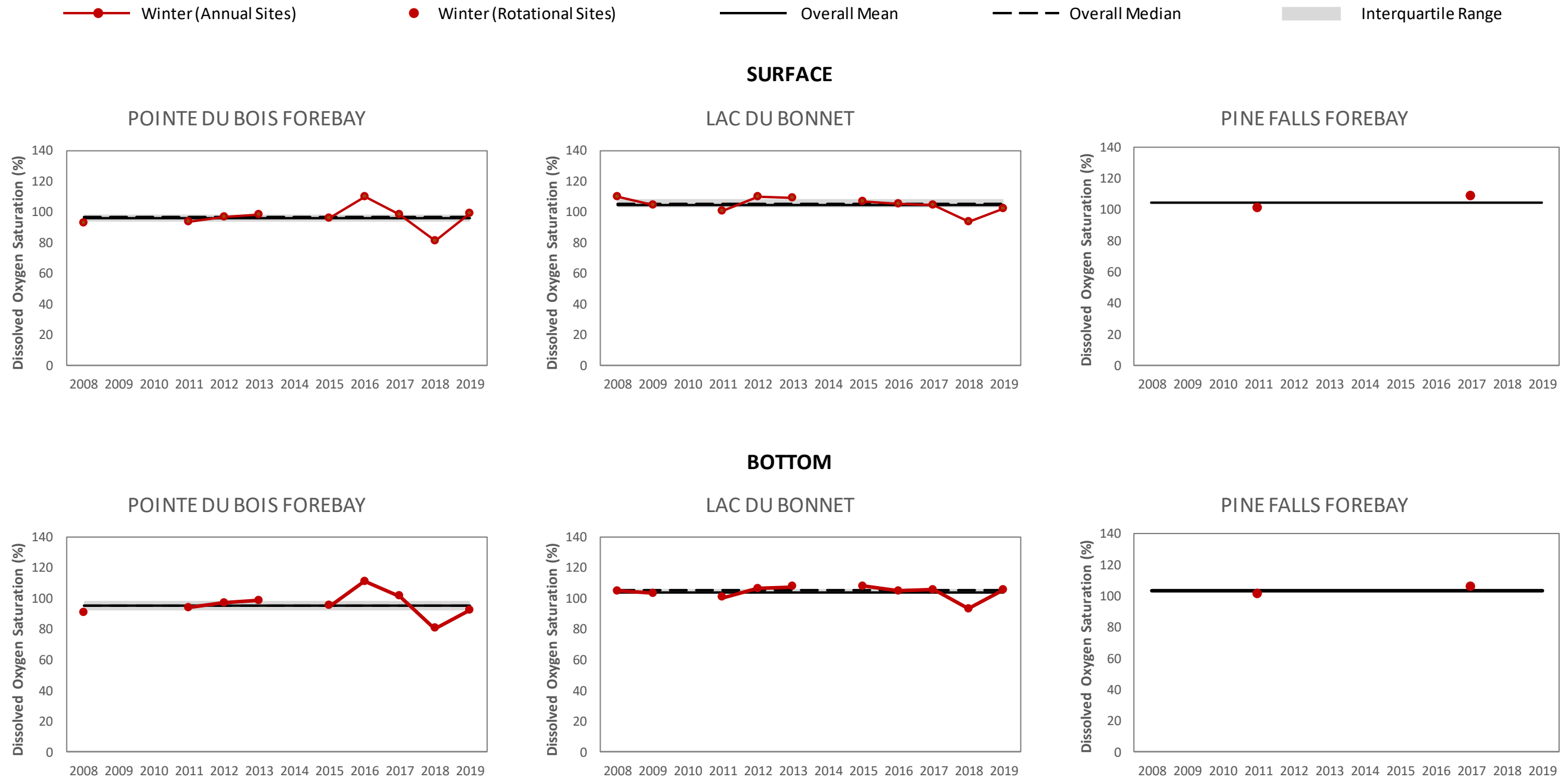


Figure 3.2-7. 2008-2019 On-system ice-cover season surface and bottom dissolved oxygen saturation.

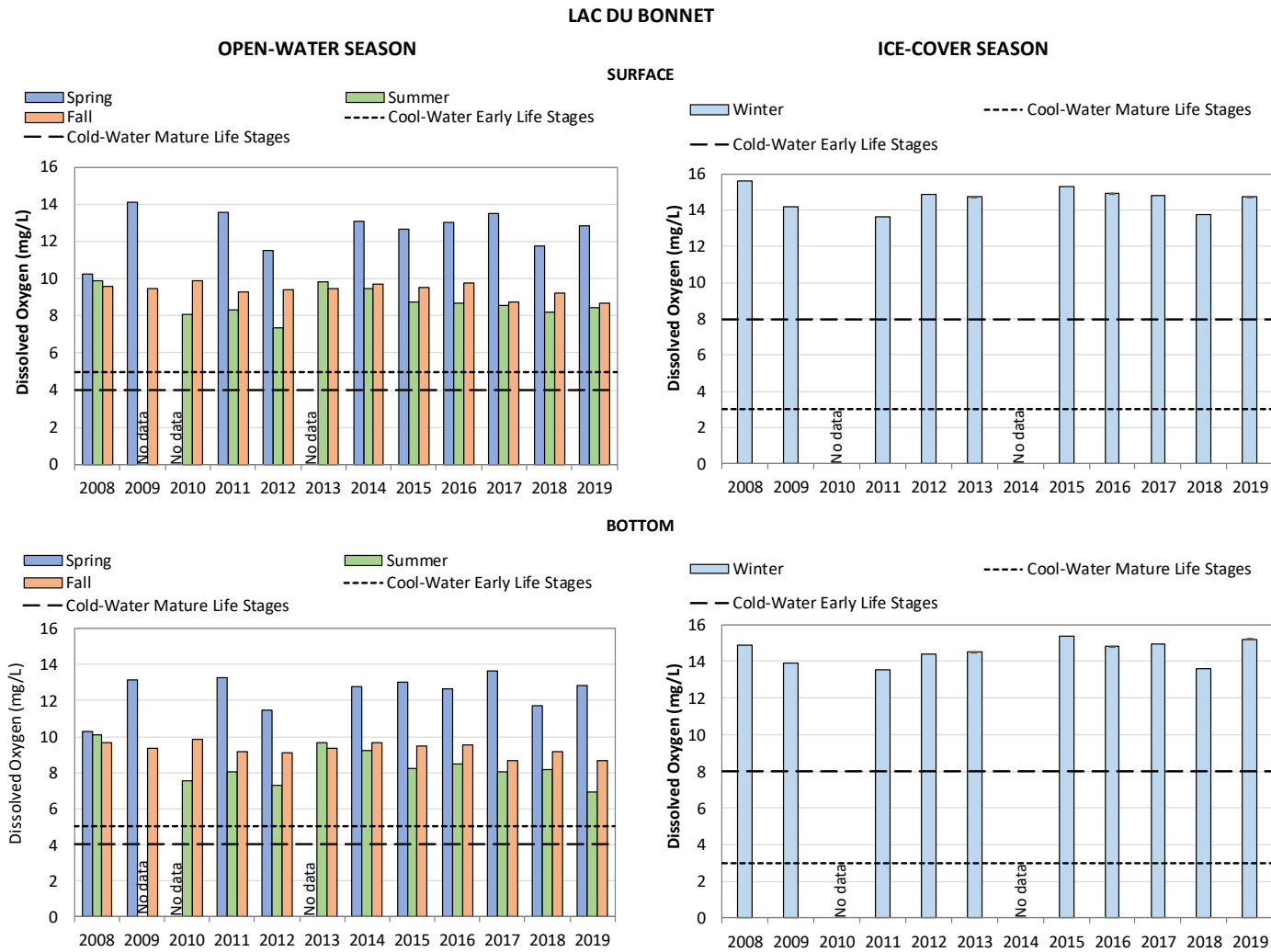


Figure 3.2-8. 2008-2019 Lac du Bonnet surface and bottom dissolved oxygen concentrations with comparison to instantaneous minimum objectives for the protection of aquatic life.

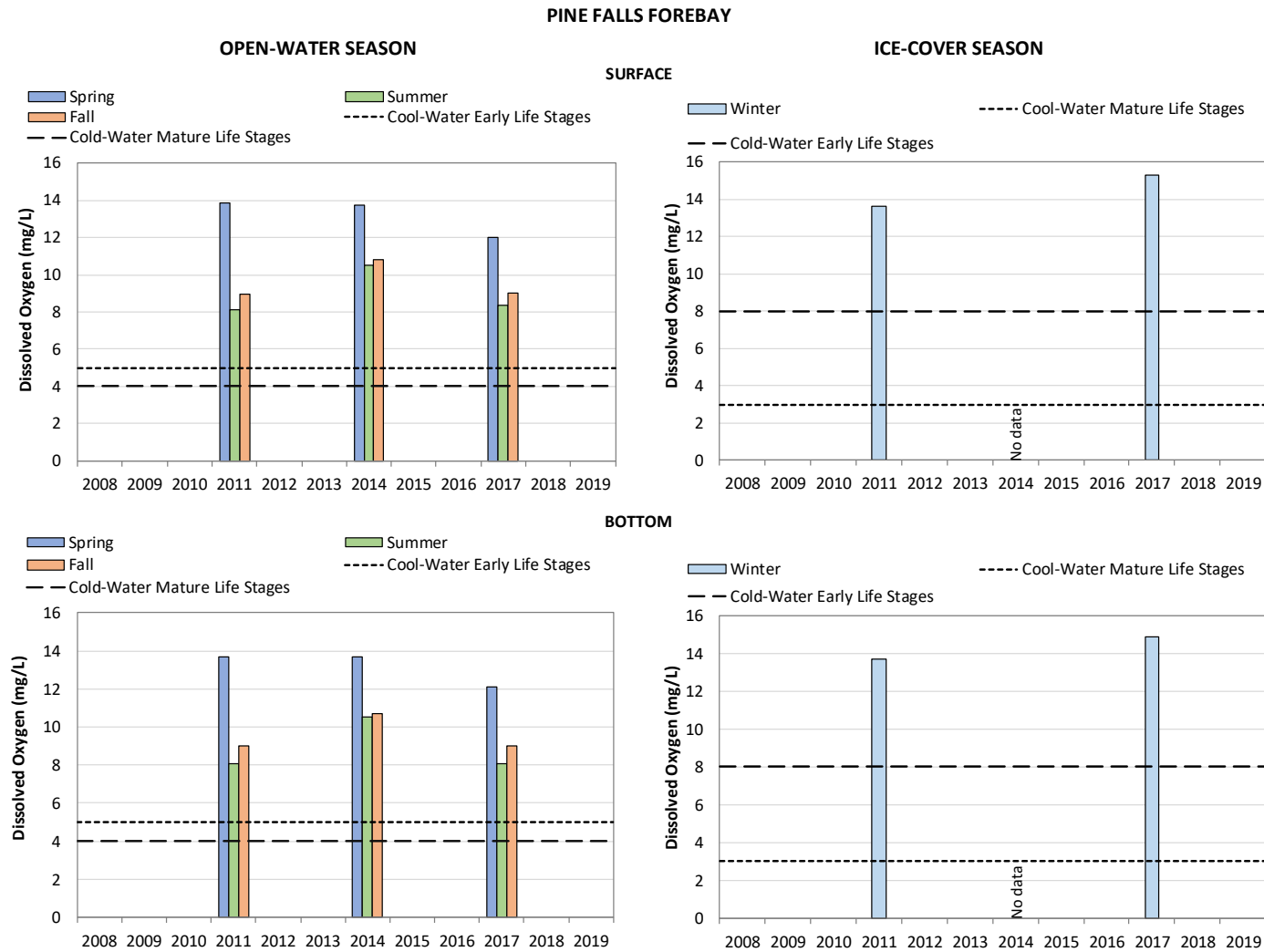


Figure 3.2-9. 2008-2019 Pine Falls Forebay surface and bottom dissolved oxygen concentrations with comparison to instantaneous minimum objectives for the protection of aquatic life.

MANIGOTAGAN LAKE

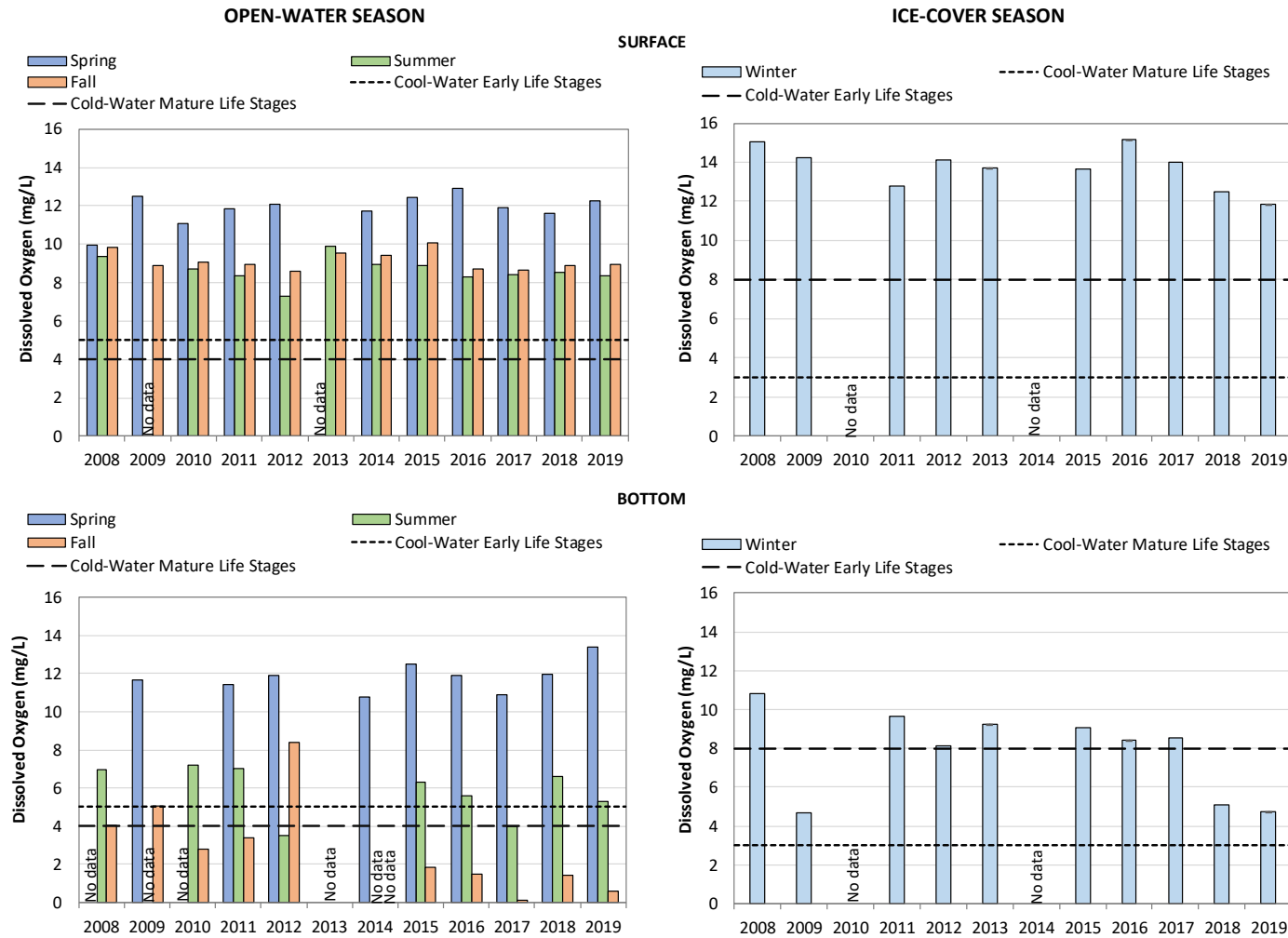


Figure 3.2-10. 2008-2019 Manigotagan Lake surface and bottom dissolved oxygen concentrations with comparison to instantaneous minimum objectives for the protection of aquatic life.

MANIGOTAGAN LAKE

----- OW: Cool-Water Early Life Stages — — — OW: Cold-Water Mature Life Stages
 ----- IC: Cool-Water Mature Life Stages — — — IC: Cold-Water Early Life Stages

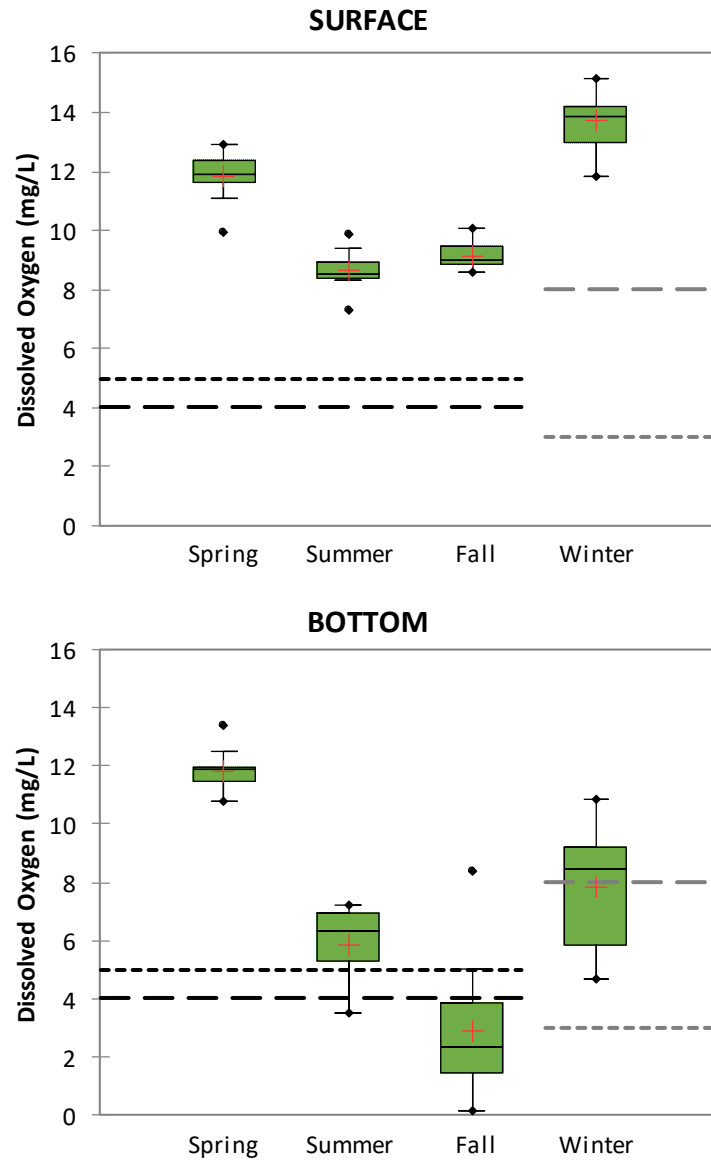
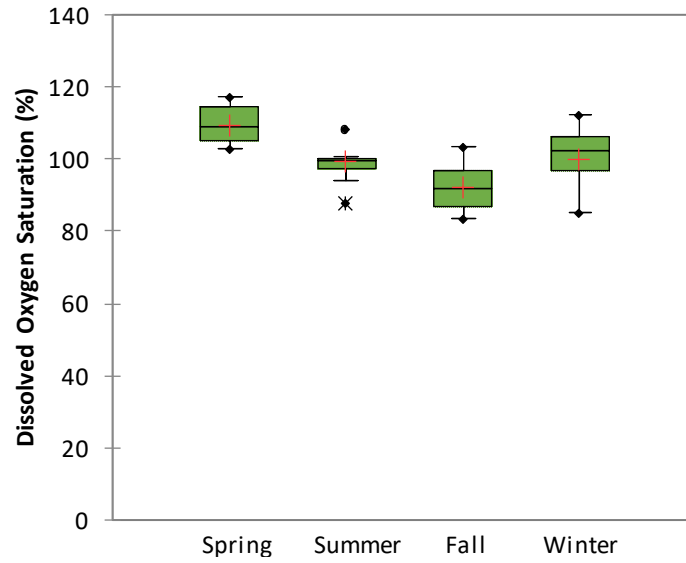


Figure 3.2-11. 2008-2019 Off-system seasonal surface and bottom dissolved oxygen concentrations with comparison to instantaneous minimum objectives for the protection of aquatic life.

MANIGOTAGAN LAKE
SURFACE



BOTTOM

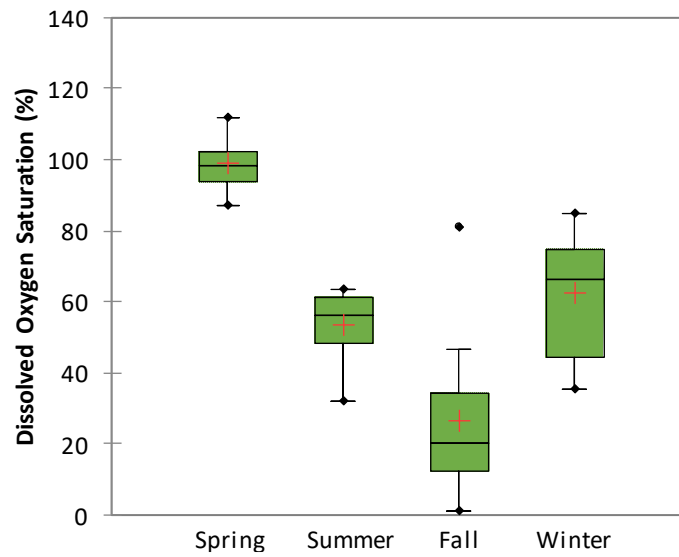
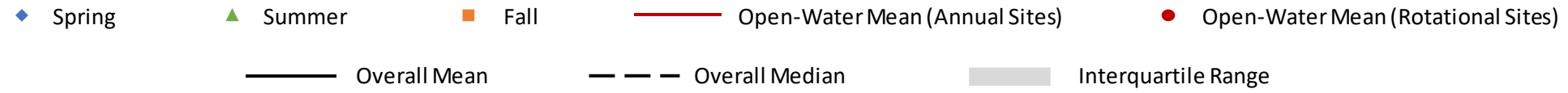
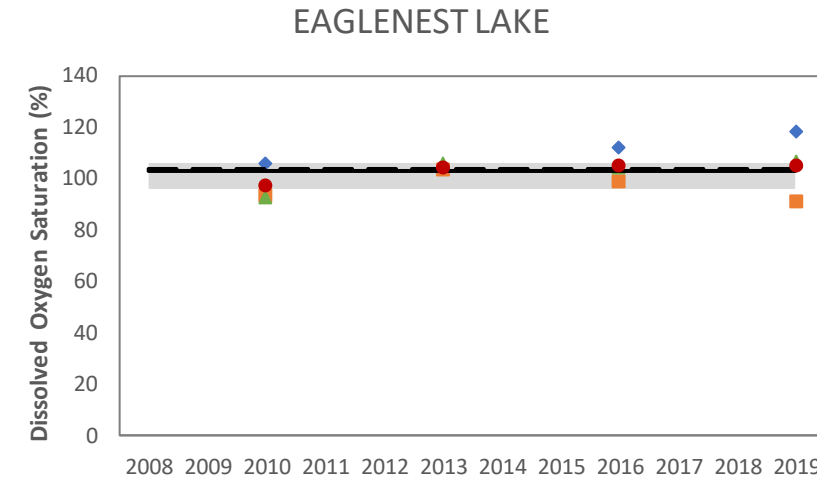
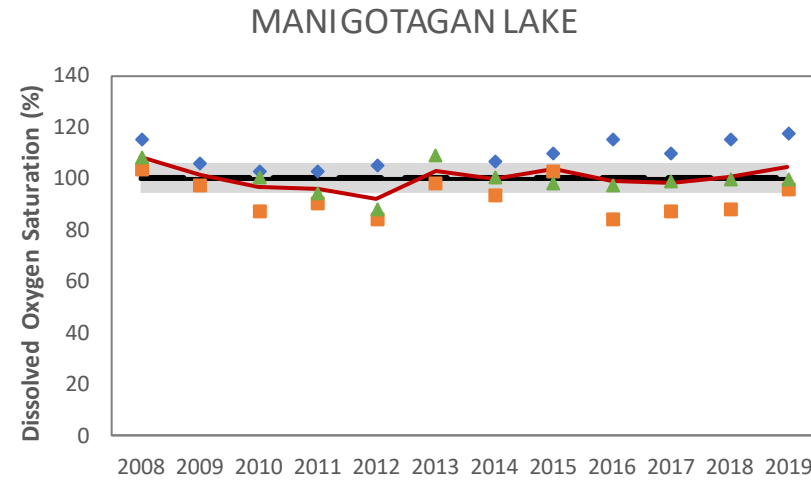


Figure 3.2-12. 2008-2019 Off-system seasonal surface and bottom dissolved oxygen saturation.



SURFACE



BOTTOM

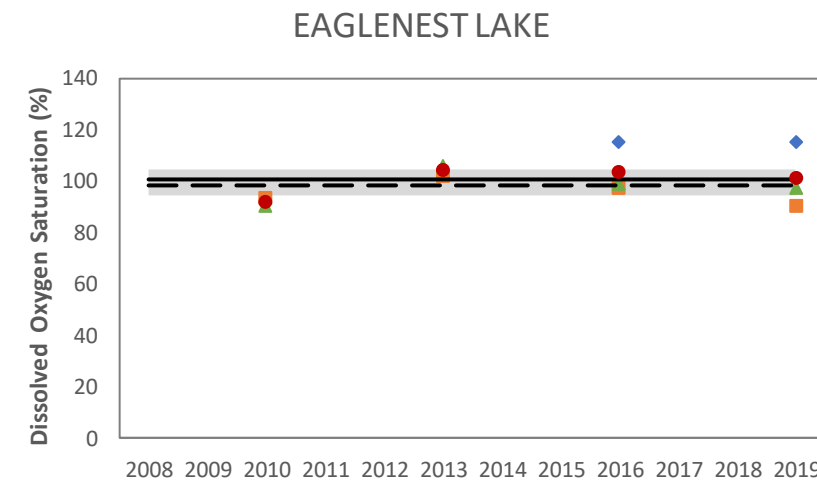
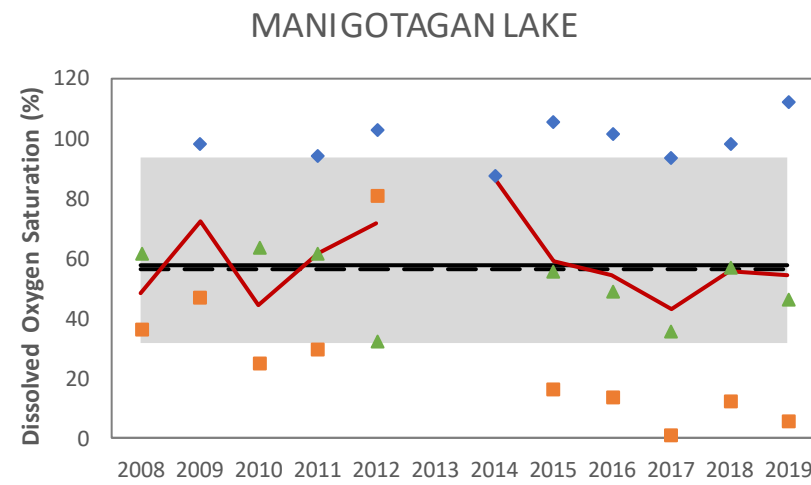


Figure 3.2-13. 2008-2019 Off-system open-water season surface and bottom dissolved oxygen saturation.

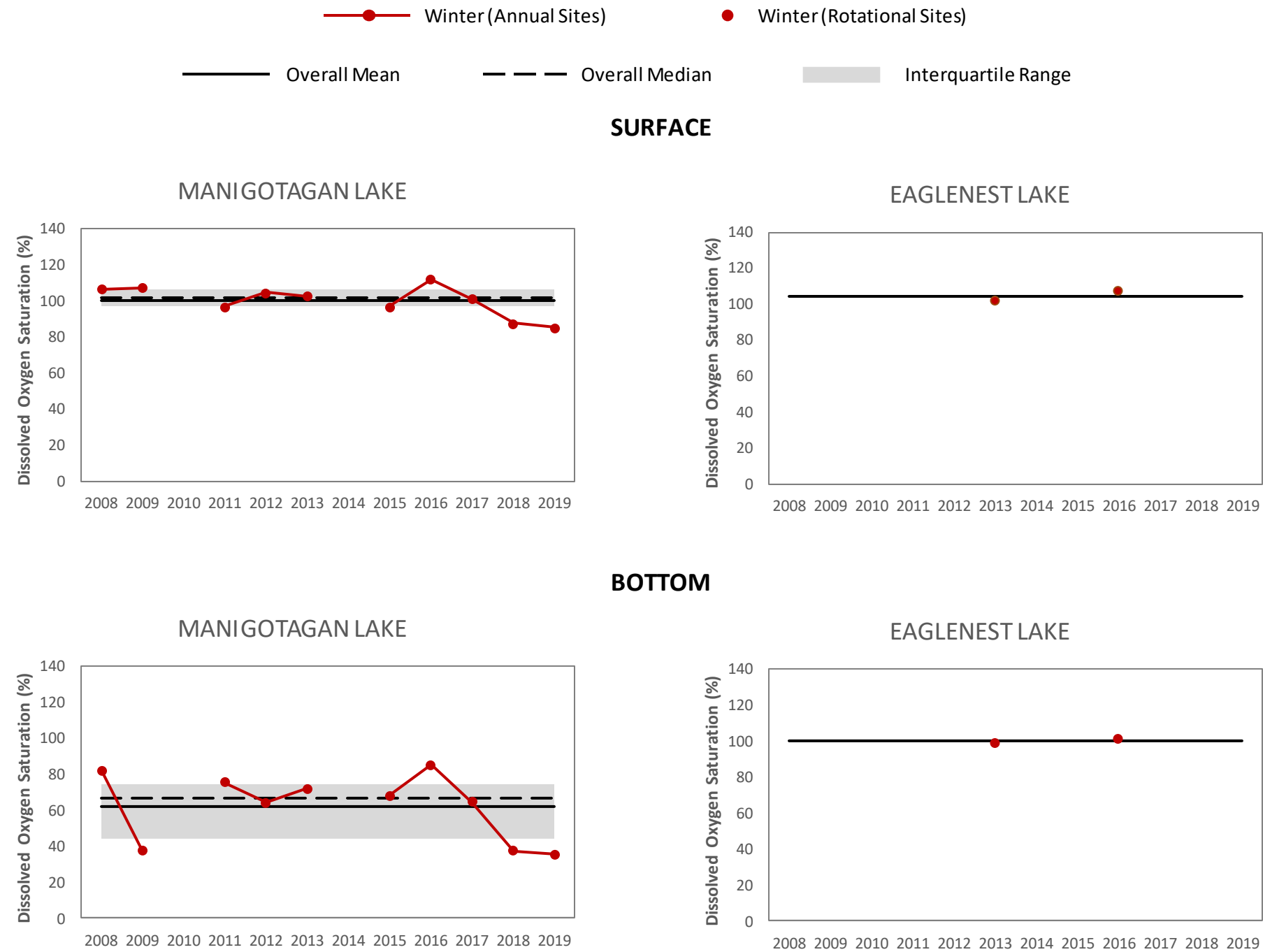


Figure 3.2-14. 2008-2019 Off-system ice-cover season surface and bottom dissolved oxygen saturation.

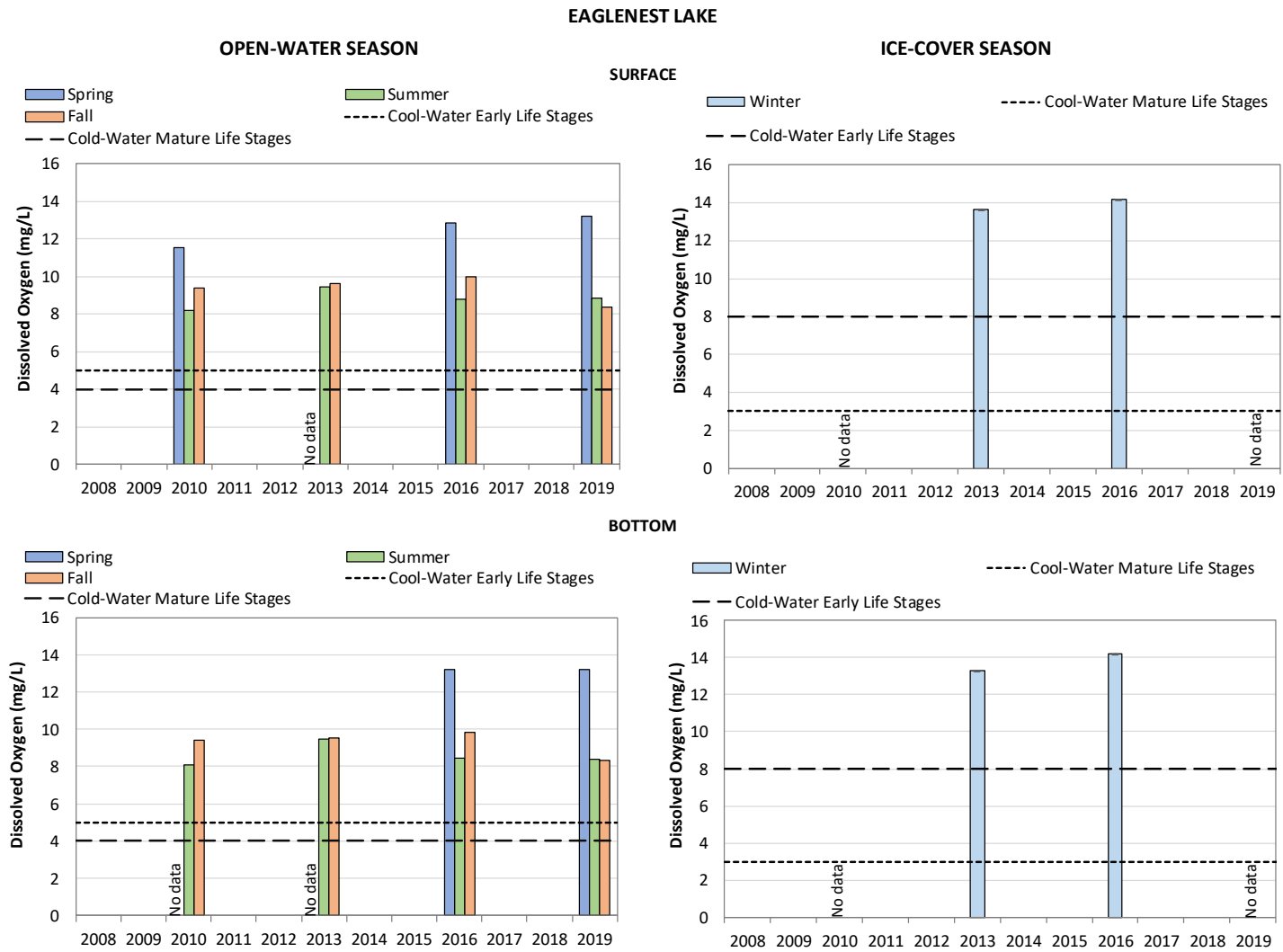


Figure 3.2-15. 2008-2019 Eaglenest Lake surface and bottom dissolved oxygen concentrations with comparison to instantaneous minimum objectives for the protection of aquatic life.

3.3 WATER CLARITY

3.3.1 SECCHI DISK DEPTH

3.3.1.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

Secchi disk depth in the Pointe du Bois Forebay ranged from 0.70 to 2.35 m during the open-water season. The mean and median measurements for the 12 years of monitoring were 1.64 and 1.65 m, respectively. Mean annual Secchi disk depths ranged from 1.21 to 2.28 m and were within the IQR (1.36 to 1.90 m) in ten of the 12 years. Mean Secchi disk depths were below the IQR in 2012 and above the IQR in 2019 (Table 3.3-1 and Figure 3.3-1).

No clear seasonality was observed for Secchi disk depth in the Pointe du Bois Forebay over the 12 years of monitoring. However, the largest mean Secchi disk depth occurred in spring (1.76 m) and the smallest in fall (1.54 m; Figure 3.3-2).

Lac du Bonnet

Secchi disk depth in Lac du Bonnet ranged from 0.55 to 2.10 m during the open-water season. The mean and median measurements for the 12 years of monitoring were 1.26 and 1.20 m, respectively. Mean annual Secchi disk depths ranged from 0.88 to 1.55 m and were within the IQR (1.00 to 1.51 m) in nine of the 12 years. Mean Secchi disk depths were below the IQR in 2009 and above the IQR in 2018, and 2019 (Table 3.3-1 and Figure 3.3-1).

On average, Secchi disk depths were higher in the spring (1.55 m) than in the summer and fall (1.14 and 1.09 m, respectively; Figure 3.3-2).

ROTATIONAL SITES

Pine Falls Forebay

Secchi disk depth in the Pine Falls Forebay ranged from 0.80 to 1.65 m during the open-water season. The mean was 1.14 m, the median was 1.05 m, and the IQR was 0.95 to 1.30 m for the three years of monitoring. Mean annual Secchi disk depths ranged from 0.93 to 1.27 m and were within the IQR in 2011 and 2017 but below the IQR in 2014 (Table 3.3-1 and Figure 3.3-1).

3.3.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

Secchi disk depth in Manigotagan Lake ranged from 1.25 to 4.90 m during the open-water season. The mean and median for the 12 years of monitoring were 2.15 and 2.03 m, respectively. Mean annual Secchi disk depths ranged from 1.60 to 3.17 m and were within the IQR (1.69 to 2.33 m) in seven of the 12 years. Mean Secchi disk depths were below the IQR in 2008, 2009, and 2010 and above the IQR in 2017, and 2019 (Table 3.3-2 and Figure 3.3-3).

No clear seasonality was observed for Secchi disk depth in Manigotagan Lake over the 12 years of monitoring. However, the largest mean Secchi disk depth occurred in summer (2.45 m) and the smallest in spring (1.86 m; Figure 3.3-4).

ROTATIONAL SITES

Eaglenest Lake

Secchi disk depth in Eaglenest Lake ranged from 1.10 to 2.60 m during the open-water season. The mean and median measurements for the four years of monitoring were 1.89 and 1.92 m, respectively, and the IQR was 1.73 to 2.14 m. Mean annual Secchi disk depths ranged from 1.59 to 2.43 m and were within the IQR in 2013 and 2016 but were below the IQR in 2010 and above the IQR in 2019 (Table 3.3-2 and Figure 3.3-3).

Table 3.3-1. 2008-2019 On-system sites water clarity summary statistics.

| Site | Statistic | Secchi Disk Depth (m) | | Turbidity (NTU) | | TSS (mg/L) | |
|------|----------------|-----------------------|----|-----------------|-------|------------|------|
| | | OW | IC | OW | IC | OW | IC |
| PDB | Mean | 1.64 | - | 3.72 | 3.65 | 3.8 | <2.0 |
| | Median | 1.65 | - | 3.57 | 3.22 | 3.6 | <2.0 |
| | Minimum | 0.70 | - | 2.38 | 2.83 | <2.0 | <2.0 |
| | Maximum | 2.35 | - | 6.10 | 4.88 | 8.0 | 4.0 |
| | SD | 0.398 | - | 0.994 | 0.749 | 1.51 | 0.92 |
| | SE | 0.067 | - | 0.166 | 0.226 | 0.25 | 0.28 |
| | Lower Quartile | 1.36 | - | 2.93 | 3.11 | 2.8 | <2.0 |
| | Upper Quartile | 1.90 | - | 4.27 | 4.15 | 4.7 | <2.0 |
| | n | 35 | - | 36 | 11 | 36 | 11 |
| | % Detections | 100 | - | 100 | 100 | 92 | 18 |
| LDB | Mean | 1.26 | - | 5.22 | 3.96 | 5.4 | <2.0 |
| | Median | 1.20 | - | 5.30 | 3.90 | 5.2 | <2.0 |
| | Minimum | 0.55 | - | 1.69 | 3.37 | <2.0 | <2.0 |
| | Maximum | 2.10 | - | 9.39 | 5.48 | 9.6 | 3.2 |
| | SD | 0.353 | - | 1.65 | 0.61 | 2.13 | 0.75 |
| | SE | 0.059 | - | 0.275 | 0.177 | 0.35 | 0.22 |
| | Lower Quartile | 1.00 | - | 4.17 | 3.49 | 3.9 | <2.0 |
| | Upper Quartile | 1.51 | - | 6.11 | 4.13 | 6.6 | 2.0 |
| | n | 36 | - | 36 | 12 | 36 | 12 |
| | % Detections | 100 | - | 100 | 100 | 97 | 33 |
| PFF | Mean | 1.14 | - | 6.35 | 4.43 | 7.0 | <2.0 |
| | Median | 1.05 | - | 6.35 | - | 6.4 | - |
| | Minimum | 0.80 | - | 3.46 | 3.27 | 2.8 | <2.0 |
| | Maximum | 1.65 | - | 9.80 | 5.22 | 11.7 | 2.0 |
| | SD | 0.261 | - | 1.98 | 1.02 | 3.55 | 0.58 |
| | SE | 0.087 | - | 0.661 | 0.592 | 1.18 | 0.33 |
| | Lower Quartile | 0.95 | - | 4.98 | - | 4.1 | - |
| | Upper Quartile | 1.30 | - | 7.69 | - | 11.2 | - |
| | n | 9 | - | 9 | 3 | 9 | 3 |
| | % Detections | 100 | - | 100 | 100 | 100 | 33 |

Notes:

1. OW = Open-water season; IC = Ice-cover season.
2. SD = standard deviation; SE = standard error; n = number of samples.

Table 3.3-2. 2008-2019 Off-system sites water clarity metric summary statistics.

| Site | Statistic | Secchi Disk Depth (m) | | Turbidity (NTU) | | TSS (mg/L) | |
|-------|----------------|-----------------------|----|-----------------|-------|------------|------|
| | | OW | IC | OW | IC | OW | IC |
| MANIG | Mean | 2.15 | - | 1.94 | 1.36 | <2.0 | <2.0 |
| | Median | 2.03 | - | 1.99 | 1.39 | <2.0 | <2.0 |
| | Minimum | 1.25 | - | 0.55 | 0.75 | <2.0 | <2.0 |
| | Maximum | 4.90 | - | 3.07 | 2.21 | 3.3 | <2.0 |
| | SD | 0.709 | - | 0.570 | 0.375 | 0.75 | - |
| | SE | 0.118 | - | 0.095 | 0.108 | 0.12 | - |
| | Lower Quartile | 1.69 | - | 1.58 | 1.05 | <2.0 | <2.0 |
| | Upper Quartile | 2.33 | - | 2.36 | 1.55 | 2.2 | <2.0 |
| | n | 36 | - | 36 | 12 | 36 | 12 |
| | % Detections | 100 | - | 100 | 100 | 31 | 0 |
| EAGLE | Mean | 1.89 | - | 2.86 | 3.66 | 2.4 | <2.0 |
| | Median | 1.92 | - | 2.86 | - | 2.5 | - |
| | Minimum | 1.10 | - | 1.86 | 3.19 | <2.0 | <2.0 |
| | Maximum | 2.60 | - | 4.86 | 4.55 | 4.0 | 3.2 |
| | SD | 0.459 | - | 0.766 | 0.766 | 1.15 | 1.27 |
| | SE | 0.133 | - | 0.221 | 0.442 | 0.33 | 0.73 |
| | Lower Quartile | 1.73 | - | 2.38 | - | <2.0 | - |
| | Upper Quartile | 2.14 | - | 3.06 | - | 3.3 | - |
| | n | 12 | - | 12 | 3 | 12 | 3 |
| | % Detections | 100 | - | 100 | 100 | 67 | 33 |

Notes:

1. OW = Open-water season; IC = Ice-cover season.
2. SD = standard deviation; SE = standard error; n = number of samples.

OPEN-WATER SEASON

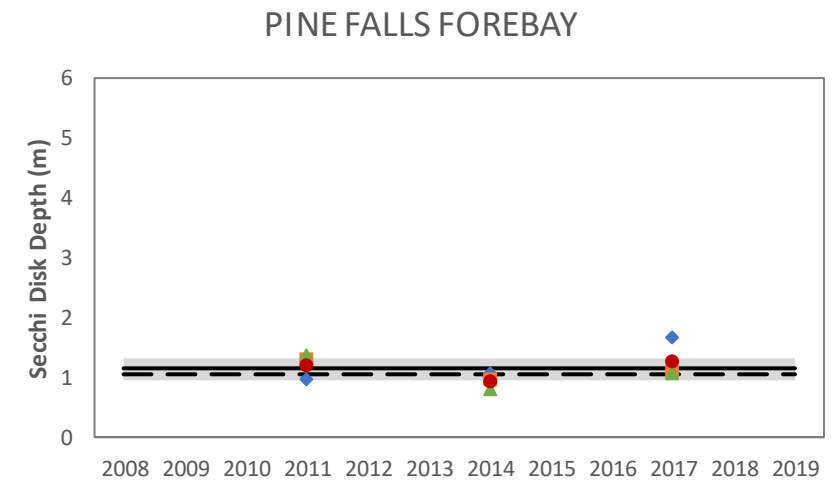
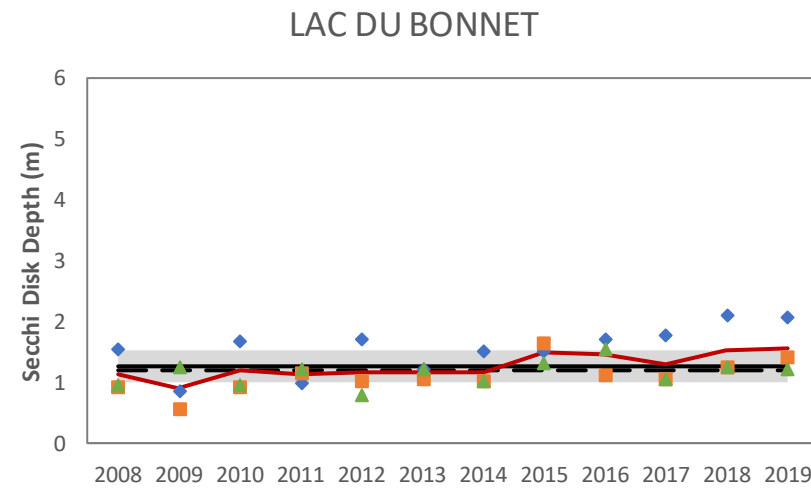
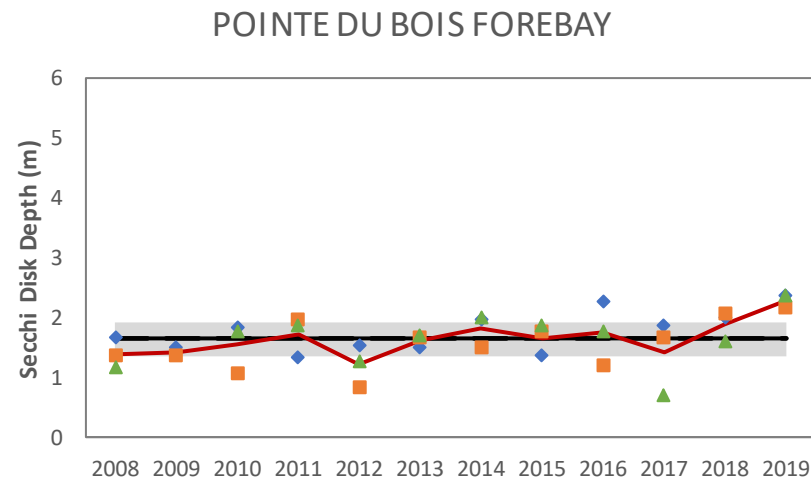
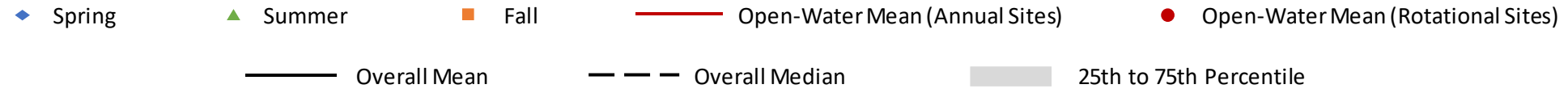


Figure 3.3-1. 2008-2019 On-system open-water season Secchi disk depths.

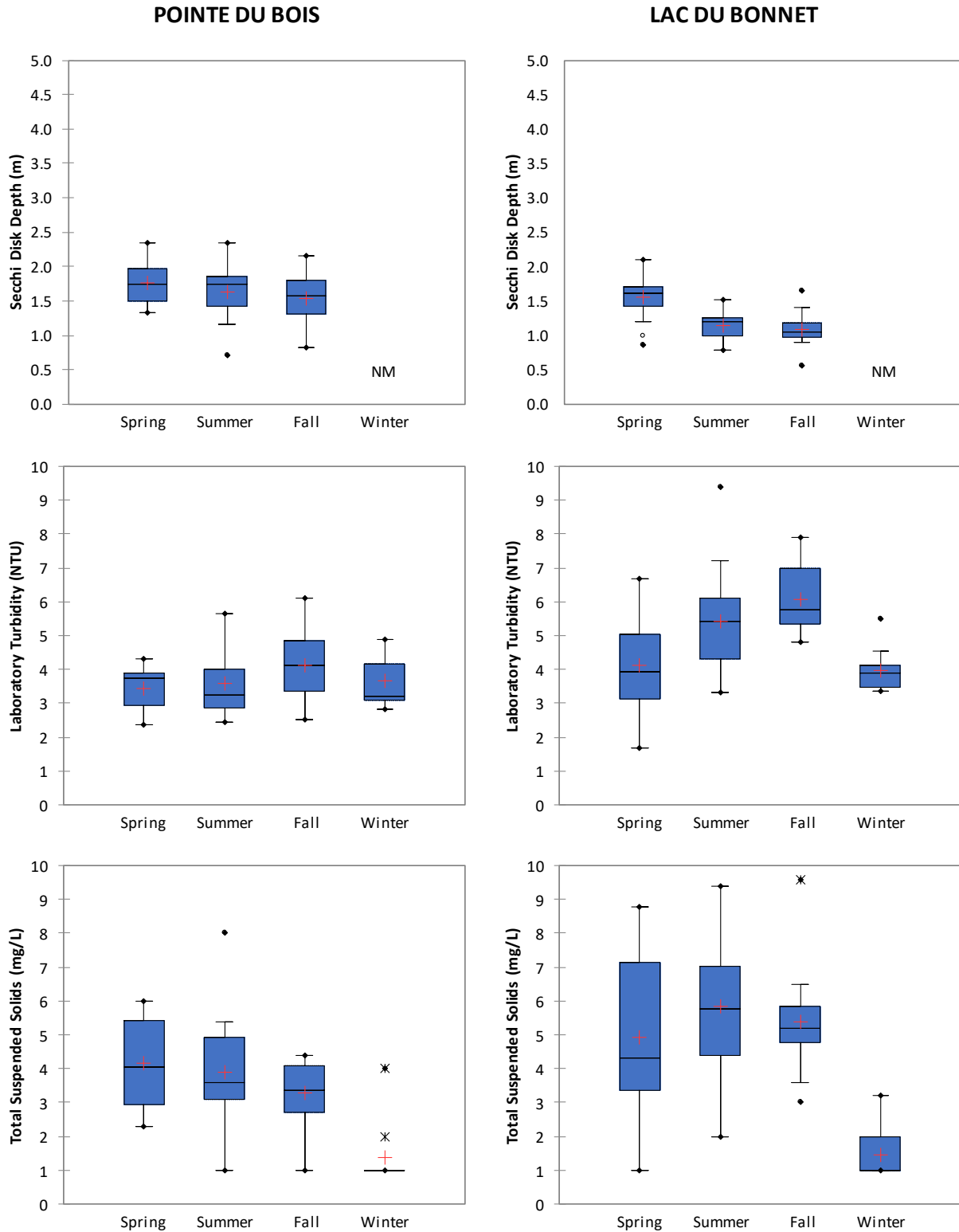


Figure 3.3-2. 2008-2019 On-system seasonal Secchi disk depth, turbidity, and TSS concentrations.

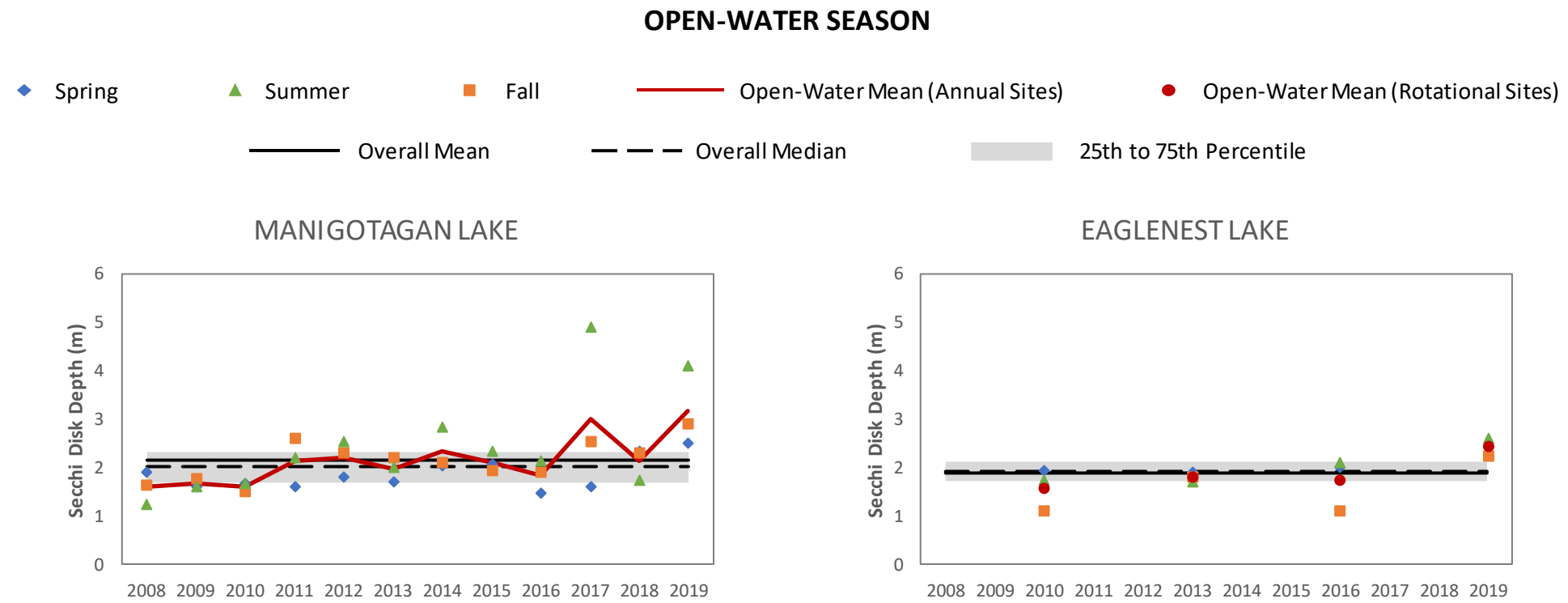


Figure 3.3-3. 2008-2019 Off-system open-water season Secchi disk depths.

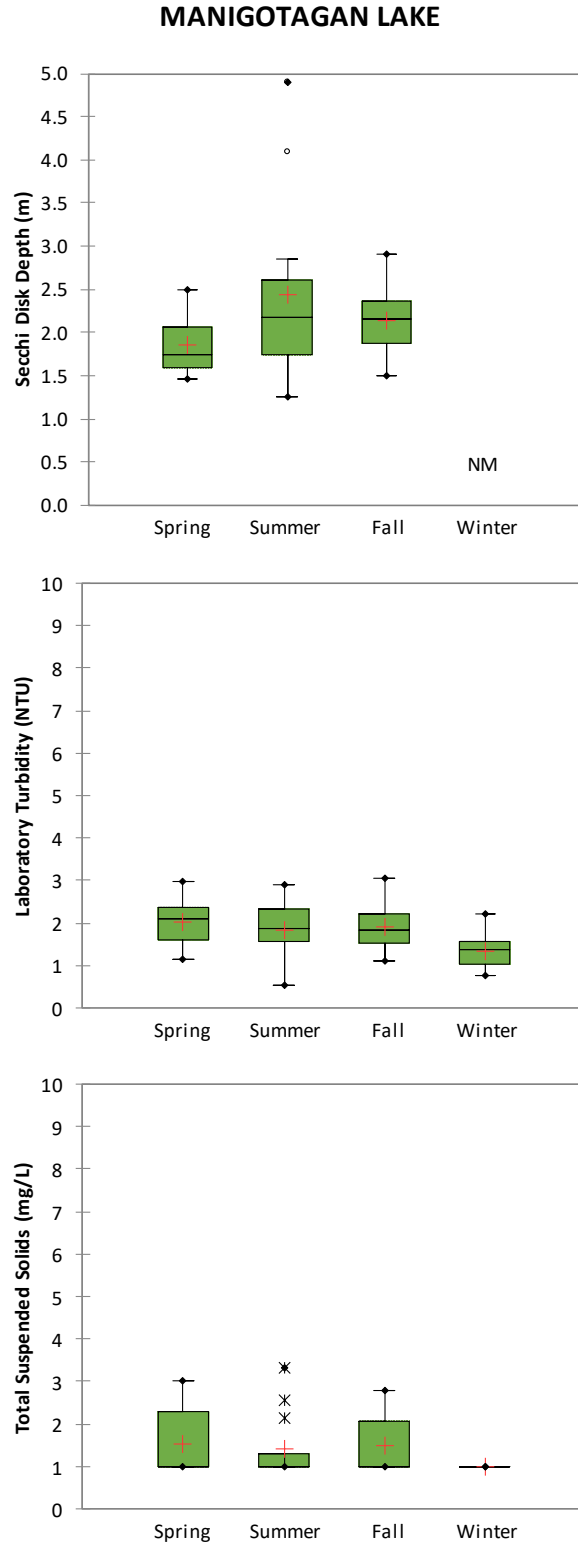


Figure 3.3-4. 2008-2019 Off-system seasonal Secchi disk depth, turbidity, and TSS concentrations.

3.3.2 TURBIDITY

3.3.2.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

Turbidity in the Pointe du Bois Forebay ranged from 2.38 to 6.10 NTU during the open-water season. The mean and median turbidity for the 12 years of monitoring were 3.72 and 3.57 NTU, respectively. Open-water season mean annual turbidity ranged from 2.45 to 4.93 NTU and was within the IQR (2.93 to 4.27 NTU) in seven of the 12 years. Mean turbidity was below the IQR in 2018, and 2019 and above the IQR in 2008, 2009, and 2012 (Table 3.3-1 and Figure 3.3-5).

Turbidity in the ice-cover season ranged from 2.83 to 4.88 NTU, with a mean of 3.65 NTU and a median of 3.22 NTU for the 11 years of monitoring. The IQR was 3.11 to 4.15 NTU (Table 3.3-1 and Figure 3.3-5).

No clear seasonality was observed for turbidity in the Pointe du Bois Forebay over the 12-year period. However, the lowest mean turbidity occurred in spring (3.45 NTU) and the highest in fall (4.13 NTU; Figure 3.3-2).

Lac du Bonnet

Turbidity in Lac du Bonnet ranged from 1.69 to 9.39 NTU during the open-water season. The mean and median turbidity for the 12 years of monitoring were 5.22 and 5.30 NTU, respectively. Open-water season mean annual turbidity ranged from 3.43 to 5.22 NTU and was within the IQR (4.17 to 6.11 NTU) in six of the 12 years. Mean turbidity was below the IQR in 2016, and 2018 and above the IQR in 2008, 2009, 2011, and 2012 (Table 3.3-1 and Figure 3.3-5).

Turbidity in the ice-cover season ranged from 3.37 to 5.48 NTU, with a mean of 3.96 NTU and 3.90 NTU for the 12 years of monitoring. The IQR was 3.49 to 4.13 NTU (Table 3.3-1 and Figure 3.3-5).

No clear seasonality was observed for turbidity in the Lac du Bonnet over the 12 years of monitoring. However, the lowest mean turbidity occurred in winter (3.96 NTU) and the highest in fall (6.09 NTU; Figure 3.3-2).

ROTATIONAL SITES

Pine Falls Forebay

Turbidity in the Pine Falls Forebay ranged from 3.46 to 9.80 NTU during the open-water season. The mean and median were both 6.35 NTU, and the IQR was 4.98 to 7.69 NTU for the three years of monitoring. Mean annual turbidity in the open-water season ranged from 5.10 to 8.08 NTU and was within the IQR in 2011, and 2017 but above the IQR in 2014 (Table 3.3-1 and Figure 3.3-5).

During the ice-cover season, turbidity was relatively similar in the three years of sampling, ranging from 3.27 to 5.22 NTU, with a mean of 4.43 NTU (Table 3.3-1 and Figure 3.3-5).

3.3.2.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

Turbidity in Manigotagan Lake ranged from 0.55 to 3.07 NTU during the open-water season. The mean was 1.94 NTU and the median was 1.99 NTU for the 12 years of monitoring. Open-water season mean annual turbidity ranged from 1.35 to 2.34 NTU and was within the IQR (1.58 to 2.36 NTU) in nine of the 12 years. Mean turbidity was below the IQR in 2008, 2017, and 2019 (Table 3.3-2 and Figure 3.3-6).

Turbidity in the ice-cover season ranged from 0.75 to 2.21 NTU, with a mean of 1.36 NTU and a median of 1.39 NTU for the 12 years of monitoring. The IQR was 1.05 to 1.55 NTU (Table 3.3-2 and Figure 3.3-6).

No clear seasonality was observed for turbidity in Manigotagan Lake over the 12 years of monitoring. However, the lowest mean turbidity occurred in winter (1.36 NTU) and the highest in spring (2.04 NTU; Figure 3.3-4).

ROTATIONAL SITES

Eaglenest Lake

Turbidity in Eaglenest Lake ranged from 1.86 to 4.86 NTU during the open-water season. The mean and median for the four years of monitoring were both 2.86 NTU, and the IQR was 2.38 to 3.06 NTU. Open-water season mean annual turbidity ranged from 2.36 to 3.45 NTU and was within

the IQR in 2013 and 2016 but was below the IQR in 2010 and above the IQR in 2019 (Table 3.3-2 and Figure 3.3-6).

During the ice-cover season, turbidity was relatively similar in the three years of sampling, ranging from 3.19 to 4.55 NTU, with a mean of 3.66 NTU (Table 3.3-2 and Figure 3.3-6).

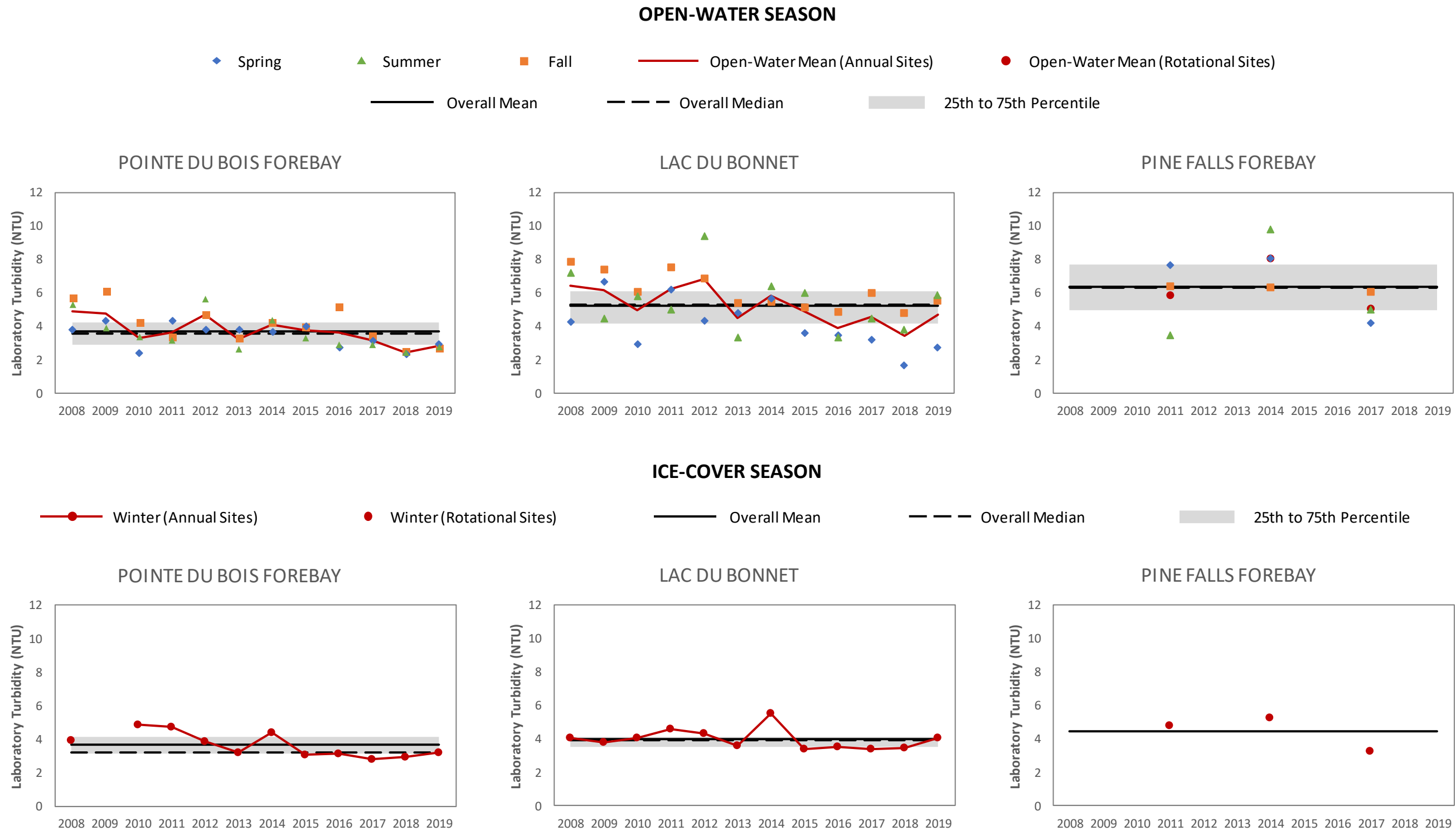
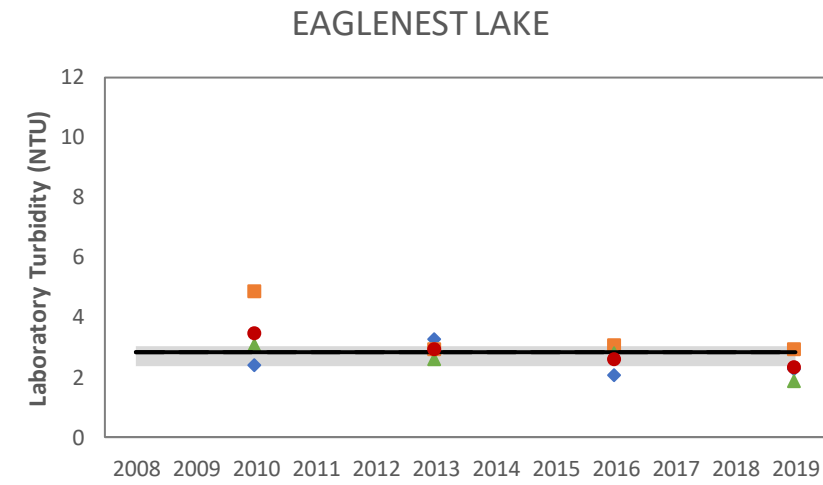
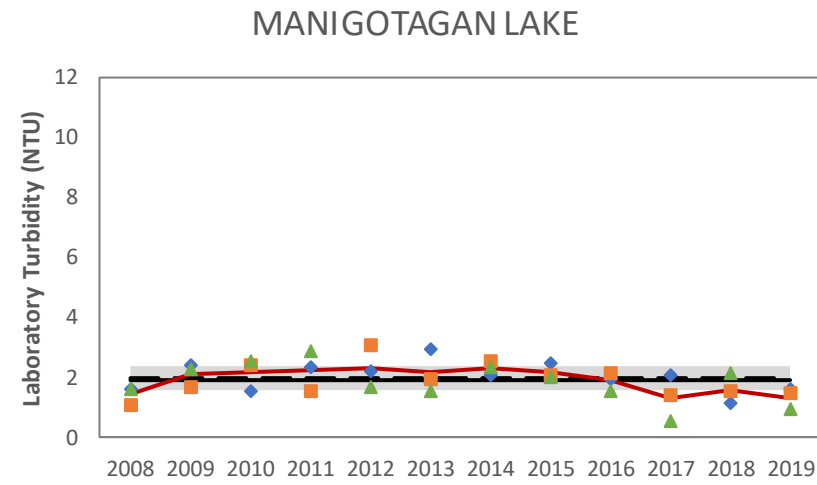
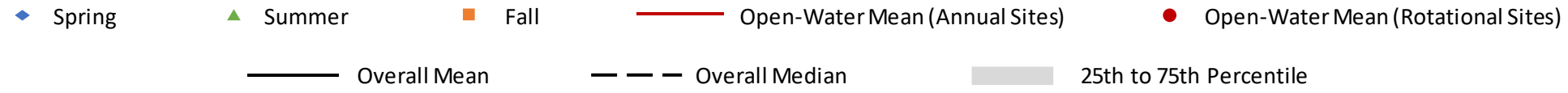


Figure 3.3-5. 2008-2019 On-system open-water and ice-cover season turbidity levels.

OPEN-WATER SEASON



ICE-COVER SEASON

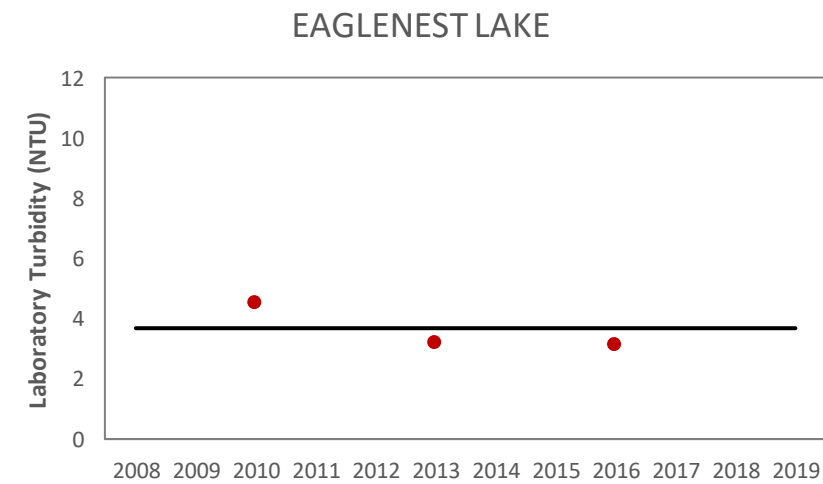
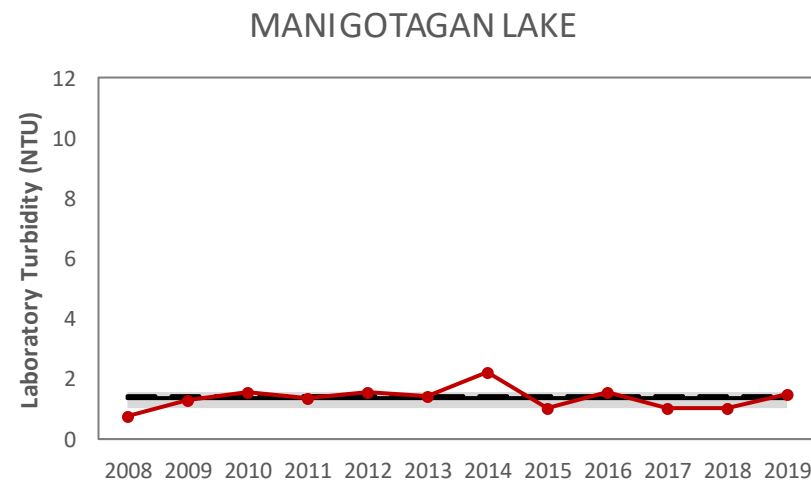
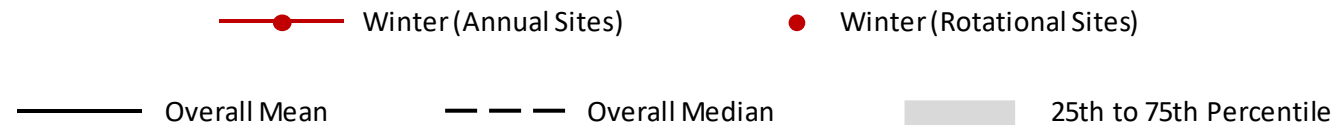


Figure 3.3-6. 2008-2019 Off-system open-water and ice-cover season turbidity levels.

3.3.3 TOTAL SUSPENDED SOLIDS

3.3.3.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

TSS concentrations in the Pointe du Bois Forebay ranged from <2.0 to 8.0 mg/L during the open-water season. The mean and median concentrations for the 12 years of monitoring were 3.8 and 3.6 mg/L, respectively. Open-water season mean annual TSS concentrations ranged from <2.0 to 5.1 mg/L and were within the IQR (2.8 to 4.7 mg/L) in nine of the 12 years. Mean TSS concentrations were below the IQR in 2018 and above the IQR in 2011, and 2014. TSS concentrations were typically above the detection limit (DL; 2.0 mg/L) during the open-water season (percent detections = 92; Table 3.3-1 and Figure 3.3-7).

TSS concentrations in the ice-cover season ranged from <2.0 to 4.0 mg/L, both the mean and median were <2.0 mg/L, and the IQR was <2.0 to <2.0 mg/L for the 11 years of monitoring. TSS concentrations were often below the DL (2.0 mg/L) during the ice-cover season (percent detections = 18; Table 3.3-1 and Figure 3.3-7).

TSS concentrations in the Point du Bois Forebay were lower in winter (mean <2.0 mg/L), often below the DL, than during the open-water season. No clear seasonality was observed for TSS concentrations in the open-water season over the 12-year period; however, the lowest mean TSS concentration occurred in fall (3.3 mg/L) and the highest in spring (4.2 mg/L; Figure 3.3-2).

Lac du Bonnet

TSS concentrations in Lac du Bonnet ranged from <2.0 to 9.6 mg/L during the open-water season. The mean and median concentrations for the 12 years of monitoring were 5.4 and 5.2 mg/L, respectively. Open-water season mean annual TSS concentrations ranged from 2.2 to 8.3 mg/L and were within the IQR (3.9 to 6.6 mg/L) in ten of the 12 years. Mean TSS concentrations were below the IQR in 2018 and above the IQR in 2014. TSS concentrations were typically above the DL (2.0 mg/L) during the open-water season (percent detections = 97; Table 3.3-1 and Figure 3.3-7).

TSS concentrations in the ice-cover season ranged from <2.0 to 3.2 mg/L, both the mean and median were <2.0 mg/L, and the IQR was <2.0 to 2.0 mg/L for the 12 years of monitoring. TSS

concentrations were often below the DL (2.0 mg/L) during the ice-cover season (percent detections = 33; Table 3.3-1 and Figure 3.3-7).

TSS concentrations in Lac du Bonnet were lower in winter (mean = <2.0 mg/L), often below the DL, than during the open-water season. No clear seasonality was observed for TSS concentrations in the open-water season over the 12 years of monitoring; however, the lowest mean TSS concentration occurred in spring (4.9 mg/L) and the highest in summer (5.8 mg/L; Figure 3.3-2).

ROTATIONAL SITES

Pine Falls Forebay

TSS concentrations in the Pine Falls Forebay ranged from 2.8 to 11.7 mg/L during the open-water season. The mean was 7.0 mg/L and median was 6.4 mg/L, and the IQR was 4.1 to 11.2 mg/L for the three years of monitoring. Mean annual TSS concentrations in the open-water season ranged from 5.3 to 9.8 mg/L and were within the IQR in all years. TSS concentrations were consistently above the DL (2.0 mg/L) during the open-water season (Table 3.3-1 and Figure 3.3-7).

During the ice-cover season, TSS concentrations ranged from <2.0 to 2.0 mg/L, with a mean of <2.0 mg/L. TSS concentrations were below the DL (2.0 mg/L) in two of three samples collected in winter (i.e., percent detections = 33; Table 3.3-1 and Figure 3.3-7).

3.3.3.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

TSS concentrations in Manigotagan Lake ranged from <2.0 to 3.3 mg/L during the open-water season. The mean and median were both <2.0 mg/L for the 12 years of monitoring. Open-water season mean annual TSS concentrations ranged from <2.0 to 2.6 mg/L and were within the IQR (<2.0 to 2.2 mg/L) in 11 of the 12 years. Mean TSS concentrations were above the IQR in 2014. TSS concentrations were above the DL (2.0 mg/L) in approximately one third of the samples collected during the open-water season (percent detections = 31; Table 3.3-2 and Figure 3.3-8).

TSS concentrations in the ice-cover season were consistently below the DL (2.0 mg/L) over the 12 years of monitoring (Table 3.3-2 and Figure 3.3-8).

Although TSS concentrations were more frequently above the DL during the open-water season than in winter, no clear seasonality was observed for TSS concentrations in Manigotagan Lake over the 12 years of monitoring. Mean TSS concentrations were below the DL (2.0 mg/L) in spring, summer, fall and winter (Figure 3.3-4).

ROTATIONAL SITES

Eaglenest Lake

TSS concentrations in Eaglenest Lake ranged from <2.0 to 4.0 mg/L during the open-water season. The mean was 2.4 mg/L, median was 2.5 mg/L, and the IQR was <2.0 to 3.3 mg/L for the four years of monitoring. Open-water season mean annual TSS concentrations ranged from 2.1 to 2.7 mg/L and were within the IQR in all years. TSS concentrations were above the DL (2.0 mg/L) in 67 percent of samples collected during the open-water season (Table 3.3-2 and Figure 3.3-8).

During the ice-cover season, TSS concentrations ranged from <2.0 to 3.2 mg/L, with a mean of <2.0 mg/L. TSS concentrations were below the DL (2.0 mg/L) in two of three samples collected in winter (percent detections = 33; Table 3.3-2 and Figure 3.3-8).

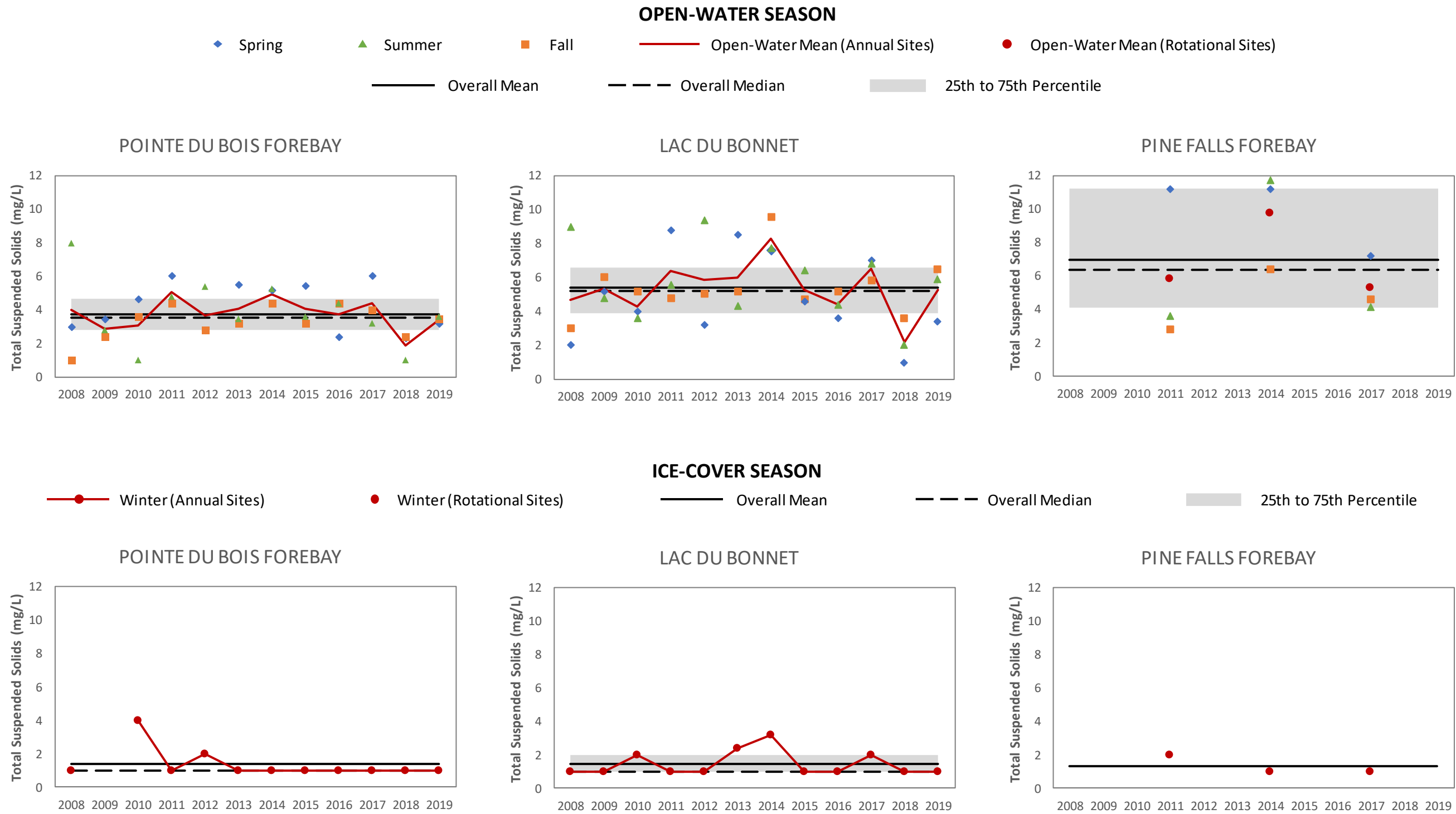


Figure 3.3-7. 2008-2019 On-system open-water and ice-cover season TSS concentrations.

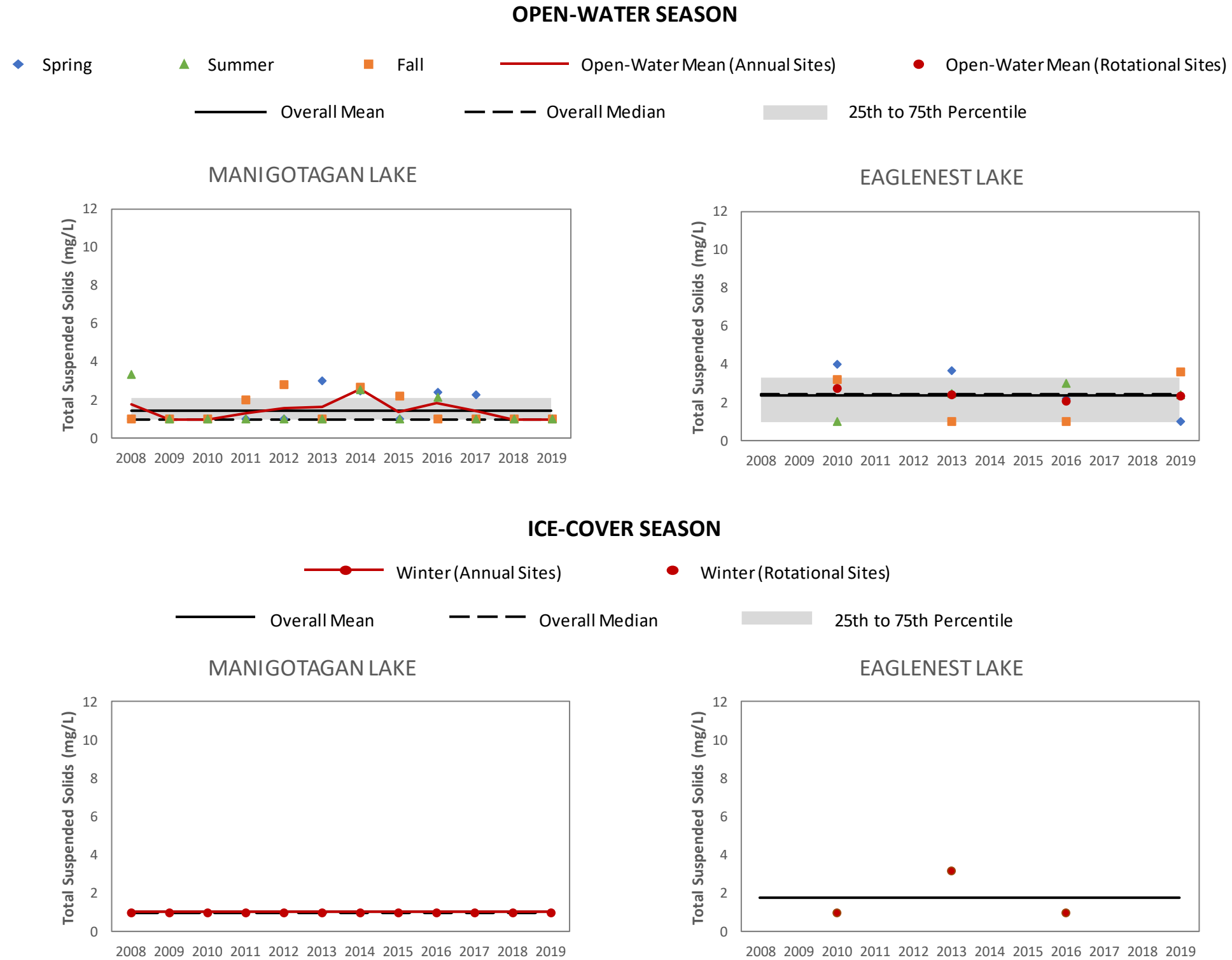


Figure 3.3-8. 2008-2019 Off-system open-water and ice-cover season TSS concentrations.

3.4 NUTRIENTS AND TROPHIC STATUS

3.4.1 TOTAL PHOSPHORUS

3.4.1.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

TP concentrations in the Pointe du Bois Forebay ranged from 0.015 to 0.043 mg/L during the open-water season. The mean and median concentrations for the 12 years of monitoring were both 0.023 mg/L. Open-water season mean annual TP concentrations ranged from 0.019 to 0.029 mg/L and were within the IQR (0.020 to 0.025 mg/L) in nine of the 12 years. Mean TP concentrations were below the IQR in 2011 and above the IQR in 2015, and 2016 (Table 3.4-1 and Figure 3.4-1).

TP concentrations in the ice-cover season ranged from 0.022 to 0.036 mg/L, with a mean of 0.028 mg/L and a median of 0.027 mg/L for the 12 years of monitoring. The IQR was 0.025 to 0.030 mg/L (Table 3.4-1 and Figure 3.4-1).

No clear seasonality was observed for TP in the Pointe du Bois Forebay over the 12-year period. However, the lowest mean TP concentration occurred in spring (0.022 mg/L) and the highest in winter (0.028 mg/L; Figure 3.4-2).

The Pointe du Bois Forebay was meso-eutrophic (0.020 to 0.035 mg/L) on the basis of the 2008-2019 mean open-water season TP concentration (0.023 mg/L). Mean annual TP concentrations (0.019 to 0.029 mg/L) in the open-water season were within the meso-eutrophic range (0.020 to 0.035 mg/L) in all years except for 2011 when the mean (0.019 mg/L) was within the mesotrophic range (0.010 to 0.020 mg/L; Table 3.4-2).

Lac du Bonnet

TP concentrations in Lac du Bonnet ranged from 0.015 to 0.035 mg/L during the open-water season. The mean and median concentrations for the 12 years of monitoring were both 0.025 mg/L. Open-water season mean annual TP concentrations ranged from 0.021 to 0.029 mg/L and were within the IQR (0.023 to 0.028 mg/L) in eight of the 12 years. Mean TP concentrations

were below the IQR in 2012, 2018, and 2019 and above the IQR in 2010 (Table 3.4-1 and Figure 3.4-1).

TP concentrations in the ice-cover season ranged from 0.024 to 0.036 mg/L, with a mean of 0.029 mg/L and a median of 0.028 mg/L for the 12 years of monitoring. The IQR was 0.027 to 0.030 mg/L (Table 3.4-1 and Figure 3.4-1).

On average, TP concentrations were lowest in spring (0.022 mg/L) and highest in winter (0.029 mg/L; Figure 3.4-2).

Lac du Bonnet was meso-eutrophic (0.020 to 0.035 mg/L) on the basis of the 2008-2019 mean open-water season TP concentration (0.025 mg/L). Mean annual TP concentrations (0.021 to 0.029 mg/L) in the open-water season were within the meso-eutrophic range in all years (Table 3.4-2).

ROTATIONAL SITES

Pine Falls Forebay

TP concentrations in the Pine Falls Forebay ranged from 0.012 to 0.031 mg/L during the open-water season. The mean was 0.025 mg/L, the median was 0.027 mg/L, and the IQR was 0.026 to 0.027 mg/L for the three years of monitoring. Mean annual TP concentrations in the open-water season ranged from 0.022 to 0.027 mg/L and were within the IQR in 2014 and 2017 but below the IQR in 2011 (Table 3.4-1 and Figure 3.4-1).

During the ice-cover season, TP concentrations ranged from 0.024 to 0.030 mg/L, with a mean of 0.027 mg/L (Table 3.4-1 and Figure 3.4-1).

The Pine Falls Forebay was meso-eutrophic (0.020 to 0.035 mg/L) based on the mean of the open-water season TP concentrations for the three years of monitoring (0.025 mg/L). Open-water season mean annual TP concentrations (0.022 to 0.027 mg/L) were also within the meso-eutrophic range in each year sampled (Table 3.4-2).

3.4.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

TP concentrations in Manigotagan Lake ranged from 0.008 to 0.034 mg/L during the open-water season. The mean and median concentrations for the 12 years of monitoring were 0.019 mg/L and 0.016 mg/L, respectively. Open-water season mean annual TP concentrations ranged from 0.012 to 0.025 mg/L but were within the IQR (0.014 to 0.023 mg/L) in nine of the 12 years. Mean TP concentrations were below the IQR in 2018 and above the IQR in 2010, and 2016 (Table 3.4-3 and Figure 3.4-3).

TP concentrations in the ice-cover season ranged from 0.022 to 0.035 mg/L, with a mean of 0.026 mg/L and a median of 0.025 mg/L for the 12 years of monitoring. The IQR was 0.023 to 0.029 mg/L (Table 3.4-3 and Figure 3.4-3).

On average, TP concentrations were higher in winter (0.026 mg/L) and spring (0.027 mg/L) than in summer (0.014 mg/L) or fall (0.017 mg/L; Figure 3.4-4).

Manigotagan Lake was mesotrophic (0.010 to 0.020 mg/L) on the basis of the 2008-2019 mean open-water season TP concentration (0.019 mg/L). Mean annual TP concentrations (0.012 to 0.025 mg/L) in the open-water season were within the mesotrophic range (0.010 to 0.020 mg/L) in eight of the 12 years of monitoring; however, they were within the meso-eutrophic range (0.020 to 0.035 mg/L) in 2010, 2011, 2014, and 2016 (Table 3.4-4).

ROTATIONAL SITES

Eaglenest Lake

TP concentrations in Eaglenest Lake ranged from 0.015 to 0.030 mg/L during the open-water season. The mean and median concentrations for the four years of monitoring were both 0.022 mg/L, and the IQR was 0.020 to 0.024 mg/L. Open-water season mean annual TP concentrations ranged from 0.019 to 0.025 mg/L and were within the IQR in 2013 and 2016 but were below the IQR in 2019 and above the IQR in 2010 (Table 3.4-3 and Figure 3.4-3).

During the ice-cover season, TP concentrations ranged from 0.026 and 0.028 mg/L, with a mean of 0.027 mg/L (Table 3.4-3 and Figure 3.4-3).

Eaglenest Lake was meso-eutrophic (0.020 to 0.035 mg/L) based on the mean of the open-water season TP concentrations for the four years of monitoring (0.022 mg/L). Open-water season mean annual TP concentrations (0.019 to 0.025 mg/L) were within the meso-eutrophic range in three of the four years of monitoring; however, the open-water season mean TP concentration was within the mesotrophic range (0.010 to 0.020 mg/L) in 2019 (Table 3.4-4).

Table 3.4-1. 2008-2019 On-system sites TP, TN, and chlorophyll *a* summary statistics.

| Site | Statistic | TP (mg/L) | | TN (mg/L) | | Chlorophyll <i>a</i> (µg/L) | |
|------|----------------|-----------|--------|-----------|-------|-----------------------------|-------|
| | | OW | IC | OW | IC | OW | IC |
| PDB | Mean | 0.023 | 0.028 | 0.42 | 0.53 | 7.56 | 0.65 |
| | Median | 0.023 | 0.027 | 0.45 | 0.52 | 6.46 | <0.60 |
| | Minimum | 0.015 | 0.022 | <0.20 | 0.42 | 1.91 | <0.60 |
| | Maximum | 0.043 | 0.036 | 0.68 | 0.65 | 26.3 | 2.06 |
| | SD | 0.0055 | 0.0043 | 0.129 | 0.072 | 4.55 | 0.524 |
| | SE | 0.0009 | 0.0013 | 0.021 | 0.022 | 0.758 | 0.158 |
| | Lower Quartile | 0.020 | 0.025 | 0.36 | 0.48 | 5.06 | <0.60 |
| | Upper Quartile | 0.025 | 0.030 | 0.50 | 0.58 | 8.05 | 0.76 |
| | n | 36 | 11 | 36 | 11 | 36 | 11 |
| | % Detections | 100 | 100 | 92 | 100 | 100 | 45 |
| LDB | Mean | 0.025 | 0.029 | 0.44 | 0.56 | 7.02 | 1.19 |
| | Median | 0.025 | 0.028 | 0.46 | 0.53 | 6.21 | 0.73 |
| | Minimum | 0.015 | 0.024 | <0.20 | 0.41 | 0.84 | <0.60 |
| | Maximum | 0.035 | 0.036 | 0.68 | 0.98 | 17.8 | 4.96 |
| | SD | 0.0045 | 0.0035 | 0.118 | 0.143 | 3.33 | 1.35 |
| | SE | 0.0008 | 0.0010 | 0.020 | 0.041 | 0.554 | 0.390 |
| | Lower Quartile | 0.023 | 0.027 | 0.39 | 0.50 | 5.10 | <0.60 |
| | Upper Quartile | 0.028 | 0.030 | 0.51 | 0.57 | 8.49 | 1.39 |
| | n | 36 | 12 | 36 | 12 | 36 | 12 |
| | % Detections | 100 | 100 | 94 | 100 | 100 | 50 |
| PFF | Mean | 0.025 | 0.027 | 0.43 | 0.47 | 7.46 | 1.72 |
| | Median | 0.027 | - | 0.46 | - | 5.54 | - |
| | Minimum | 0.012 | 0.024 | <0.20 | 0.40 | 2.10 | 1.53 |
| | Maximum | 0.031 | 0.030 | 0.57 | 0.51 | 18.1 | 1.91 |
| | SD | 0.0056 | 0.0030 | 0.133 | 0.064 | 5.00 | 0.190 |
| | SE | 0.0019 | 0.0017 | 0.044 | 0.037 | 1.67 | 0.110 |
| | Lower Quartile | 0.026 | - | 0.44 | - | 4.28 | - |
| | Upper Quartile | 0.027 | - | 0.49 | - | 9.93 | - |
| | n | 9 | 3 | 9 | 3 | 9 | 3 |
| | % Detections | 100 | 100 | 89 | 100 | 100 | 100 |

Notes:

1. OW = Open-water season; IC = Ice-cover season.
2. SD = standard deviation; SE = standard error; n = number of samples.

Table 3.4-2. 2008-2019 On-system trophic status based on TP, TN, and chlorophyll *a* open-water season mean concentrations.

| Trophic Categories | Total Phosphorus (mg/L) | | | Total Nitrogen (mg/L) | | | Chlorophyll <i>a</i> (µg/L) | | |
|---------------------|------------------------------|-------|-------|-----------------------|------|------|-----------------------------|------|------|
| Ultra-oligotrophic | <0.004 | | | | | | | | |
| Oligotrophic | 0.004-0.010 | | | <0.350 | | | <2.5 | | |
| Mesotrophic | 0.010-0.020 | | | 0.350-0.650 | | | 2.5-8 | | |
| Meso-eutrophic | 0.020-0.035 | | | | | | | | |
| Eutrophic | 0.035-0.100 | | | 0.651-1.20 | | | 8-25 | | |
| Hypereutrophic | > 0.100 | | | >1.20 | | | >25 | | |
| Reference | CCME (1999; updated to 2024) | | | Nürnberg (1996) | | | OECD (1982) | | |
| Sampling Year | PDB | LDB | PFF | PDB | LDB | PFF | PDB | LDB | PFF |
| 2008 | 0.025 | 0.028 | - | 0.52 | 0.52 | - | 6.33 | 6.00 | - |
| 2009 | 0.024 | 0.025 | - | 0.40 | 0.40 | - | 5.60 | 6.63 | - |
| 2010 | 0.025 | 0.029 | - | 0.30 | 0.44 | - | 3.52 | 3.64 | - |
| 2011 | 0.019 | 0.026 | 0.022 | 0.50 | 0.50 | 0.48 | 9.73 | 10.4 | 9.56 |
| 2012 | 0.020 | 0.021 | - | 0.47 | 0.48 | - | 6.95 | 4.82 | - |
| 2013 | 0.020 | 0.025 | - | 0.43 | 0.43 | - | 9.54 | 8.70 | - |
| 2014 | 0.025 | 0.026 | 0.027 | 0.46 | 0.49 | 0.49 | 6.08 | 7.58 | 7.11 |
| 2015 | 0.029 | 0.024 | - | 0.43 | 0.42 | - | 12.2 | 7.95 | - |
| 2016 | 0.027 | 0.027 | - | 0.55 | 0.55 | - | 5.28 | 5.54 | - |
| 2017 | 0.024 | 0.026 | 0.026 | 0.32 | 0.35 | 0.33 | 6.44 | 6.73 | 5.73 |
| 2018 | 0.021 | 0.021 | - | 0.31 | 0.28 | - | 8.14 | 6.56 | - |
| 2019 | 0.021 | 0.022 | - | 0.38 | 0.39 | - | 10.9 | 9.77 | - |
| Overall (2008-2019) | 0.023 | 0.025 | 0.025 | 0.42 | 0.44 | 0.43 | 7.56 | 7.02 | 7.46 |

Notes:

1. CCME = Canadian Council of Ministers of the Environment
2. OECD = Organization for Economic Cooperation and Development

Table 3.4-3. 2008-2019 Off-system sites TP, TN, and chlorophyll *a* summary statistics.

| Site | Statistic | TP (mg/L) | | TN (mg/L) | | Chlorophyll <i>a</i> (µg/L) | |
|-------|----------------|-----------|--------|-----------|-------|-----------------------------|-------|
| | | OW | IC | OW | IC | OW | IC |
| MANIG | Mean | 0.019 | 0.026 | 0.50 | 0.55 | 6.73 | 1.85 |
| | Median | 0.016 | 0.025 | 0.52 | 0.55 | 6.06 | 1.31 |
| | Minimum | 0.008 | 0.022 | <0.20 | 0.38 | 1.53 | 0.68 |
| | Maximum | 0.034 | 0.035 | 0.96 | 0.66 | 19.6 | 6.11 |
| | SD | 0.0072 | 0.0041 | 0.156 | 0.077 | 3.50 | 1.53 |
| | SE | 0.0012 | 0.0012 | 0.026 | 0.022 | 0.584 | 0.442 |
| | Lower Quartile | 0.014 | 0.023 | 0.45 | 0.51 | 4.34 | 0.99 |
| | Upper Quartile | 0.023 | 0.029 | 0.58 | 0.61 | 8.16 | 1.96 |
| | n | 36 | 12 | 36 | 12 | 36 | 12 |
| | % Detections | 100 | 100 | 94 | 100 | 100 | 100 |
| EAGLE | Mean | 0.022 | 0.027 | 0.43 | 0.59 | 6.48 | <0.60 |
| | Median | 0.022 | - | 0.41 | - | 6.03 | - |
| | Minimum | 0.015 | 0.026 | 0.27 | 0.46 | 1.40 | <0.60 |
| | Maximum | 0.030 | 0.028 | 0.70 | 0.77 | 15.1 | <0.60 |
| | SD | 0.0041 | 0.0013 | 0.118 | 0.159 | 3.60 | - |
| | SE | 0.0012 | 0.0007 | 0.034 | 0.092 | 1.04 | - |
| | Lower Quartile | 0.020 | - | 0.35 | - | 4.78 | - |
| | Upper Quartile | 0.024 | - | 0.49 | - | 7.45 | - |
| | n | 12 | 3 | 12 | 3 | 12 | 3 |
| | % Detections | 100 | 100 | 100 | 100 | 100 | 0 |

Notes:

1. OW = Open-water season; IC = Ice-cover season.
2. SD = standard deviation; SE = standard error; n = number of samples.

Table 3.4-4. 2008-2019 Off-system trophic status based on TP, TN, and chlorophyll *a* open-water season mean concentrations.

| Trophic Categories | Total Phosphorus (mg/L) | | Total Nitrogen (mg/L) | | Chlorophyll <i>a</i> (µg/L) | |
|---------------------|------------------------------|-------|-----------------------|-------|-----------------------------|-------|
| Ultra-oligotrophic | <0.004 | | | | | |
| Oligotrophic | 0.004-0.010 | | <0.350 | | <2.5 | |
| Mesotrophic | 0.010-0.020 | | 0.350-0.650 | | 2.5-8 | |
| Meso-eutrophic | 0.020-0.035 | | | | | |
| Eutrophic | 0.035-0.100 | | 0.651-1.20 | | 8-25 | |
| Hypereutrophic | > 0.100 | | >1.20 | | >25 | |
| References | CCME (1999; updated to 2024) | | Nürnberg (1996) | | OECD (1982) | |
| Sampling Year | MANIG | EAGLE | MANIG | EAGLE | MANIG | EAGLE |
| 2008 | 0.017 | - | 0.58 | - | 5.33 | - |
| 2009 | 0.020 | - | 0.46 | - | 6.11 | - |
| 2010 | 0.025 | 0.025 | 0.48 | 0.36 | 3.64 | 4.22 |
| 2011 | 0.020 | - | 0.55 | - | 5.16 | - |
| 2012 | 0.019 | - | 0.66 | - | 5.70 | - |
| 2013 | 0.020 | 0.021 | 0.43 | 0.39 | 7.56 | 6.91 |
| 2014 | 0.021 | - | 0.55 | - | 8.97 | - |
| 2015 | 0.019 | - | 0.52 | - | 6.67 | - |
| 2016 | 0.025 | 0.024 | 0.62 | 0.57 | 6.62 | 5.52 |
| 2017 | 0.017 | - | 0.36 | - | 8.52 | - |
| 2018 | 0.012 | - | 0.34 | - | 8.87 | - |
| 2019 | 0.016 | 0.019 | 0.47 | 0.38 | 7.63 | 9.29 |
| Overall (2008-2019) | 0.019 | 0.022 | 0.50 | 0.43 | 6.73 | 6.48 |

Notes:

1. CCME = Canadian Council of Ministers of the Environment.
2. OECD = Organization for Economic Cooperation and Development.



Figure 3.4-1. 2008-2019 On-system open-water and ice-cover season TP concentrations.

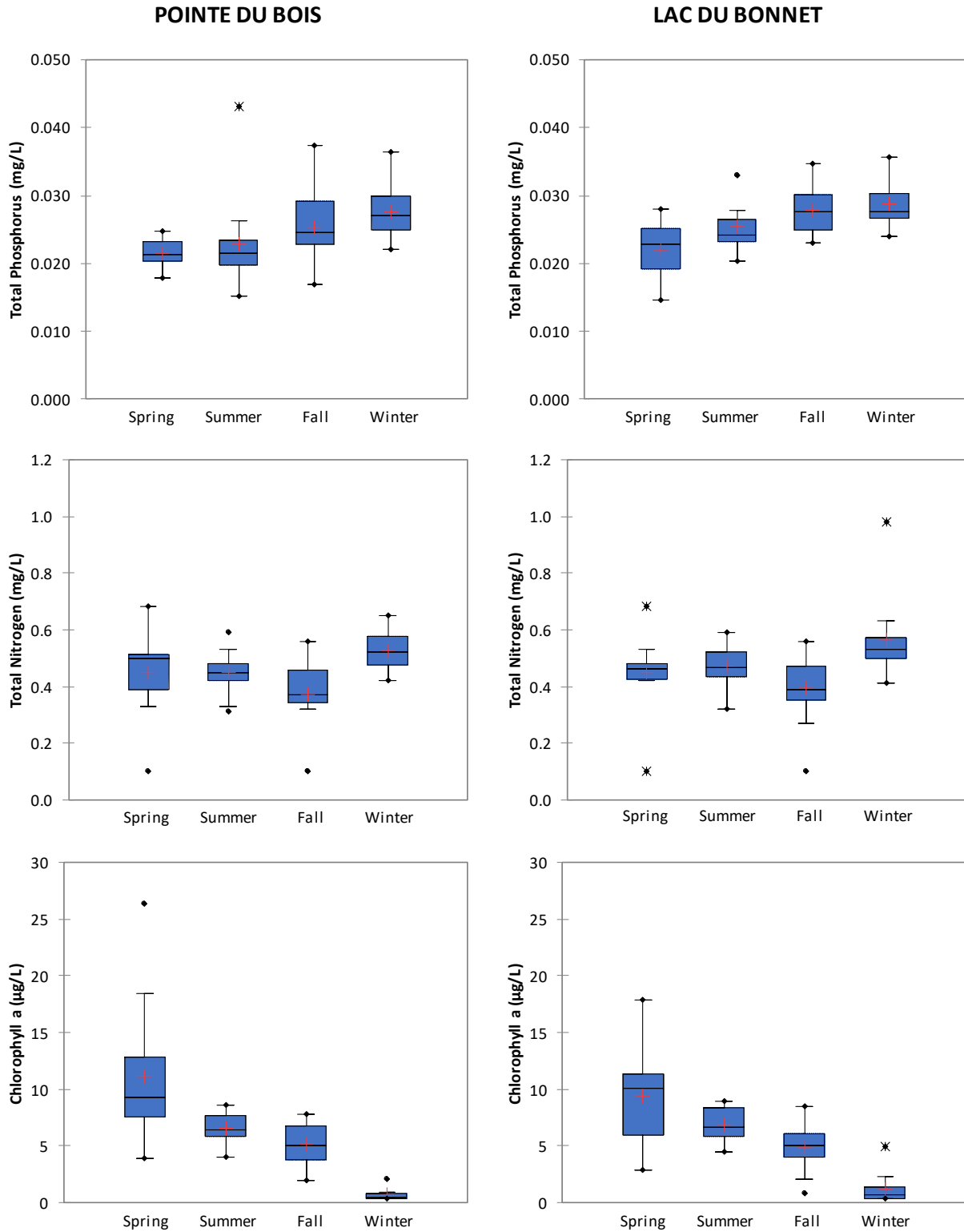


Figure 3.4-2. 2008-2019 On-system seasonal total phosphorus, total nitrogen, and chlorophyll *a* concentrations.

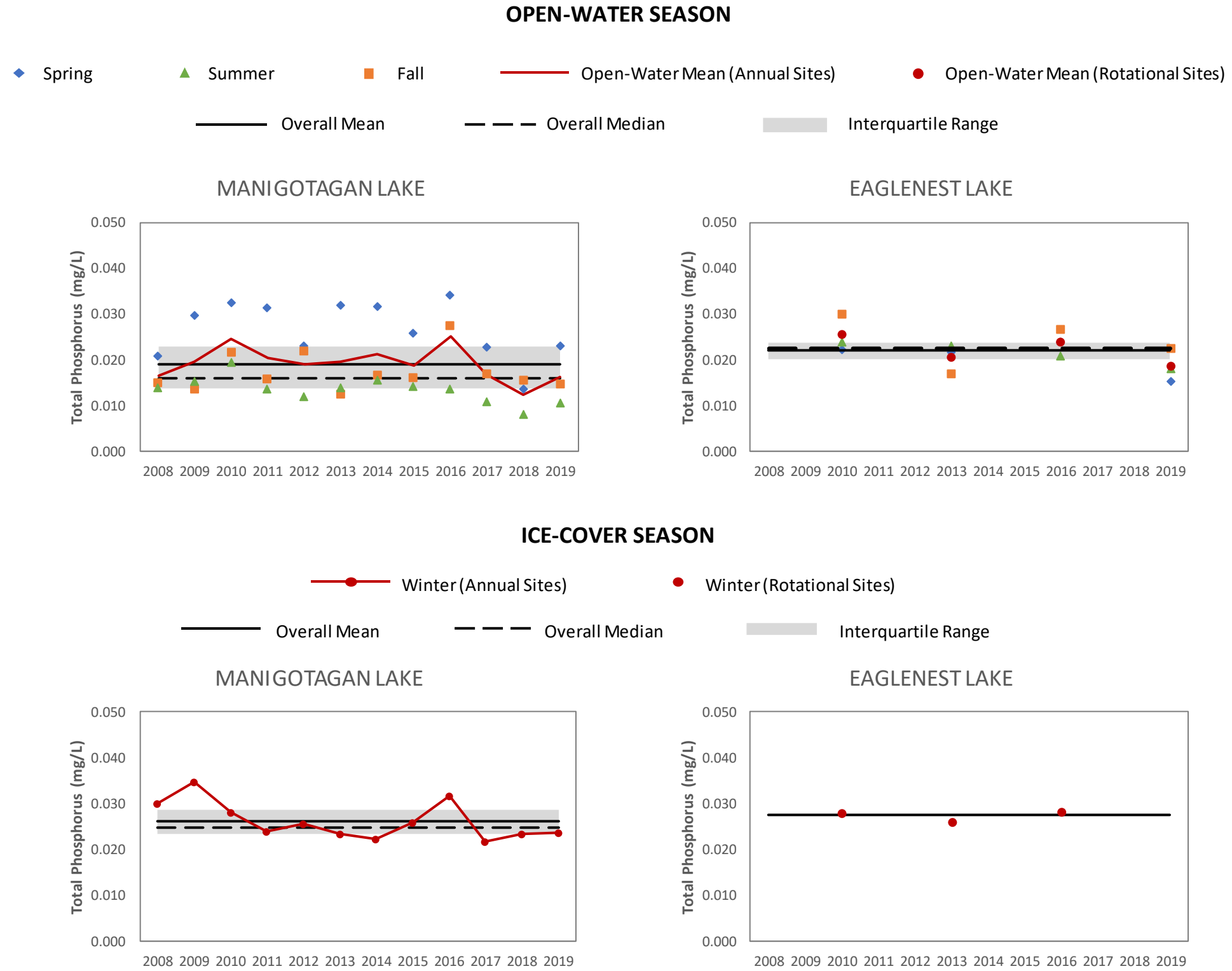


Figure 3.4-3. 2008-2019 Off-system open-water and ice-cover season TP concentrations.

MANIGOTAGAN LAKE

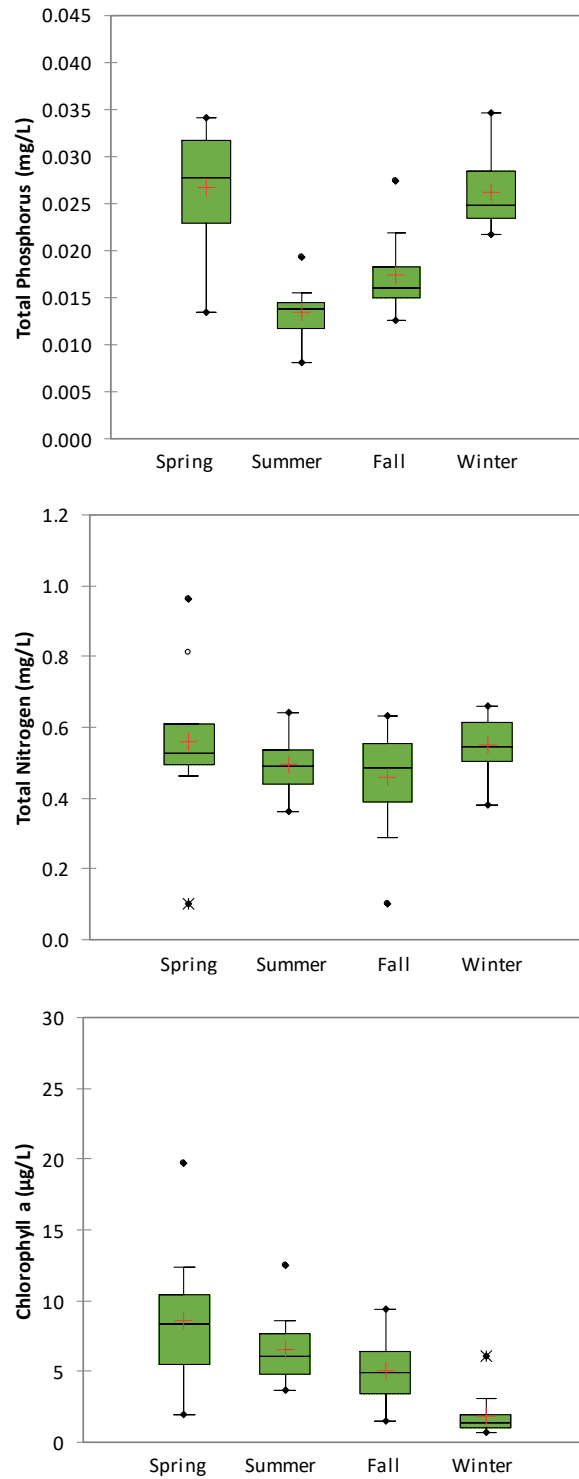


Figure 3.4-4. 2008-2019 Off-system seasonal total phosphorus, total nitrogen, and chlorophyll *a* concentrations.

3.4.2 TOTAL NITROGEN

3.4.2.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

TN concentrations in the Pointe du Bois Forebay ranged from <0.20 to 0.68 mg/L during the open-water season. The mean and median for the 12 years of monitoring were 0.42 mg/L and 0.45 mg/L, respectively. Open-water season mean annual TN concentrations ranged from 0.30 to 0.55 mg/L and were within or near the IQR (0.36 to 0.50 mg/L) in seven of the 12 years. Mean TN concentrations were below the IQR in 2010, 2017, and 2018 and above the IQR in 2008, and 2016 (Table 3.4-1 and Figure 3.4-5).

TN concentrations in the ice-cover season ranged from 0.42 to 0.65 mg/L, with a mean of 0.53 mg/L and a median of 0.52 mg/L for the 12 years of monitoring. The IQR was 0.48 to 0.58 mg/L (Table 3.4-1 and Figure 3.4-5).

No clear seasonality was observed for TN in the Pointe du Bois Forebay over the 12-year period. However, the lowest mean TN concentration occurred in fall (0.37 mg/L) and the highest in winter (0.53 mg/L; Figure 3.4-2).

The Pointe du Bois Forebay was mesotrophic (0.350 to 0.650 mg/L) on the basis of the 2008-2019 mean open-water season TN concentration (0.42 mg/L). Mean annual TN concentrations (0.30 to 0.55 mg/L) in the open-water season were within the mesotrophic range (0.350 to 0.650 mg/L) in all years except for 2010, 2017 and 2018 when mean TN concentrations were in the oligotrophic range (i.e., <0.350 mg/L; Table 3.4-2).

Lac du Bonnet

TN concentrations in Lac du Bonnet ranged from <0.20 to 0.68 mg/L during the open-water season. The mean and median for the 12 years of monitoring were 0.44 mg/L and 0.46 mg/L, respectively. Open-water season mean annual TN concentrations ranged from 0.28 to 0.55 mg/L and were within the IQR (0.39 to 0.51 mg/L) in eight of the 12 years. Mean TN concentrations were below the IQR in 2017, and 2018 and above the IQR in 2008, and 2016 (Table 3.4-1 and Figure 3.4-5).

TN concentrations in the ice-cover season ranged from 0.41 to 0.98 mg/L, with a mean of 0.56 mg/L and a median of 0.53 mg/L for the 12 years of monitoring. The IQR was 0.50 to 0.57 mg/L. TN concentrations were within or near the IQR except in 2013 when it was above the IQR (Table 3.4-1 and Figure 3.4-5).

No clear seasonality was observed for TN in Lac du Bonnet over the 12 years of monitoring. However, the lowest mean TN concentration occurred in fall (0.39 mg/L) and the highest in winter (0.56 mg/L; Figure 3.4-2).

Lac du Bonnet was mesotrophic (0.350 to 0.650 mg/L) on the basis of the 2008-2019 mean open-water season TN concentration (0.44 mg/L). Mean annual TN concentrations (0.28 to 0.55 mg/L) in the open-water season were within the mesotrophic range (0.350 to 0.650 mg/L) in all years except for 2018 when the mean TN concentration was in the oligotrophic range (i.e., <0.350 mg/L; Table 3.4-2).

ROTATIONAL SITES

Pine Falls Forebay

TN concentrations in the Pine Falls Forebay ranged from <0.20 to 0.57 mg/L during the open-water season. The mean was 0.43 mg/L, the median was 0.46 mg/L, and the IQR was 0.44 to 0.49 mg/L for the three years of monitoring. Mean annual TN concentrations in the open-water season ranged from 0.33 to 0.49 mg/L and were within the IQR in 2011 and 2014, but below the IQR in 2017 (Table 3.4-1 and Figure 3.4-5).

During the ice-cover season, TN concentrations ranged from 0.40 to 0.51 mg/L, with a mean of 0.47 mg/L (Table 3.4-1 and Figure 3.4-5).

The Pine Falls Forebay was mesotrophic (0.350 to 0.650 mg/L) based on the mean of the open-water season TN concentrations for the three years of monitoring (0.43 mg/L). Open-water season mean annual TN concentrations (0.33 to 0.49 mg/L) were also within the meso-eutrophic range in 2011 and 2014 but were within the oligotrophic range (i.e., <0.350 mg/L) in 2017 (Table 3.4-2).

3.4.2.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

TN concentrations in Manigotagan Lake ranged from <0.20 to 0.96 mg/L during the open-water season. The mean and median concentrations for the 12 years of monitoring were 0.50 mg/L and 0.52 mg/L, respectively. Open-water season mean annual TN concentrations ranged from 0.34 to 0.66 mg/L but were within the IQR (0.45 to 0.58 mg/L) in seven of the 12 years. Mean TN concentrations were below the IQR in 2013, 2017, and 2018 and above the IQR in 2012, and 2016 (Table 3.4-3 and Figure 3.4-6).

TN concentrations in the ice-cover season ranged from 0.38 to 0.66 mg/L, and the mean and median were both 0.55 mg/L for the 12 years of monitoring. The IQR was 0.51 to 0.61 mg/L (Table 3.4-3 and Figure 3.4-6).

No clear seasonality was observed for TN in Manigotagan Lake over the 12 years of monitoring. However, the lowest mean TN concentration occurred in fall (0.46 mg/L) and the highest in spring (0.56 mg/L; Figure 3.4-4).

Manigotagan Lake was mesotrophic (0.350 to 0.650 mg/L) on the basis of the 2008-2019 mean open-water season TN concentration (0.50 mg/L). Mean annual TN concentrations (0.34 to 0.66 mg/L) in the open-water season were also within the mesotrophic range in ten of the 12 years of monitoring. However, the mean TN concentration was within the oligotrophic range (i.e., <0.350 mg/L) in 2018 and within the eutrophic range (0.651-1.20 mg/L) in 2012 (Table 3.4-4).

ROTATIONAL SITES

Eaglenest Lake

TN concentrations in Eaglenest Lake ranged from 0.27 to 0.70 mg/L during the open-water season. The mean and median concentrations for the four years of monitoring were 0.43 mg/L and 0.41 mg/L, respectively. Open-water season mean annual TN concentrations ranged from 0.36 to 0.57 mg/L and were within the IQR (0.35 to 0.49 mg/L) in all years sampled except for 2016 when it was above the IQR (Table 3.4-3 and Figure 3.4-6).

During the ice-cover season, TN concentrations ranged from 0.46 to 0.77 mg/L, with a mean of 0.59 mg/L (Table 3.4-3 and Figure 3.4-6).

Eaglenest Lake was mesotrophic (0.350 to 0.650 mg/L) based on the mean of the open-water season TN concentrations for the four years of monitoring (0.43 mg/L). Open-water season mean annual TN concentrations (0.36 to 0.57 mg/L) were also within the mesotrophic range in each year of monitoring (Table 3.4-4).

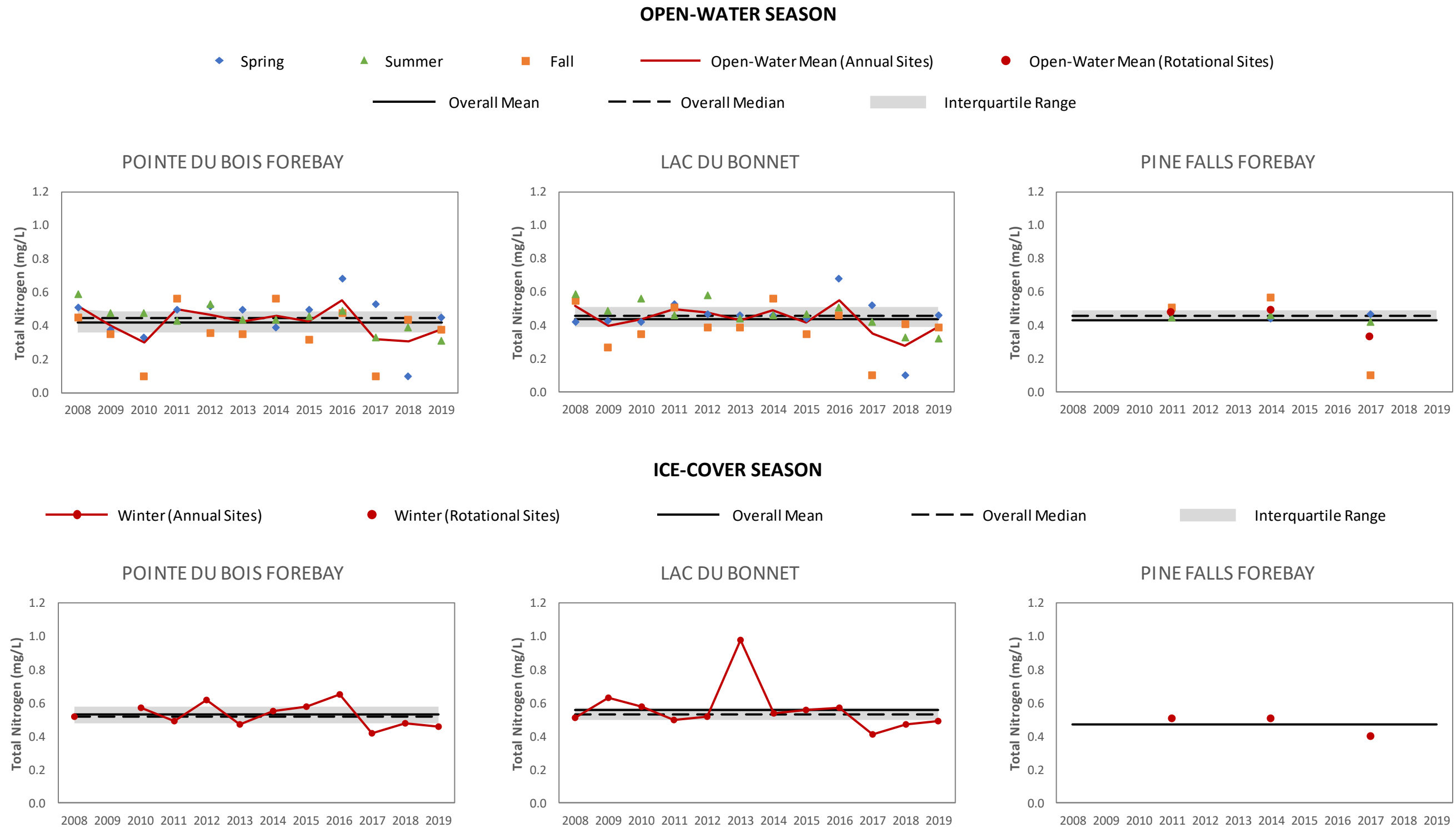


Figure 3.4-5. 2008-2019 On-system open-water and ice-cover season TN concentrations.

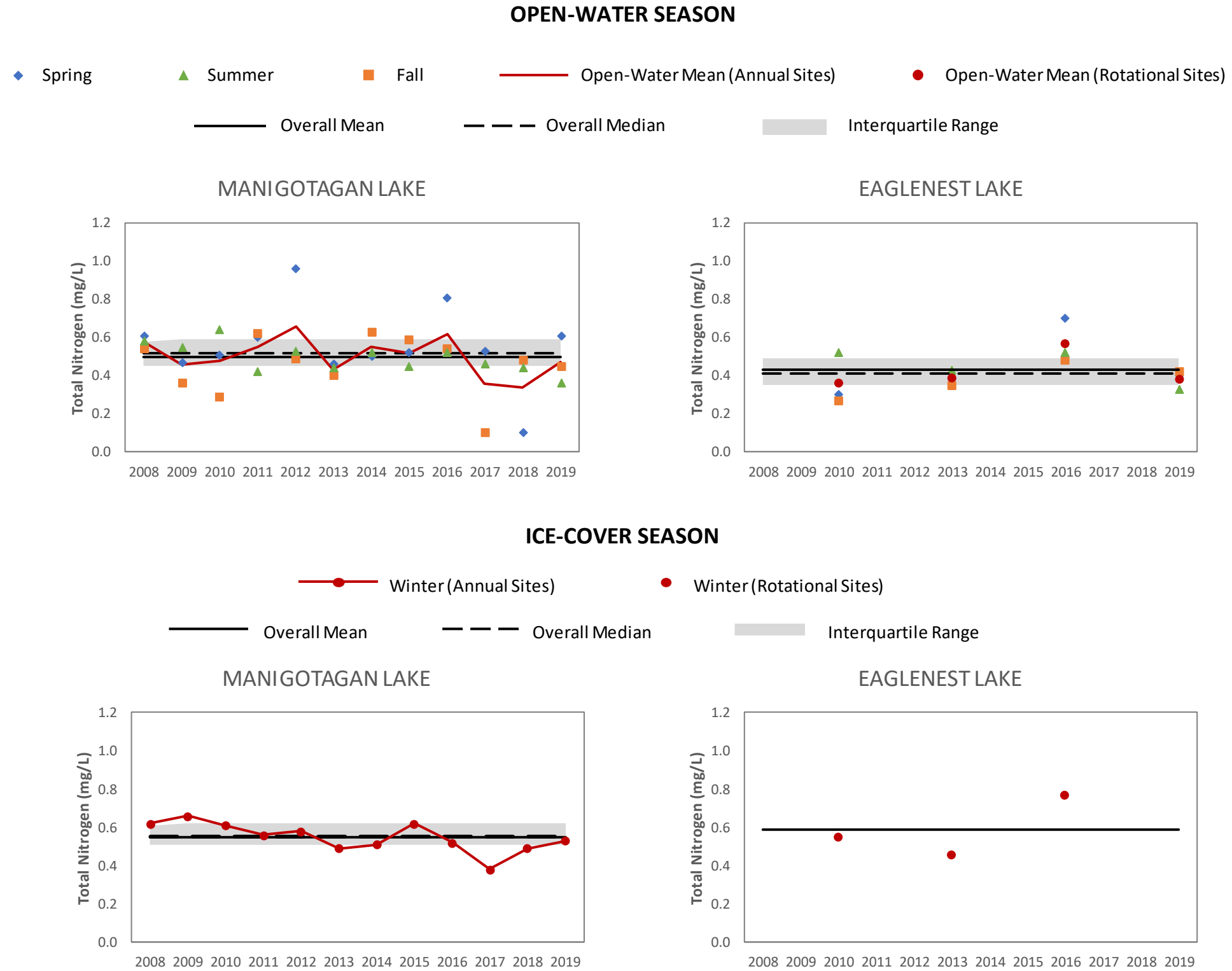


Figure 3.4-6. 2008-2019 Off-system open-water and ice-cover season TN concentrations.

3.4.3 CHLOROPHYLL A

3.4.3.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

Chlorophyll *a* concentrations in the Pointe du Bois Forebay ranged from 1.91 to 26.3 µg/L during the open-water season. The mean and median for the 12 years of monitoring were 7.56 µg/L and 6.46 µg/L, respectively. Open-water season mean annual chlorophyll *a* concentrations ranged from 3.52 to 12.2 µg/L and were within the IQR (5.06 to 8.05 µg/L) in six of the 12 years. Mean chlorophyll *a* concentrations were below the IQR in 2010 and above the IQR in 2011, 2013, 2015, 2018, and 2019 (Table 3.4-1 and Figure 3.4-7).

Chlorophyll *a* concentrations in the ice-cover season ranged from <0.60 to 2.06 µg/L, with a mean of 0.65 µg/L and a median of <0.60 µg/L for the 12 years of monitoring. The IQR was <0.60 to 0.76 µg/L (Table 3.4-1 and Figure 3.4-7).

Chlorophyll *a* concentrations were lower in the winter, often less than the DL (0.60-1.0 µg/L; percent detection = 45), compared to the open-water season (Table 3.4-1). On average, chlorophyll *a* concentrations during the open-water season were lowest in fall (5.12 µg/L) and highest in spring (11.1 µg/L; Figure 3.4-2).

The Pointe du Bois Forebay was mesotrophic (2.5 to 8 µg/L) on the basis of the 2008-2019 mean open-water season chlorophyll *a* concentration (7.56 µg/L). Mean annual chlorophyll *a* concentrations (3.52 to 12.2 µg/L) in the open-water season were within the mesotrophic range (2.5 to 8 µg/L) in 2008-2010, 2012, 2014, 2016, and 2017; and within the eutrophic range (8 to 25 µg/L) in 2011, 2013, 2015, 2018, and 2019 (Table 3.4-2).

Lac du Bonnet

Chlorophyll *a* concentrations in Lac du Bonnet ranged from 0.84 to 17.8 µg/L during the open-water season. The mean and median for the 12 years of monitoring were 7.02 µg/L and 6.21 µg/L, respectively. Open-water season mean annual chlorophyll *a* concentrations ranged from 3.64 to 10.4 µg/L and were within the IQR (5.10 to 8.49 µg/L) in seven of the 12 years. Mean chlorophyll *a*

concentrations were below the IQR in 2010 and 2012 and above the IQR in 2011, 2013, and 2019 (Table 3.4-1 and Figure 3.4-7).

Chlorophyll *a* concentrations in the ice-cover season ranged from <0.60 to 4.96 µg/L, with a mean of 1.19 µg/L and a median of 0.73 µg/L for the 12 years of monitoring. Chlorophyll *a* concentrations were within or near the IQR (0.60 to 1.39 µg/L) in most years, the exception was winter 2017 when it was above the IQR. (Table 3.4-1 and Figure 3.4-7).

Chlorophyll *a* concentrations were lower in the winter, often less than the DL (0.60-1.0 µg/L; percent detection = 50), compared the open-water season (Table 3.4-1). On average, chlorophyll *a* concentrations during the open-water season were lowest in fall (4.80 µg/L) and highest in spring (9.37 µg/L; Figure 3.4-2).

Lac du Bonnet was mesotrophic (2.5 to 8 µg/L) on the basis of the 2008-2019 mean open-water season chlorophyll *a* concentration (7.02 µg/L). Mean annual chlorophyll *a* concentrations (3.64 to 10.4 µg/L) in the open-water season were within the mesotrophic range (2.5 to 8 µg/L) in most years except for 2011, 2013, and 2019 when the mean chlorophyll *a* concentrations were within the eutrophic range (8 to 25 µg/L; Table 3.4-2).

ROTATIONAL SITES

Pine Falls Forebay

Chlorophyll *a* concentrations in the Pine Falls Forebay ranged from 2.10 to 18.1 µg/L during the open-water season. The mean was 7.46 µg/L, the median was 5.54 µg/L, and the IQR was 4.28 to 9.93 µg/L for the three years of monitoring. Mean annual chlorophyll *a* concentrations in the open-water season ranged from 5.73 to 9.56 µg/L and were within the IQR in all years (Table 3.4-1 and Figure 3.4-7).

During the ice-cover season, chlorophyll *a* concentrations ranged from 1.53 to 1.91 µg/L, with a mean of 1.72 µg/L (Table 3.4-1 and Figure 3.4-7).

The Pine Falls Forebay was mesotrophic (2.5 to 8 µg/L) based on the mean of the open-water season chlorophyll *a* concentrations for the three years of monitoring (7.46 µg/L). Open-water season mean annual chlorophyll *a* concentrations (5.73 to 9.56 µg/L) were also within the mesotrophic range in 2014 and 2017 but were within the eutrophic range (8 to 25 µg/L) in 2011 (Table 3.4-2).

3.4.3.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

Chlorophyll *a* concentrations in Manigotagan Lake ranged from 1.53 to 19.6 µg/L during the open-water season. The mean and median concentrations for the 12 years of monitoring were 6.73 µg/L and 6.06 µg/L, respectively. Open-water season mean annual chlorophyll *a* concentrations ranged from 3.64 to 8.97 µg/L and were within the IQR (4.34 to 8.16 µg/L) in eight of the 12 years. Mean chlorophyll *a* concentrations were below the IQR in 2010 and above the IQR in 2014, 2017, and 2018 (Table 3.4-3 and Figure 3.4-8).

Chlorophyll *a* concentrations in the ice-cover season ranged from 0.68 to 6.11 µg/L, with a mean of 1.85 µg/L and a median of 1.31 µg/L for the 12 years of monitoring. Chlorophyll *a* concentrations during the ice-cover season were within or near the IQR (0.99 to 1.96 µg/L) in all years except 2016 when the concentration was above the IQR (Table 3.4-3 and Figure 3.4-8).

Chlorophyll *a* concentrations were lower under ice-cover (mean = 1.85 µg/L) than during the open-water season (mean = 6.73 µg/L). On average, chlorophyll *a* concentrations during the open-water season were lowest in fall (5.06 µg/L) and highest in spring (8.62 µg/L; Figure 3.4-4).

Manigotagan Lake was mesotrophic (2.5 to 8 µg/L) on the basis of the 2008-2019 mean open-water season chlorophyll *a* concentration (6.73 µg/L). Mean annual chlorophyll *a* concentrations (3.64 to 8.97 µg/L) in the open-water season were also within the mesotrophic range in nine of the 12 years. However, the mean chlorophyll *a* concentration was within the eutrophic range (8 to 25 µg/L) in 2014, 2017, and 2018 (Table 3.4-4).

ROTATIONAL SITES

Eaglenest Lake

Chlorophyll *a* concentrations in Eaglenest Lake ranged from 1.40 to 15.1 µg/L during the open-water season. The mean and median concentrations for the four years of monitoring were 6.48 µg/L and 6.03 µg/L, respectively. Open-water season mean annual chlorophyll *a* concentrations ranged from 4.22 to 9.29 µg/L and were within the IQR (4.78 to 7.45 µg/L) in 2013 and 2016 but were below the IQR in 2010 and above the IQR in 2019 (Table 3.4-3 and Figure 3.4-8).

During the ice-cover season, chlorophyll *a* concentrations were consistently below the DL (0.60 µg/L; Table 3.4-3 and Figure 3.4-8).

Eaglenest Lake was mesotrophic (2.5 to 8 µg/L) based on the mean of the open-water season chlorophyll *a* concentrations for the four years of monitoring (6.48 µg/L). Open-water season mean annual chlorophyll *a* concentrations (4.22 to 9.29 µg/L) were also within the meso-eutrophic range in 2010, 2013, and 2016 but were within the eutrophic range (8 to 25 µg/L) in 2019 (Table 3.4-4).

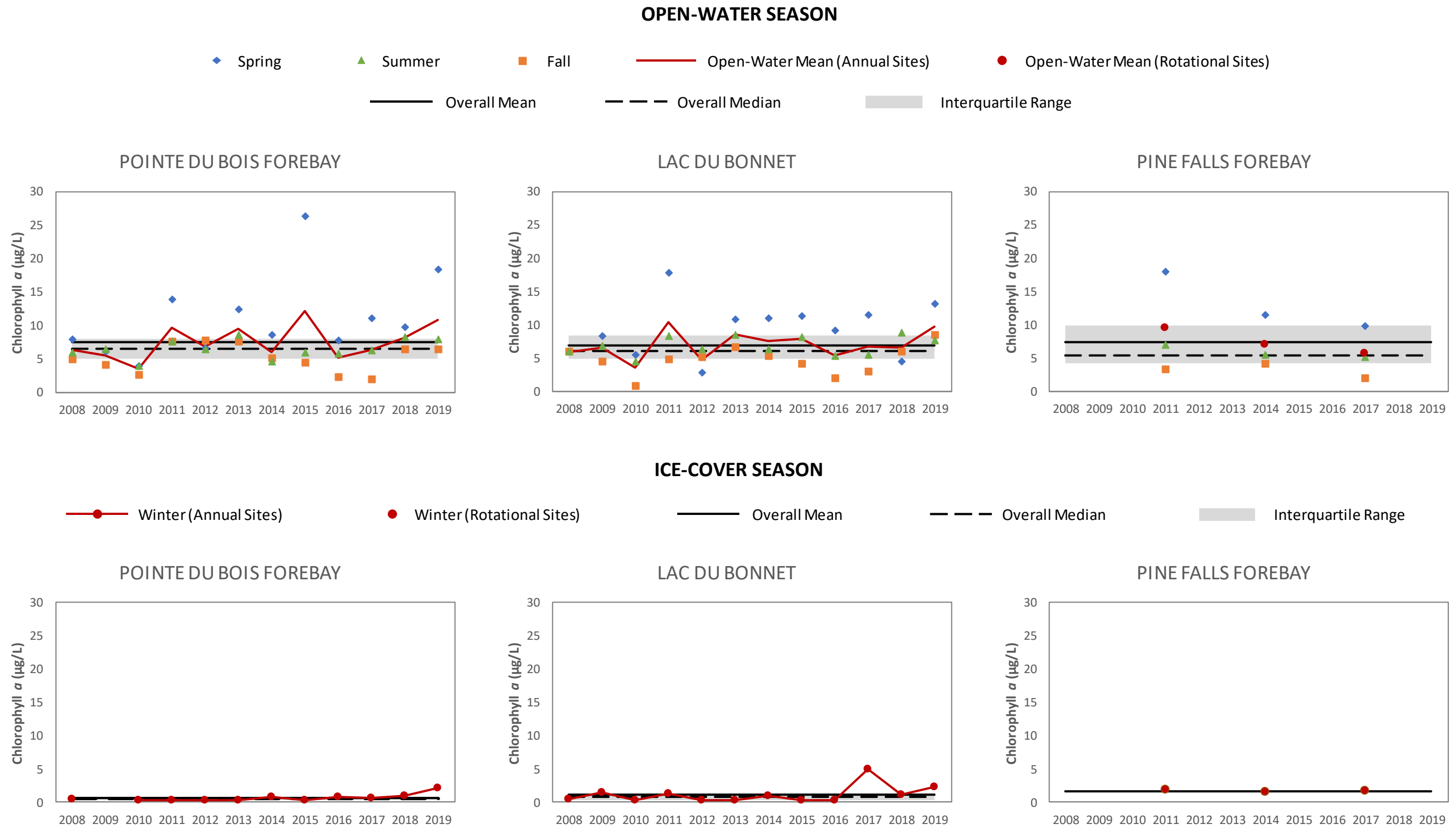


Figure 3.4-7. 2008-2019 On-system open-water and ice-cover season chlorophyll *a* concentrations.

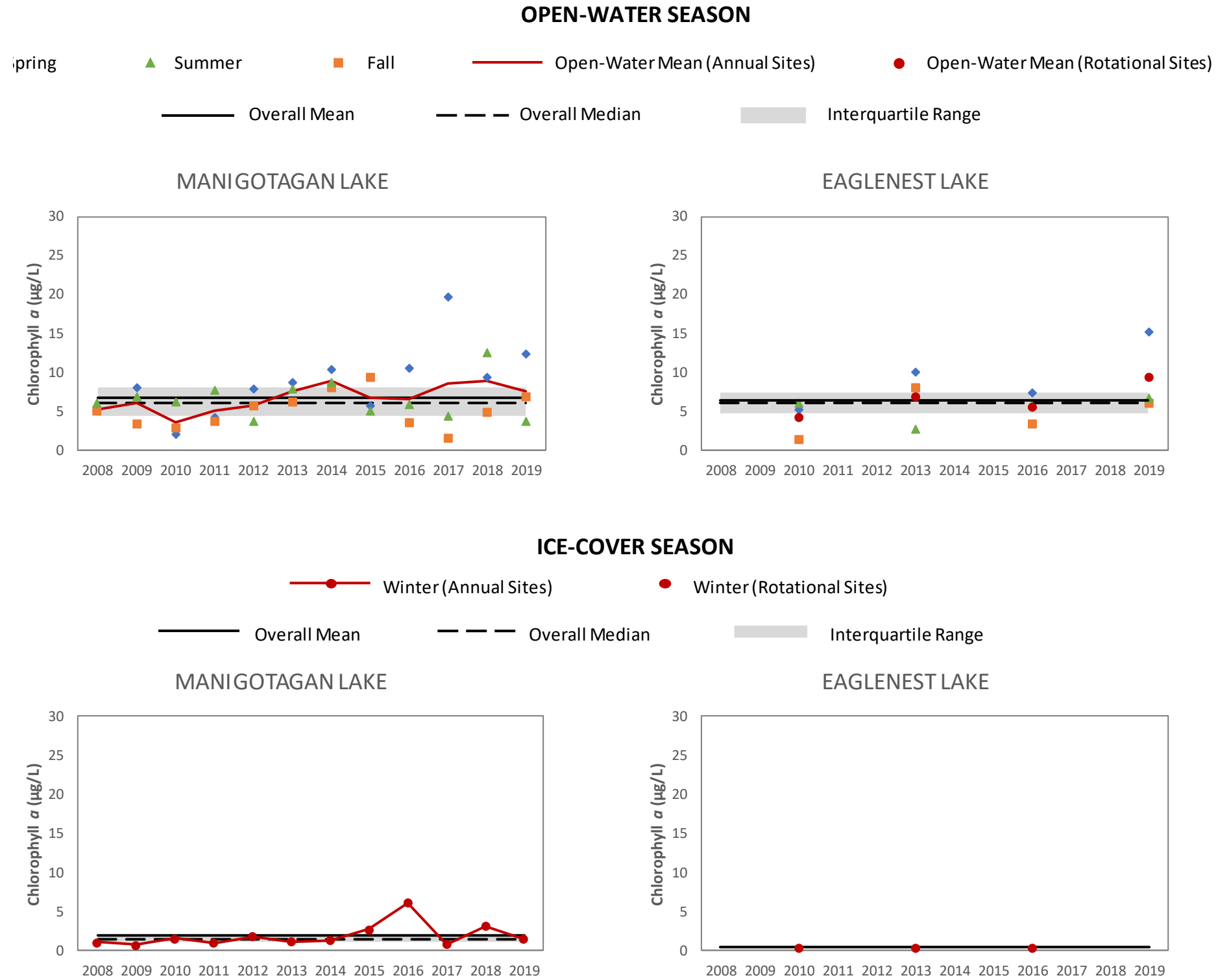


Figure 3.4-8. 2008-2019 Off-system open-water and ice-cover season chlorophyll *a* concentrations.

APPENDIX 3-1. WATER QUALITY SAMPLING SITES: 2008-2019

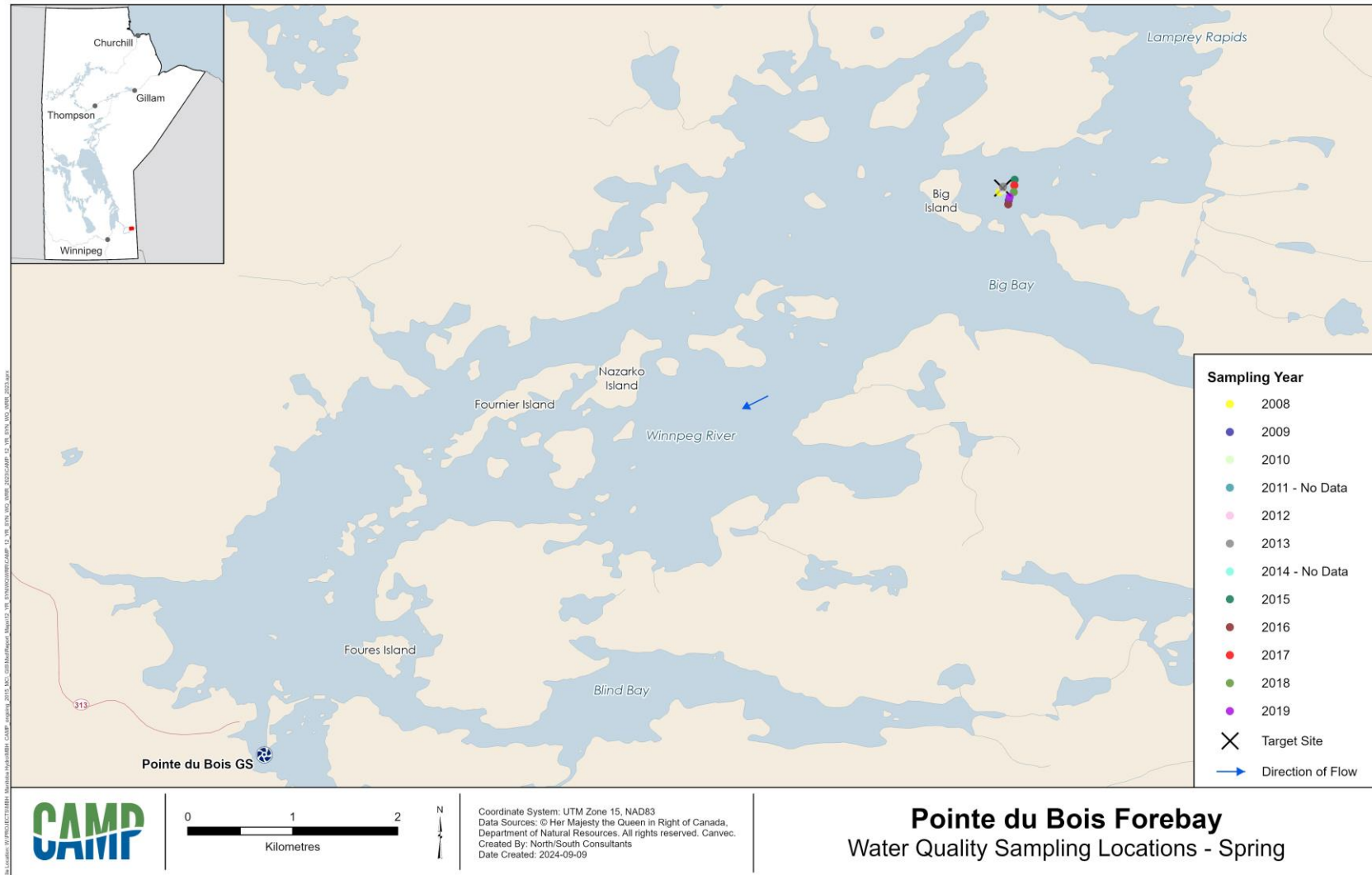


Figure A3-1-1. Spring water quality sampling locations: Pointe du Bois Forebay.

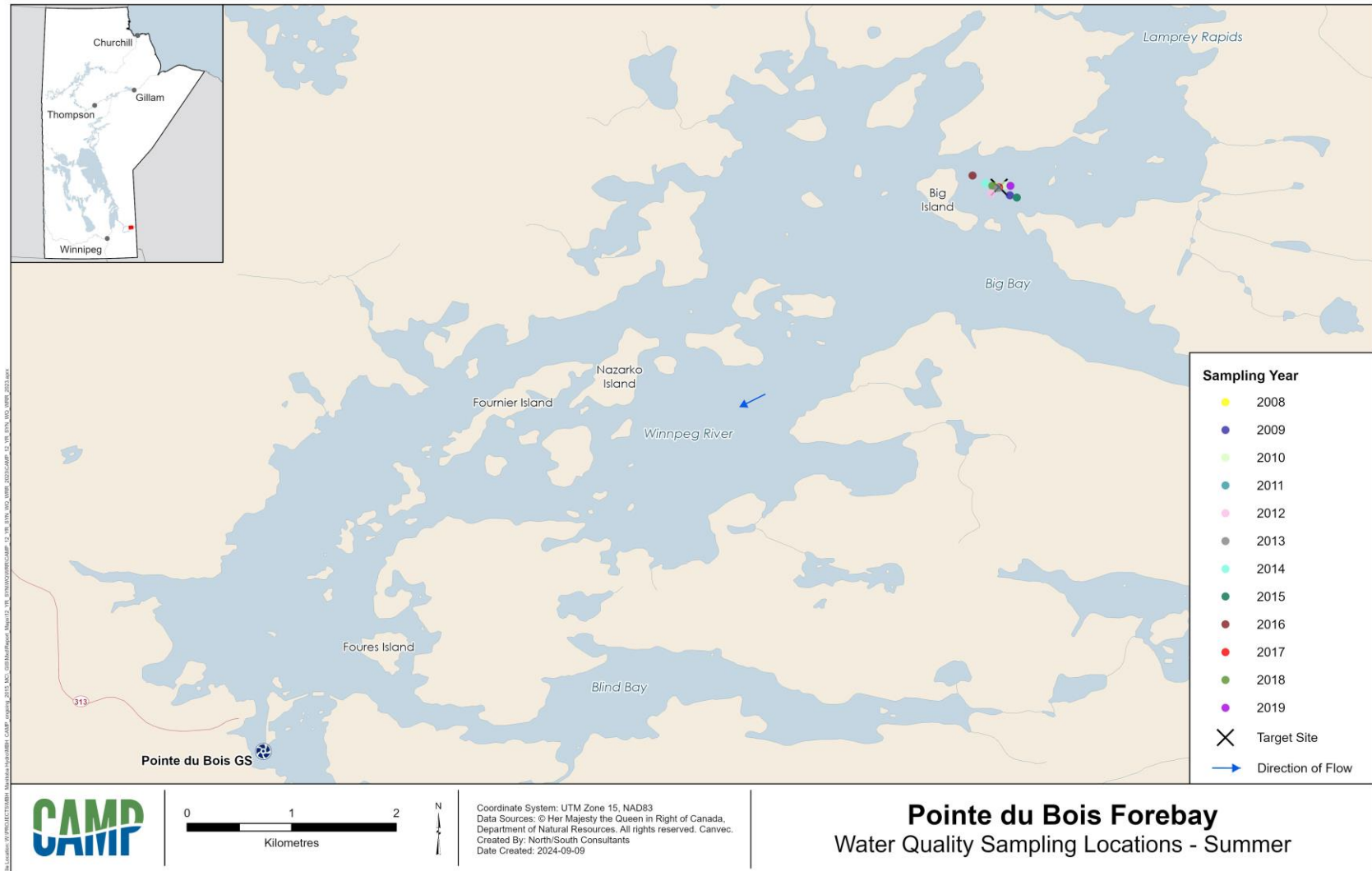


Figure A3-1-2. Summer water quality sampling locations: Pointe du Bois Forebay.

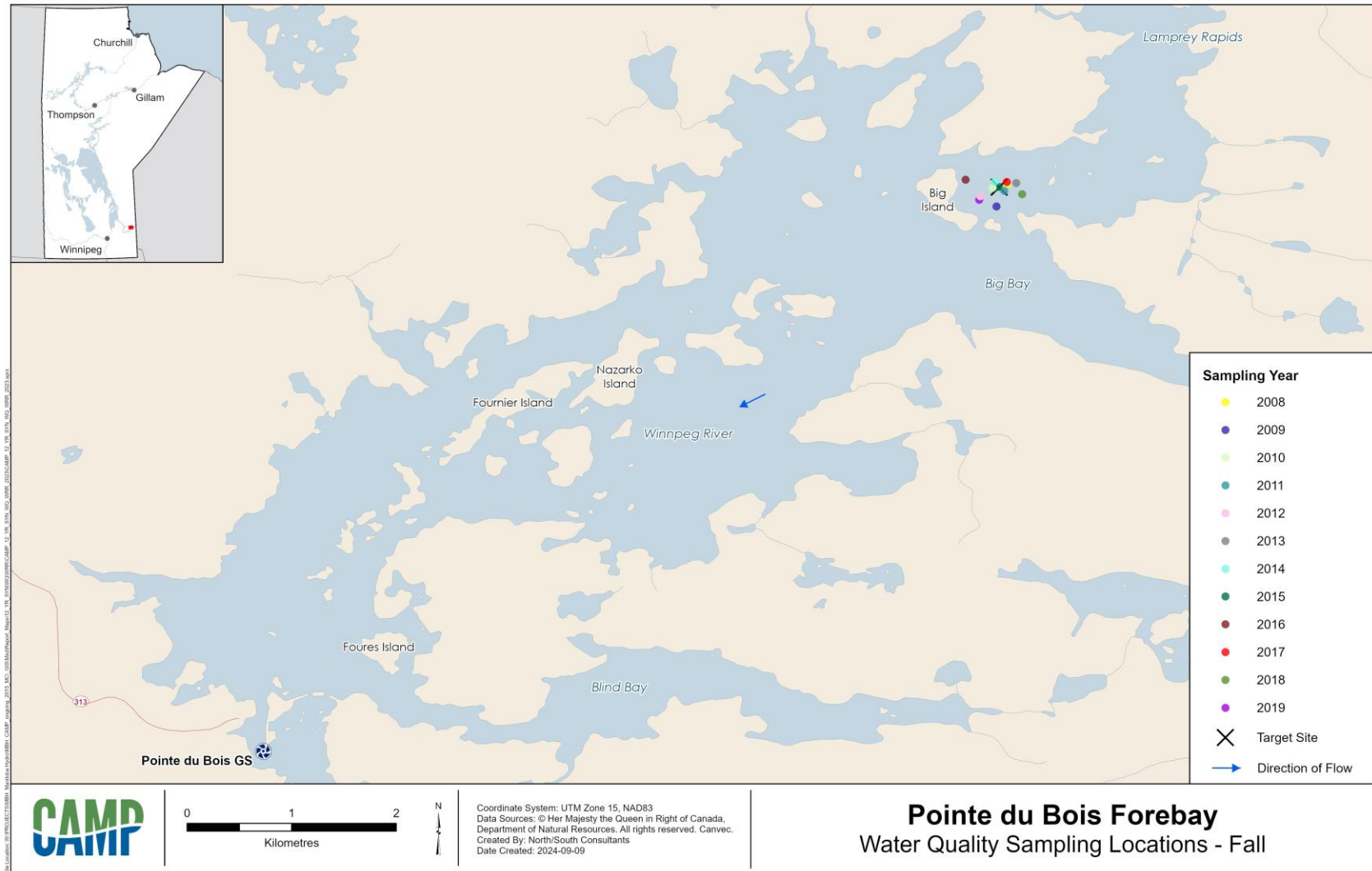


Figure A3-1-3. Fall water quality sampling locations: Pointe du Bois Forebay.

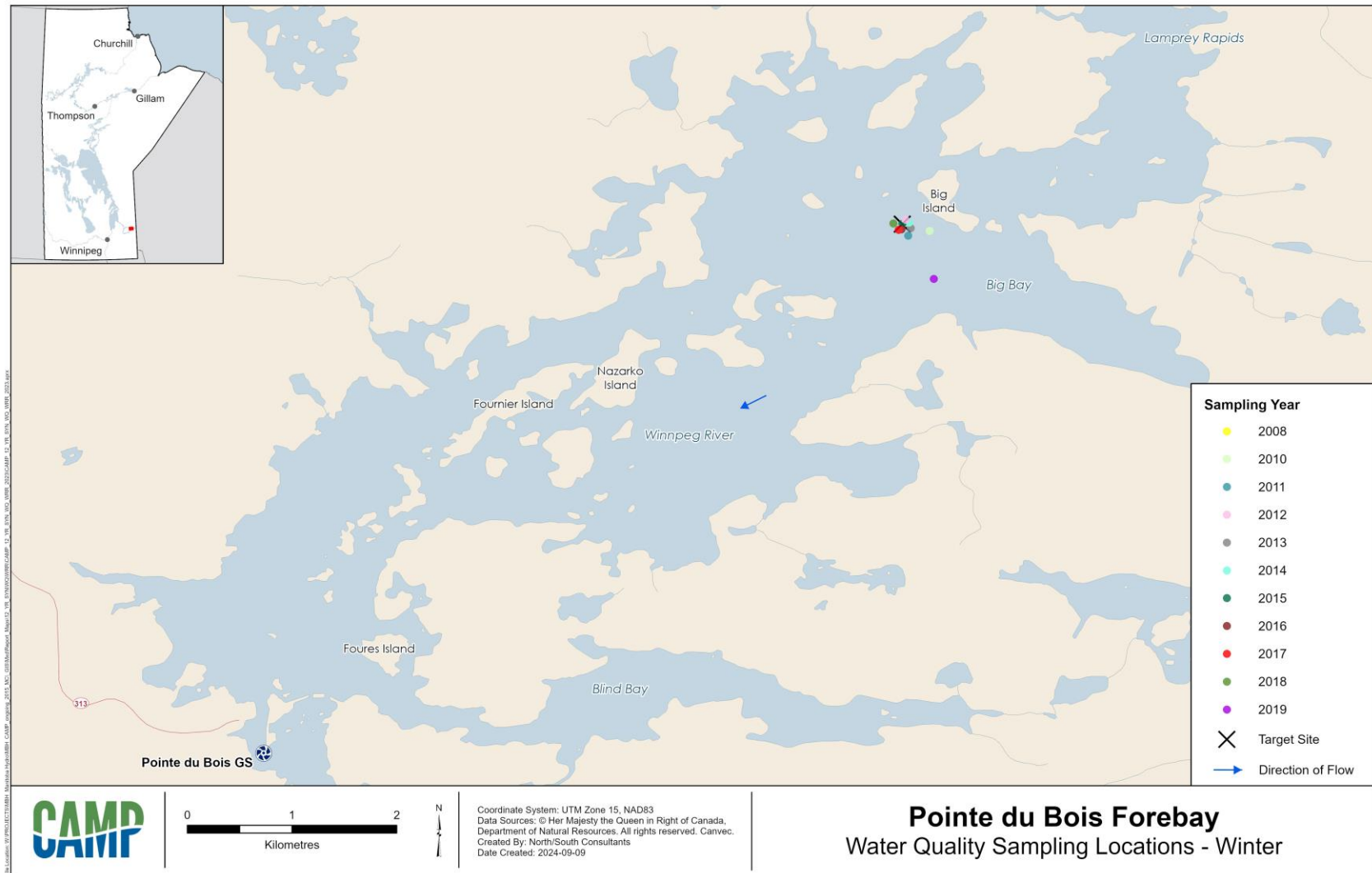


Figure A3-1-4. Winter water quality sampling locations: Pointe du Bois Forebay.

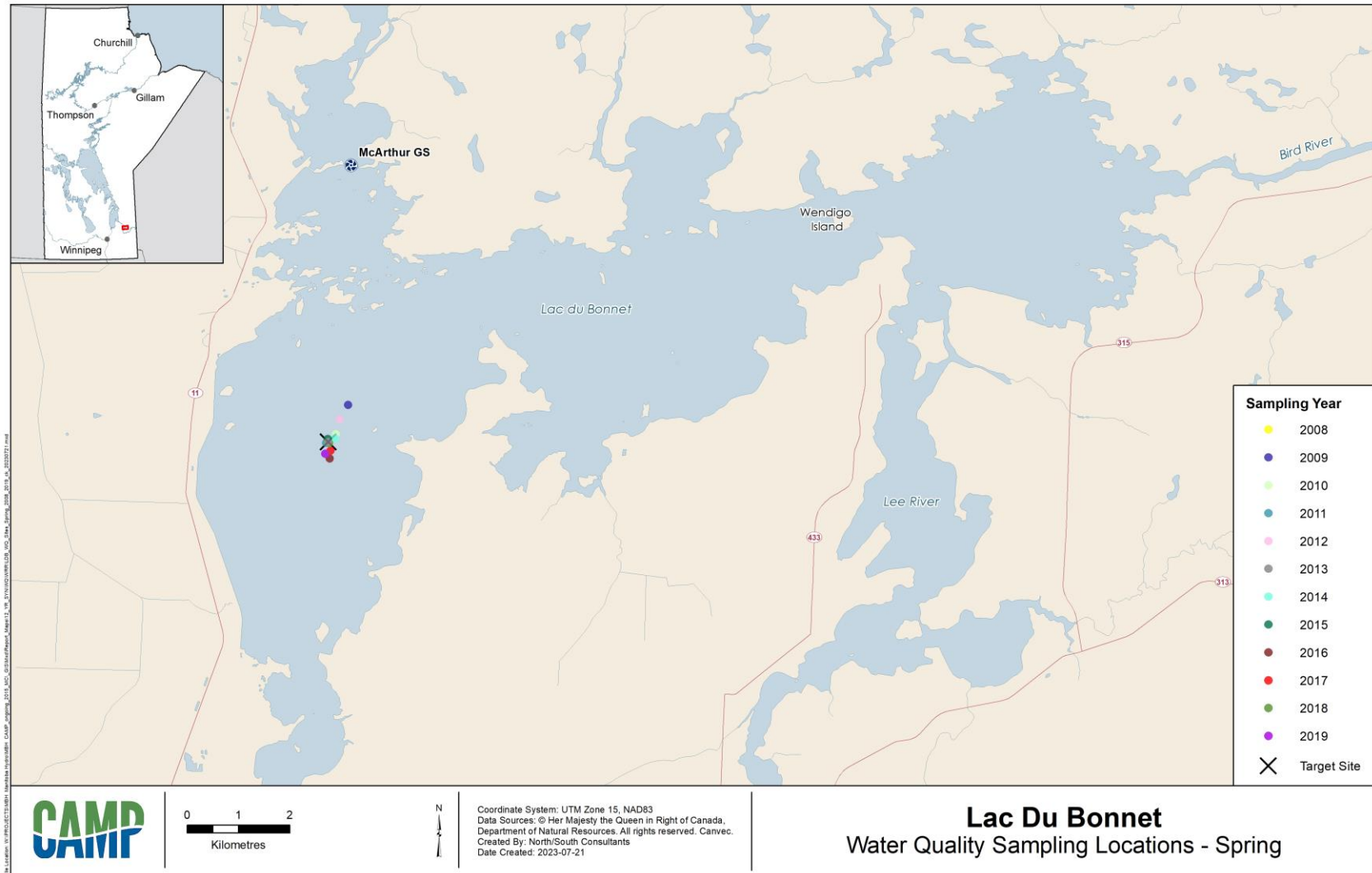


Figure A3-1-5. Spring water quality sampling locations: Lac du Bonnet.

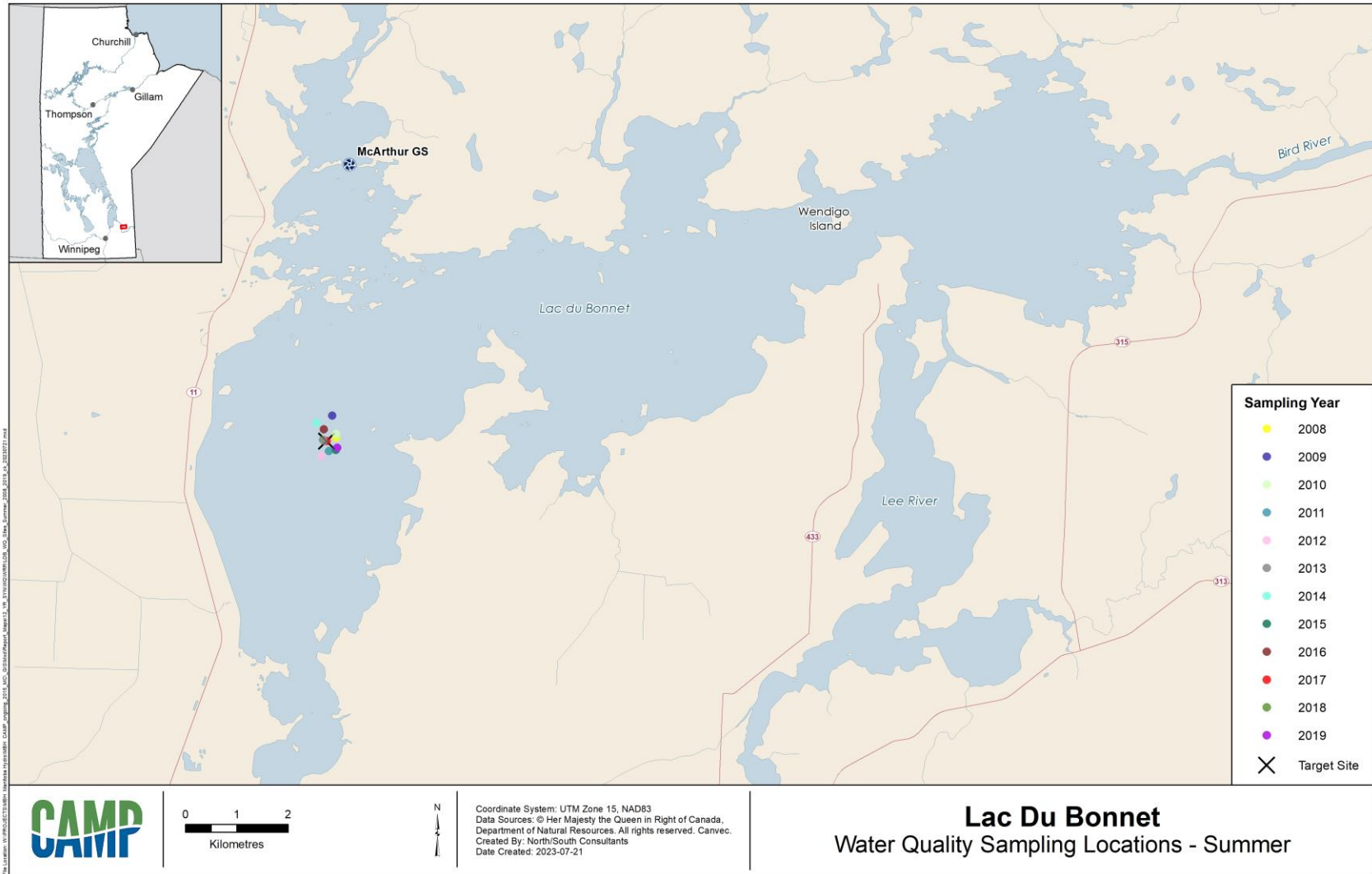


Figure A3-1-6. Summer water quality sampling locations: Lac du Bonnet.

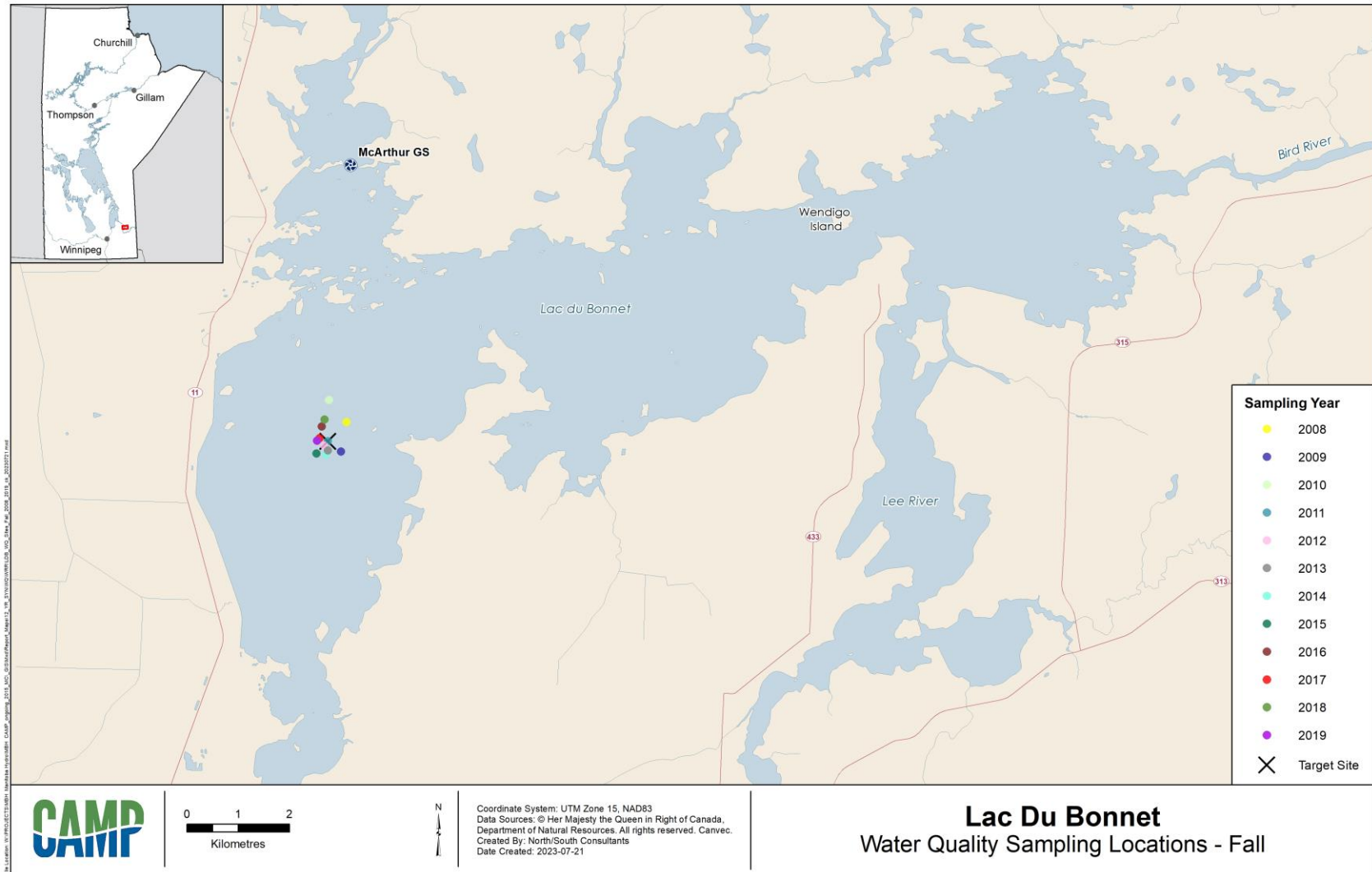


Figure A3-1-7. Fall water quality sampling locations: Lac du Bonnet.

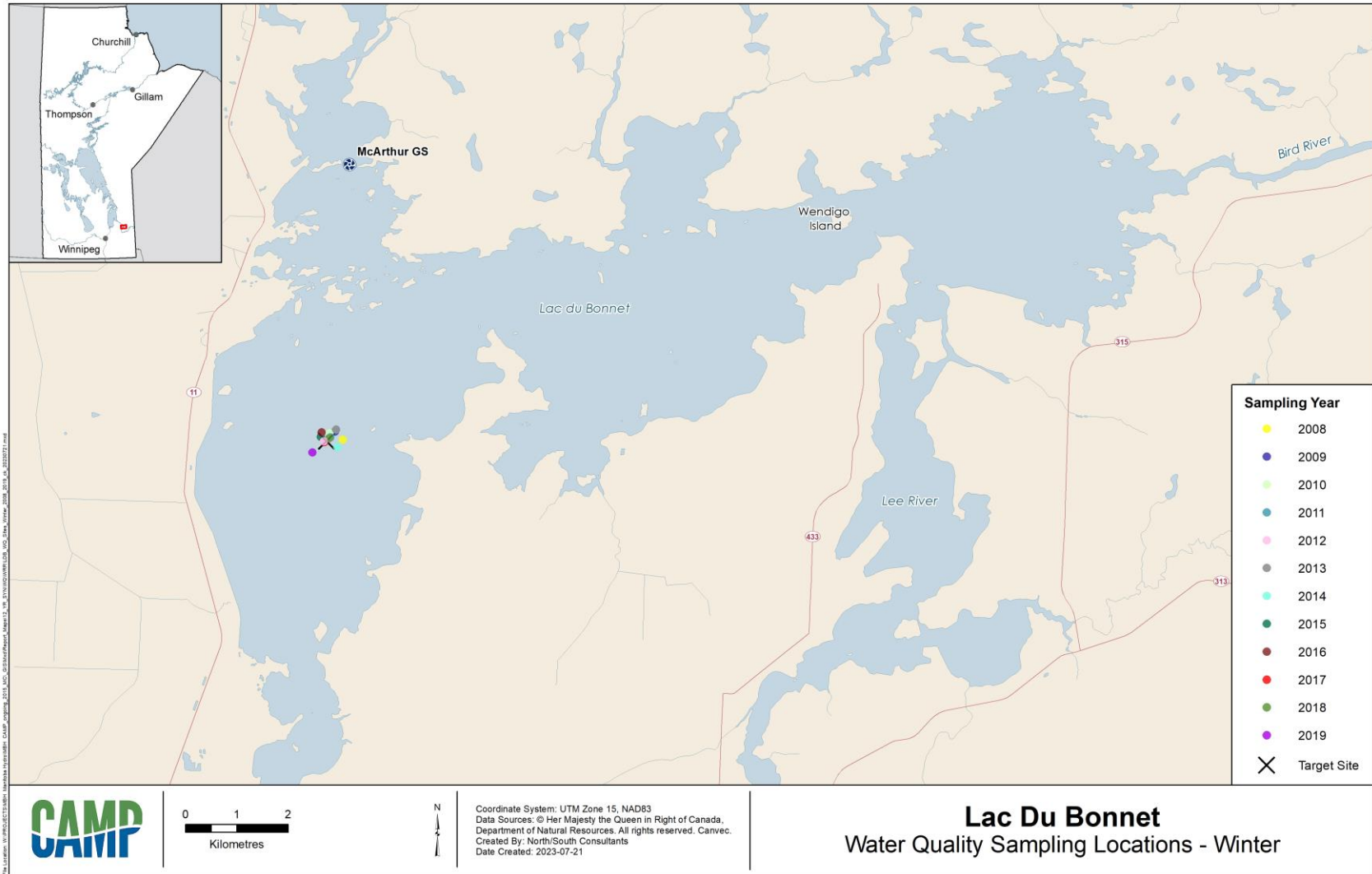


Figure A3-1-8. Winter water quality sampling locations: Lac du Bonnet.

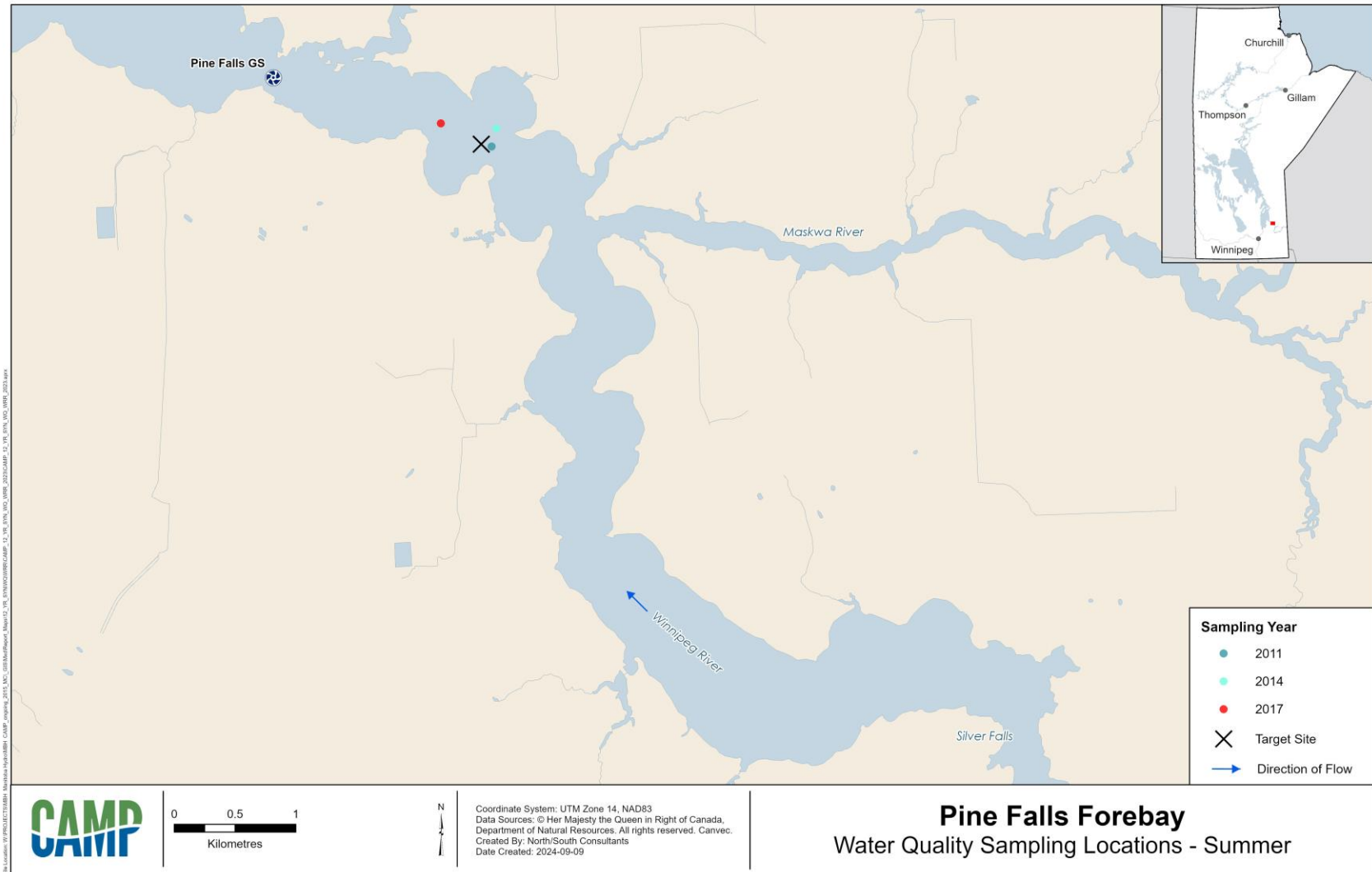


Figure A3-1-10. Summer water quality sampling locations: Pine Falls Forebay.

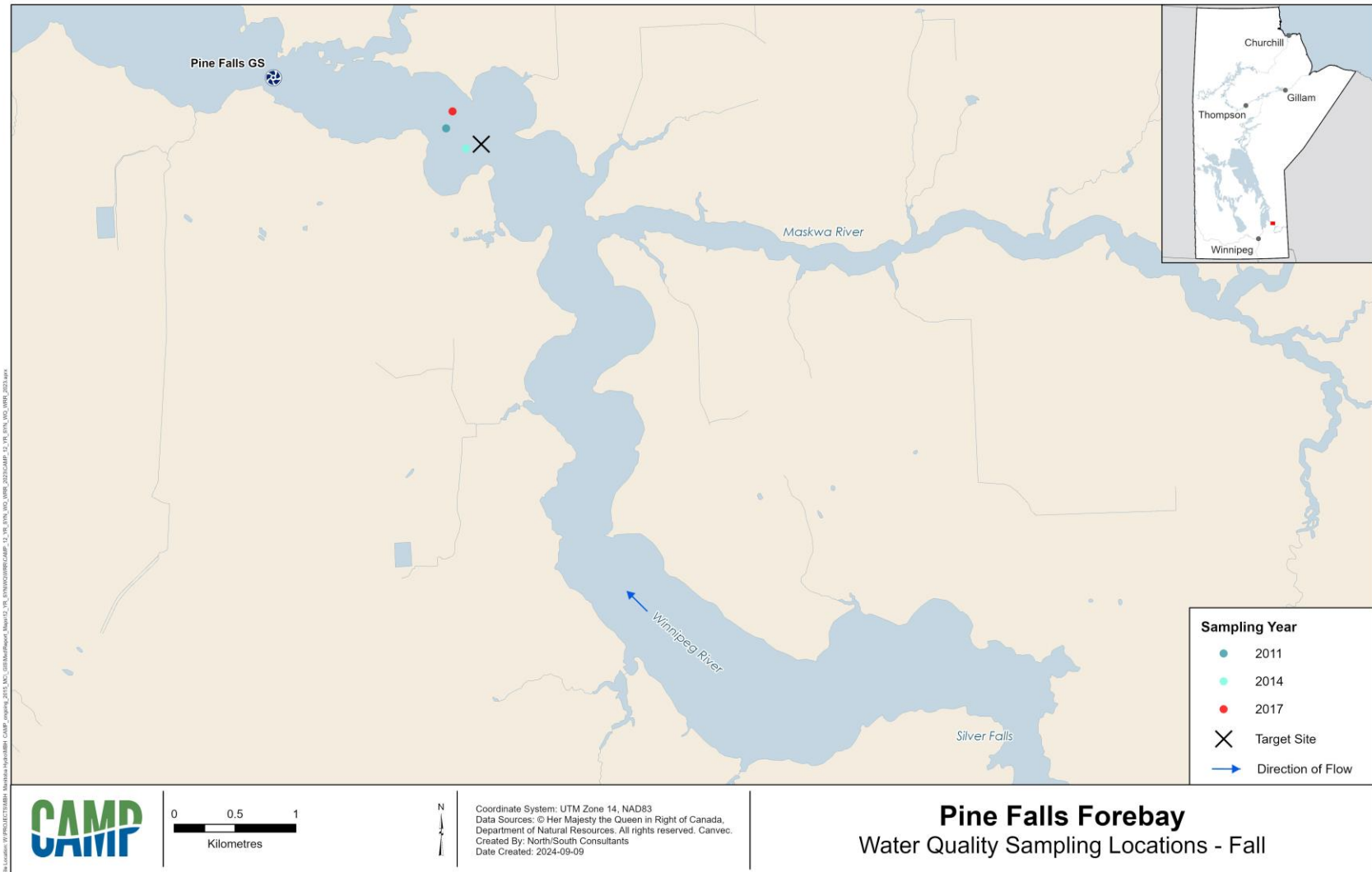


Figure A3-1-11. Fall water quality sampling locations: Pine Falls Forebay.

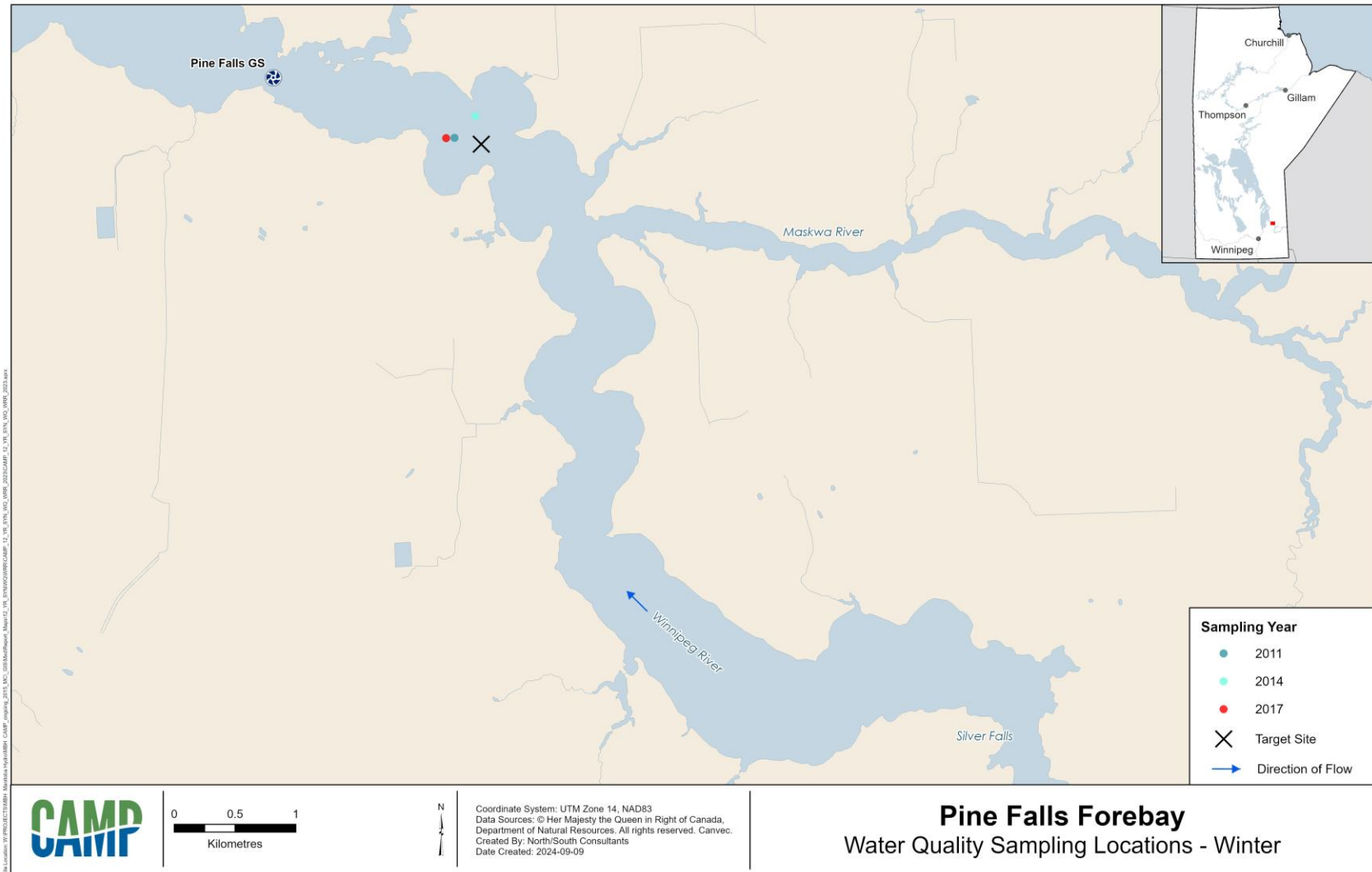


Figure A3-1-12. Winter water quality sampling locations: Pine Falls Forebay.

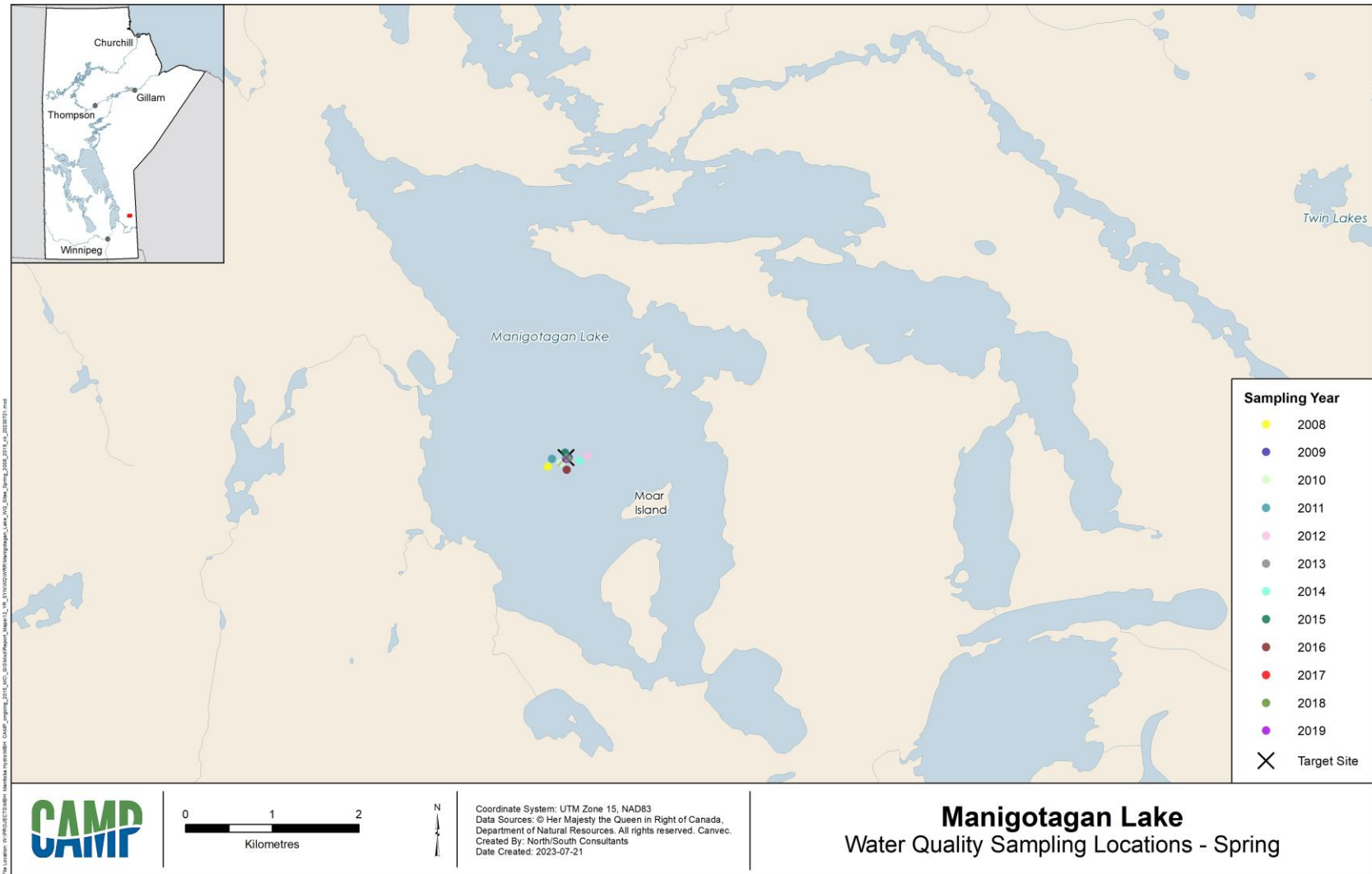


Figure A3-1-13. Spring water quality sampling locations: Manigotagan Lake.

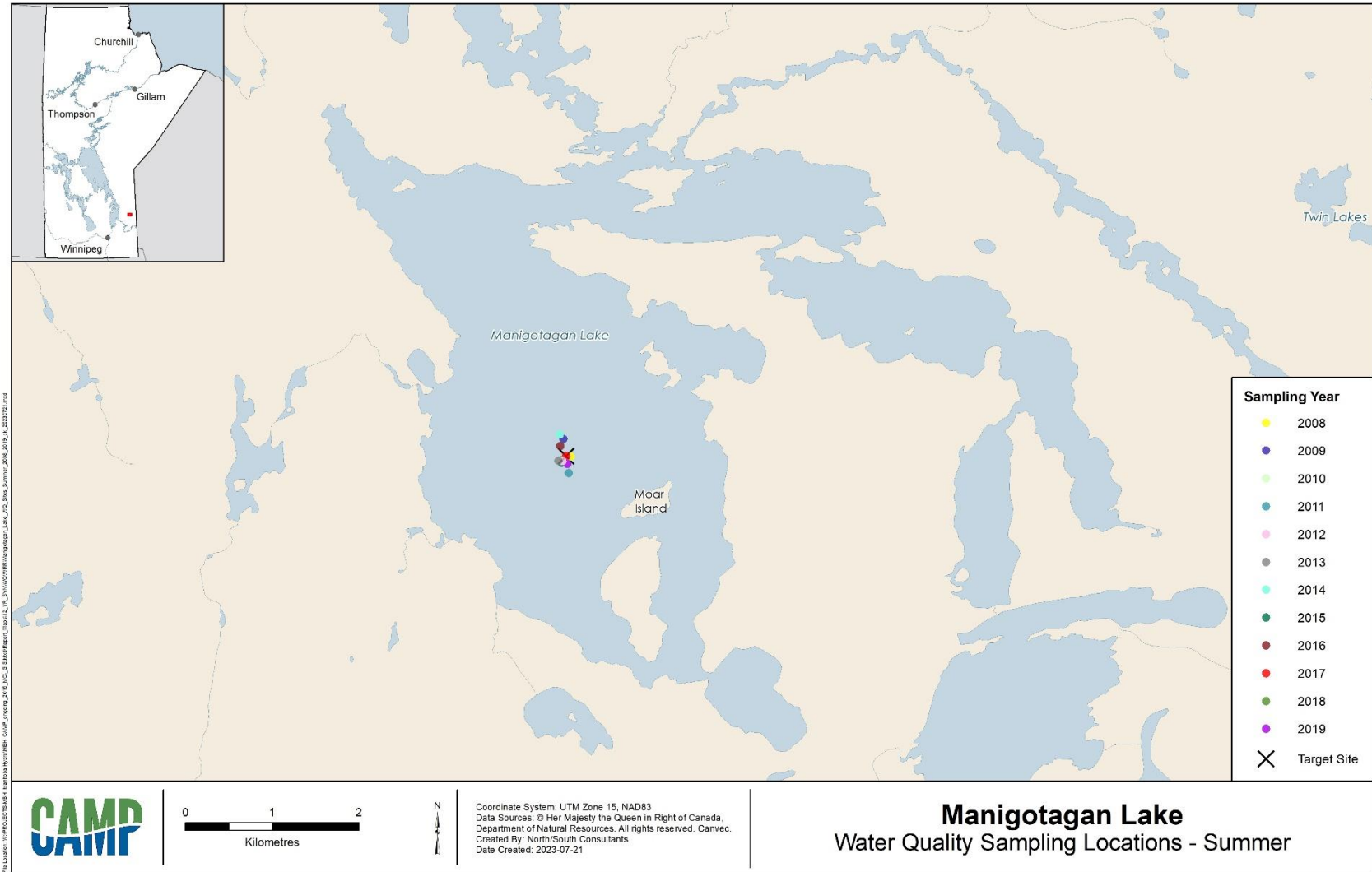


Figure A3-1-14. Summer water quality sampling locations: Manigotagan Lake.

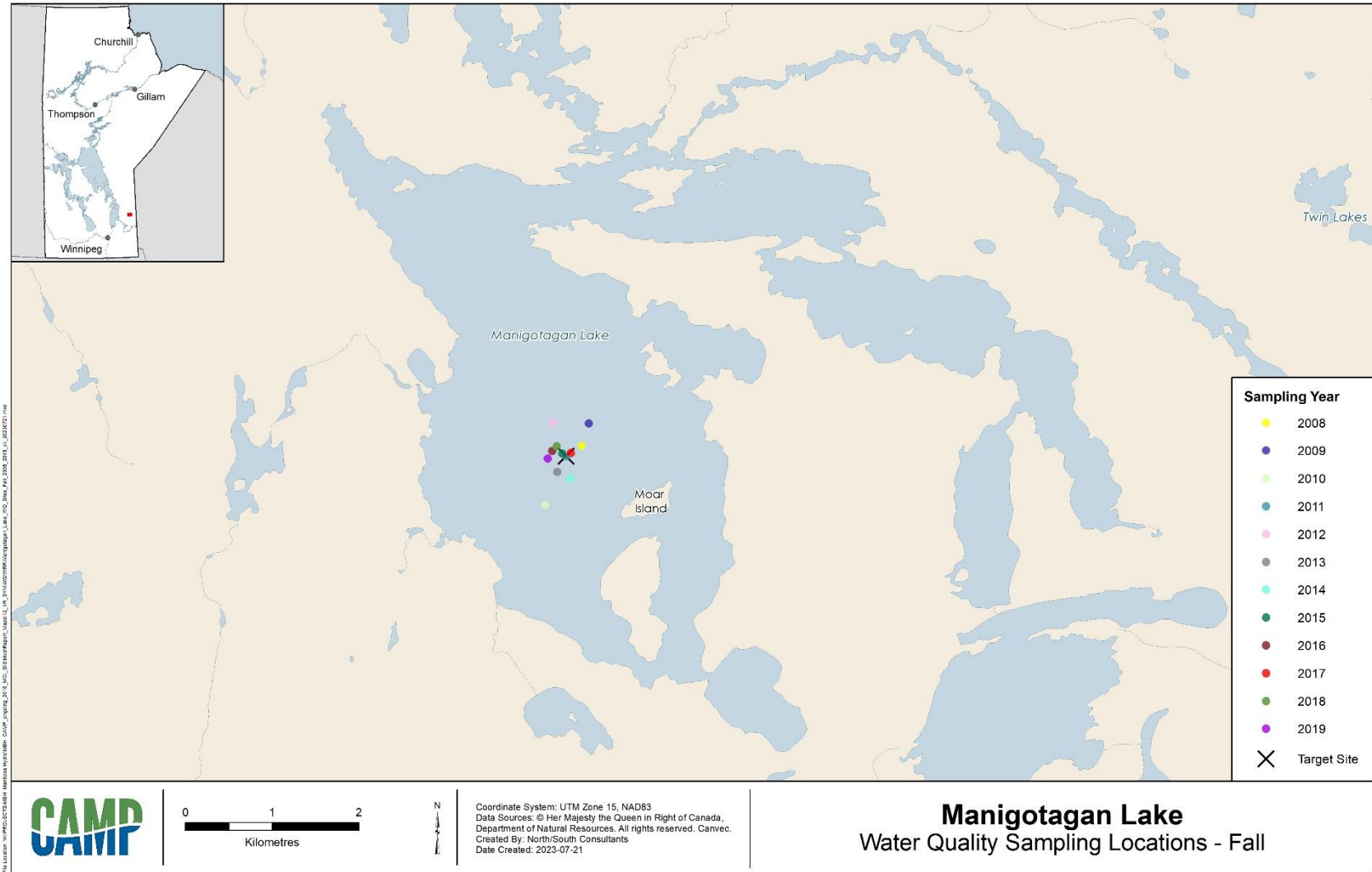


Figure A3-1-15. Fall water quality sampling locations: Manigotagan Lake.

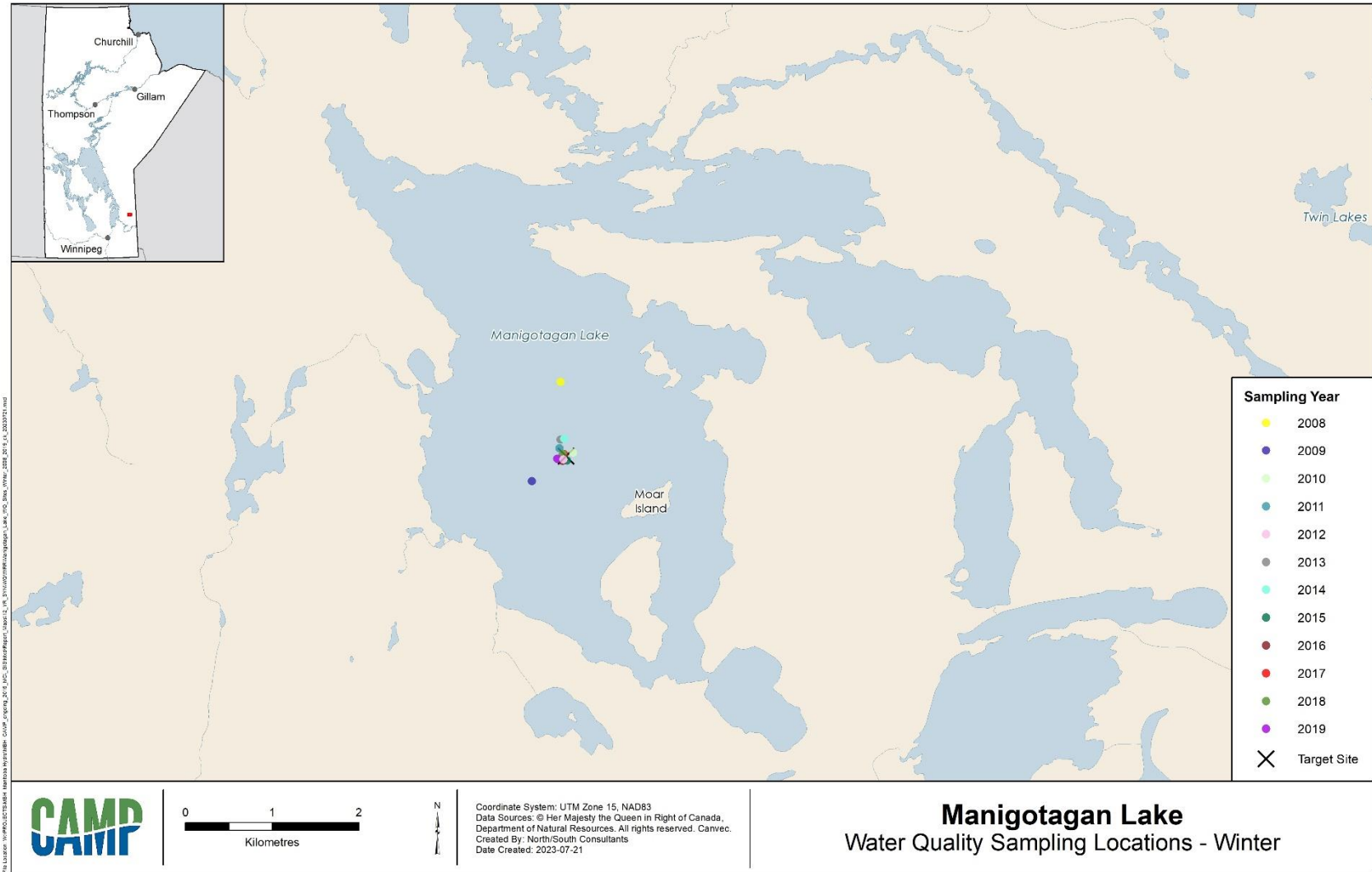


Figure A3-1-16. Winter water quality sampling locations: Manigotagan Lake.



Figure A3-1-17. Spring water quality sampling locations: Eaglenest Lake.

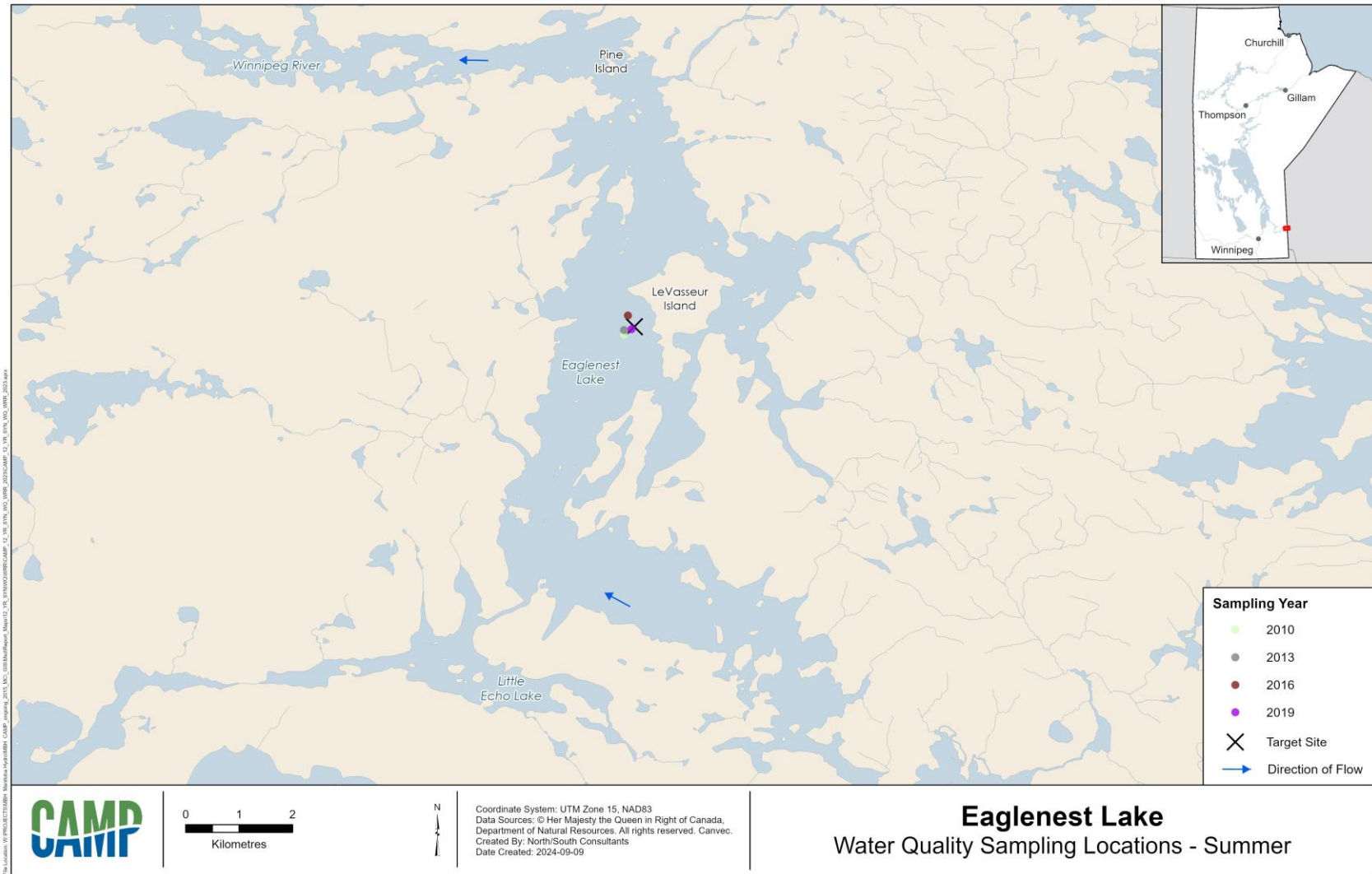


Figure A3-1-18. Summer water quality sampling locations: Eaglenest Lake.

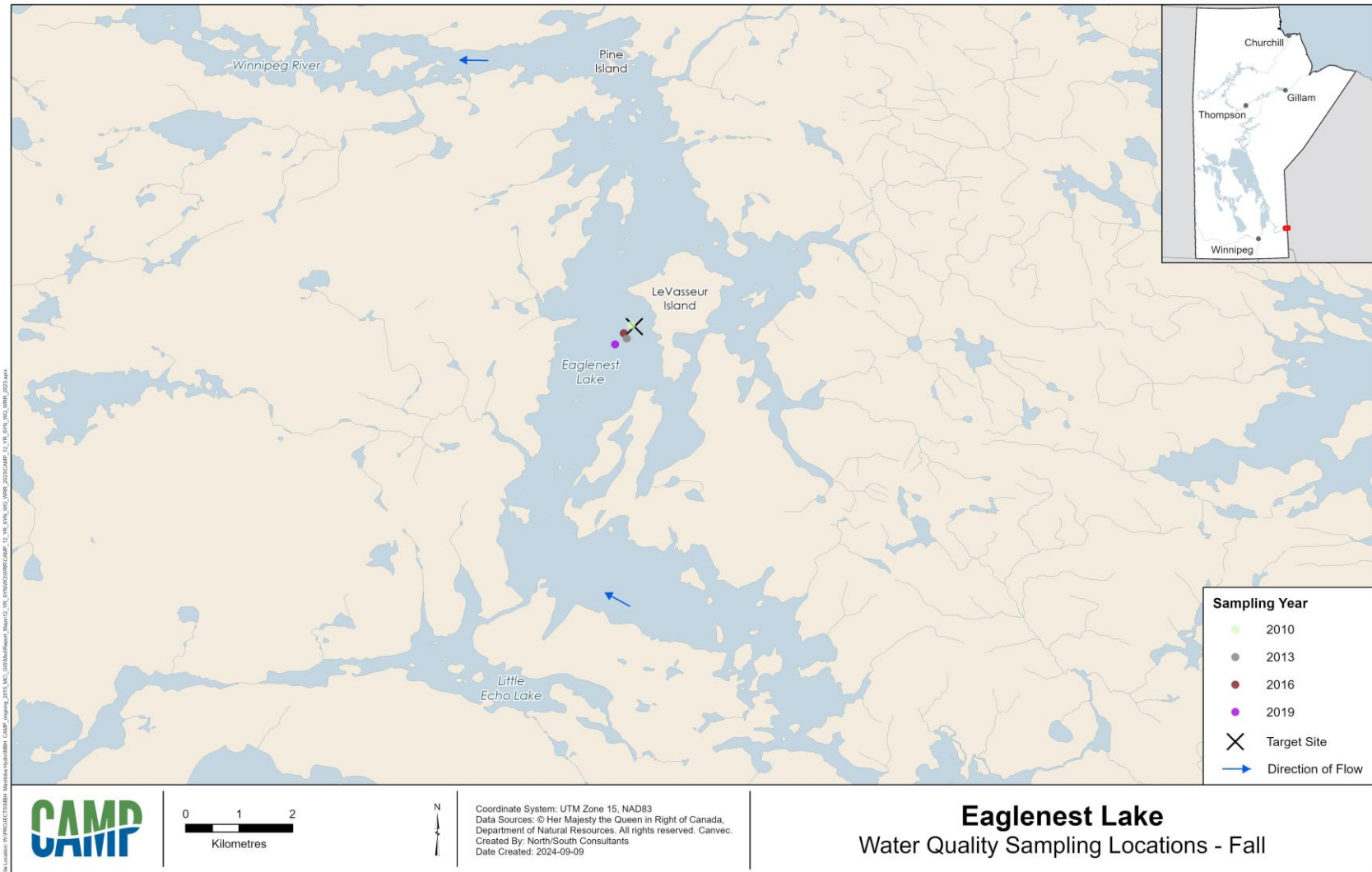


Figure A3-1-19. Fall water quality sampling locations: Eglenest Lake.



Figure A3-1-20. Winter water quality sampling locations: Eaglenest Lake.

4.0 BENTHIC INVERTEBRATES

4.1 INTRODUCTION

The following presents the results of the benthic invertebrate community monitoring conducted from 2010-2019 in the Winnipeg River Region. The 2008 and 2009 benthic invertebrate datasets were excluded due to a significant change in the sampling design in 2010.

Five waterbodies were monitored in the Winnipeg River Region: two on-system annual sites (the Pointe du Bois Forebay and Lac du Bonnet) and one on-system rotational site (the Pine Falls Forebay); and one off-system annual site (Manigotagan Lake) and one off-system rotational site (Eaglenest Lake; Table 4.1-1 and Figure 4.1-1). Eaglenest Lake is located on the Winnipeg River upstream of the Pointe du Bois Forebay and is not affected by MH's hydraulic operating system and is therefore considered an "off-system" site.

Two sampling polygons (nearshore [NS] and offshore [OS]) defined by water depth, flow, and substrate composition were sampled in each waterbody in late summer/fall per year (Appendix A4-1). Five benthic invertebrate samples were collected in each polygon for a total of ten invertebrate samples per waterbody per year. Five sediment samples were also collected in each polygon (where possible) to provide supporting information on substrate composition, total organic carbon (TOC), and texture. Dominant substrate type(s) and sediment analysis results are presented in Appendix 4-2. Sampling was completed at all sites as planned over the period of 2010-2019.

Four benthic invertebrate indicators (abundance, community composition, taxonomic richness, and diversity) were selected for detailed reporting (Table 4.1-2). Metrics for these indicators that are presented herein include: total invertebrate abundance or total invertebrate density; the Ephemeroptera, Plecoptera, and Trichoptera (EPT) Index; the Oligochaeta and Chironomidae (O+C) Index; total taxa richness; EPT taxa richness; and Hill's effective richness (Hill's Index). A detailed description of these indicators is provided in CAMP (2024).

A detailed description of the program design and sampling methods are provided Technical Document 1, Section 2.4.

Table 4.1-1. 2010 to 2019 Benthic invertebrate sampling inventory.

| Waterbody/ Area | Sampling Year | | | | | | | | | | | |
|--------------------|---------------|------|------|------|------|------|------|------|------|------|------|------|
| | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
| PDB | -1 | -1 | • | • | • | • | • | • | • | • | • | • |
| LDB | -1 | -1 | • | • | • | • | • | • | • | • | • | • |
| PFF | | | | • | | | • | | | • | | |
| MANIG | -1 | -1 | • | • | • | • | • | • | • | • | • | • |
| EAGLE | | | • | | | • | | | • | | | • |

Notes:

1. Dataset excluded from analysis and reporting due to change in sampling design in 2010.

Table 4.1-2. Benthic invertebrate indicators and metrics.

| Indicator | Metric | Units |
|-----------------------|---|--|
| Abundance | • Total Invertebrate Abundance | Number (no.) per sample |
| | • Total Invertebrate Density | no. per square metre (m ²) |
| Community Composition | • Relative Proportions of Major Invertebrate Groups | % |
| | • EPT Index | % |
| | • O+C Index | % |
| Taxonomic Richness | • Total Taxa Richness | no. of families |
| | • EPT Taxa Richness | no. of families |
| Diversity | • Hill's Effective Richness (Hill's Index) | value |

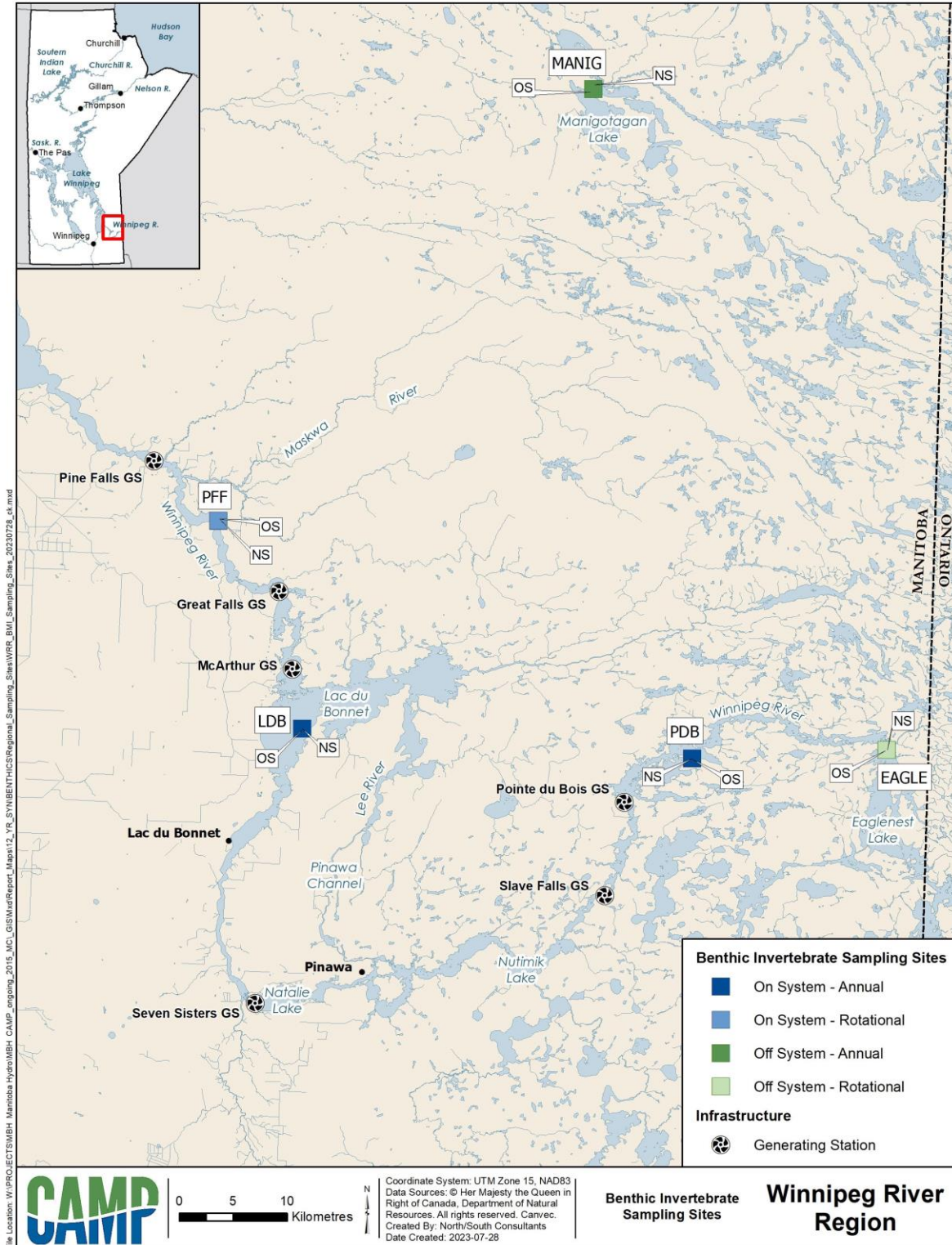


Figure 4.1-1. 2010 to 2019 Benthic invertebrate nearshore (NS) and offshore (OS) sampling sites.

4.2 ABUNDANCE

4.2.1 TOTAL INVERTEBRATE ABUNDANCE

4.2.1.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

Nearshore Habitat

Annual mean abundance over the ten years of monitoring ranged from 517 invertebrates per sample (2010) to 16,506 invertebrates per sample (2014; Figure 4.2-1). The overall mean was 3,539 invertebrates per sample, the overall median was 1,724 invertebrates per sample, and the interquartile range was 971 to 2,829 invertebrates per sample. Annual means were below the IQR in 2010 and 2016, and above the IQR in 2012, 2014, 2015, and 2017.

Offshore Habitat

Annual mean abundance (density) over the ten years of monitoring ranged from 606 invertebrates per m² (2010) to 4,129 invertebrates per m² (2011; Figure 4.2-2). The overall mean was 1,972 invertebrates per m², the overall median was 1,565 invertebrates per m², and the IQR was 905 to 2,622 invertebrates per m². Annual means were below the IQR in 2010 and 2014, and above the IQR in 2011, 2013, 2017, and 2019.

Lac du Bonnet

Nearshore Habitat

Annual mean abundance over the ten years of monitoring ranged from 247 invertebrates per sample (2019) to 6,581 invertebrates per sample (2012; Figure 4.2-1). The overall mean was 1,965 invertebrates per sample, the overall median was 717 invertebrates per sample, and the IQR was 318 to 1,733 invertebrates per sample. Annual means were below the IQR in 2016 and 2019, and above the IQR in 2012, 2014, 2015, and 2017.

Offshore Habitat

Annual mean abundance (density) over the ten years of monitoring ranged from 929 invertebrates per m² (2019) to 10,047 invertebrates per m² (2016; Figure 4.2-2). The overall mean was 4,212

invertebrates per m², the overall median was 3,044 invertebrates per m², and the IQR was 1,666 to 5,955 invertebrates per m². Annual means were below the IQR in 2013, 2014, and 2019, and above the IQR in 2011 and 2016.

ROTATIONAL SITES

Pine Falls Forebay

Nearshore Habitat

Annual mean abundance over the three years of monitoring ranged from 1,791 (2017) to 8,189 invertebrates per sample (2014; Figure 4.2-1). The overall mean was 4,273 invertebrates per sample, the overall median was 2,691 invertebrates per sample, and the IQR was 1,422 to 5,829 invertebrates per sample. Annual means were within the IQR, except for in 2014 (above).

Offshore Habitat

Annual mean abundance (density) over the three years of monitoring ranged from 1,423 invertebrates per m² (2014) to 2,678 invertebrates per m² (2011; Figure 4.2-2). The overall mean was 2,176 invertebrates per m², the overall median was 2,121 invertebrates per m², and the IQR was 1,147 to 3,189 invertebrates per m². Annual means for all years fell within the IQR.

4.2.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

Nearshore Habitat

Annual mean abundance over the ten years of monitoring ranged from 64 invertebrates per sample (2014) to 2,140 invertebrates per sample (2018; Figure 4.2-1). The overall mean was 536 invertebrates per sample, the overall median was 163 invertebrates per sample, and the IQR was 70 to 539 invertebrates per sample. Annual means were below the IQR in 2014, and above the IQR in 2012 and 2018.

Offshore Habitat

Annual mean abundance (density) over the ten years of monitoring ranged from 193 invertebrates per m² (2013) to 649 invertebrates per m² (2014; Figure 4.2-2). The overall mean was 380 invertebrates per m², the overall median was 310 invertebrates per m², and the IQR was 216 to

447 invertebrates per m². Annual means were below the IQR in 2013, and above the IQR in 2014, 2015, and 2019.

ROTATIONAL SITES

Eaglenest Lake

Nearshore Habitat

Annual mean abundance over the four years of monitoring ranged from 1,108 (2016) to 2,033 invertebrates per sample (2019; Figure 4.2-1). The overall mean was 1,466 invertebrates per sample, the overall median was 1,238 invertebrates per sample, and the IQR was 623 to 2,200 invertebrates per sample. Annual means for all years fell within the IQR.

Offshore Habitat

Annual mean abundance (density) over the four years of monitoring ranged from 718 invertebrates per m² (2010) to 2,753 invertebrates per m² (2013; Figure 4.2-2). The overall mean was 1,502 invertebrates per m², the overall median was 1,190 invertebrates per m², and the IQR was 700 to 1,490 invertebrates per m². Annual means were within the IQR, except for in 2013 (above).

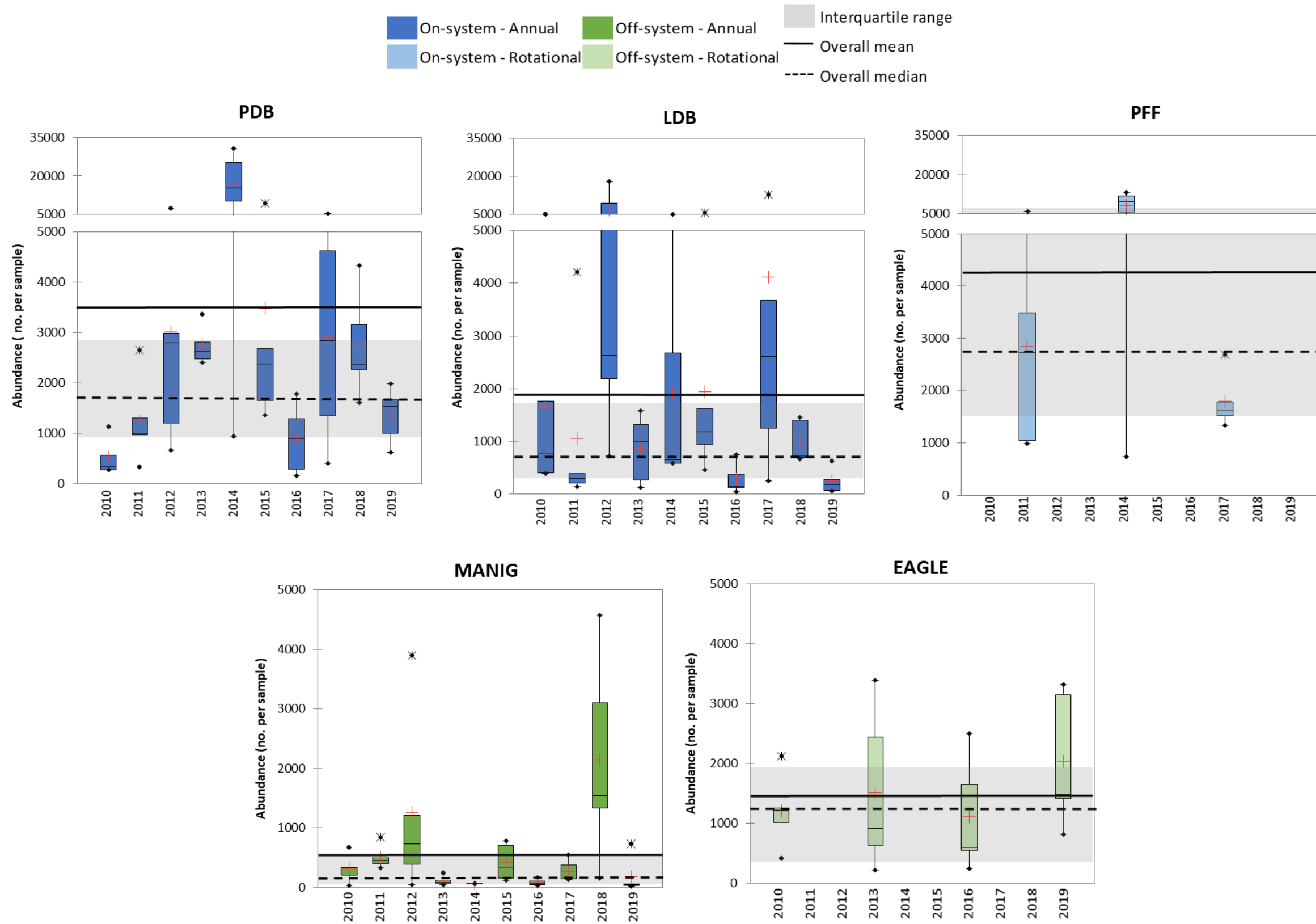


Figure 4.2-1. 2010 to 2019 Nearshore benthic invertebrate abundance (total no. per sample).

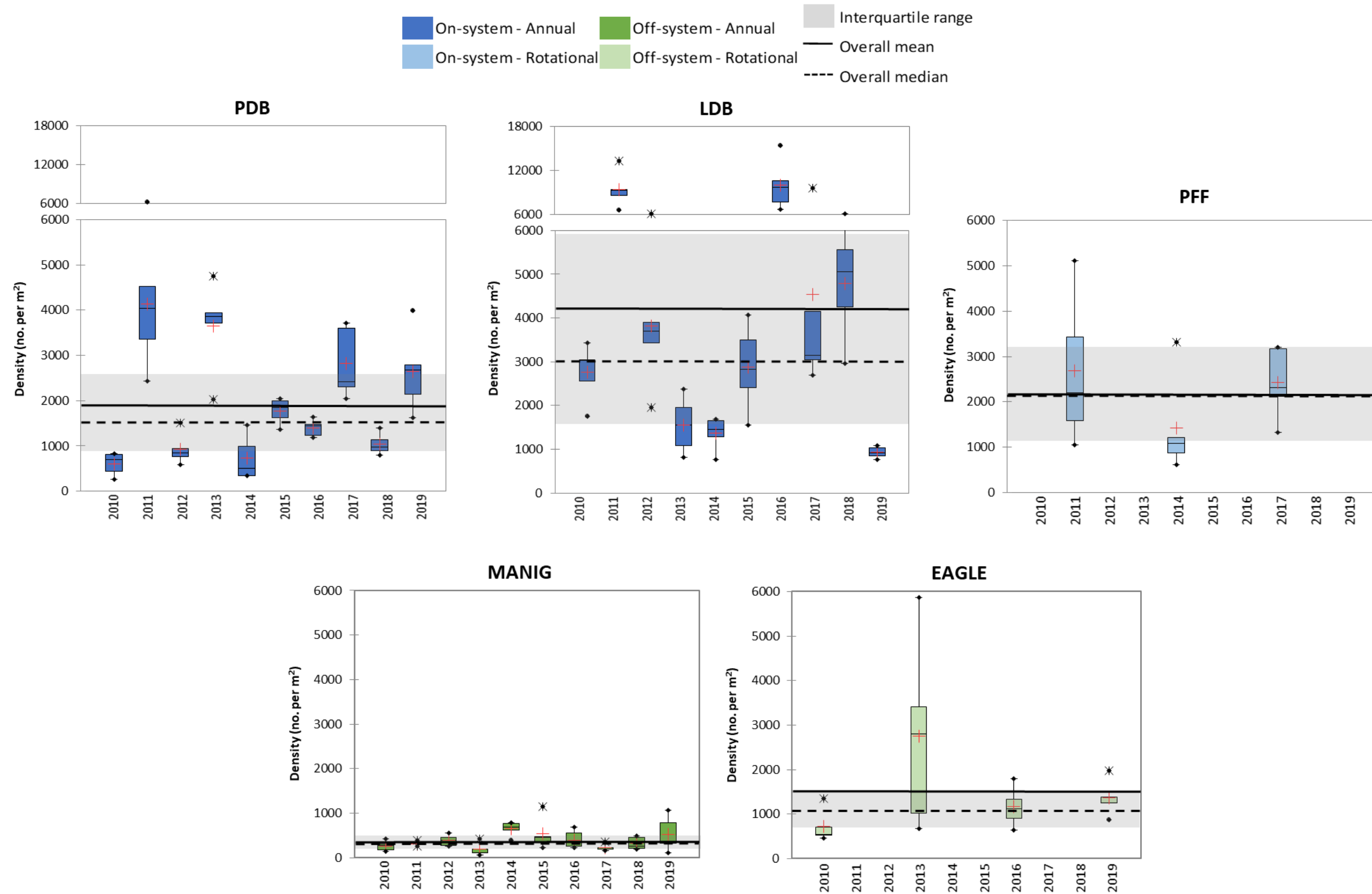


Figure 4.2-2. 2010 to 2019 Offshore benthic invertebrate abundance (density, total no. per m²).

4.3 COMMUNITY COMPOSITION

4.3.1 RELATIVE ABUNDANCE

4.3.1.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

Nearshore Habitat

Amphipoda (freshwater shrimps, mainly Hyalellidae) dominated the benthic invertebrate community in seven of the ten years of monitoring (2011 to 2017; Table 4.3-1). Among those years, mean annual relative abundances of Amphipoda ranged between 32% (2015) and 76% (2014). Amphipoda (28%, Hyalellidae) and Corixidae (water boatmen, 21%) were the dominant invertebrate taxa in 2010. Amphipoda (23%, mainly Hyalellidae), Gastropoda (snails, 24%, mainly Hydrobiidae, Planorbidae, and Lymnaeidae), and Ephemeroptera (22%, mainly Caenidae and Ephemeridae) were co-dominant in 2018. Gastropoda (31%) was the most abundant taxon in 2019.

Offshore Habitat

Ephemeroptera (mayflies, mainly Ephemeridae) dominated the benthic invertebrate community over the ten years of monitoring (2010 to 2019; Table 4.3-2). Excluding 2012, mean annual relative abundances of Ephemeroptera ranged between 50% (2010, 2015, and 2018) and 89% (2013). Ephemeroptera (37%), Oligochaeta (29%), and Chironomidae (29%) were dominant in 2012.

Lac du Bonnet

Nearshore Habitat

Amphipoda (freshwater shrimps, mainly Hyalellidae) dominated the benthic invertebrate community in nine of the ten years of monitoring (2010 to 2018; Table 4.3-3). Among those years, mean annual relative abundances of Amphipoda ranged between 38% (2015) and 78% (2010). Chironomidae (30%), Oligochaeta (22%) and Amphipoda (18%, Hyalellidae) were dominant in 2019.

Offshore Habitat

Benthic invertebrate community composition varied over the ten years of monitoring (2010 to 2019; Table 4.3-4). Oligochaeta (aquatic segmented worms) was the dominant taxon in 2010 (51%), 2012 (40%), and 2017 (37%). Ephemeroptera (mayflies, mainly Ephemeridae) was the dominant group in 2011 (46%) and 2018 (66%). Oligochaeta (39%) and Ephemeroptera (31%) dominated in 2013. Gastropoda (snails, 35%, mainly Hydrobiidae) dominated in 2014. Bivalvia (clams, 34%, mainly Sphaeriidae) and Gastropoda (30%) dominated in 2015. Bivalvia was dominant in 2016 (41%). Chironomidae (non-biting midges, 23%) and Ephemeroptera (24%) were co-dominant in 2019.

ROTATIONAL SITES

Pine Falls Forebay

Nearshore Habitat

Benthic invertebrate community composition varied over the over the three years of monitoring (2011, 2014, and 2017; Table 4.3-5). Amphipoda (freshwater shrimps, mainly Hyalellidae) was the dominant taxon in 2011 (44%). Oligochaeta (aquatic segmented worms, 24%) and Amphipoda (29%) were dominant in 2014. Amphipoda (29%) and Ephemeroptera (mayflies, 38%, mainly Caenidae and Ephemeridae) were the dominant invertebrate groups in 2017.

Offshore Habitat

Benthic invertebrate community composition varied over the over the three years of monitoring (2011, 2014, and 2017; Table 4.3-6). Chironomidae (non-biting midges) was the dominant taxon in 2011 (39%) and 2017 (39%). Oligochaeta (aquatic segmented worms, 19%) and Ephemeroptera (mayflies, Ephemeridae, 19%) were co-dominant in 2014.

4.3.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

Nearshore Habitat

Benthic invertebrate community composition varied over the ten years of monitoring (2010 to 2019; Table 4.3-7). Amphipoda (freshwater shrimps, mainly Hyalellidae) was the dominant taxon in 2010 (50%). Oligochaeta (aquatic segmented worms) (23%), Amphipoda (24%), and

Chironomidae (21%) were dominant in 2011. Ephemeroptera (mayflies, mainly Caenidae) was the dominant group in 2012 (60%), 2013 (36%), and 2014 (41%). Oligochaeta was the dominant taxon in 2015 (49%). Oligochaeta (23%) and Ephemeroptera (24%) were co-dominant in 2016. Amphipoda (21%) and Chironomidae (23%) were dominant in 2017. Oligochaeta (24%) and Bivalvia (clams, 22%) were dominant in 2018. Oligochaeta (20%), Amphipoda (19%), Chironomidae (21%), and Ephemeroptera (23%) were the dominant groups in 2019.

Offshore Habitat

Benthic invertebrate community composition varied over the ten years of monitoring (2010 to 2019; Table 4.3-8). Chironomidae was the dominant taxon in 2010 (33%), 2014 (32%), 2015 (36%), and 2018 (51%). Oligochaeta (aquatic segmented worms) was the dominant group in 2011 (35%) and 2012 (31%). Oligochaeta (24%) and Bivalvia (clams, Sphaeriidae, 22%) were dominant in 2013. Bivalvia was dominant in 2016 (40%) and 2017 (31%). Oligochaeta (20%), Chironomidae (19%), and Ephemeroptera (mayflies, Ephemeridae, 23%) were dominant in 2019.

ROTATIONAL SITES

Eaglenest Lake

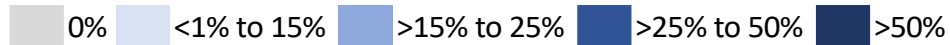
Nearshore Habitat

Benthic invertebrate community composition varied over the four years of monitoring (2010, 2013, 2016, and 2019; Table 4.3-9). Amphipoda (freshwater shrimps, 21%, mainly Hyalellidae), Gastropoda (snails, 19%, mainly Planorbidae and Lymnaeidae), Chironomidae (23%), and Ephemeroptera (mayflies, 15%, mainly Caenidae) were dominant in 2010. Oligochaeta (36%) and Chironomidae (32%) were the dominant groups in 2013. Gastropoda (36%) was the most dominant taxon in 2016. Oligochaeta (37%) and Chironomidae (41%) were the dominant invertebrate groups in 2019.

Offshore Habitat

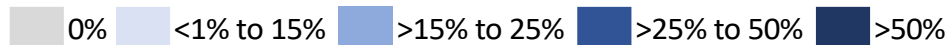
Benthic invertebrate community composition varied over the four years of monitoring (2010, 2013, 2016, and 2019; Table 4.3-10). Bivalvia (clams, 27%, mainly Sphaeriidae), Chironomidae (33%) and Ephemeroptera (Ephemeridae, 31%) were the dominant invertebrate taxa in 2010. Ephemeroptera (mainly Ephemeridae) was the most dominant group in 2013 (68%), 2016 (54%), and 2019 (57%).

Table 4.3-1. 2010 to 2019 Pointe du Bois Forebay nearshore benthic invertebrate relative abundance.



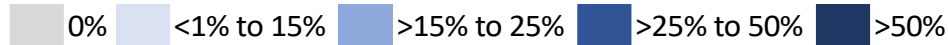
| Invertebrate Taxa | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
|-------------------|------|------|------|------|------|------|------|------|------|------|
| Oligochaeta | 9% | 7% | 20% | 12% | 3% | 9% | 17% | 9% | 10% | 15% |
| Amphipoda | 28% | 41% | 39% | 48% | 76% | 32% | 29% | 25% | 23% | 19% |
| Bivalvia | 1% | 1% | 3% | 2% | 2% | 3% | 2% | 9% | 6% | 3% |
| Gastropoda | 10% | 15% | 13% | 15% | 6% | 10% | 8% | 19% | 24% | 31% |
| Ceratopogonidae | 1% | 0% | <1% | 1% | <1% | <1% | <1% | 1% | 1% | 1% |
| Chironomidae | 16% | 5% | 8% | 7% | 3% | 24% | 6% | 6% | 8% | 14% |
| Other Diptera | <1% | <1% | <1% | <1% | <1% | 0% | <1% | <1% | <1% | <1% |
| Ephemeroptera | 5% | 8% | 10% | 9% | 5% | 13% | 27% | 22% | 22% | 8% |
| Plecoptera | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Trichoptera | 3% | 3% | 3% | 4% | 2% | 5% | 3% | 4% | 4% | 5% |
| Corixidae | 21% | 17% | 1% | 1% | 2% | 1% | 2% | 1% | <1% | <1% |
| Coleoptera | 3% | 1% | 1% | 1% | 1% | 1% | 3% | 4% | 2% | 2% |
| All other taxa | 2% | 1% | 1% | 1% | <1% | 1% | 2% | 1% | 1% | 2% |

Table 4.3-2. 2010 to 2019 Pointe du Bois Forebay offshore benthic invertebrate relative abundance.



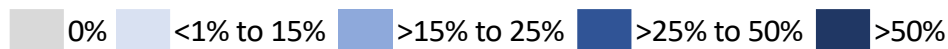
| Invertebrate Taxa | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
|-------------------|------|------|------|------|------|------|------|------|------|------|
| Oligochaeta | 9% | 3% | 29% | 4% | 15% | 17% | 14% | 10% | 12% | 8% |
| Amphipoda | 0% | <1% | <1% | <1% | 2% | 1% | 0% | <1% | 0% | <1% |
| Bivalvia | 16% | 4% | 2% | 1% | 14% | 8% | 14% | 4% | 9% | 13% |
| Gastropoda | <1% | <1% | 1% | <1% | 2% | 1% | <1% | <1% | 0% | 1% |
| Ceratopogonidae | 3% | <1% | 1% | <1% | 3% | 1% | 3% | 1% | 1% | 1% |
| Chironomidae | 18% | 8% | 29% | 4% | 7% | 19% | 15% | 12% | 20% | 19% |
| Other Diptera | 0% | <1% | 0% | 0% | 0% | 0% | <1% | 0% | 1% | 1% |
| Ephemeroptera | 50% | 83% | 37% | 89% | 53% | 50% | 52% | 70% | 50% | 56% |
| Plecoptera | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Trichoptera | 4% | 2% | <1% | 1% | 1% | 2% | 2% | 1% | 3% | 1% |
| Corixidae | 0% | <1% | 0% | 0% | 1% | 0% | <1% | 0% | <1% | <1% |
| Coleoptera | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| All other taxa | 1% | <1% | <1% | <1% | 2% | <1% | <1% | 0% | 2% | <1% |

Table 4.3-3. 2010 to 2019 Lac du Bonnet nearshore benthic invertebrate relative abundance.



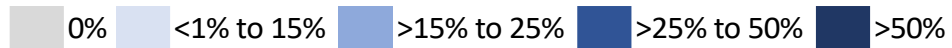
| Invertebrate Taxa | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
|-------------------|------|------|------|------|------|------|------|------|------|------|
| Oligochaeta | 6% | 5% | 4% | 6% | 12% | 3% | 13% | 5% | 18% | 22% |
| Amphipoda | 78% | 62% | 49% | 76% | 45% | 38% | 50% | 63% | 62% | 18% |
| Bivalvia | 0% | 6% | 7% | 1% | 2% | 14% | 2% | 6% | 1% | 1% |
| Gastropoda | 1% | 10% | 7% | 3% | 9% | 16% | 3% | 6% | 3% | 4% |
| Ceratopogonidae | 0% | <1% | <1% | <1% | <1% | <1% | 0% | <1% | <1% | <1% |
| Chironomidae | 5% | 4% | 1% | 4% | 4% | 6% | 5% | 2% | 6% | 30% |
| Other Diptera | <1% | <1% | <1% | <1% | 1% | <1% | <1% | 1% | 1% | 1% |
| Ephemeroptera | 6% | 7% | 23% | 4% | 18% | 8% | 9% | 11% | 3% | 13% |
| Plecoptera | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Trichoptera | 2% | 4% | 7% | 4% | 5% | 8% | 10% | 4% | 2% | 2% |
| Corixidae | 1% | 2% | <1% | 1% | 2% | 5% | 8% | <1% | 4% | <1% |
| Coleoptera | <1% | <1% | 0% | <1% | <1% | 0% | <1% | <1% | <1% | 0% |
| All other taxa | <1% | 1% | <1% | 1% | 1% | 1% | 1% | 2% | 1% | 8% |

Table 4.3-4. 2010 to 2019 Lac du Bonnet offshore benthic invertebrate relative abundance.



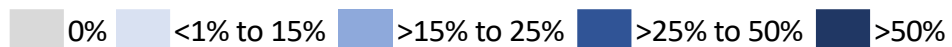
| Invertebrate Taxa | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
|-------------------|------|------|------|------|------|------|------|------|------|------|
| Oligochaeta | 51% | 11% | 40% | 39% | 14% | 17% | 21% | 37% | 5% | 2% |
| Amphipoda | 1% | 1% | 3% | 5% | 8% | 5% | 1% | 3% | 1% | 7% |
| Bivalvia | 31% | 25% | 15% | 3% | 26% | 34% | 41% | 19% | 7% | 18% |
| Gastropoda | 8% | 7% | 2% | 9% | 35% | 30% | 5% | 7% | 4% | 12% |
| Ceratopogonidae | 1% | 1% | 3% | 1% | 0% | <1% | 1% | 5% | <1% | 2% |
| Chironomidae | 4% | 8% | 16% | 11% | 3% | 5% | 8% | 19% | 11% | 23% |
| Other Diptera | 0% | <1% | 0% | 0% | 0% | 0% | 3% | 0% | 2% | 0% |
| Ephemeroptera | 3% | 46% | 18% | 31% | 5% | 5% | 17% | 7% | 66% | 24% |
| Plecoptera | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Trichoptera | 1% | 3% | 3% | 1% | 7% | 5% | 2% | 4% | 3% | 11% |
| Corixidae | 0% | <1% | <1% | 0% | 0% | 0% | 0% | 0% | <1% | <1% |
| Coleoptera | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| All other taxa | 1% | 0% | <1% | <1% | 1% | 1% | <1% | <1% | <1% | <1% |

Table 4.3-5. 2010 to 2019 Pine Falls Forebay nearshore benthic invertebrate relative abundance.



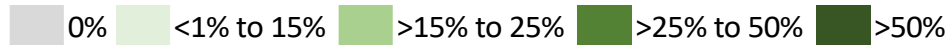
| Invertebrate Taxa | 2011 | 2014 | 2017 |
|-------------------|------|------|------|
| Oligochaeta | 21% | 24% | 19% |
| Amphipoda | 44% | 29% | 29% |
| Bivalvia | <1% | 1% | 1% |
| Gastropoda | 4% | 8% | 2% |
| Ceratopogonidae | <1% | <1% | <1% |
| Chironomidae | 5% | 11% | 6% |
| Other Diptera | <1% | <1% | <1% |
| Ephemeroptera | 17% | 19% | 38% |
| Plecoptera | <1% | 0% | 0% |
| Trichoptera | 5% | 3% | 2% |
| Corixidae | 1% | 1% | <1% |
| Coleoptera | 2% | 1% | 2% |
| All other taxa | <1% | 2% | 2% |

Table 4.3-6. 2010 to 2019 Pine Falls Forebay offshore benthic invertebrate relative abundance.



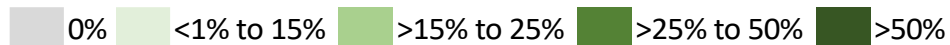
| Invertebrate Taxa | 2011 | 2014 | 2017 |
|-------------------|------|------|------|
| Oligochaeta | 26% | 19% | 25% |
| Amphipoda | 0% | 2% | <1% |
| Bivalvia | 4% | 16% | 7% |
| Gastropoda | 1% | 6% | 3% |
| Ceratopogonidae | <1% | <1% | 0% |
| Chironomidae | 39% | 13% | 39% |
| Other Diptera | 1% | 1% | 2% |
| Ephemeroptera | 23% | 19% | 19% |
| Plecoptera | 0% | 0% | 0% |
| Trichoptera | 2% | 11% | <1% |
| Corixidae | 0% | <1% | 0% |
| Coleoptera | <1% | 0% | 0% |
| All other taxa | 3% | 13% | 5% |

Table 4.3-7. 2010 to 2019 Manigotagan Lake nearshore benthic invertebrate relative abundance.



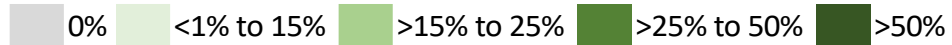
| Invertebrate Taxa | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
|-------------------|------|------|------|------|------|------|------|------|------|------|
| Oligochaeta | 16% | 23% | 17% | 13% | 2% | 49% | 23% | 7% | 24% | 20% |
| Amphipoda | 50% | 24% | 15% | 3% | 4% | 3% | 9% | 21% | 8% | 19% |
| Bivalvia | 0% | 3% | <1% | <1% | 0% | 0% | <1% | 7% | 22% | 1% |
| Gastropoda | <1% | 4% | <1% | 3% | 17% | 2% | 8% | 13% | 15% | 4% |
| Ceratopogonidae | 0% | <1% | 1% | 0% | 0% | 0% | 0% | <1% | <1% | 0% |
| Chironomidae | 22% | 21% | 5% | 20% | 7% | 11% | 16% | 23% | 10% | 21% |
| Other Diptera | 1% | 1% | 1% | 6% | 0% | <1% | 2% | <1% | <1% | <1% |
| Ephemeroptera | 7% | 15% | 60% | 36% | 41% | 19% | 24% | 16% | 17% | 23% |
| Plecoptera | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Trichoptera | 3% | 7% | 2% | 8% | 11% | 5% | 12% | 8% | 3% | 5% |
| Corixidae | 1% | 1% | <1% | <1% | 1% | <1% | 0% | <1% | 0% | 0% |
| Coleoptera | <1% | 2% | 1% | 8% | 11% | 1% | 3% | 2% | 1% | 1% |
| All other taxa | <1% | <1% | <1% | 2% | 7% | 9% | 3% | 3% | 1% | 5% |

Table 4.3-8. 2010 to 2019 Manigotagan Lake offshore benthic invertebrate relative abundance.



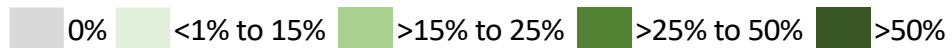
| Invertebrate Taxa | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
|-------------------|------|------|------|------|------|------|------|------|------|------|
| Oligochaeta | 25% | 35% | 31% | 24% | 23% | 9% | 6% | 8% | 25% | 20% |
| Amphipoda | 5% | 0% | 2% | 1% | 8% | 1% | 7% | 1% | 3% | 1% |
| Bivalvia | 10% | 21% | 10% | 22% | 11% | 23% | 40% | 31% | 2% | 13% |
| Gastropoda | 2% | 1% | 2% | 9% | <1% | 5% | 3% | 3% | 1% | 4% |
| Ceratopogonidae | 0% | 0% | 0% | 0% | 0% | 1% | 0% | 1% | 0% | 1% |
| Chironomidae | 33% | 26% | 20% | 15% | 32% | 36% | 26% | 8% | 51% | 19% |
| Other Diptera | 20% | 12% | 25% | 1% | 0% | 5% | 4% | 24% | 5% | 7% |
| Ephemeroptera | 3% | 5% | 8% | 18% | 24% | 15% | 10% | 16% | 13% | 23% |
| Plecoptera | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Trichoptera | 1% | 1% | 0% | 0% | 1% | 0% | 1% | 1% | 0% | 1% |
| Corixidae | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 1% | 1% | 0% |
| Coleoptera | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| All other taxa | 0% | 0% | 2% | 9% | 1% | 6% | 3% | 6% | 1% | 12% |

Table 4.3-9. 2010 to 2019 Eaglenest Lake nearshore benthic invertebrate relative abundance.



| Invertebrate Taxa | 2010 | 2013 | 2016 | 2019 |
|-------------------|------|------|------|------|
| Oligochaeta | 5% | 36% | 24% | 37% |
| Amphipoda | 21% | 5% | 18% | 3% |
| Bivalvia | <1% | 0% | <1% | 3% |
| Gastropoda | 19% | 15% | 36% | 9% |
| Ceratopogonidae | 0% | 0% | 0% | <1% |
| Chironomidae | 23% | 32% | 6% | 41% |
| Other Diptera | <1% | 0% | <1% | <1% |
| Ephemeroptera | 15% | 1% | 2% | 3% |
| Plecoptera | 0% | 0% | 0% | 0% |
| Trichoptera | 8% | 5% | 3% | 3% |
| Corixidae | 7% | 5% | 8% | <1% |
| Coleoptera | 2% | <1% | 1% | 1% |
| All other taxa | <1% | 0% | 1% | <1% |

Table 4.3-10. 2010 to 2019 Eaglenest Lake offshore benthic invertebrate relative abundance.



| Invertebrate Taxa | 2010 | 2013 | 2016 | 2019 |
|-------------------|------|------|------|------|
| Oligochaeta | 1% | 5% | 10% | 7% |
| Amphipoda | 2% | 1% | 2% | 1% |
| Bivalvia | 27% | 3% | 5% | 6% |
| Gastropoda | 0% | <1% | 2% | 1% |
| Ceratopogonidae | <1% | <1% | <1% | <1% |
| Chironomidae | 33% | 21% | 21% | 25% |
| Other Diptera | <1% | <1% | <1% | 1% |
| Ephemeroptera | 31% | 68% | 54% | 57% |
| Plecoptera | 0% | 0% | 0% | 0% |
| Trichoptera | 4% | 1% | 1% | <1% |
| Corixidae | 2% | 0% | <1% | <1% |
| Coleoptera | 0% | 0% | 0% | 0% |
| All other taxa | <1% | <1% | 3% | 1% |

4.3.2 EPT INDEX

4.3.2.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

Nearshore Habitat

Annual mean EPT Index over the ten years of monitoring ranged from 7% (2010 and 2014) to 26% (2018; Figure 4.3-1). The overall mean was 15%, the overall median was 12%, and the IQR was 6% to 18%. Annual means were above the IQR in 2016, 2017, and 2018.

Offshore Habitat

Annual mean EPT Index over the ten years of monitoring ranged from 40% (2012) to 89% (2013; Figure 4.3-2). The overall mean and median were 60% and the IQR was 49% to 72%. Annual means were below the IQR in 2012 and 2014, and above the IQR in 2011, 2013, and 2017.

Lac du Bonnet

Nearshore Habitat

Annual mean EPT Index over the ten years of monitoring ranged from 5% (2018) to 33% (2012; Figure 4.3-1). The overall mean was 14%, the overall median was 10%, and the IQR was 6% to 22%. Annual means were below the IQR in 2018, and above the IQR in 2012.

Offshore Habitat

Annual mean EPT Index over the ten years of monitoring ranged from 4% (2010) to 68% (2018; Figure 4.3-2). The overall mean was 27%, the overall median was 22%, and the IQR was 11% to 38%. Annual means were below the IQR in 2010, and above the IQR in 2011 and 2018.

ROTATIONAL SITES

Pine Falls Forebay

Nearshore Habitat

Annual mean EPT Index over the three years of monitoring ranged from 21% (2011) to 35% (2017; Figure 4.3-1). The overall mean was 29%, the overall median was 25%, and the IQR was 17% to 32%. Annual means were within the IQR, except in 2017 (above).

Offshore Habitat

Annual mean EPT Index over the three years of monitoring ranged from 24% (2017) to 41% (2011; Figure 4.3-2). The overall mean was 31%, the overall median was 29%, and the IQR was 8% to 46%. Annual means for all years fell within the IQR.

4.3.2.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

Nearshore Habitat

Annual mean EPT Index over the ten years of monitoring ranged from 11% (2010) to 60% (2012; Figure 4.3-1). The overall mean was 32%, the overall median was 25%, and the IQR was 20% to 43%. Annual means were below the IQR in 2010 and 2018, and above the IQR in 2012 and 2014.

Offshore Habitat

Annual mean EPT Index over the ten years of monitoring ranged from 4% (2010) to 26% (2014; Figure 4.3-2). The overall mean was 15%, the overall median was 13%, and the IQR was 4% to 23%. Annual means were above the IQR in 2014 and 2019.

ROTATIONAL SITES

Eaglenest Lake

Nearshore Habitat

Annual mean EPT Index over the four years of monitoring ranged from 6% (2016 and 2019) to 21% (2010; Figure 4.3-1). The overall mean was 10%, the overall median was 8%, and the IQR was 5% to 13%. Annual means were within the IQR, except in 2010 (above).

Offshore Habitat

Annual mean EPT Index over the four years of monitoring ranged from 45% (2010) to 73% (2013; Figure 4.3-2). The overall mean EPT Index was 58%, the overall median EPT Index was 61%, and the IQR was 50% to 71%. Annual means were below the IQR in 2010, and above the IQR in 2013.

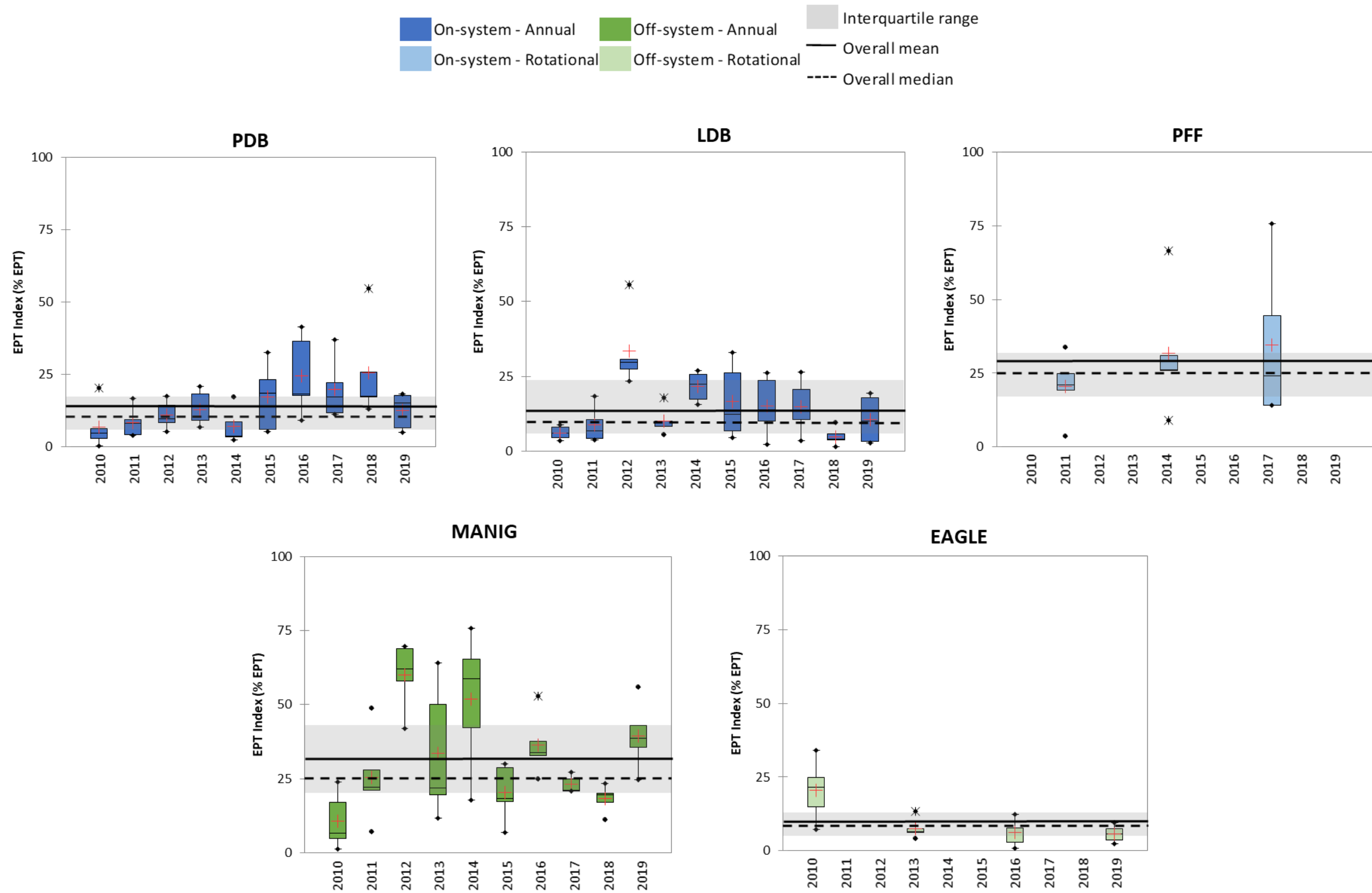


Figure 4.3-1. 2010 to 2019 Nearshore benthic invertebrate EPT Index.

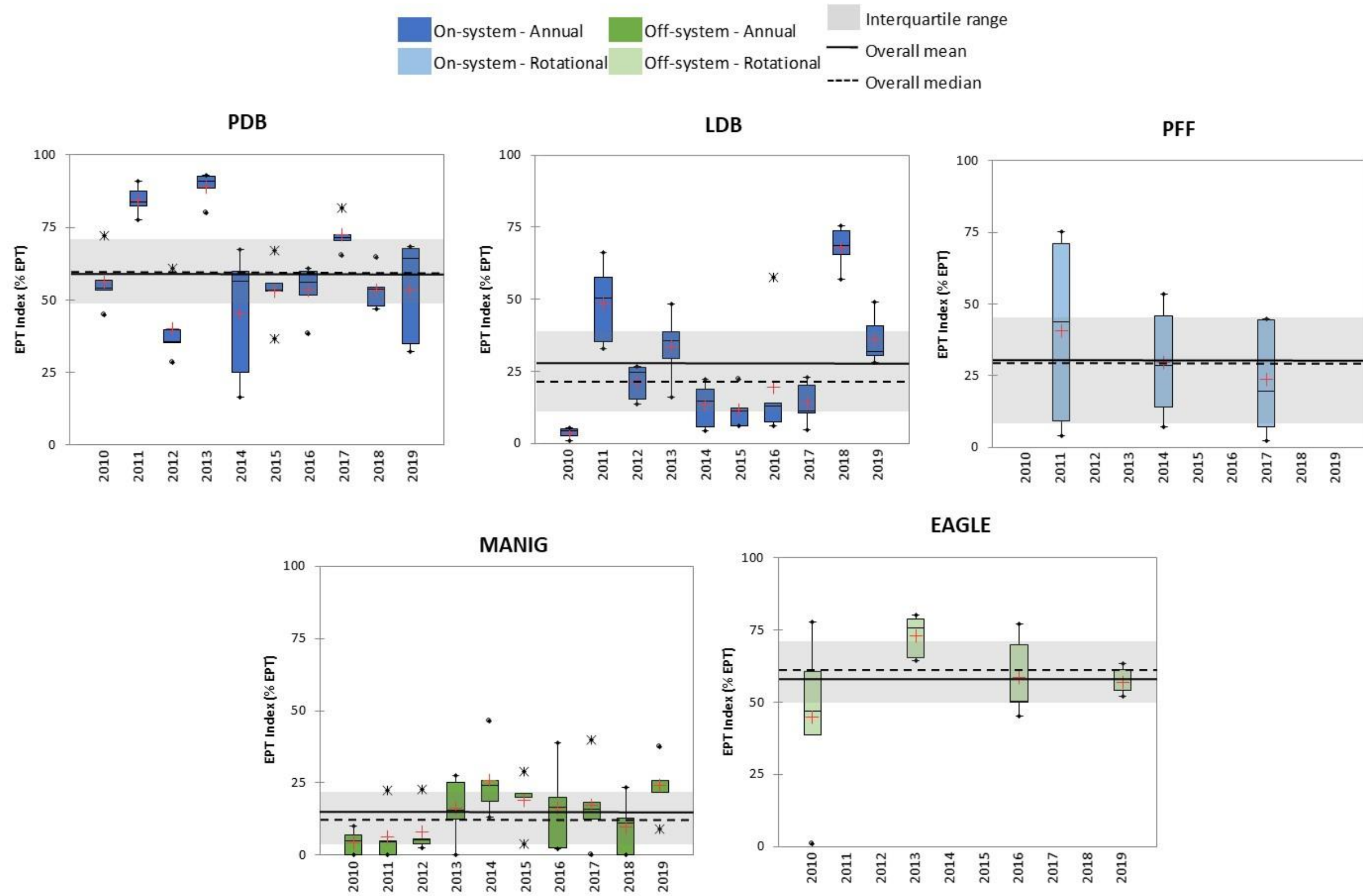


Figure 4.3-2. 2010 to 2019 Offshore benthic invertebrate EPT Index.

4.3.3 O+C INDEX

4.3.3.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

Nearshore Habitat

Annual mean O+C Index over the ten years of monitoring ranged from 6% (2014) to 30% (2015 and 2019; Figure 4.3-3). The overall mean was 21%, the overall median was 20%, and the IQR was 13% to 29%. Annual means were below the IQR in 2011 and 2014, and above the IQR in 2015 and 2019.

Offshore Habitat

Annual mean O+C Index over the ten years of monitoring ranged from 9% (2013) to 55% (2012; Figure 4.3-4). The overall mean was 27%, the overall median was 26%, and the IQR was 19% to 35%. Annual means were below the IQR in 2011 and 2013, and above the IQR in 2012 and 2015.

Lac du Bonnet

Nearshore Habitat

Annual mean O+C Index over the ten years of monitoring ranged from 10% (2010) to 53% (2019; Figure 4.3-3). The overall mean was 20%, the overall median was 13%, and the IQR was 8% to 24%. Annual means were above the IQR in 2018 and 2019.

Offshore Habitat

Annual mean O+C Index over the ten years of monitoring ranged from 15% (2018) to 55% (2012; Figure 4.3-4). The overall mean was 33%, the overall median was 30%, and the IQR was 17% to 48%. Annual means were below the IQR in 2014 and 2018, and above the IQR in 2010, 2012, and 2017.

ROTATIONAL SITES

Pine Falls Forebay

Nearshore Habitat

Annual mean O+C Index over the three years of monitoring ranged from 26% (2017) to 31% (2011 and 2014; Figure 4.3-3). The overall mean was 29%, the overall median was 27%, and the IQR was 24% to 33%. Annual means for all years fell within the IQR.

Offshore Habitat

Annual mean O+C Index over the three years of monitoring ranged from 31% (2014) to 59% (2017; Figure 4.3-4). The overall mean was 46%, the overall median was 39%, and the IQR was 26% to 70%. Annual means for all years fell within the IQR.

4.3.3.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

Nearshore Habitat

Annual mean O+C Index over the ten years of monitoring ranged from 9% (2014) to 54% (2015; Figure 4.3-3). The overall mean was 36%, the overall median was 35%, and the IQR was 24% to 47%. Annual means were below the IQR in 2014, and above the IQR in 2010 and 2015.

Offshore Habitat

Annual mean O+C Index over the ten years of monitoring ranged from 17% (2017) to 77% (2018; Figure 4.3-4). The overall mean was 47%, the overall median was 48%, and the IQR was 32% to 59%. Annual means were below the IQR in 2017, and above the IQR in 2018.

ROTATIONAL SITES

Eaglenest Lake

Nearshore Habitat

Annual mean O+C Index over the four years of monitoring ranged from 30% (2010) to 72% (2019; Figure 4.3-3). The overall mean was 48%, the overall median was 46%, and the IQR was 22% to 73%. Annual means for all years fell within the IQR.

Offshore Habitat

Annual mean O+C Index over the four years of monitoring ranged from 24% (2013) to 33% (2019; Figure 4.3-4). The overall mean was 30%, the overall median was 29%, and the IQR was 24% to 35%. Annual means were within the IQR, except in 2013 (below).

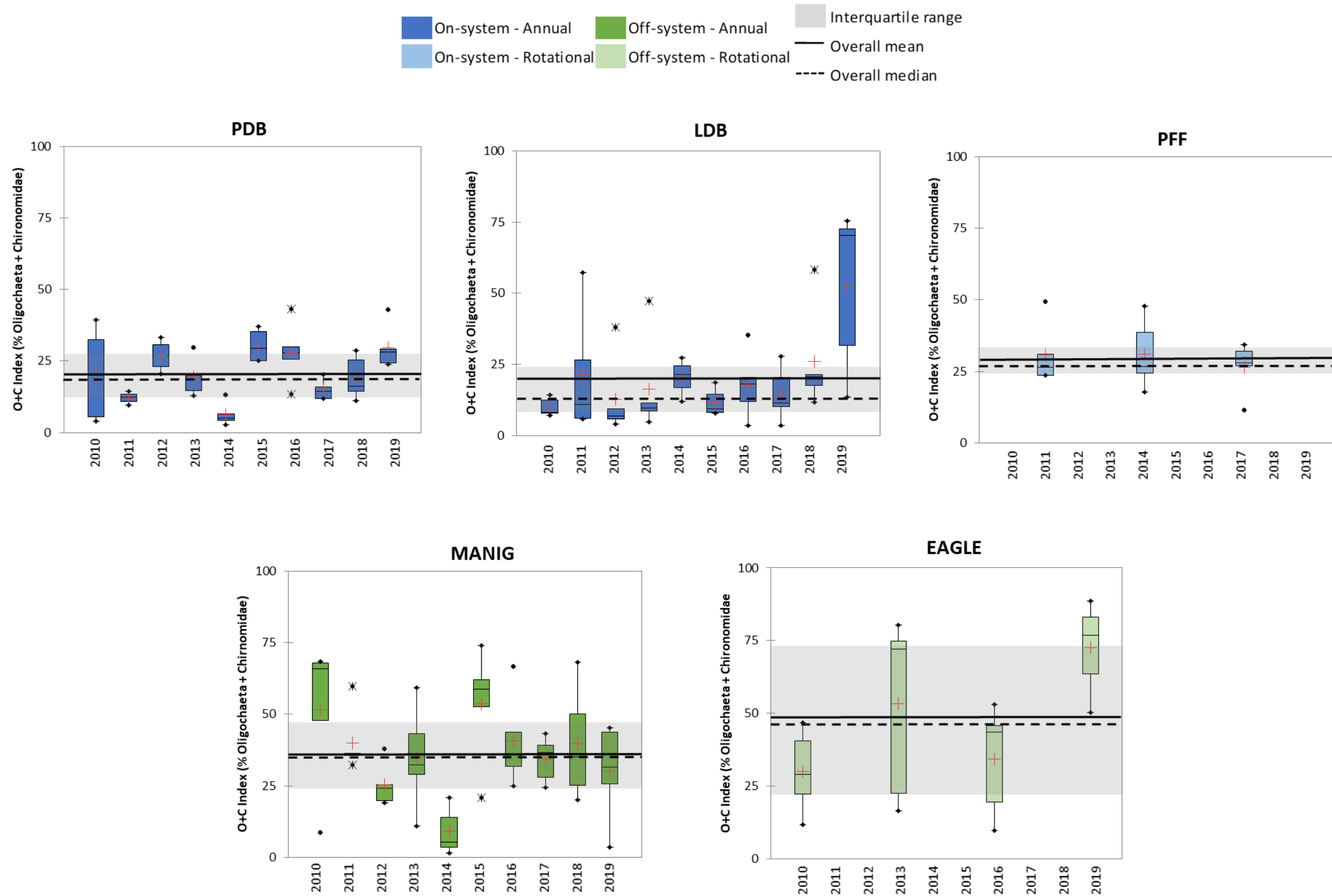


Figure 4.3-3. 2010 to 2019 Nearshore benthic invertebrate O+C Index.

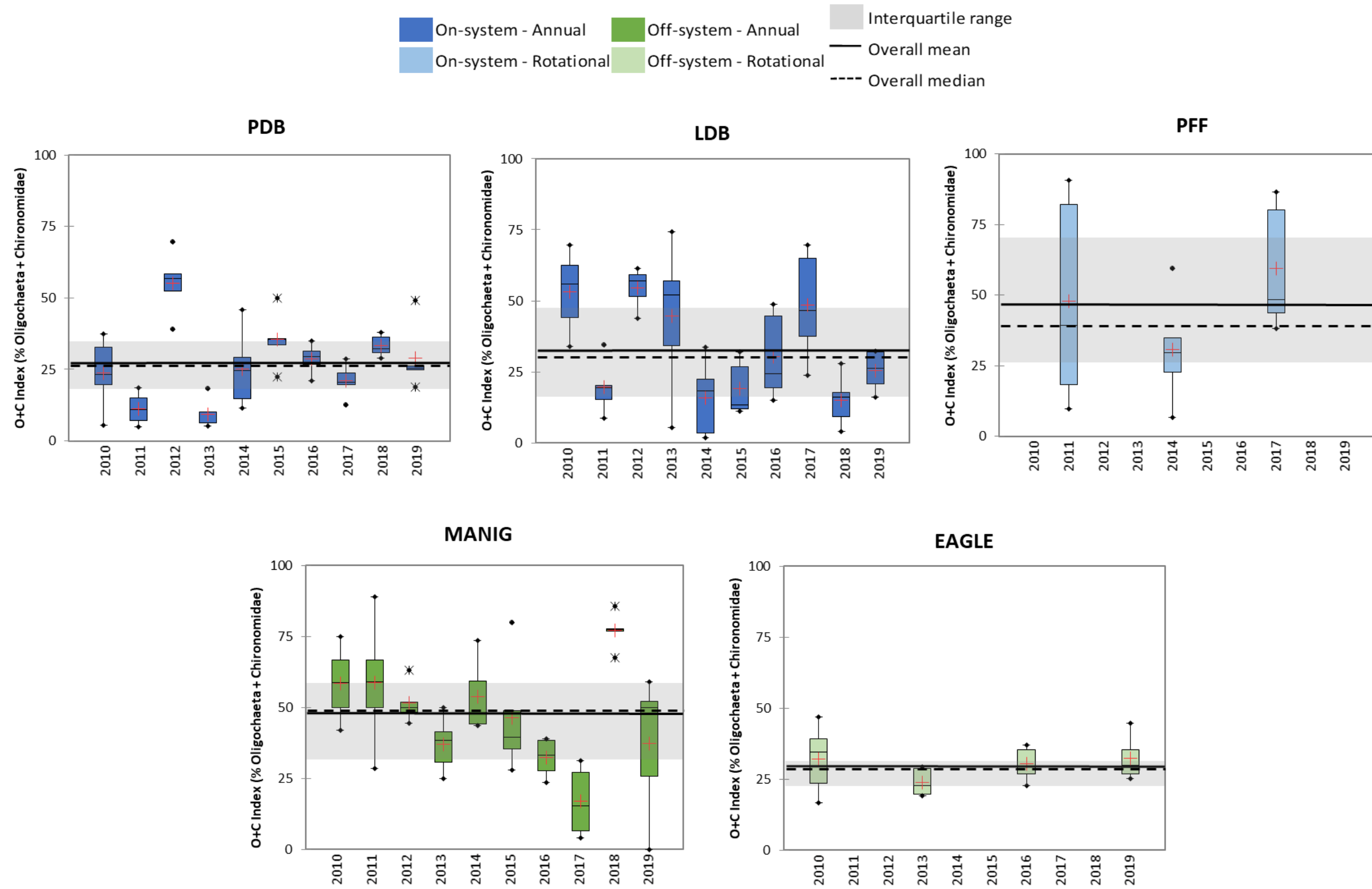


Figure 4.3-4. 2010 to 2019 Offshore benthic invertebrate O+C Index.

4.4 RICHNESS

4.4.1 TOTAL TAXA RICHNESS

4.4.1.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

Nearshore Habitat

Annual mean total taxa richness over the ten years of monitoring ranged from 17 families (2010) to 26 families (2013 and 2018; Figure 4.4-1). The overall mean and median were 23 families, and the IQR was 21 to 25 families. Annual means were below the IQR in 2010, and above the IQR in 2013, 2017, and 2018.

Offshore Habitat

Annual mean total taxa richness over the ten years of monitoring ranged from five families (2012) to eight families (2011, 2014, 2015, and 2019; Figure 4.4-2). The overall mean and median were seven families, and the IQR was 6 to 8 families. Annual means were below the IQR in 2010 and 2012.

Lac du Bonnet

Nearshore Habitat

Annual mean total taxa richness over the ten years of monitoring ranged from 13 families (2019) to 22 families (2012, 2015, and 2017; Figure 4.4-1). The overall mean and median were 18 families, and the IQR was 14 to 21 families. Annual means were below the IQR in 2019, and above the IQR in 2012, 2015, and 2017.

Offshore Habitat

Annual mean total taxa richness over the ten years of monitoring ranged from seven families (2013) to 12 families (2014; Figure 4.4-2). The overall mean and median were nine families, and the IQR was 8 to 10 families. Annual means were below the IQR in 2013 and 2019, and above the IQR in 2014 and 2015.

ROTATIONAL SITES

Pine Falls Forebay

Nearshore Habitat

Annual mean total taxa richness over the three years of monitoring ranged from 24 families (2017) to 26 families (2014; Figure 4.4-1). The overall mean was 25 families, the overall median was 24 families, and the IQR was 23 to 28 families. Annual means for all years were within the IQR.

Offshore Habitat

Annual mean total taxa richness over the three years of monitoring ranged from eight families (2017) to ten families (2014; Figure 4.4-2). The overall mean and median were nine families, and the IQR was 8 to 10 families. Annual means for all years were within the IQR.

4.4.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

Nearshore Habitat

Annual mean total taxa richness over the ten years of monitoring ranged from 11 families (2010) to 24 families (2011; Figure 4.4-1). The overall mean was 17 families, the overall median was 16 families, and the IQR was 14 to 21 families. Annual means were below the IQR in 2010 and 2014, and above the IQR in 2011.

Offshore Habitat

Annual mean total taxa richness over the ten years of monitoring ranged from five families (2011, 2012, 2013, 2014, and 2018) to seven families (2019; Figure 4.4-2). The overall mean was six families, the overall median was five families, and the IQR was 5 to 6 families. Annual means were below the IQR in 2011 and 2018, and above the IQR in 2016 and 2019.

ROTATIONAL SITES

Eaglenest Lake

Nearshore Habitat

Annual mean total taxa richness over the four years of monitoring ranged from 14 families (2014) to 17 families (2010, 2016, and 2019; Figure 4.4-1). The overall mean was 16 families, the overall median was 17 families, and the IQR was 15 to 18 families. Annual means were within the IQR, except in 2013 (below).

Offshore Habitat

Annual mean total taxa richness over the four years of monitoring ranged from six families (2010) to eight families (2016 and 2019; Figure 4.4-2). The overall mean and median were seven families, and the IQR was 5 to 9 families. Annual means for all years were within the IQR.

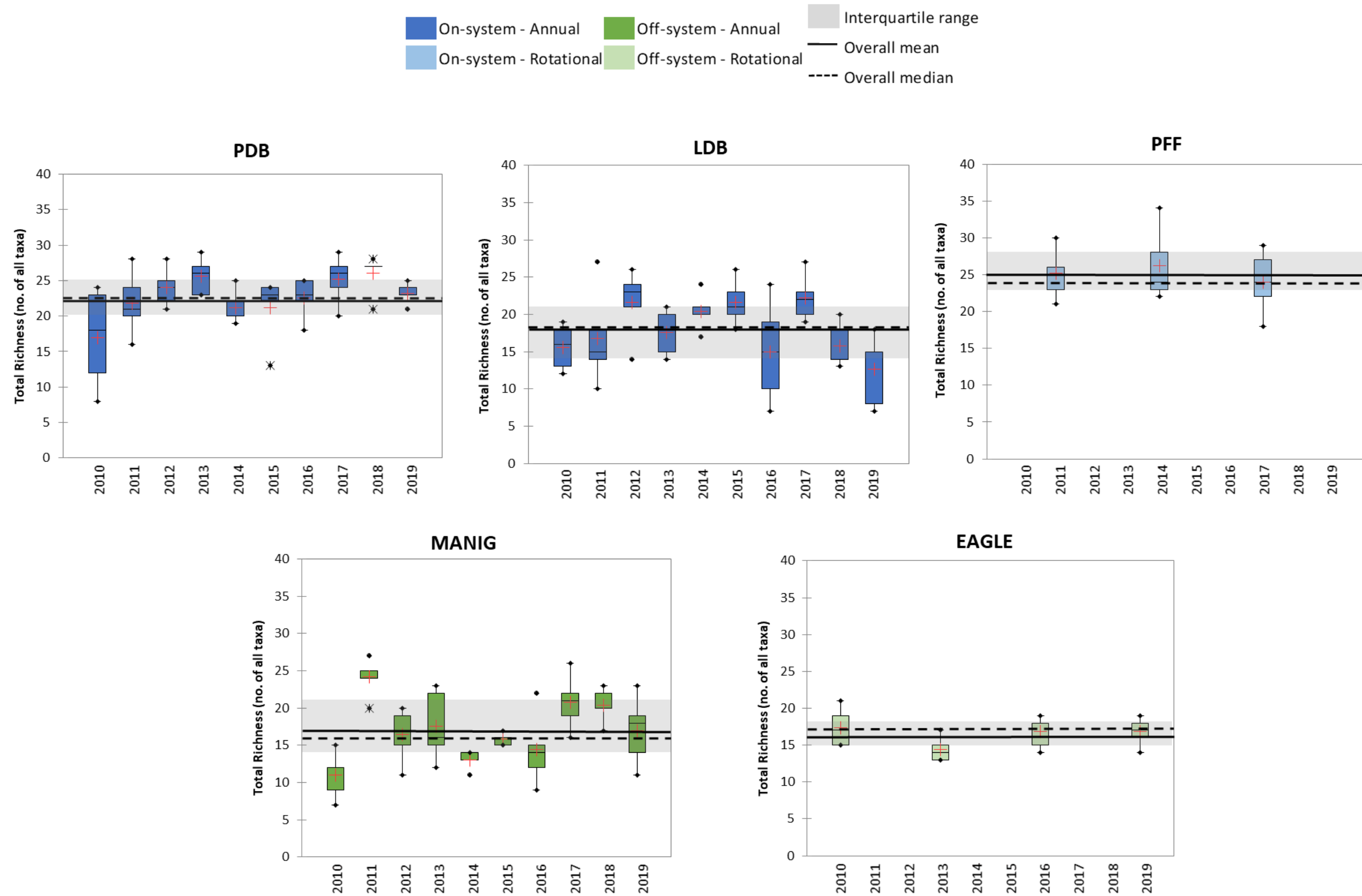


Figure 4.4-1. 2010 to 2019 Nearshore benthic invertebrate total richness (family-level).

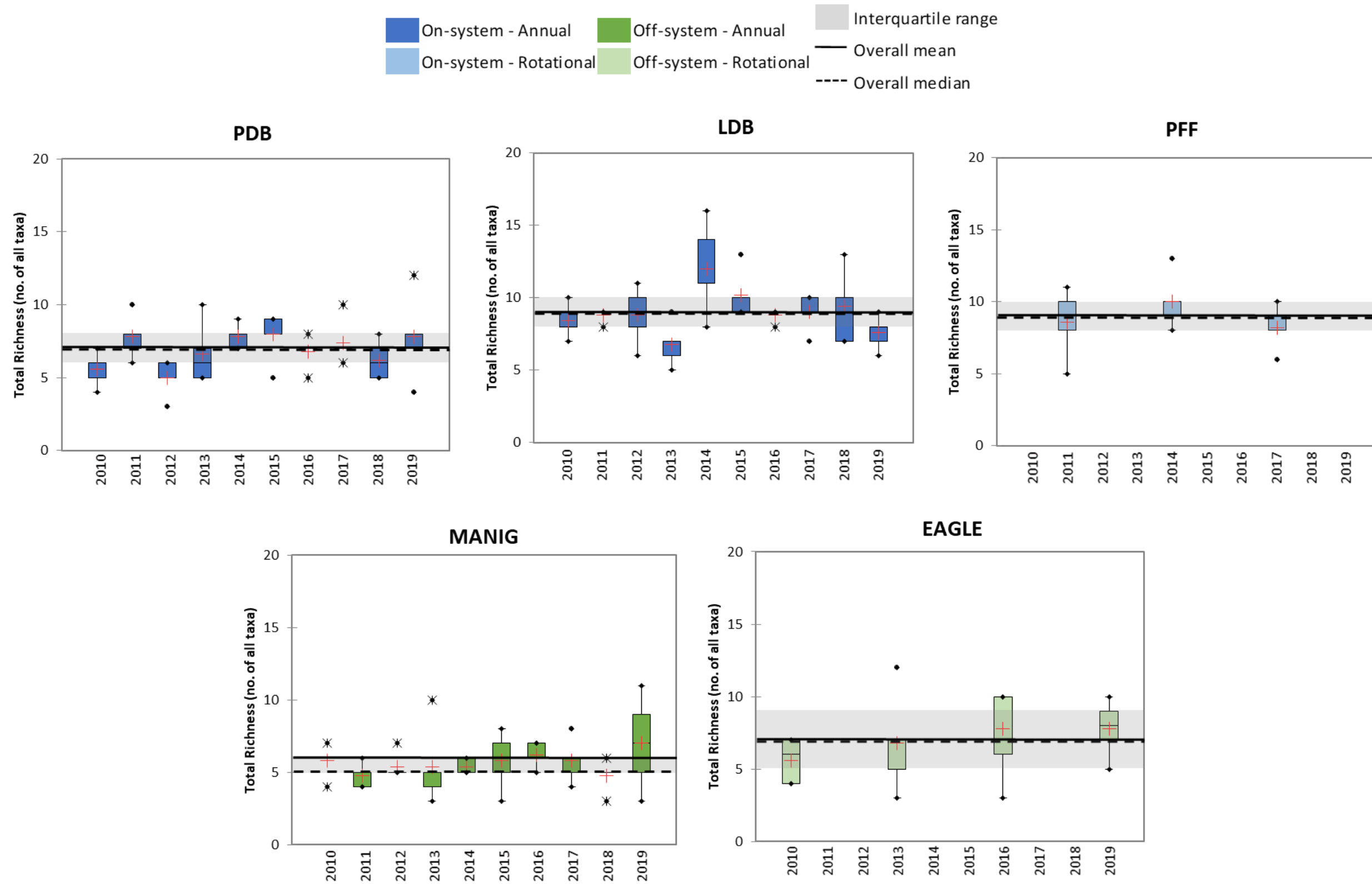


Figure 4.4-2. 2010 to 2019 Offshore benthic invertebrate total richness (family-level).

4.4.2 EPT TAXA RICHNESS

4.4.2.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

Nearshore Habitat

Annual mean EPT taxa richness over the ten years of monitoring ranged from six families (2010) to nine families (2012, 2013, and 2018; Figure 4.4-3). The overall mean and median were eight families, the IQR was 6 to 9 families. Annual means were within the IQR, except in 2010 (below).

Offshore Habitat

Annual mean EPT taxa richness over the ten years of monitoring ranged from one family (2012 and 2014) to three families (2011 and 2017; Figure 4.4-4). The overall mean and median were two families, and the IQR was 1 to 2 families. Annual means were above the IQR in 2011, 2015, and 2017.

Lac du Bonnet

Nearshore Habitat

Annual mean EPT taxa richness over the ten years of monitoring ranged from four families (2018 and 2019) to eight families (2012, 2014, 2015, and 2017; Figure 4.4-3). The overall mean was seven families, the overall median was eight families, and the IQR was 5 to 9 families. Annual means were below the IQR in 2018 and 2019.

Offshore Habitat

Annual mean EPT taxa richness over the ten years of monitoring ranged from two families (2010 to 2013 and 2015 to 2019) to three families (2014; Figure 4.4-4). The overall mean, median, and IQR were two families. Annual means were below the IQR in 2013, and above the IQR in 2012, and 2014 to 2017.

ROTATIONAL SITES

Pine Falls Forebay

Nearshore Habitat

Annual mean EPT taxa richness over the three years of monitoring ranged from eight families (2017) to ten families (2011 and 2014; Figure 4.4-3). The overall mean and median were nine families, and the IQR was 8 to 11 families. Annual means for all years were within the IQR.

Offshore Habitat

Annual mean EPT taxa richness over the three years of monitoring ranged from two families (2017) to three families (2011 and 2014; Figure 4.4-4). The overall mean and median were three families, and the IQR was 2 to 3 families. Annual means for all years were within the IQR.

4.4.2.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

Nearshore Habitat

Annual mean EPT taxa richness over the ten years of monitoring ranged from five families (2010 and 2014) to 11 families (2011; Figure 4.4-3). The overall mean was eight families, the overall median was seven families, and the IQR was 6 to 9 families. Annual means were below the IQR in 2010 and 2014, and above the IQR in 2011 and 2017.

Offshore Habitat

Annual mean EPT taxa richness over the ten years of monitoring was one family for all years (Figure 4.4-4). The overall mean, median, and IQR were also one family. Annual means were below the IQR in 2010, 2011, and 2018, and above the IQR in 2014, 2016, and 2019.

ROTATIONAL SITES

Eaglenest Lake

Nearshore Habitat

Annual mean EPT taxa richness over the four years of monitoring was five families for all years (Figure 4.4-3). The overall mean and median were five families, and the IQR was 4 to 6 families. Annual means for all years were within the IQR.

Offshore Habitat

Annual mean EPT taxa richness over the four years of monitoring ranged from one family (2010 and 2019) to two families (2013 and 2016; Figure 4.4-4). The overall mean and median were two families, and the IQR was 1 to 2 families. Annual means were within the IQR, except in 2013 (above).

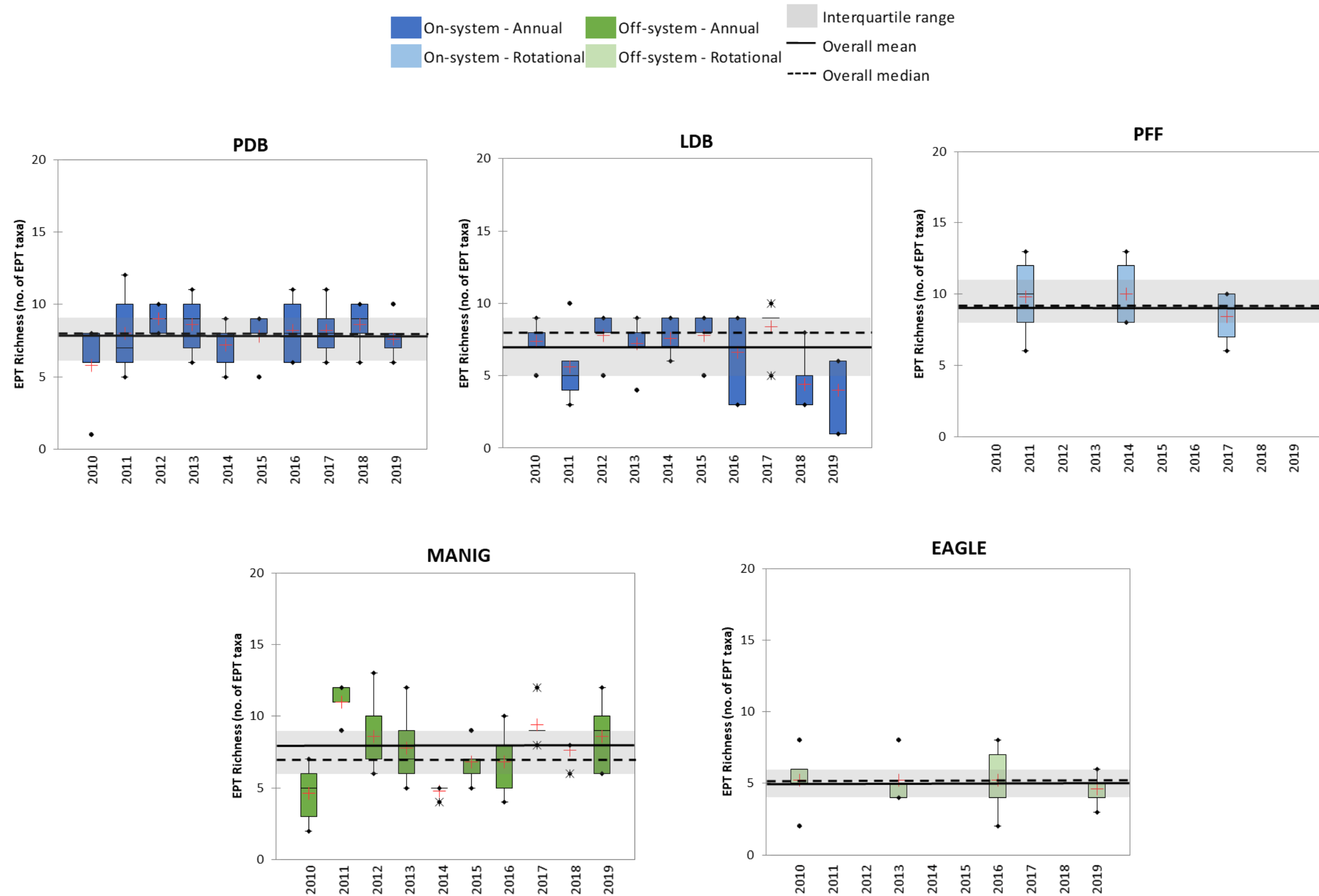


Figure 4.4-3. 2010 to 2019 Nearshore benthic invertebrate EPT richness (family-level).

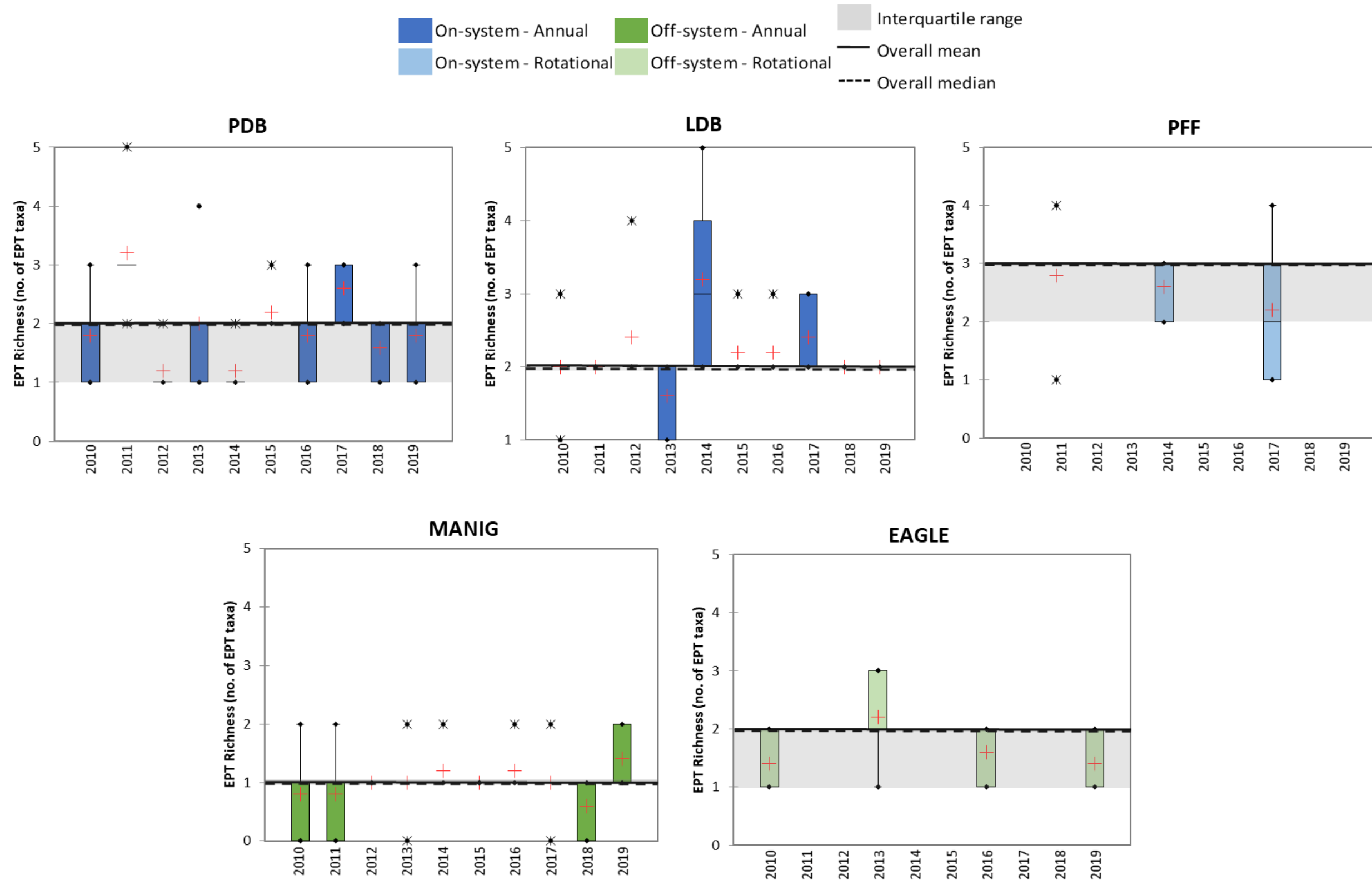


Figure 4.4-4. 2010 to 2019 Offshore benthic invertebrate EPT richness (family-level).

4.5 DIVERSITY

4.5.1 HILL'S EFFECTIVE RICHNESS

4.5.1.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

Nearshore Habitat

Annual mean Hill's effective richness (Hill's index) over the ten years of monitoring ranged from three (2014) to nine (2016 to 2019; Figure 4.5-1). The overall mean and median were seven and the IQR was 5 to 10. Annual means were within the IQR, except for in 2014 (below).

Offshore Habitat

Annual mean Hill's index over the ten years of monitoring ranged from two (2011, 2013) to four (2010, 2014, 2015, 2016, 2018, and 2019; Figure 4.5-2). The overall mean and median were three and the IQR was 3 to 4. Annual means were below the IQR in 2011, 2013 and 2017, and above the IQR in 2014.

Lac du Bonnet

Nearshore Habitat

Annual mean Hill's index over the ten years of monitoring ranged from three (2010) to nine (2014; Figure 4.5-1). The overall mean and median were five and the IQR was 3 to 6. Annual means were below the IQR in 2010, and above the IQR in 2014 and 2015.

Offshore Habitat

Annual mean Hill's index over the ten years of monitoring ranged from three (2018) to six (2014 and 2017; Figure 4.5-2). The overall mean and median were five and the IQR was 4 to 5. Annual means were below the IQR in 2010, 2013 and 2018, and above the IQR in 2014, 2015, 2017 and 2019.

ROTATIONAL SITES

Pine Falls Forebay

Nearshore Habitat

Annual mean Hill's index over the three years of monitoring ranged from six (2011 and 2017) to ten (2014; Figure 4.5-1). The overall mean and median were seven and the IQR was 6 to 8. Annual means were within the IQR, except for in 2014 (above).

Offshore Habitat

Annual mean Hill's index over the three years of monitoring ranged from three (2011) to six (2014; Figure 4.5-2). The overall mean was five, the overall median was four, and the IQR was 4 to 5. Annual means were below the IQR in 2011, and above the IQR in 2014.

4.5.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

Nearshore Habitat

Annual mean Hill's index over the ten years of monitoring ranged from four (2010) to 11 (2019; Figure 4.5-1). The overall mean was eight, the overall median was nine, and the IQR was 6 to 10. Annual means were below the IQR in 2010 and 2015, and above the IQR in 2017 and 2019.

Offshore Habitat

Annual mean Hill's index over the ten years of monitoring ranged from three (2018) to five (2015, 2017, and 2019; Figure 4.5-2). The overall mean and median were four and the IQR was 3 to 5. Annual means were within the IQR, except for in 2019 (above).

ROTATIONAL SITES

Eaglenest Lake

Nearshore Habitat

Annual mean Hill's index over the four years of monitoring ranged from five (2019) to eight (2010; Figure 4.5-1). The overall mean and median were six and the IQR was 5 to 7. Annual means were within the IQR, except for in 2010 (above).

Offshore Habitat

Annual mean Hill's index over the four years of monitoring ranged from two (2013) to four (2016; Figure 4.5-2). The overall mean and median were three and the IQR was 3 to 4. Annual means were within the IQR, except for in 2013 (below).

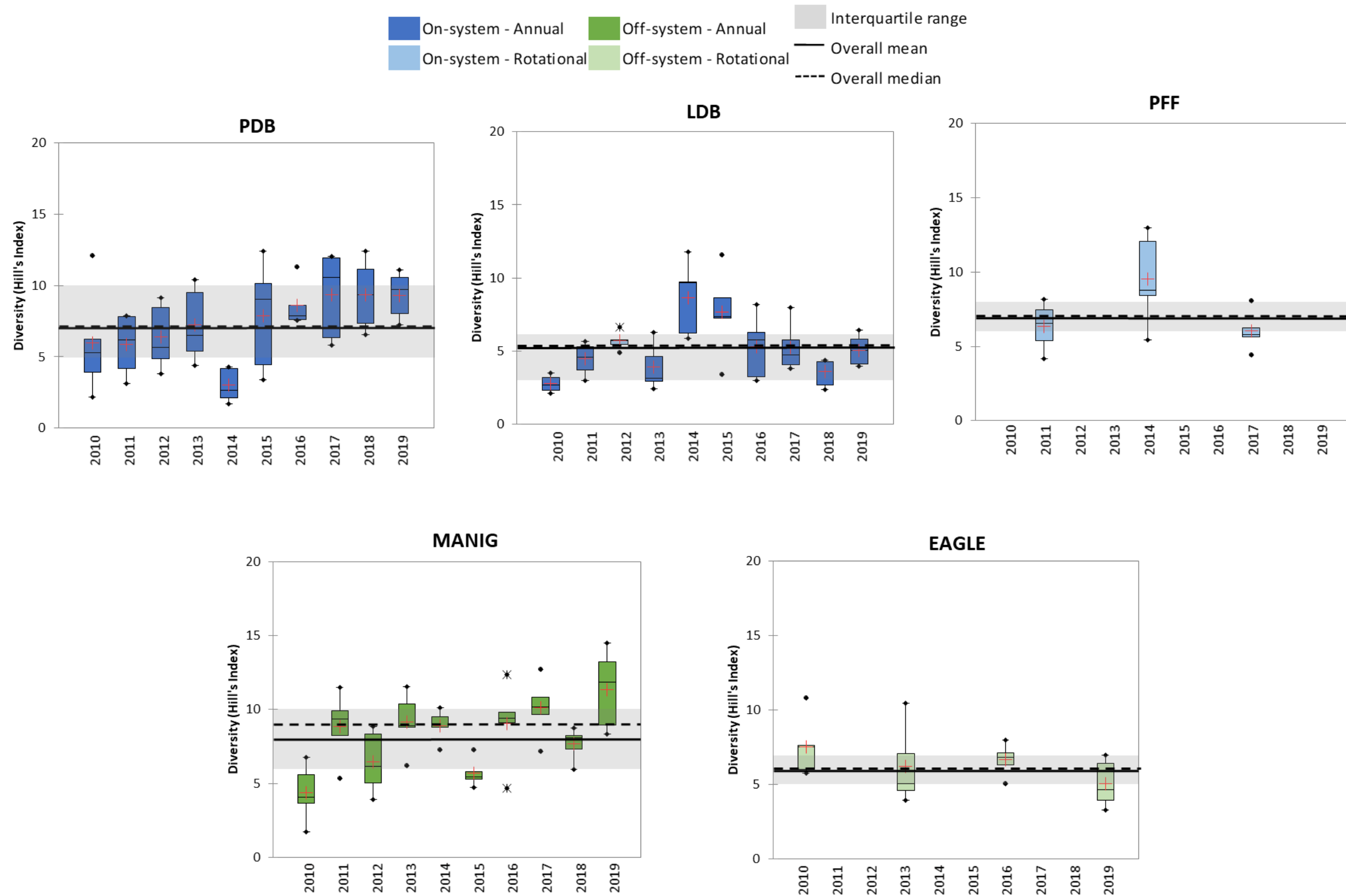


Figure 4.5-1. 2010 to 2019 Nearshore benthic invertebrate diversity (family-level).

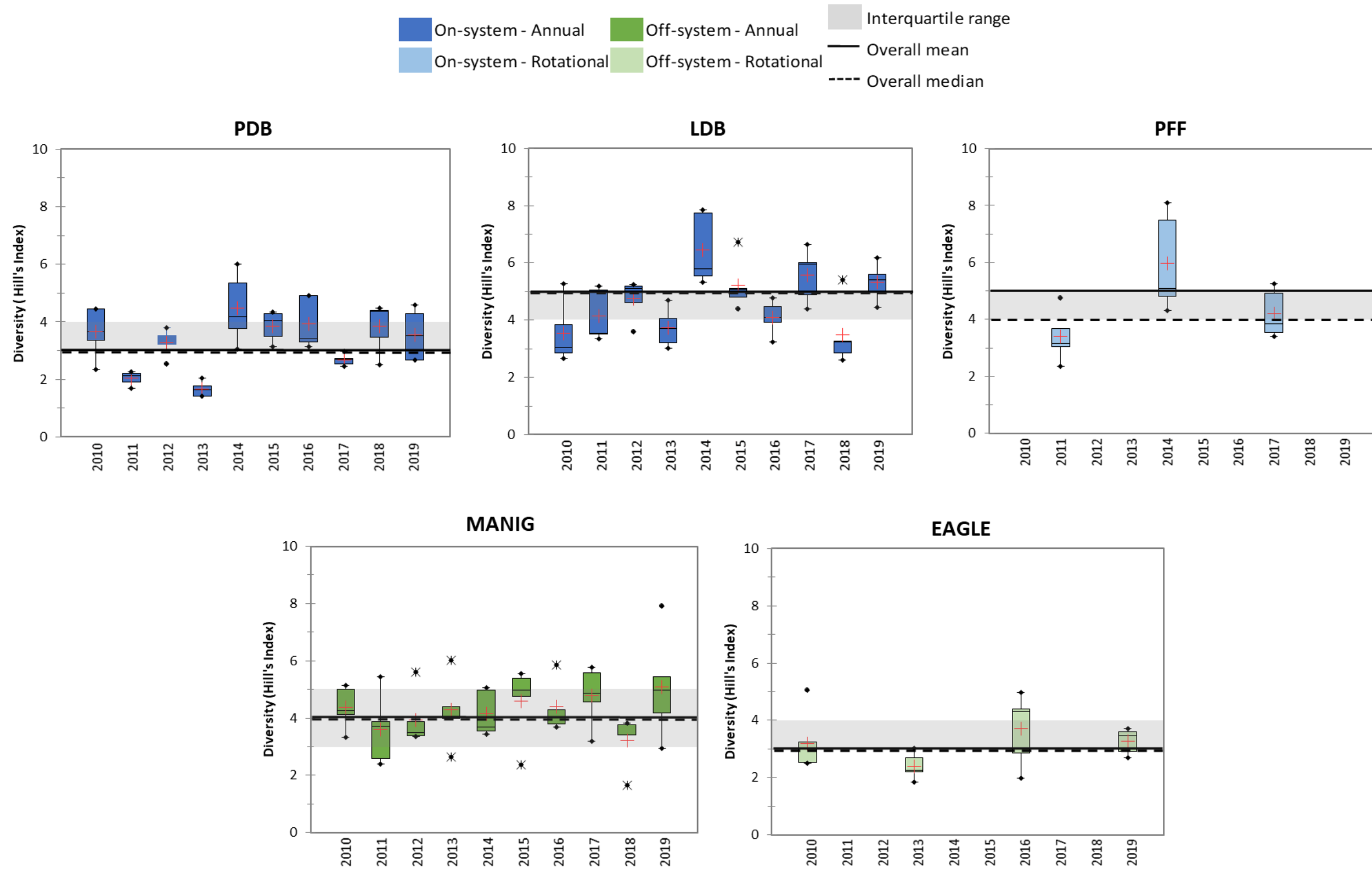


Figure 4.5-2. 2010 to 2019 Offshore benthic invertebrate diversity (family-level).

**APPENDIX 4-1. BENTHIC INVERTEBRATE NEARSHORE AND
OFFSHORE SAMPLING SITES: 2008-2019**

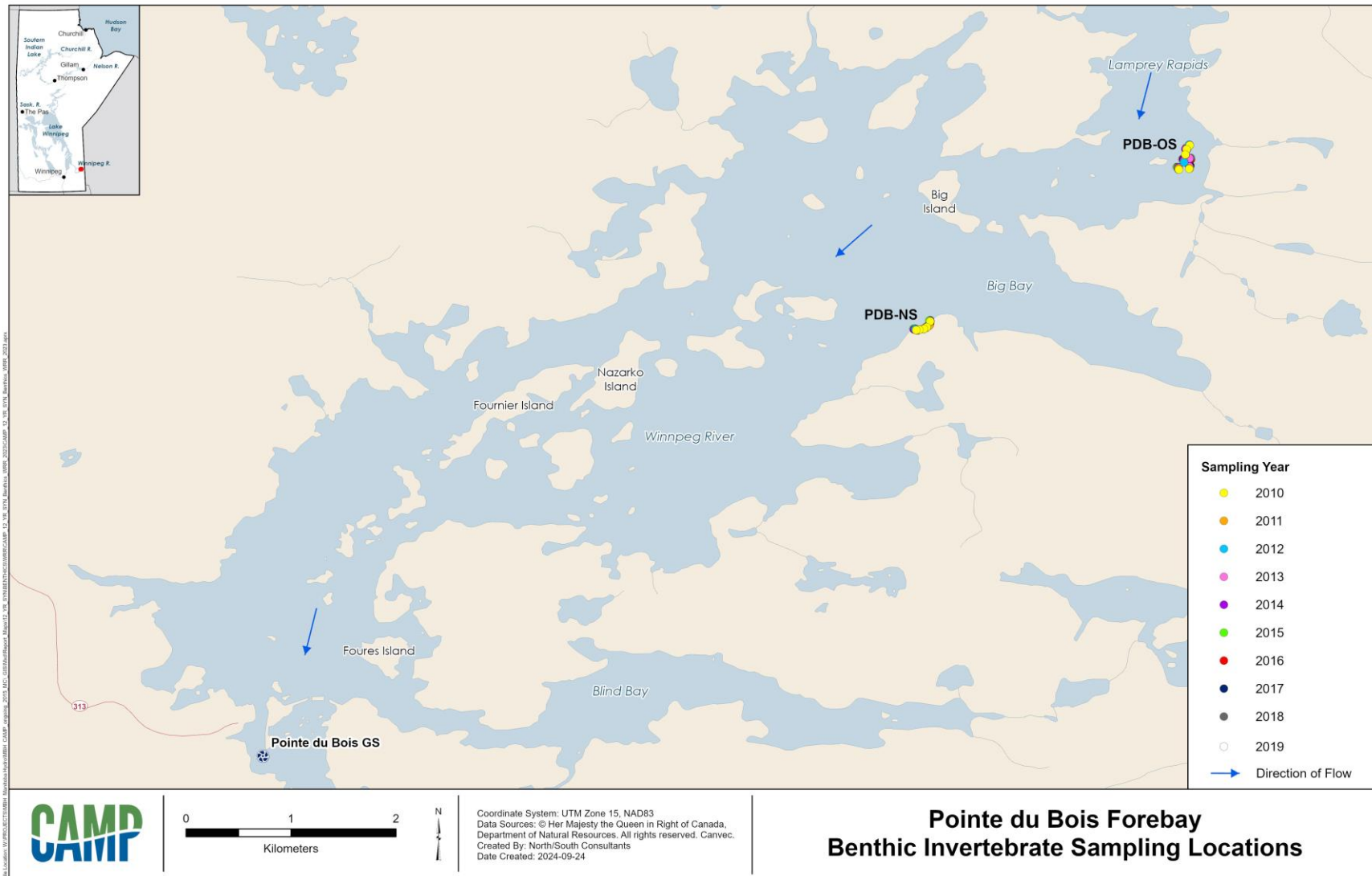


Figure A4-1-1. 2010 to 2019 Pointe du Bois Forebay nearshore (NS) and offshore (OS) benthic invertebrate sampling sites.

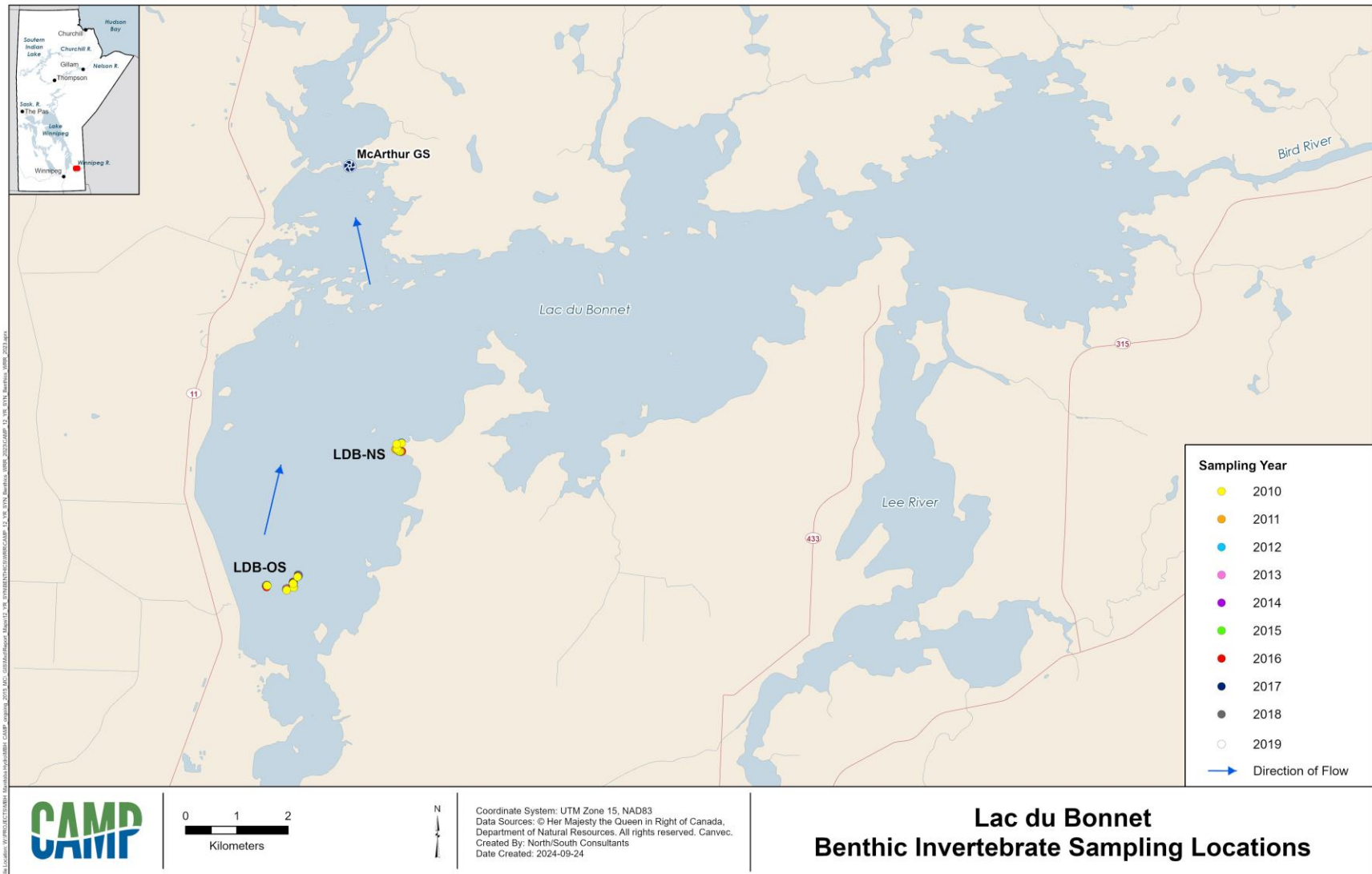


Figure A4-1-2. 2010 to 2019 Lac du Bonnet nearshore (NS) and offshore (OS) benthic invertebrate sampling sites.

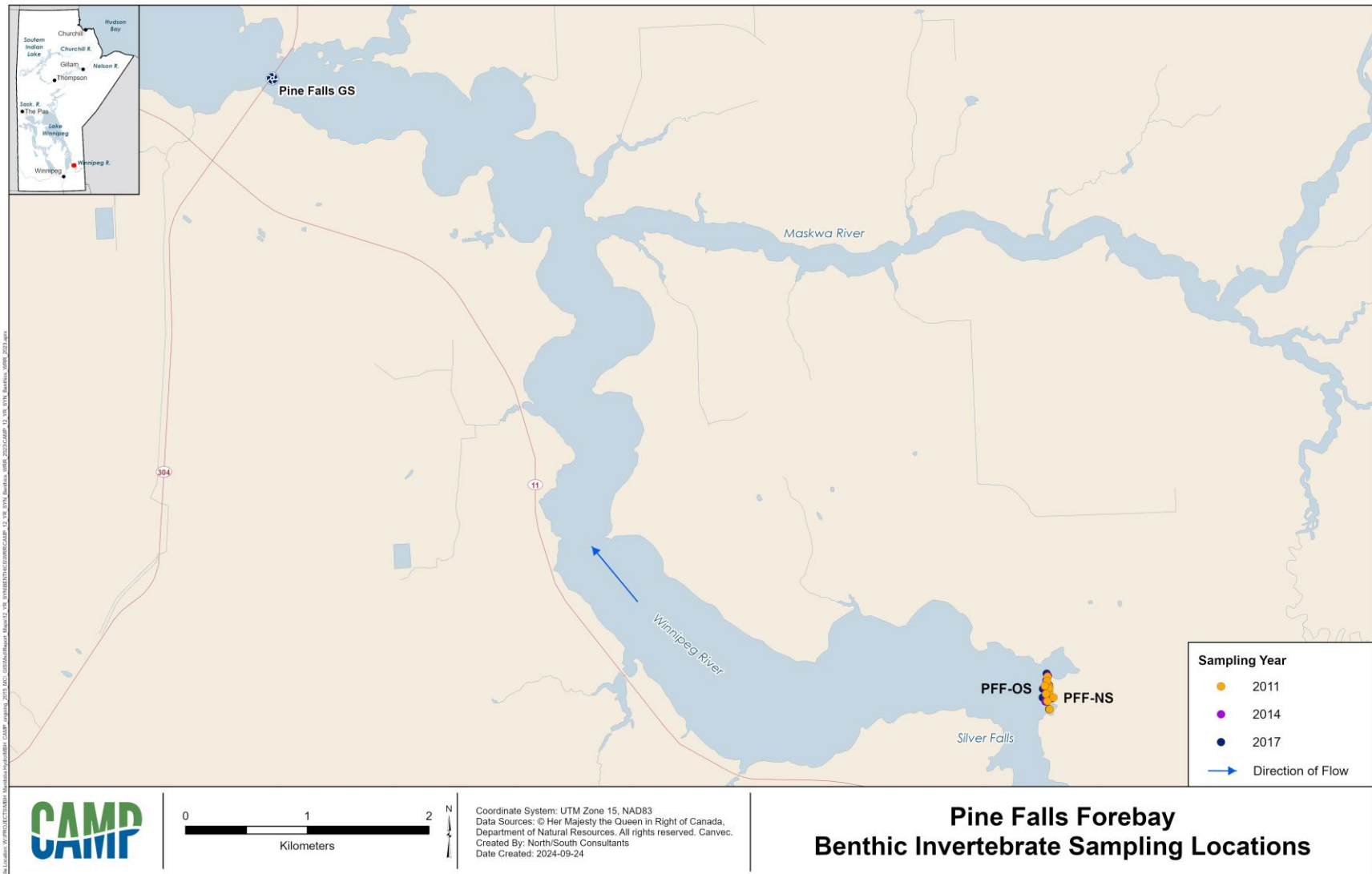


Figure A4-1-3. 2010 to 2019 Pine Falls Forebay nearshore (NS) and offshore (OS) benthic invertebrate sampling sites.

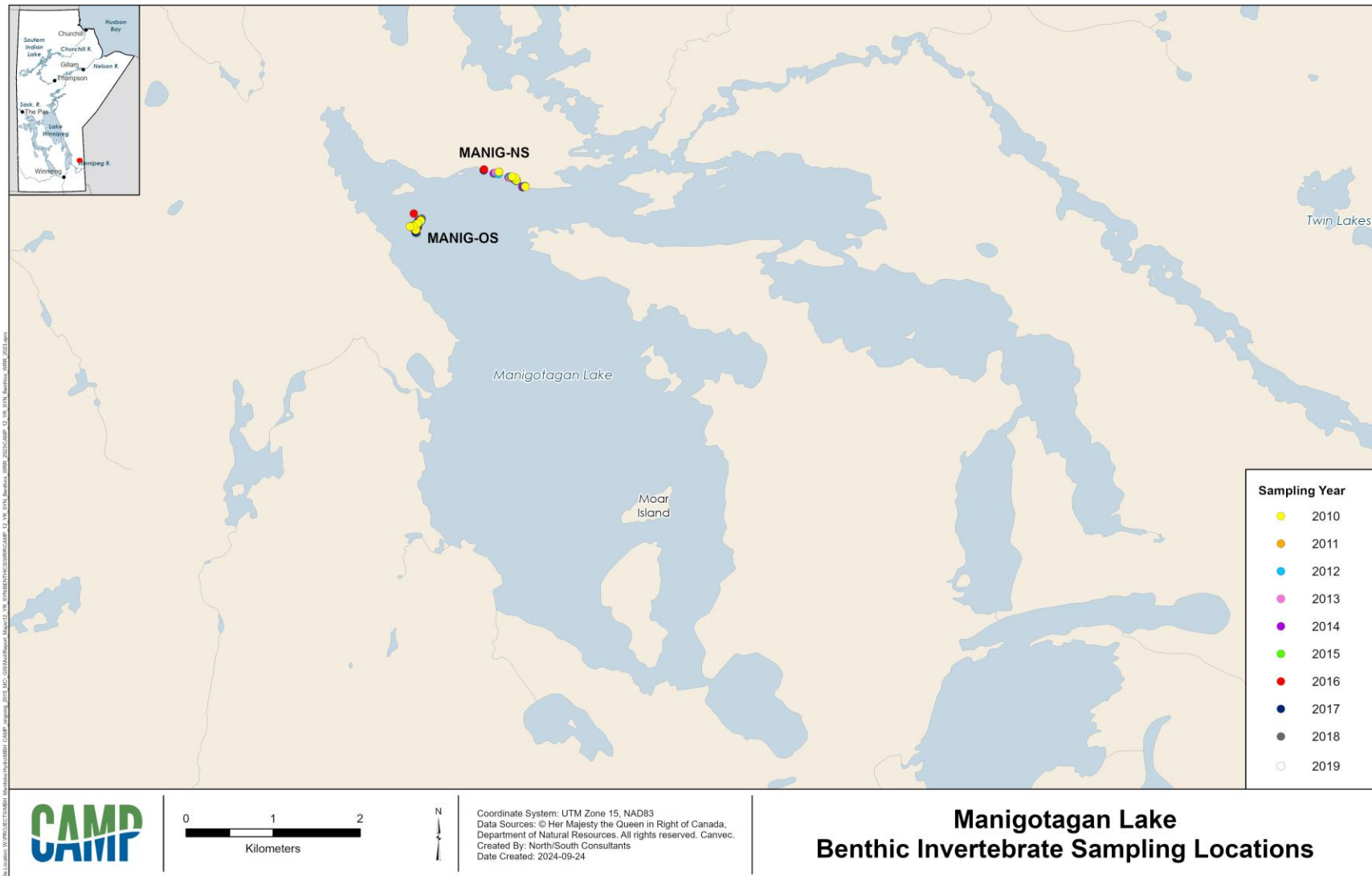


Figure A4-1-4. 2010 to 2019 Manigotagan Lake nearshore (NS) and offshore (OS) benthic invertebrate sampling sites.



Figure A4-1-5. 2010 to 2019 Eaglenest Lake nearshore (NS) and offshore (OS) benthic invertebrate sampling sites.

**APPENDIX 4-2. BENTHIC INVERTEBRATE NEARSHORE AND
OFFSHORE SUPPORTING SUBSTRATE DATA
BY YEAR**

Table A4-2-1. 2010 to 2019 Pointe du Bois Forebay nearshore supporting benthic substrate data.

| Year | Dominant Substrate | Sample Water Depth (m) | Supporting Substrate Analysis | | | | |
|------|-------------------------------------|------------------------|-------------------------------|------|------|--------------|-------------------|
| | | | Mean Particle Size (%) | | | Mean TOC (%) | Texture |
| | | | Sand | Silt | Clay | | |
| 2010 | finer with coarse | 0.5 | 49.1 | 36.2 | 14.7 | 0.3 | Sandy loam |
| 2011 | coarse | 0.9 | 85.5 | 10.5 | 4.0 | 5.6 | Sandy loam |
| 2012 | coarse with fines | 0.4 | 89.7 | 5.6 | 4.7 | 1.6 | Sand / Loamy sand |
| 2013 | finer | 0.4 | 77.4 | 10.2 | 12.4 | 0.9 | Sandy clay loam |
| 2014 | finer with coarse | 0.3 | 76.7 | 7.1 | 16.3 | 0.3 | Sandy loam |
| 2015 | finer, coarse, hard | 0.8 | 78.8 | 16.7 | 4.5 | 0.5 | Sandy loam |
| 2016 | finer | 0.5 | 75.4 | 15.4 | 8.9 | 0.6 | Sandy loam |
| 2017 | coarse with fines | 0.5 | 73.8 | 17.1 | 9.1 | 0.5 | Sandy loam |
| 2018 | finer with coarse | 0.7 | 63.6 | 15.7 | 20.7 | 0.9 | Clay |
| 2019 | finer with coarse, hard, + organics | 0.4 | 94.1 | 2.5 | 3.1 | 0.3 | Sand |

Notes:

1. TOC = Total organic carbon.

Table A4-2-2. 2010 to 2019 Pointe du Bois Forebay offshore supporting benthic substrate data.

| Year | Dominant Substrate | Sample Water Depth (m) | Supporting Substrate Analysis | | | | |
|------|---------------------|------------------------|-------------------------------|------|------|--------------|-------------------------|
| | | | Mean Particle Size (%) | | | Mean TOC (%) | Texture |
| | | | Sand | Silt | Clay | | |
| 2010 | finer with organics | 7.0 | 84.8 | 13.3 | 1.8 | 0.7 | Loamy sand |
| 2011 | finer | 7.4 | 81.9 | 15.2 | 3.0 | 0.7 | Sandy loam |
| 2012 | finer | 6.6 | 76.0 | 16.8 | 7.3 | 0.9 | Sandy loam |
| 2013 | finer | 7.2 | 81.3 | 14.9 | 3.8 | 0.6 | Loamy sand |
| 2014 | finer | 7.3 | 80.3 | 15.6 | 4.0 | 0.6 | Loamy sand |
| 2015 | finer with organics | 7.2 | 73.4 | 22.7 | 3.9 | 0.7 | Loamy sand |
| 2016 | finer | 7.2 | 76.5 | 20.3 | 3.1 | 0.5 | Loamy sand |
| 2017 | finer | 7.2 | 82.1 | 15.0 | 2.9 | 0.6 | Sandy loam / Loamy sand |
| 2018 | finer | 7.1 | 80.6 | 17.1 | 2.3 | 0.5 | Loamy sand |
| 2019 | finer with organics | 7.5 | 77.6 | 19.2 | 3.1 | 0.6 | Loamy sand |

Notes:

1. TOC = Total organic carbon.

Table A4-2-3. 2010 to 2019 Lac du Bonnet nearshore supporting benthic substrate data.

| Year | Dominant Substrate | Sample Water Depth (m) | Supporting Substrate Analysis | | | | |
|------|---------------------------------|------------------------|-------------------------------|------|------|--------------|-------------------------|
| | | | Mean Particle Size (%) | | | Mean TOC (%) | Texture |
| | | | Sand | Silt | Clay | | |
| 2010 | hard, coarse, organics, + fines | 0.4 | 92.8 | 2.4 | 4.7 | 0.4 | Sand |
| 2011 | hard + coarse | 0.9 | 83.9 | 6.2 | 9.9 | 0.4 | Loamy sand |
| 2012 | fines + coarse | no sample | - | - | - | - | - |
| 2013 | coarse | 0.6 | 63.9 | 7.8 | 28.3 | 0.6 | Sand |
| 2014 | coarse + fines | 0.3 | 81.3 | 8.6 | 10.1 | 0.5 | Sandy loam / Loamy sand |
| 2015 | hard, coarse + fines | 0.7 | 80.8 | 7.2 | 11.9 | 0.4 | Sand |
| 2016 | coarse | 0.5 | 74.5 | 10.3 | 15.0 | 1.1 | Sand |
| 2017 | coarse | 0.5 | 69.5 | 10.9 | 19.6 | 1.5 | Sand |
| 2018 | coarse + fines | 0.3 | 95.8 | 1.0 | 2.9 | 0.4 | Sand |
| 2019 | coarse, fines + organics | 0.4 | 81.1 | 4.6 | 14.0 | 0.6 | Sand |

Notes:

1. TOC = Total organic carbon.

Table A4-2-4. 2010 to 2019 Lac du Bonnet offshore supporting benthic substrate data.

| Year | Dominant Substrate | Sample Water Depth (m) | Supporting Substrate Analysis | | | | |
|------|--------------------|------------------------|-------------------------------|------|------|--------------|------------|
| | | | Mean Particle Size (%) | | | Mean TOC (%) | Texture |
| | | | Sand | Silt | Clay | | |
| 2010 | fines | 7.3 | 79.8 | 15.1 | 5.1 | 0.3 | Loamy sand |
| 2011 | fines | 7.0 | 74.0 | 16.4 | 9.7 | 0.4 | Clay loam |
| 2012 | fines | 7.0 | 74.4 | 18.4 | 7.2 | 0.9 | Loamy sand |
| 2013 | fines | 7.1 | 80.0 | 14.8 | 5.3 | 0.6 | Loamy sand |
| 2014 | fines with coarse | 7.1 | 82.9 | 11.0 | 6.1 | 0.6 | Sandy loam |
| 2015 | fines | 7.0 | 77.3 | 17.4 | 5.3 | 0.6 | Loamy sand |
| 2016 | fines | 7.5 | 79.4 | 15.6 | 5.0 | 0.6 | Sandy loam |
| 2017 | fines | 7.3 | 79.1 | 14.5 | 6.5 | 0.7 | Sandy loam |
| 2018 | fines | 7.3 | 85.1 | 10.9 | 4.0 | 0.4 | Sand |
| 2019 | fines with coarse | 7.4 | 81.6 | 12.6 | 5.9 | 0.5 | Sandy loam |

Notes:

1. TOC = Total organic carbon.

Table A4-2-5. 2010 to 2019 Pine Falls Forebay nearshore supporting benthic substrate data.

| Year | Dominant Substrate | Sample Water Depth (m) | Supporting Substrate Analysis | | | | |
|------|--------------------|------------------------|-------------------------------|------|------|--------------|------------|
| | | | Mean Particle Size (%) | | | Mean TOC (%) | Texture |
| | | | Sand | Silt | Clay | | |
| 2011 | coarse | 0.7 | 80.9 | 6.9 | 12.2 | 0.2 | Sand |
| 2014 | coarse + fines | 0.4 | 70.0 | 10.3 | 19.7 | 0.7 | Loamy sand |
| 2017 | coarse | 0.5 | 94.7 | 2.9 | 1.9 | 0.7 | - |

Notes:

1. TOC = Total organic carbon.

Table A4-2-6. 2010 to 2019 Pine Falls Forebay offshore supporting benthic substrate data.

| Year | Dominant Substrate | Sample Water Depth (m) | Supporting Substrate Analysis | | | | |
|------|-------------------------|------------------------|-------------------------------|------|------|--------------|-----------|
| | | | Mean Particle Size (%) | | | Mean TOC (%) | Texture |
| | | | Sand | Silt | Clay | | |
| 2011 | fines | 7.1 | 21.6 | 36.0 | 42.3 | 2.2 | Clay |
| 2014 | fines, coarse, organics | 6.6 | 18.0 | 34.5 | 47.6 | 2.5 | Clay |
| 2017 | fines | 8.1 | 28.6 | 40.5 | 30.9 | 2.7 | Clay loam |

Notes:

1. TOC = Total organic carbon.

Table A4-2-7. 2010 to 2019 Manigotagan Lake nearshore supporting benthic substrate data.

| Year | Dominant Substrate | Sample Water Depth (m) | Supporting Substrate Analysis | | | | |
|------|--------------------|------------------------|-------------------------------|------|------|--------------|---------|
| | | | Mean Particle Size (%) | | | Mean TOC (%) | Texture |
| | | | Sand | Silt | Clay | | |
| 2010 | coarse + fines | no sample | - | - | - | - | - |
| 2011 | coarse | no sample | - | - | - | - | - |
| 2012 | coarse | no sample | - | - | - | - | - |
| 2013 | coarse | no sample | - | - | - | - | - |
| 2014 | coarse | no sample | - | - | - | - | - |
| 2015 | coarse + fines | no sample | - | - | - | - | - |
| 2016 | coarse | no sample | - | - | - | - | - |
| 2017 | coarse + fines | 0.5 | 97.7 | 1.4 | 0.1 | 0.1 | Sand |
| 2018 | coarse + fines | 0.4 | 97.7 | 1.2 | 0.8 | 0.2 | Sand |
| 2019 | coarse + fines | no sample | - | - | - | - | - |

Notes:

1. TOC = Total organic carbon.

Table A4-2-8. 2010 to 2019 Manigotagan Lake offshore supporting benthic substrate data.

| Year | Dominant Substrate | Sample Water Depth (m) | Supporting Substrate Analysis | | | | |
|------|--------------------|------------------------|-------------------------------|------|------|--------------|---------|
| | | | Mean Particle Size (%) | | | Mean TOC (%) | Texture |
| | | | Sand | Silt | Clay | | |
| 2010 | fines | 7.9 | 92.5 | 6.1 | 1.5 | 0.5 | Sand |
| 2011 | fines | 6.7 | 93.2 | 4.1 | 2.6 | 0.5 | Sand |
| 2012 | fines | 7.2 | 88.3 | 8.6 | 3.1 | 0.6 | Sand |
| 2013 | fines | 6.6 | 89.3 | 7.8 | 3.0 | 0.5 | Sand |
| 2014 | fines | 7.7 | 89.9 | 8.1 | 2.0 | 0.5 | Sand |
| 2015 | fines | 7.9 | 87.1 | 9.7 | 3.2 | 0.8 | Sand |
| 2016 | fines | 7.9 | 78.4 | 16.2 | 5.4 | 0.9 | Sand |
| 2017 | fines | 7.0 | 89.0 | 8.3 | 2.7 | 0.7 | Sand |
| 2018 | fines, gravel | 7.5 | 86.0 | 10.3 | 3.6 | 0.7 | Sand |
| 2019 | fines | 8.6 | 81.3 | 14.5 | 4.2 | 1.0 | Sand |

Notes:

1. TOC = Total organic carbon.

Table A4-2-9. 2010 to 2019 Eaglenest Lake nearshore supporting benthic substrate data.

| Year | Dominant Substrate | Sample Water Depth (m) | Supporting Substrate Analysis | | | | |
|------|--------------------------|------------------------|-------------------------------|------|------|--------------|---------|
| | | | Mean Particle Size (%) | | | Mean TOC (%) | Texture |
| | | | Sand | Silt | Clay | | |
| 2010 | hard, fines, organics | 0.8 | 46.3 | 30.8 | 22.9 | 7.3 | Loam |
| 2013 | hard | no sample | - | - | - | - | - |
| 2016 | hard | no sample | - | - | - | - | - |
| 2019 | hard + coarse with fines | no sample | - | - | - | - | - |

Notes:

1. TOC = Total organic carbon.

Table A4-2-10. 2010 to 2019 Eaglenest Lake offshore supporting benthic substrate data.

| Year | Dominant Substrate | Sample Water Depth (m) | Supporting Substrate Analysis | | | | |
|------|--------------------|------------------------|-------------------------------|------|------|--------------|-------------------------|
| | | | Mean Particle Size (%) | | | Mean TOC (%) | Texture |
| | | | Sand | Silt | Clay | | |
| 2010 | fines, organics | 7.2 | 85.1 | 13.2 | 1.6 | 0.4 | Loamy sand |
| 2013 | fines | 6.9 | 78.2 | 18.1 | 3.7 | 0.5 | Sandy loam / Loamy sand |
| 2016 | fines | 7.0 | 78.4 | 18.4 | 3.1 | 0.5 | Loamy sand |
| 2019 | fines, organics | 6.4 | 73.3 | 24.4 | 2.2 | 0.6 | Sandy loam |

Notes:

1. TOC = Total organic carbon

5.0 FISH COMMUNITY

5.1 INTRODUCTION

The following presents the results of fish community monitoring conducted from 2008 to 2019 in the Winnipeg River Region. Five waterbodies were monitored in the Winnipeg River Region: two on-system annual sites (the Pointe du Bois Forebay and Lac du Bonnet); one on-system rotational site (the Pine Falls Forebay); one off-system annual site (Manigotagan Lake); and one off-system rotational site (Eaglenest Lake; Table 5.1-1 and Figure 5.1-1). Eaglenest Lake is located on the Winnipeg River upstream of the Pointe du Bois Forebay and is not affected by Manitoba Hydro's hydraulic operating system and is therefore identified as an "off-system" site. There were no departures from the planned field sampling during the 12-year period.

Monitoring targets both small-bodied fish species (i.e., forage fish) and large-bodied fish species (e.g., fish targeted in subsistence, commercial, and/or recreational fisheries). Within a given waterbody, sampling was conducted at approximately the same time of year during each year of monitoring. Standard gang index gill nets (GN; 51, 76, 95, 108, and 127 mm stretched mesh panels) were set at each site and a small mesh index gillnet gang (SN; 16, 20, and 25 mm bar measure panels) was attached to the end of the standard gang at approximately every third site (Appendix 5-1). Gill nets were set for approximately 24 hours. All fish captured at each site were counted by mesh size and species. Individual metrics (e.g., length, weight, deformities, erosion, lesions, and tumours [DELTs], sex and maturity, age) were collected for species of management interest (i.e., "target" species). These include: Walleye (*Sander vitreus*) and Northern Pike (*Esox lucius*) from all waterbodies in all years; Sauger (*S. canadensis*) from Lac du Bonnet and the Pointe du Bois Forebay in all years; Sauger from 2016-2019 in Eaglenest Lake and 2017-2019 in the Pine Falls Forebay; and White Sucker (*Catostomus commersonii*) from all waterbodies starting in 2010. All other species were bulk weighed.

Five fish community indicators (abundance, condition, growth, recruitment, and community diversity) were selected for detailed reporting (Table 5.1-2). Metrics for these indicators that are presented herein include: catch-per-unit-effort (CPUE); Fulton's condition factor (KF); relative weight (Wr); fork length-at-age (FLA); relative year-class strength (RYCS); Hill's effective species richness (Hill's index); and relative species abundance (RSA; Table 5.1-2).

A detailed description of the program design and sampling methods is provided in Technical Document 1, Section 2.5.

Table 5.1-1. 2008-2019 Inventory of fish community sampling.

| Waterbody/Area | Sampling Year | | | | | | | | | | | |
|----------------|---------------|------|------|------|------|------|------|------|------|------|------|------|
| | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
| PDB | • | • | • | • | • | • | • | • | • | • | • | • |
| LDB | • | • | • | • | • | • | • | • | • | • | • | • |
| PFF | | | | • | | | • | | | • | | |
| MANIG | • | • | • | • | • | • | • | • | • | • | • | • |
| EAGLE | | | • | | | • | | | • | | | • |

Table 5.1-2. Fish community indicators and metrics.

| Indicator | Metric | Units |
|-------------|---|--|
| Abundance | • Catch-Per-Unit-Effort (CPUE) | # fish/30 m/24 hour (h) # fish/100 m/24 h |
| Condition | • Fulton’s Condition Factor (KF) | - |
| | • Relative Weight (Wr) | - |
| Growth | • Fork Length-At-Age (FLA) | mm |
| Recruitment | • Relative Year-Class Strength (RYCS) | - |
| Diversity | • Hill’s Effective Species Richness | species |
| | • Relative Species Abundance (RSA) ¹ | % |

Notes:

1. Supporting metric.

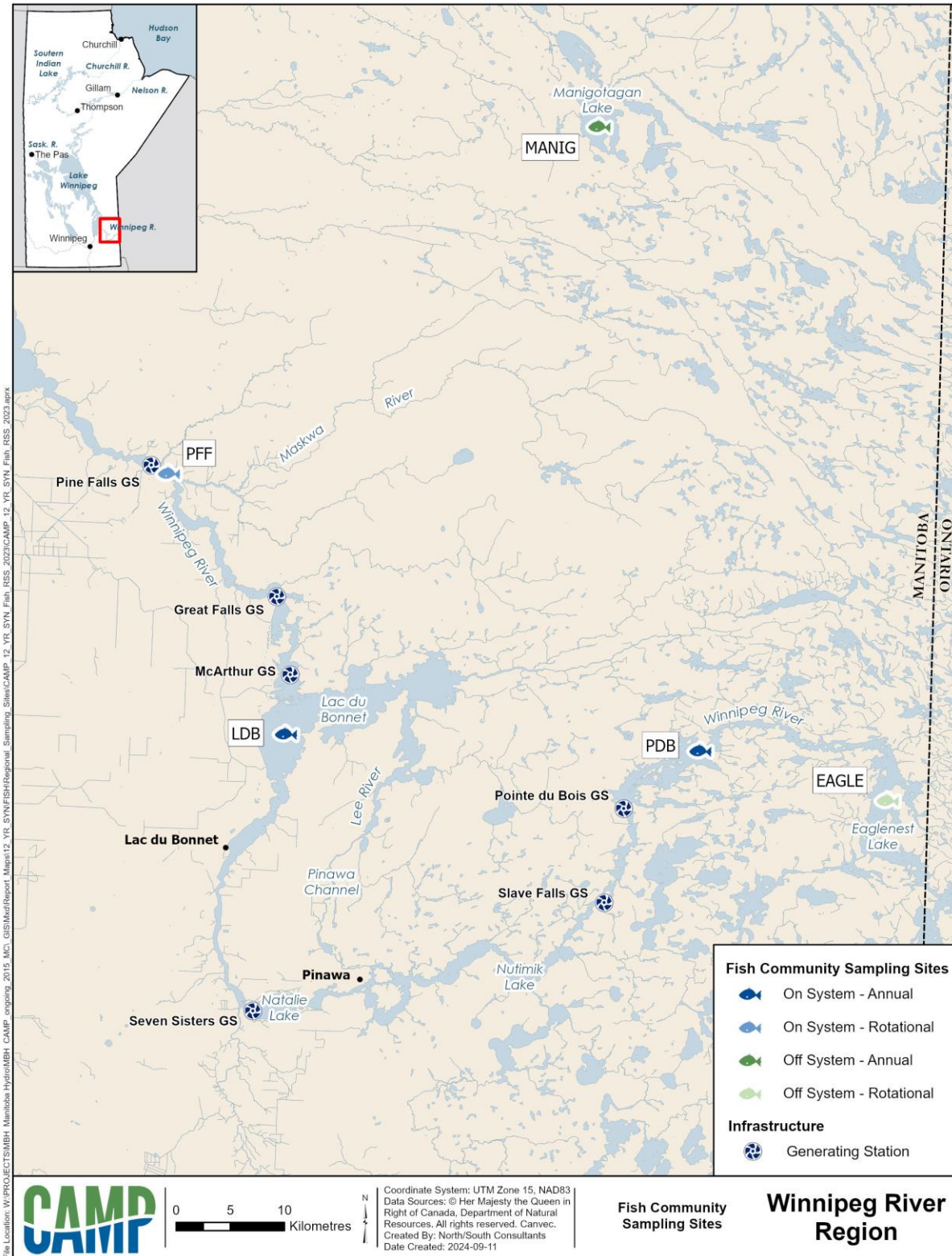


Figure 5.1-1. 2008-2019 Fish community sampling sites.

5.2 ABUNDANCE

5.2.1 CATCH-PER-UNIT-EFFORT

5.2.1.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

Standard Gang Index Gill Nets

The annual mean CPUE varied over the 12 years of monitoring from a low of 21.2 in 2010 to a maximum of 52.1 fish/100 m/24 h in 2019 (Table 5.2-1; Figure 5.2-1).

The overall mean CPUE was 38.7, the median was 36.2, and the IQR was 33.6-49.2 fish/100 m/24 h (Figure 5.2-1). The annual mean CPUE fell within the overall IQR except in 2010 and 2014 when it was below the IQR and in 2011 and 2019 when it was above the IQR.

Small Mesh Index Gill Nets

The annual mean CPUE in the small mesh gangs over the 12 years of monitoring was more variable than in the standard gangs, with the mean ranging from a low of 10.6 in 2009 to a high of 144.6 fish/30 m/24 h in 2012 (Table 5.2-1; Figure 5.2-2).

The overall mean CPUE was 76.5, the median was 73.6, and the IQR was 45.2-114.3 fish/30 m/24 h (Figure 5.2-2). The annual mean CPUE fell within the overall IQR except in 2009, 2010, and 2016 when it was below the IQR and in 2011, 2012, and 2013 when it was above the IQR.

Lake Whitefish

Lake Whitefish (*Coregonus clupeaformis*) was not selected as a target species in any of the Winnipeg River Region waterbodies even though it is a target species in other CAMP regions.

Northern Pike

Catches of Northern Pike were relatively low in the Pointe du Bois Forebay over the 12 years of monitoring, with the annual mean CPUE ranging from a low of 0.7 in 2012 to a high of 3.3 fish/100 m/24 h in 2013 (Table 5.2-1; Figure 5.2-3).

The overall mean CPUE was 1.6, the median was 1.5, and the IQR was 1.3-1.8 fish/100 m/24 h (Figure 5.2-3). The annual mean CPUE fell within the overall IQR except in 2010 and 2012 when it was below the IQR and in 2008, 2013, and 2014 when it was above the IQR.

Sauger

The annual mean CPUE over the 12 years of monitoring varied up to about four-fold from year-to-year, ranging from a low of 2.8 in 2010 to a high of 10.4 fish/100 m/24 h in 2009 (Table 5.2-1; Figure 5.2-4).

The overall mean CPUE was 7.2, the median was 8.9, and the IQR was 5.1-8.9 fish/100 m/24 h (Figure 5.2-4). The annual mean CPUE fell within the overall IQR except in 2010 when it was below the IQR and in 2008 and 2009 when it was above the IQR.

Walleye

The annual mean CPUE over the 12 years of monitoring varied by up to about two-fold from year-to-year, ranging from a low of 2.6 in 2016 to a high of 6.7 fish/100 m/24 h in 2009 (Table 5.2-1; Figure 5.2-5).

The overall mean CPUE was 4.1, the median was 3.7, and the IQR was 3.1-5.0 fish/100 m/24 h (Figure 5.2-5). The annual mean CPUE fell within the overall IQR except in 2008, 2009, and 2019 when it was above the IQR.

White Sucker

The annual mean CPUE over the 12 years of monitoring varied up to about three-fold from year-to-year, with the mean ranging from a low of 5.6 in 2016 to a high of 19.3 fish/100 m/24 h in 2009 (Table 5.2-1; Figure 5.2-6).

The overall mean CPUE was 10.3, the median was 8.9, and the IQR was 8.0-11.7 fish/100 m/24 h (Figure 5.2-6). The annual mean CPUE fell within the overall IQR except in 2014 and 2016 when it was below the IQR and in 2008, 2009, and 2011 when it was above the IQR.

Lac du Bonnet

Standard Gang Index Gill Nets

The annual mean CPUE over the 12 years of monitoring varied up to about two-fold from year-to-year, with the mean ranging from a low of 32.4 in 2008 to a high of 51.1 fish/100 m/24 h in 2019 (Table 5.2-1; Figure 5.2-1).

The overall mean CPUE was 41.5, the median was 42.3, and the IQR was 37.2-45.3 fish/100 m/24 h (Figure 5.2-1). The annual mean CPUE fell within the overall IQR except in 2008, 2010, and 2015 when it was below the IQR and in 2012, 2013, and 2019 when it was above the IQR.

Small Mesh Index Gill Nets

The annual mean CPUE in the small mesh gangs over the 12 years of monitoring was more variable than in the standard gangs, with the mean ranging from a low of 25.4 in 2016 to a high of 424.6 fish/30 m/24 h in 2013 (Table 5.2-1; Figure 5.2-2).

The overall mean CPUE was 166.2, the median was 144.8, and the IQR was 51.8-216.8 fish/30 m/24 h (Figure 5.2-2). The annual mean CPUE fell within the overall IQR except in 2009, 2010, and 2016 when it was below the IQR and in 2011, 2013, and in 2015 when it was above the IQR.

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

Catches of Northern Pike were relatively low in Lac du Bonnet over the 12 years of monitoring, with the annual mean ranging from a low of 1.2 in 2009 to a high of 5.9 fish/100 m/24 in 2012 (Table 5.2-1; Figure 5.2-3).

The overall mean and median CPUE were 3.8 and the IQR was 3.4-4.5 fish/100 m/24 (Figure 5.2-3). The annual mean CPUE fell within the overall IQR except in 2008, 2009, and 2011 when it was below the IQR and in 2012, 2013, and 2014 when it was above the IQR.

Sauger

The annual mean CPUE over the 12 years of monitoring varied up to about two-fold from year-to-year, with the mean ranging from a low of 6.4 in 2010 to a high of 12.7 fish/100 m/24 h in 2009 (Table 5.2-1; Figure 5.2-4).

The overall mean CPUE was 8.8, the median was 8.6, and the IQR was 7.5-9.6 fish/100 m/24 h (Figure 5.2-4). The annual mean CPUE fell within the overall IQR except in 2010, 2015, and 2019 when it was below the IQR and in 2009 and 2016 when it was above the IQR.

Walleye

The annual mean CPUE over the 12 years of monitoring varied up to about two-fold from year-to-year, with the mean ranging from a low of 4.8 in 2008 to a high of 12.4 fish/100 m/24 h in 2012 (Table 5.2-1; Figure 5.2-5).

The overall mean CPUE was 7.0, the median was 6.3, and the IQR was 5.8-7.8 fish/100 m/24 h (Figure 5.2-5). The annual mean CPUE fell within the overall IQR except in 2008, 2010, and 2016 when it was below the IQR and in 2012 and 2014 when it was above the IQR.

White Sucker

The annual mean CPUE over the 12 years of monitoring varied up to about two-fold from year-to-year, with the mean ranging from a low of 3.1 in 2012 to a high of 6.7 fish/100 m/24 h in 2019 (Table 5.2-1; Figure 5.2-6).

The overall mean CPUE was 4.3, the median was 3.8, and the IQR was 3.7-5.0 fish/100 m/24 h (Figure 5.2-6). The annual mean CPUE fell within the overall IQR except in 2009, 2012, and 2015 when it was below the IQR and in 2008, 2013, and 2019 when it was above the IQR.

ROTATIONAL SITES

Pine Falls Forebay

Standard Gang Index Gill Nets

The annual mean CPUE over the three years of monitoring varied by up to about two-fold, with the mean ranging from a low of 18.4 in 2011 to a high of 35.3 fish/100 m/24 h in 2014 (Table 5.2-1; Figure 5.2-1).

The overall mean CPUE was 26.1, the median was 24.6, and the IQR was 21.5-29.9 fish/100 m/24 h (Figure 5.2-1). The annual mean CPUE was below the IQR in 2011 and was above the IQR in 2014.

Small Mesh Index Gill Nets

The annual mean CPUE over the three years of monitoring ranged from a low of 3.5 in 2011 to a high of 9.0 fish/30 m/24 h in 2017 (Table 5.2-1; Figure 5.2-2).

The overall mean CPUE was 6.9, the median was 8.3, and the IQR was 5.9-8.6 fish/30 m/24 h (Figure 5.2-2). The annual mean CPUE was below the IQR in 2011 and was above the IQR in 2014.

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

Catches of Northern Pike were relatively low in the Pine Falls Forebay over the three years of monitoring, with the annual mean ranging from a low of 0.5 in 2011 to a high of 1.1 fish/100 m/24 h in 2017 (Table 5.2-1; Figure 5.2-3).

The overall mean and median CPUE were 0.7 and the IQR was 0.6-0.9 fish/100 m/24 h (Figure 5.2-3). The annual mean CPUE was below the IQR in 2011 and above the IQR in 2017.

Sauger

Catches of Sauger were relatively low in the Pine Falls Forebay over the three years of monitoring, with the annual mean ranging from a low of 0.9 in 2011 to a high of 2.5 fish/100 m/24 h in 2017 (Table 5.2-1; Figure 5.2-4).

The overall mean CPUE was 1.9, the median was 2.3, and the IQR was 1.6-2.4 fish/100 m/24 h (Figure 5.2-4). The annual mean CPUE was below the IQR in 2011 and was marginally above the IQR in 2017.

Walleye

The annual mean CPUE over the three years of monitoring varied by up to about three-fold, with the mean ranging from a low of 2.0 in 2011 to a high of 5.8 fish/100 m/24 h in 2014 (Table 5.2-1; Figure 5.2-5).

The overall mean CPUE was 3.5, the median was 2.7, and the IQR was 2.3-4.3 fish/100 m/24 h (Figure 5.2-5). The annual mean CPUE was above the IQR in 2014.

White Sucker

Catches of White Sucker were relatively low in the Pine Falls Forebay over the three years of monitoring, with the annual mean ranging from a low of 1.5 in 2011 to a high of 2.9 fish/100 m/24 h in 2014 (Table 5.2-1; Figure 5.2-6).

The overall mean CPUE was 2.3, the median was 2.7, and the IQR was 2.1-2.8 fish/100 m/24 h (Figure 5.2-6). The annual mean CPUE was below the IQR in 2011 and was marginally above the IQR in 2014.

5.2.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

Standard Gang Index Gill Nets

The annual mean CPUE over the 12 years of monitoring varied by up to about two-fold, with the mean ranging from a low of 43.0 in 2013 to a high of 68.0 fish/100 m/24 h in 2008 (Table 5.2-1; Figure 5.2-1).

The overall mean CPUE was 55.6, the median was 53.6, and the IQR was 49.3-64.2 fish/100 m/24 h (Figure 5.2-1). The annual mean CPUE fell within the overall IQR except in 2011, 2013, and 2019 when it was below the IQR and in 2008, 2015, and 2018 when it was above the IQR.

Small Mesh Index Gill Nets

The annual mean CPUE in the small mesh gangs over the 12 years of monitoring was more variable than in the standard gangs, with the mean ranging from a low of 2.6 in 2011 to a high of 71.4 fish/30 m/24 h in 2012 (Table 5.2-1; Figure 5.2-2).

The overall mean CPUE was 20.0, the median was 12.7, and the IQR was 7.2-25.2 fish/30 m/24 h (Figure 5.2-2). The annual mean CPUE fell within the overall IQR except in 2009, 2011, and 2017 when it was below the IQR and in 2012, 2013, and 2018 when it was above the IQR.

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

Catches of Northern Pike were relatively low in Manigotagan Lake over the 12 years of monitoring, with the annual mean CPUE ranging from a low of 2.0 in 2009 to a high of 4.4 fish/100 m/24 h in 2016 (Table 5.2-1; Figure 5.2-3).

The overall mean and median CPUE were 3.1 and the IQR was 2.3-3.8 fish/100 m/24 h (Figure 5.2-3). The annual mean CPUE fell within the overall IQR except in 2009, 2010, and 2019 when it was below the IQR and in 2014, 2016, and 2018 when it was above the IQR.

Sauger

Sauger were not captured in Manigotagan Lake over the 12 years of monitoring (Table 5.2-1).

Walleye

The annual mean CPUE over the 12 years of monitoring varied by up to about two-fold, with the mean ranging from a low of 11.0 in 2013 to a high of 23.5 fish/100 m/24 h in 2008 (Table 5.2-1; Figure 5.2-5).

The overall mean CPUE was 17.1, the median was 16.2, and the IQR was 15.5-18.7 fish/100 m/24 h (Figure 5.2-5). The annual mean CPUE fell within the overall IQR except in in 2013, 2014, and 2015 when it was below the IQR and in 2008, 2010, and 2018 when it was above the IQR.

White Sucker

The annual mean CPUE over the 12 years of monitoring varied by up to about five-fold, with the mean ranging from a low of 1.7 in 2011 to a high of 8.0 fish/100 m/24 h in 2018 (Table 5.2-1; Figure 5.2-6).

The overall mean CPUE was 4.2, the median was 3.8, and the IQR was 3.0-4.9 fish/100 m/24 h (Figure 5.2-6). The annual mean CPUE fell within the overall IQR except in in 2010 and 2011 when it was below the IQR and 2018 and 2019 when it was above the IQR.

ROTATIONAL SITES

Eaglenest Lake

Standard Gang Index Gill Nets

The annual mean CPUE over the four years of monitoring was generally similar among years, with the mean ranging from a low of 35.5 in 2010 to a high of 45.0 fish/100 m/24 h in 2019 (Table 5.2-1; Figure 5.2-1).

The overall mean CPUE was 42.0, the median was 43.7, and the IQR was 41.1-44.6 fish/100 m/24 h (Figure 5.2-1). The annual mean CPUE fell within the overall IQR except in 2010 when it was below the IQR and in 2019 when it was above the IQR.

Small Mesh Index Gill Nets

The annual mean CPUE over the four years of monitoring ranged from a low of 38.2 in 2016 to a high of 109.1 fish/30 m/24 h in 2019 (Table 5.2-1; Figure 5.2-2).

The overall mean CPUE was 71.0, the median was 68.3, and the IQR was 50.8-88.4 fish/30 m/24 h (Figure 5.2-2). The annual mean CPUE fell within the overall IQR except in 2016 when it was below the IQR and in 2019 when it was above the IQR.

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

Catches of Northern Pike were relatively low in Eaglenest Lake over the four years of monitoring, with the annual mean ranging from a low of 2.5 in 2010 to a high of 4.3 fish/100 m/24 in 2019 (Table 5.2-1; Figure 5.2-3).

The overall mean CPUE was 3.6, the median was 3.8, and the IQR was 3.3-4.1 fish/100 m/24 (Figure 5.2-3). The annual mean CPUE fell within the overall IQR except in 2010 when it was below the IQR and in 2019 when it was above the IQR.

Sauger

The annual mean CPUE over the four years of monitoring was generally similar among years, with the mean ranging from a low of 4.2 in 2016 to a high of 4.8 fish/100 m/24 h in 2010 (Table 5.2-1; Figure 5.2-4).

The overall mean CPUE was 4.5, the median was 4.6, and the IQR was 4.5-4.6 fish/100 m/24 h (Figure 5.2-4). The annual mean CPUE fell within the overall IQR except in 2016 when it was below the IQR and in 2010 when it was above the IQR.

Walleye

The annual mean CPUE over the 12 years of monitoring varied by up to about two-fold, with the mean ranging from a low of 4.7 in 2016 to a high of 8.0 fish/100 m/24 h in 2019 (Table 5.2-1; Figure 5.2-5).

The overall mean CPUE was 6.9, the median was 7.5, and the IQR was 6.7-7.7 fish/100 m/24 h (Figure 5.2-5). The annual mean CPUE fell within the overall IQR except in 2016 when it was below the IQR and in 2019 when it was above the IQR.

White Sucker

The annual mean CPUE over the four years of monitoring was generally similar among years, with the mean ranging from a low of 5.0 in 2013 to a high of 6.2 fish/100 m/24 h in 2010 (Table 5.2-1; Figure 5.2-6).

The overall mean CPUE was 5.7, the median was 5.8, and the IQR was 5.6-5.9 fish/100 m/24 h (Figure 5.2-6). The annual mean CPUE fell within the overall IQR except in 2013 when it was below the IQR and in 2010 when it was above the IQR.

Table 5.2-1. 2008-2019 Catch-per-unit-effort.

| Waterbody | Year | Small Mesh Catch ¹ | | | | Total Catch ² | | | | NRPK | | | SAUG | | | WALL | | | WHSC | | |
|-----------|------|-------------------------------|-----------------------------|-------|-----------------|--------------------------|----------------|------|------|----------------|------|-----|----------------|------|-----|----------------|------|-----|----------------|------|------|
| | | n _S ³ | n _F ⁴ | Mean | SE ⁵ | n _S | n _F | Mean | SE | n _F | Mean | SE | n _F | Mean | SE | n _F | Mean | SE | n _F | Mean | SE |
| PDB | 2008 | 3 | 124 | 46.6 | 18.9 | 9 | 413 | 39.8 | 5.8 | 20 | 1.9 | 0.7 | 101 | 9.4 | 2.6 | 62 | 6.2 | 1.5 | 157 | 15.3 | 3.7 |
| | 2009 | 3 | 33 | 10.6 | 3.1 | 7 | 411 | 49.7 | 14.3 | 12 | 1.5 | 0.4 | 87 | 10.4 | 3.3 | 55 | 6.7 | 1.6 | 161 | 19.3 | 10.3 |
| | 2010 | 4 | 100 | 22.8 | 5.9 | 10 | 265 | 21.2 | 2.7 | 13 | 1.1 | 0.3 | 35 | 2.8 | 0.5 | 35 | 2.8 | 0.7 | 100 | 8.1 | 1.7 |
| | 2011 | 4 | 568 | 135.2 | 55.1 | 10 | 613 | 50.4 | 8.7 | 18 | 1.4 | 0.4 | 61 | 4.9 | 1.7 | 48 | 3.9 | 1.1 | 194 | 15.8 | 3.9 |
| | 2012 | 4 | 617 | 144.6 | 39.0 | 10 | 401 | 33.2 | 2.6 | 9 | 0.7 | 0.2 | 106 | 8.8 | 1.3 | 43 | 3.5 | 0.9 | 108 | 8.9 | 1.1 |
| | 2013 | 4 | 573 | 121.6 | 47.6 | 10 | 631 | 49.0 | 7.1 | 42 | 3.3 | 0.9 | 111 | 8.6 | 2.5 | 61 | 4.6 | 1.6 | 114 | 8.8 | 2.4 |
| | 2014 | 3 | 236 | 74.2 | 40.5 | 9 | 286 | 28.9 | 5.3 | 22 | 2.1 | 0.7 | 80 | 8.1 | 1.9 | 30 | 3.1 | 0.9 | 62 | 6.4 | 2.3 |
| | 2015 | 4 | 242 | 56.8 | 22.9 | 10 | 415 | 34.6 | 3.4 | 15 | 1.3 | 0.4 | 88 | 7.4 | 1.8 | 46 | 3.9 | 0.8 | 97 | 8.0 | 1.6 |
| | 2016 | 4 | 183 | 41.1 | 10.0 | 10 | 415 | 33.7 | 5.4 | 21 | 1.8 | 0.5 | 66 | 5.1 | 1.5 | 31 | 2.6 | 0.5 | 71 | 5.6 | 1.1 |
| | 2017 | 4 | 352 | 80.0 | 29.8 | 10 | 413 | 33.8 | 4.2 | 15 | 1.2 | 0.4 | 83 | 6.8 | 1.2 | 42 | 3.4 | 0.8 | 110 | 9.0 | 1.7 |
| | 2018 | 4 | 314 | 73.0 | 21.7 | 10 | 461 | 37.8 | 3.3 | 18 | 1.4 | 0.5 | 62 | 5.0 | 0.9 | 35 | 2.9 | 0.8 | 98 | 8.0 | 1.6 |
| 2019 | 4 | 493 | 111.8 | 40.7 | 10 | 612 | 52.1 | 6.6 | 21 | 1.7 | 0.5 | 104 | 9.0 | 2.6 | 70 | 6.1 | 1.9 | 122 | 10.5 | 1.7 | |
| LDB | 2008 | 3 | 133 | 54.3 | 24.4 | 10 | 305 | 32.4 | 6.4 | 22 | 2.3 | 0.5 | 71 | 7.6 | 1.7 | 45 | 4.8 | 1.3 | 49 | 5.2 | 1.4 |
| | 2009 | 3 | 102 | 40.6 | 14.7 | 10 | 349 | 37.4 | 4.8 | 11 | 1.2 | 0.2 | 116 | 12.7 | 1.8 | 55 | 5.9 | 0.9 | 34 | 3.6 | 0.8 |
| | 2010 | 3 | 110 | 44.5 | 25.0 | 10 | 311 | 32.7 | 5.8 | 38 | 3.9 | 0.6 | 65 | 6.4 | 1.1 | 52 | 5.3 | 1.2 | 35 | 3.8 | 1.2 |
| | 2011 | 3 | 1131 | 388.3 | 151.2 | 10 | 412 | 42.2 | 7.5 | 32 | 3.1 | 0.7 | 78 | 8.2 | 2.0 | 65 | 6.5 | 1.2 | 48 | 4.9 | 1.0 |
| | 2012 | 3 | 553 | 194.3 | 153.2 | 10 | 488 | 50.1 | 4.4 | 58 | 5.9 | 1.0 | 86 | 8.9 | 1.7 | 122 | 12.4 | 2.2 | 31 | 3.1 | 0.7 |
| | 2013 | 3 | 1235 | 424.6 | 314.0 | 10 | 477 | 48.2 | 5.0 | 50 | 4.9 | 0.9 | 95 | 9.6 | 1.5 | 59 | 6.1 | 1.2 | 54 | 5.4 | 1.2 |
| | 2014 | 3 | 187 | 65.0 | 37.6 | 10 | 418 | 43.2 | 4.9 | 52 | 5.3 | 1.1 | 73 | 7.6 | 1.7 | 96 | 9.7 | 2.4 | 36 | 3.8 | 0.8 |
| | 2015 | 3 | 744 | 284.3 | 170.1 | 10 | 361 | 36.7 | 4.4 | 37 | 3.7 | 0.8 | 72 | 7.3 | 1.1 | 77 | 7.8 | 2.7 | 31 | 3.2 | 0.8 |
| | 2016 | 3 | 68 | 25.4 | 3.8 | 10 | 413 | 42.4 | 5.7 | 35 | 3.5 | 0.6 | 113 | 11.4 | 2.3 | 57 | 5.7 | 1.1 | 36 | 3.7 | 0.7 |
| | 2017 | 3 | 311 | 110.9 | 68.5 | 8 | 320 | 37.7 | 3.5 | 36 | 4.2 | 0.8 | 80 | 9.4 | 2.1 | 54 | 6.4 | 2.0 | 32 | 3.8 | 0.6 |
| | 2018 | 3 | 497 | 178.8 | 61.5 | 10 | 443 | 44.4 | 8.1 | 35 | 3.5 | 0.6 | 96 | 9.6 | 2.5 | 64 | 6.3 | 1.2 | 45 | 4.7 | 1.2 |
| 2019 | 3 | 525 | 183.9 | 113.8 | 10 | 531 | 51.1 | 7.4 | 45 | 4.3 | 0.6 | 72 | 6.9 | 1.3 | 81 | 7.8 | 2.2 | 69 | 6.7 | 1.9 | |
| PFF | 2011 | 3 | 10 | 3.5 | 2.2 | 9 | 194 | 18.4 | 2.7 | 5 | 0.5 | 0.2 | 10 | 0.9 | 0.3 | 22 | 2.0 | 0.7 | 16 | 1.5 | 0.3 |
| | 2014 | 3 | 20 | 8.3 | 1.7 | 9 | 312 | 35.3 | 11.8 | 6 | 0.7 | 0.4 | 21 | 2.3 | 0.9 | 49 | 5.8 | 1.5 | 24 | 2.9 | 0.8 |
| | 2017 | 3 | 27 | 9.0 | 4.1 | 9 | 256 | 24.6 | 4.4 | 11 | 1.1 | 0.4 | 26 | 2.5 | 0.8 | 28 | 2.7 | 0.5 | 28 | 2.7 | 0.5 |

Table 5.2-1. continued.

| Waterbody | Year | Small Mesh Catch ¹ | | | | Total Catch ² | | | | NRPK ² | | | SAUG ² | | | WALL ² | | | WHSC ² | | |
|-----------|------|-------------------------------|-----------------------------|-------|-----------------|--------------------------|----------------|------|------|-------------------|------|-----|-------------------|------|-----|-------------------|------|------|-------------------|------|-----|
| | | n _S ³ | n _F ⁴ | Mean | SE ⁵ | n _S | n _F | Mean | SE | n _F | Mean | SE | n _F | Mean | SE | n _F | Mean | SE | n _F | Mean | SE |
| MANIG | 2008 | 2 | 13 | 8.4 | 5.9 | 6 | 343 | 68.0 | 8.0 | 13 | 2.4 | 1.1 | - | - | - | 130 | 23.5 | 9.8 | 26 | 4.9 | 2.2 |
| | 2009 | 2 | 5 | 3.0 | 1.1 | 6 | 281 | 50.3 | 6.1 | 13 | 2.0 | 1.3 | - | - | - | 96 | 16.2 | 6.6 | 20 | 3.4 | 1.7 |
| | 2010 | 2 | 40 | 24.4 | 12.2 | 5 | 271 | 59.5 | 19.4 | 11 | 2.1 | 1.1 | - | - | - | 114 | 23.1 | 10.2 | 10 | 2.1 | 0.8 |
| | 2011 | 2 | 5 | 2.6 | 0.4 | 6 | 289 | 46.5 | 5.6 | 19 | 2.9 | 0.9 | - | - | - | 97 | 15.6 | 5.9 | 11 | 1.7 | 0.7 |
| | 2012 | 2 | 142 | 71.4 | 50.5 | 5 | 290 | 51.2 | 5.2 | 22 | 3.8 | 1.7 | - | - | - | 105 | 18.6 | 7.3 | 24 | 4.3 | 2.0 |
| | 2013 | 2 | 56 | 27.5 | 18.7 | 6 | 294 | 43.0 | 7.8 | 23 | 3.3 | 1.0 | - | - | - | 77 | 11.0 | 3.3 | 22 | 3.1 | 1.1 |
| | 2014 | 2 | 25 | 12.6 | 2.3 | 6 | 366 | 53.7 | 6.2 | 27 | 3.9 | 1.5 | - | - | - | 106 | 15.3 | 6.0 | 21 | 3.0 | 1.4 |
| | 2015 | 2 | 39 | 19.3 | 12.9 | 6 | 437 | 65.2 | 11.0 | 18 | 2.6 | 1.0 | - | - | - | 97 | 14.1 | 5.0 | 21 | 3.0 | 1.4 |
| | 2016 | 2 | 16 | 8.1 | 3.6 | 6 | 437 | 63.9 | 4.7 | 30 | 4.4 | 1.6 | - | - | - | 104 | 15.5 | 6.0 | 33 | 4.9 | 2.0 |
| | 2017 | 2 | 10 | 4.6 | 3.3 | 6 | 387 | 53.6 | 11.0 | 27 | 3.8 | 1.4 | - | - | - | 117 | 16.3 | 5.2 | 31 | 4.4 | 1.9 |
| | 2018 | 2 | 102 | 45.5 | 32.2 | 6 | 479 | 66.7 | 9.5 | 29 | 3.9 | 1.4 | - | - | - | 139 | 19.2 | 5.3 | 58 | 8.0 | 3.1 |
| 2019 | 2 | 25 | 12.8 | 9.0 | 6 | 308 | 45.3 | 9.0 | 15 | 2.2 | 1.0 | - | - | - | 118 | 17.0 | 5.0 | 52 | 7.4 | 2.8 | |
| EAGLE | 2010 | 3 | 185 | 55.0 | 22.8 | 11 | 483 | 35.5 | 3.9 | 34 | 2.5 | 0.4 | 65 | 4.8 | 0.9 | 102 | 7.6 | 1.8 | 84 | 6.2 | 1.3 |
| | 2013 | 4 | 331 | 81.6 | 3.5 | 11 | 579 | 44.4 | 3.3 | 52 | 4.0 | 0.5 | 60 | 4.6 | 1.0 | 95 | 7.3 | 1.2 | 65 | 5.0 | 1.3 |
| | 2016 | 4 | 156 | 38.2 | 15.4 | 12 | 593 | 43.0 | 5.0 | 51 | 3.6 | 0.7 | 58 | 4.2 | 1.1 | 64 | 4.7 | 0.6 | 79 | 5.9 | 1.7 |
| | 2019 | 4 | 450 | 109.1 | 35.1 | 12 | 640 | 45.0 | 4.2 | 61 | 4.3 | 0.9 | 65 | 4.6 | 1.2 | 114 | 8.0 | 1.6 | 83 | 5.8 | 0.9 |

Notes:

1. fish/30 m/24 h.
2. fish/100 m/24 h.
3. n_S = number of sites fished (excludes sets > 36 h).
4. n_F = number of fish caught.
5. SE = standard error.

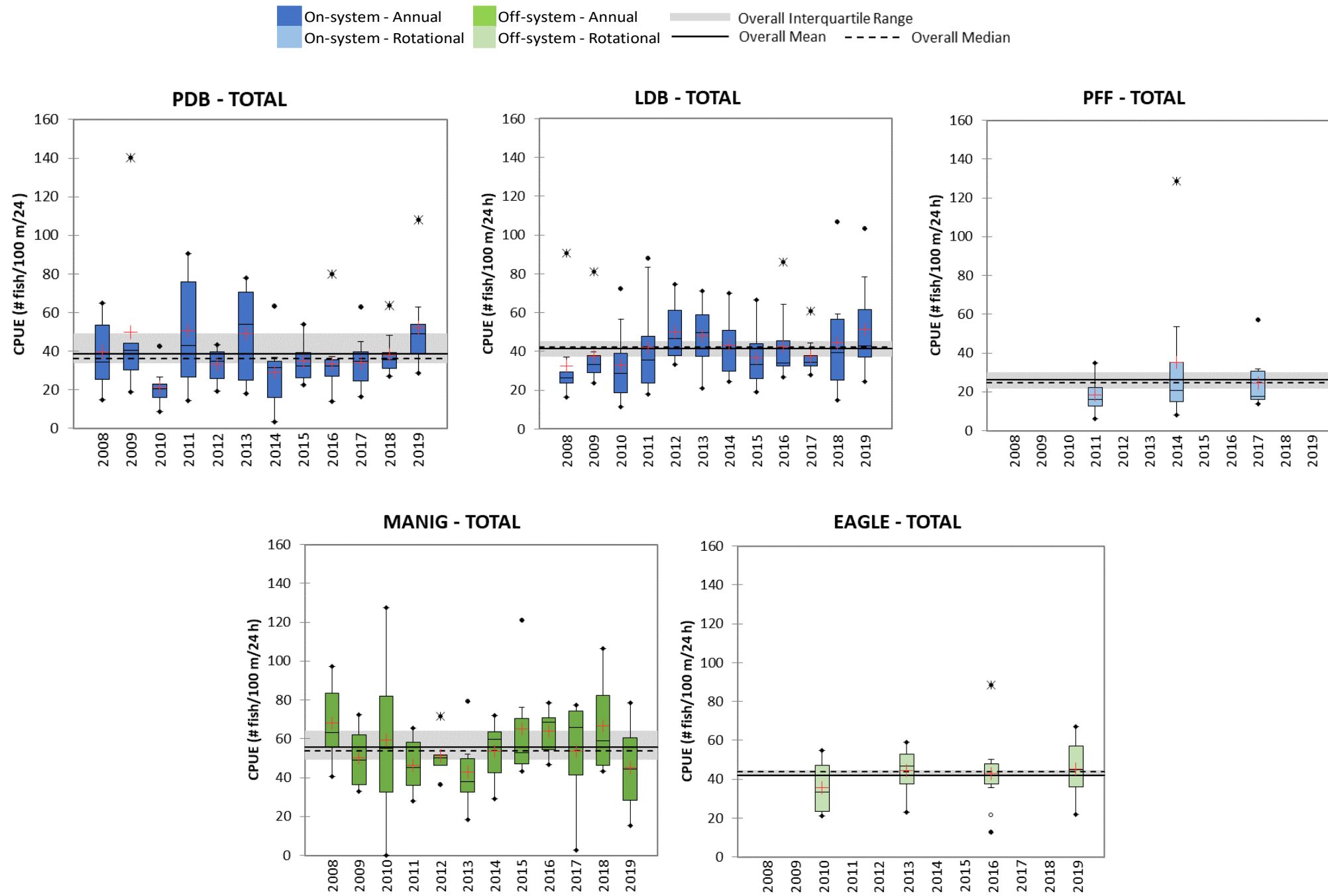


Figure 5.2-1. 2008-2019 Catch-per-unit-effort (CPUE) of standard gang index gill nets.

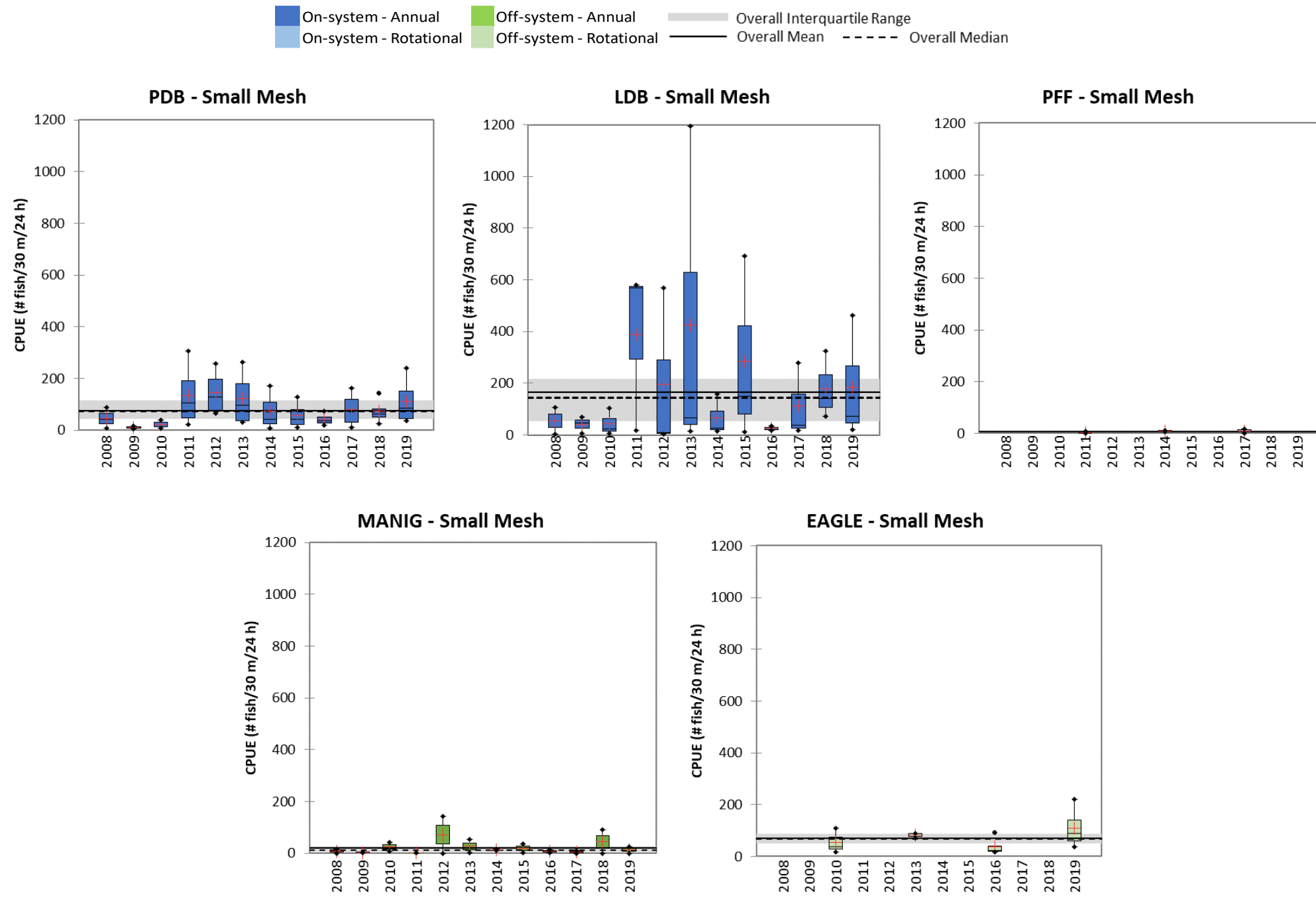


Figure 5.2-2. 2008-2019 Catch-per-unit-effort (CPUE) of small mesh index gill nets.

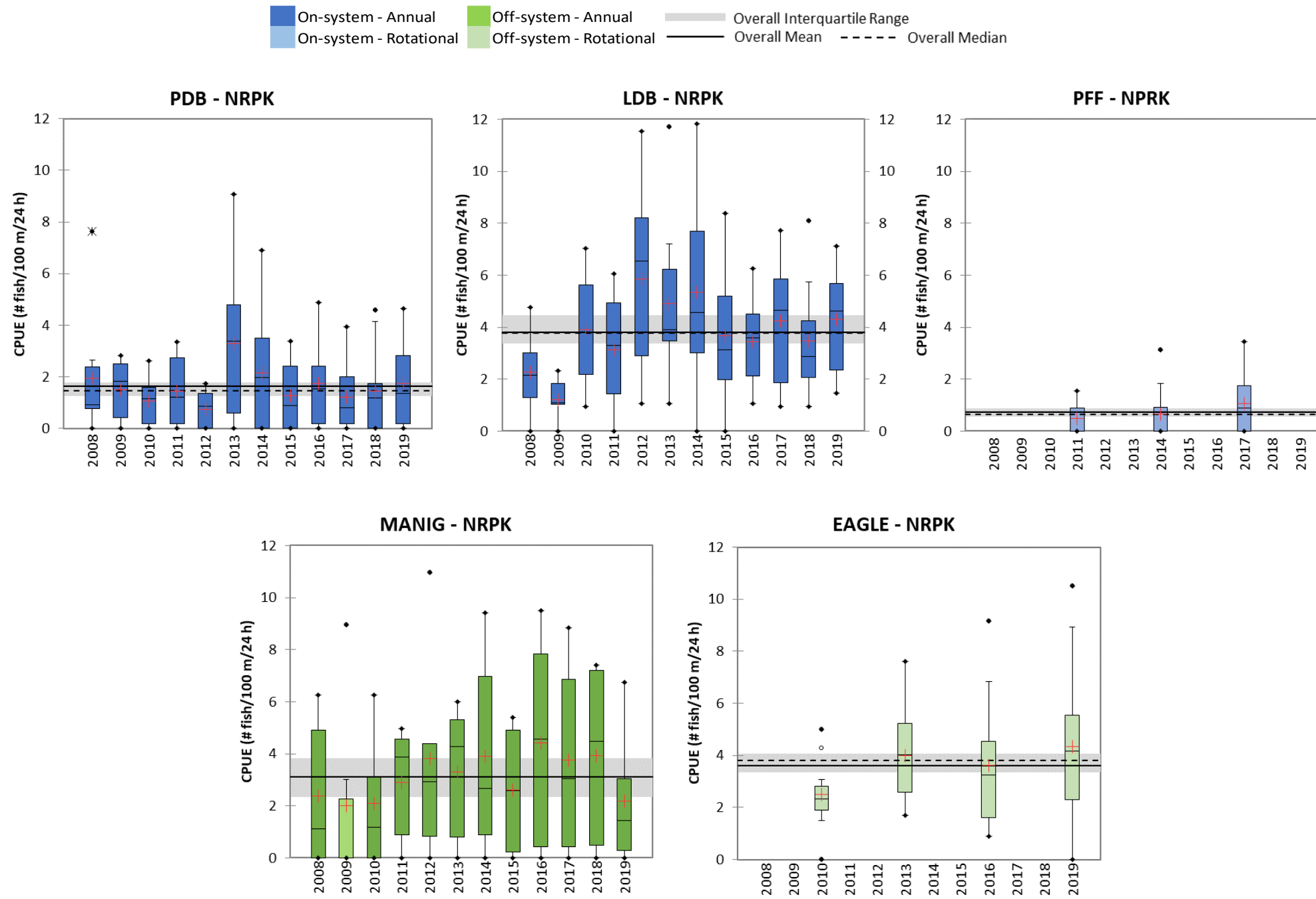


Figure 5.2-3. 2008-2019 Catch-per-unit-effort (CPUE) of Northern Pike.

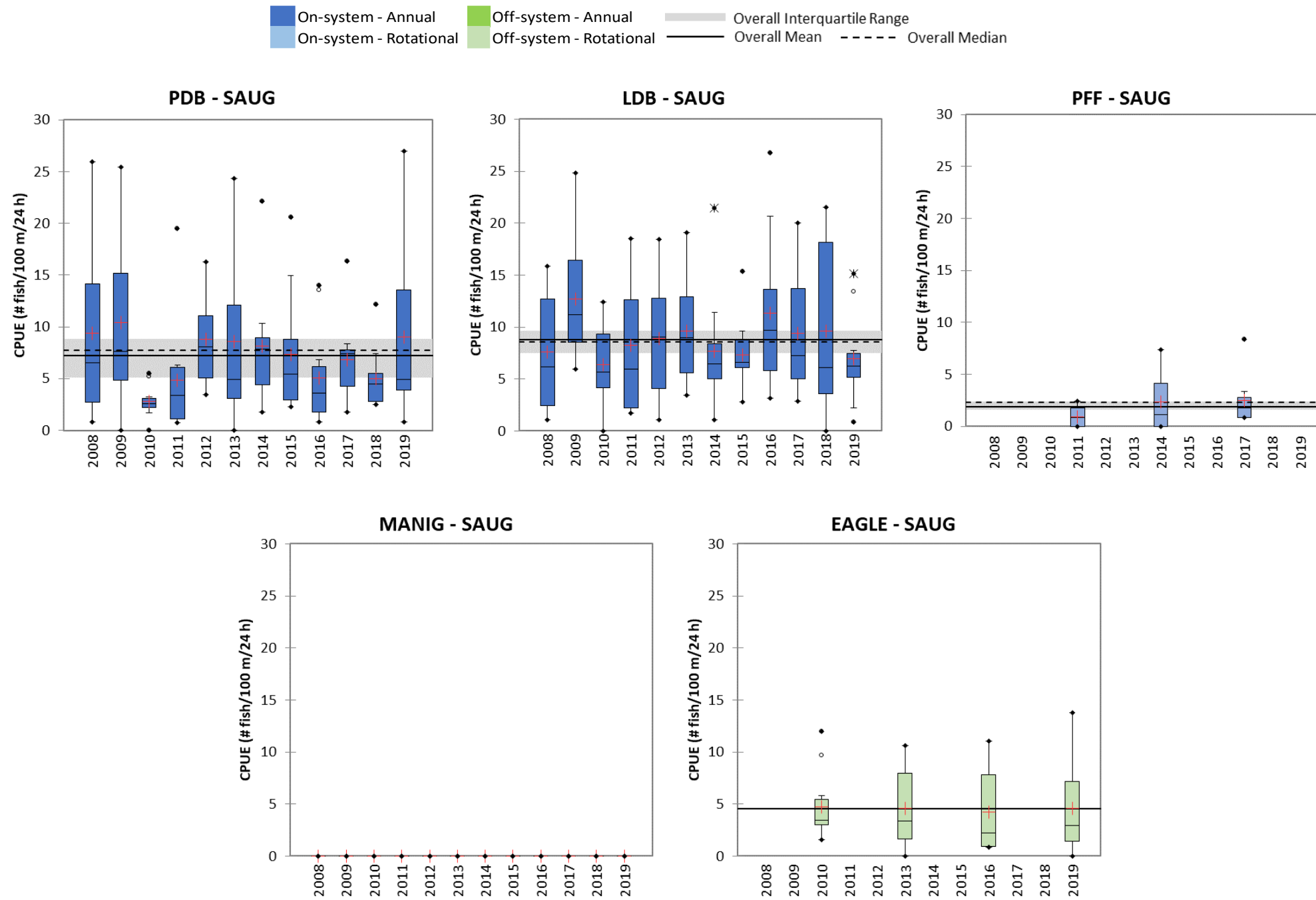


Figure 5.2-4. 2008-2019 Catch-per-unit-effort (CPUE) of Sauger.

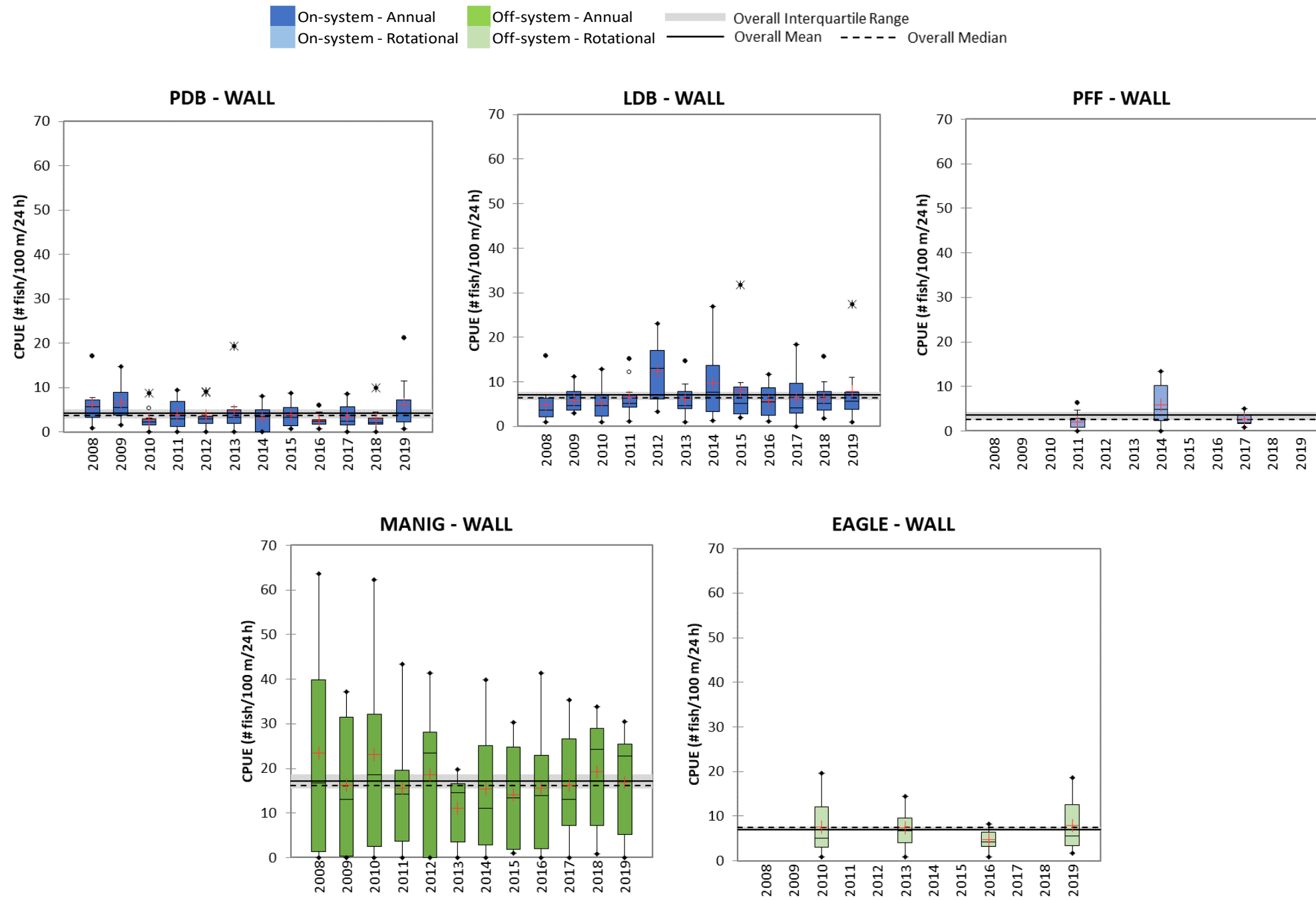


Figure 5.2-5. 2008-2019 Catch-per-unit-effort (CPUE) of Walleye.

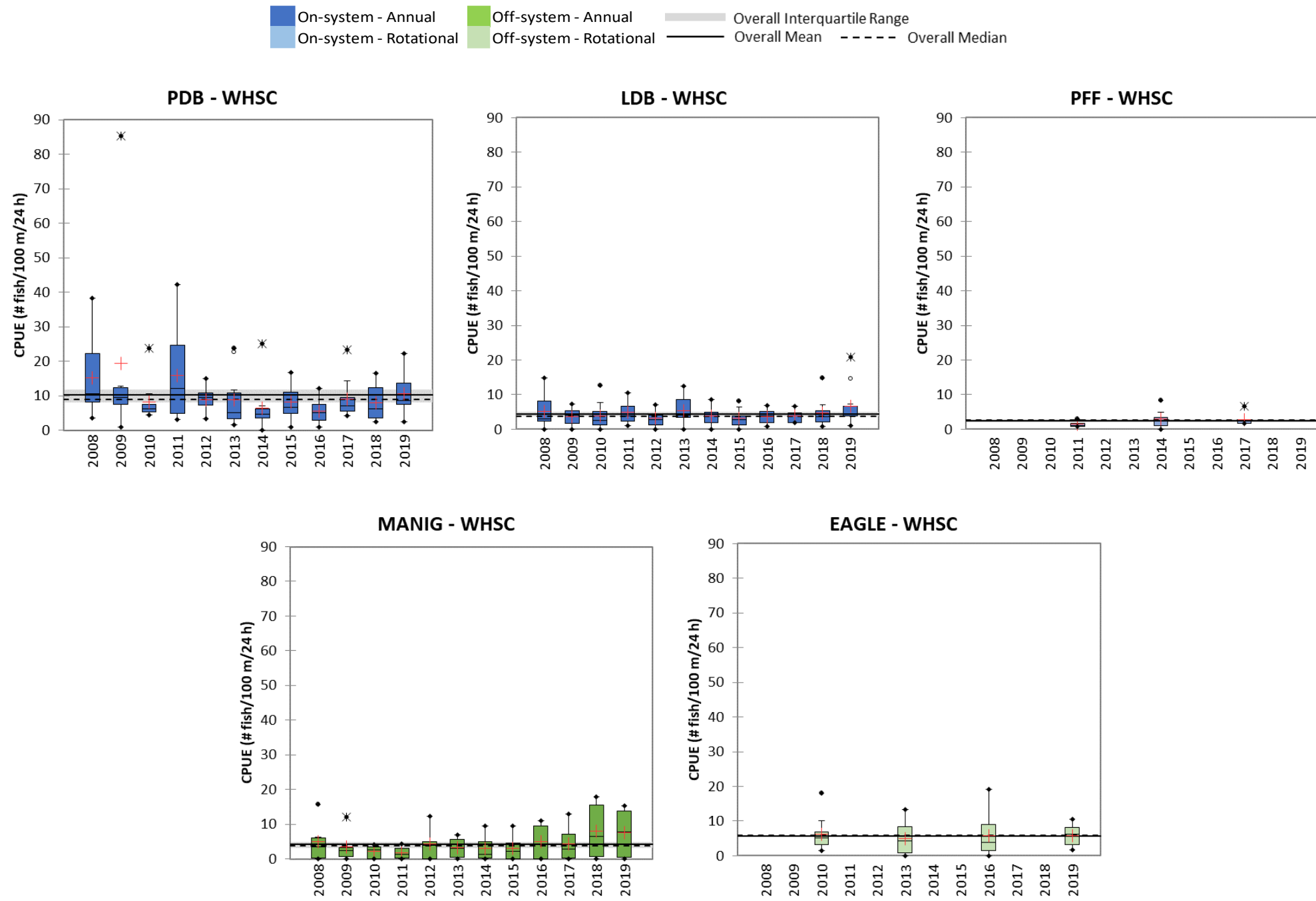


Figure 5.2-6. 2008-2019 Catch-per-unit-effort (CPUE) of White Sucker.

5.3 CONDITION

5.3.1 FULTON'S CONDITION FACTOR

5.3.1.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

The annual mean KF of Northern Pike between 400 and 699 mm in fork length over the 12 years of monitoring ranged from a low of 0.60 in 2010 to a high of 0.71 in 2016 (Table 5.3-1; Figure 5.3-1).

The overall mean KF was 0.67, the median was 0.68, and the IQR was 0.65-0.69 (Figure 5.3-1). The annual mean KF fell within the overall IQR except in 2009, 2010, and 2011 when it was below the IQR and in 2016 when it was above the IQR.

Sauger

The annual mean KF of Sauger between 200 and 349 mm in fork length over the 12 years of monitoring ranged from a low of 0.87 in 2009 to a high of 1.05 in 2011 (Table 5.3-1; Figure 5.3-2).

The overall mean and median KF were 0.95 and the IQR was 0.89-0.99 (Figure 5.3-2). The annual mean KF fell within the overall IQR except in 2009 and 2018 when it was below the IQR and in 2008, 2011, and 2012 when it was above the IQR.

Walleye

The annual mean KF of Walleye between 300 and 499 mm in fork length over the 12 years of monitoring ranged from a low of 1.09 in 2010 to a high of 1.20 in 2008 (Table 5.3-1; Figure 5.3-3).

The overall mean and median KF were 1.14 and the IQR was 1.11-1.16 (Figure 5.3-3). The annual mean KF fell within the overall IQR except in 2009, 2010, and 2018 when it was below the IQR and in 2008, 2011, and 2016 when it was above the IQR.

White Sucker

The annual mean KF of White Sucker between 300 and 499 mm in fork length over the 10 years of monitoring that it was a target species ranged from a low of 1.46 in 2010 to a high of 1.63 in 2011 and 2012 (Table 5.3-1; Figure 5.3-4).

The overall mean KF was 1.59, the median was 1.60, and the IQR was 1.59-1.63 (Figure 5.3-4). The annual mean KF fell within the overall IQR except in 2010 and 2018 when it was below the IQR.

Lac du Bonnet

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

The annual mean KF of Northern Pike between 400 and 699 mm in fork length over the 12 years of monitoring ranged from a low of 0.64 in 2009 to a high of 0.75 in 2011 (Table 5.3-1; Figure 5.3-1).

The overall mean and median KF were 0.70 and the IQR was 0.67-0.73 (Figure 5.3-1). The annual mean KF fell within the overall IQR except in 2008, 2009, and 2017 when it was below the IQR and in 2011 and 2012 when it was above the IQR.

Sauger

The annual mean KF of Sauger between 200 and 349 mm in fork length over the 12 years of monitoring ranged from a low of 0.90 in 2015 to a high of 1.00 in 2012 (Table 5.3-1; Figure 5.3-2).

The overall mean KF was 0.94, the median was 0.93, and the IQR was 0.91-0.94 (Figure 5.3-2). The annual mean KF fell within the overall IQR except in 2015 when it was below the IQR and in 2012 and 2014 when it was above the IQR.

Walleye

The annual mean KF of Walleye between 300 and 499 mm in fork length over the 12 years of monitoring ranged from a low of 1.10 in 2016 to a high of 1.17 in 2013 (Table 5.3-1; Figure 5.3-3).

The overall mean and median KF were 1.14 and the IQR was 1.12-1.16 (Figure 5.3-3). The annual mean KF fell within the overall IQR except in 2008, 2015, and 2016 when it was below the IQR and in 2013 when it was above the IQR.

White Sucker

Individual White Sucker were only measured for length and weight sporadically at Lac du Bonnet and KF is only available for 2010, 2016, 2018, and 2019 (Table 5.3-1). The annual mean KF of White Sucker between 300 and 499 mm in fork length ranged from a low of 1.59 in 2010 and 2018 to a high of 1.65 in 2019 (Table 5.3-1; Figure 5.3-4).

The overall mean KF was 1.62, the median was 1.63, and the IQR was 1.59-1.65 (Figure 5.3-4). The annual mean KF fell within the overall IQR except in 2008 when it was above the IQR.

ROTATIONAL SITES

Pine Falls Forebay

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

The annual mean KF of Northern Pike between 400 and 699 mm in fork length over the three years of monitoring ranged from a low of 0.64 in 2014 to a high of 0.66 in 2017 (Table 5.3-1; Figure 5.3-1).

The overall mean and median KF were 0.65 and the IQR was 0.64-0.66 (Figure 5.3-1). The annual mean KF was equal to or fell within the overall IQR in all three years.

Sauger

Sauger between 200 and 349 mm in fork length had an annual mean KF of 0.97 in 2017 (Table 5.3-1; Figure 5.3-2). Sauger was not a target species in the Pine Falls forebay prior to 2017.

Walleye

The annual mean KF of Walleye between 300 and 499 mm in fork length over the three years of monitoring ranged from a low of 1.01 in 2014 to a high of 1.13 in 2017 (Table 5.3-1; Figure 5.3-3).

The overall mean KF was 1.07, the median was 1.12, and the IQR was 1.01-1.12 (Figure 5.3-3). The annual mean KF fell within or was equal to the overall IQR except in 2017 when it was above the IQR.

White Sucker

The annual mean KF of White Sucker between 300 and 499 mm in fork length over the three years of monitoring was 1.48 in 2011 and 2014 and 1.59 in 2017 (Table 5.3-1; Figure 5.3-4).

The overall mean KF was 1.52, the median was 1.48, and the IQR was 1.48-1.59 (Figure 5.3-4). The annual mean KF was equal to or fell within the overall IQR in all three years.

5.3.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

The annual mean KF of Northern Pike between 400 and 699 mm in fork length over the 12 years of monitoring ranged from a low of 0.55 in 2010 to a high of 0.72 in 2019 (Table 5.3-1; Figure 5.3-1).

The overall mean KF was 0.67, the median was 0.68, and the IQR was 0.66-0.69 (Figure 5.3-1). The annual mean KF fell within the overall IQR except in 2008, 2010, and 2011 when it was below the IQR and in 2017 and 2019 when it was above the IQR.

Sauger

Sauger were not captured in Manigotagan Lake over the 12 years of monitoring (Table 5.3-1).

Walleye

The annual mean KF of Walleye between 300 and 499 mm in fork length over the 12 years of monitoring ranged from a low of 1.02 in 2014 to a high of 1.10 in 2012 (Table 5.3-1; Figure 5.3-3).

The overall mean and median KF were 1.07 and the IQR was 1.07-1.08 (Figure 5.3-3). The annual mean KF fell within the overall IQR except in 2011, 2013, and 2014 when it was below the IQR and in 2012 when it was above the IQR.

White Sucker

The annual mean KF of White Sucker between 300 and 499 mm in fork length over the 10 years of monitoring that it was a target species ranged from a low of 1.47 in 2010 to a high of 1.65 in 2015 (Table 5.3-1; Figure 5.3-4).

The overall mean and median KF were 1.57 and the IQR was 1.57-1.62 (Figure 5.3-4). The annual mean KF fell within the overall IQR except in 2010, 2011, 2013, and 2014 when it was below the IQR and in 2015 when it was above the IQR.

ROTATIONAL SITES

Eaglenest Lake

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

The annual mean KF of Northern Pike between 400 and 699 mm in fork length over the four years of monitoring ranged from a low of 0.68 in 2019 to a high of 0.70 in 2013 (Table 5.3-1; Figure 5.3-1).

The overall mean and median KF were 0.69 and the IQR was 0.68-0.70 (Figure 5.3-1). The annual mean KF was equal to or fell within the overall IQR in all three years.

Sauger

The annual mean KF of Sauger between 200 and 349 mm in fork length was 0.97 in 2016 and 0.90 in 2019 (Table 5.3-1; Figure 5.3-2). Sauger was not a target species in Eaglenest Lake prior to 2016.

Walleye

The annual mean KF of Walleye between 300 and 499 mm in fork length over the four years of monitoring ranged from a low of 1.10 in 2010 to a high of 1.17 in 2013 (Table 5.3-1; Figure 5.3-3).

The overall mean KF was 1.13, the median was 1.12, and the IQR was 1.10-1.17 (Figure 5.3-3). The annual mean KF was equal to or fell within the overall IQR in all four years.

White Sucker

The annual mean KF of White Sucker between 300 and 499 mm in fork length over the four years of monitoring ranged from a low of 1.48 in 2010 to a high of 1.72 in 2016 (Table 5.3-1; Figure 5.3-4).

The overall mean KF was 1.61, the median was 1.63, and the IQR was 1.48-1.72 (Figure 5.3-4). The annual mean KF was equal to or fell within the overall IQR in all four years.

Table 5.3-1. 2008-2019 Fulton’s condition factor of target species.

| Waterbody | Year | NRPK | | | SAUG | | | WALL | | | WHSC | | |
|-----------|------|-----------------------------|------|-----------------|----------------|------|------|----------------|------|------|----------------|------|------|
| | | n _F ¹ | Mean | SE ² | n _F | Mean | SE | n _F | Mean | SE | n _F | Mean | SE |
| PDB | 2008 | 12 | 0.66 | 0.03 | 78 | 1.05 | 0.01 | 27 | 1.20 | 0.02 | | | |
| | 2009 | 20 | 0.63 | 0.02 | 110 | 0.87 | 0.01 | 40 | 1.10 | 0.02 | | | |
| | 2010 | 10 | 0.60 | 0.03 | 33 | 0.92 | 0.01 | 19 | 1.09 | 0.02 | 95 | 1.46 | 0.01 |
| | 2011 | 14 | 0.63 | 0.03 | 57 | 1.05 | 0.02 | 23 | 1.17 | 0.02 | 183 | 1.63 | 0.01 |
| | 2012 | 5 | 0.68 | 0.05 | 120 | 1.02 | 0.01 | 21 | 1.16 | 0.02 | 95 | 1.63 | 0.01 |
| | 2013 | 39 | 0.69 | 0.01 | 115 | 0.99 | 0.01 | 33 | 1.14 | 0 | 93 | 1.59 | 0.01 |
| | 2014 | 19 | 0.69 | 0.02 | 90 | 0.90 | 0.01 | 16 | 1.11 | 0 | 58 | 1.59 | 0.02 |
| | 2015 | 14 | 0.68 | 0.01 | 117 | 0.89 | 0.01 | 25 | 1.14 | 0 | 87 | 1.60 | 0.01 |
| | 2016 | 16 | 0.71 | 0.02 | 80 | 0.94 | 0.01 | 22 | 1.18 | 0 | 62 | 1.60 | 0.02 |
| | 2017 | 9 | 0.65 | 0.02 | 87 | 0.97 | 0.01 | 22 | 1.16 | 0 | 89 | 1.60 | 0.01 |
| | 2018 | 9 | 0.68 | 0.03 | 77 | 0.88 | 0.01 | 16 | 1.09 | 0.02 | 78 | 1.55 | 0.02 |
| 2019 | 14 | 0.66 | 0.01 | 109 | 0.95 | 0.01 | 43 | 1.13 | 0.02 | 110 | 1.61 | 0.01 | |
| LDB | 2008 | 10 | 0.66 | 0.02 | 84 | 0.91 | 0.02 | 17 | 1.11 | 0.02 | | | |
| | 2009 | 9 | 0.64 | 0.02 | 133 | 0.93 | 0.01 | 18 | 1.12 | 0.01 | | | |
| | 2010 | 27 | 0.73 | 0.01 | 72 | 0.92 | 0.01 | 23 | 1.15 | 0.02 | 32 | 1.59 | 0.02 |
| | 2011 | 18 | 0.75 | 0.02 | 81 | 0.91 | 0.01 | 24 | 1.14 | 0.02 | | | |
| | 2012 | 35 | 0.74 | 0.01 | 90 | 1.00 | 0.01 | 83 | 1.16 | 0.01 | | | |
| | 2013 | 32 | 0.72 | 0.01 | 102 | 0.94 | 0.01 | 31 | 1.17 | 0.02 | | | |
| | 2014 | 33 | 0.73 | 0.02 | 74 | 1.00 | 0.01 | 48 | 1.15 | 0.01 | | | |
| | 2015 | 27 | 0.68 | 0.02 | 76 | 0.90 | 0.02 | 43 | 1.11 | 0.02 | | | |
| | 2016 | 28 | 0.67 | 0.01 | 64 | 0.91 | 0.01 | 35 | 1.10 | 0.01 | 32 | 1.63 | 0.02 |
| | 2017 | 29 | 0.66 | 0.01 | 80 | 0.93 | 0.01 | 40 | 1.13 | 0.01 | | | |
| | 2018 | 23 | 0.68 | 0.01 | 106 | 0.94 | 0.01 | 40 | 1.12 | 0.01 | 36 | 1.59 | 0.03 |
| 2019 | 32 | 0.70 | 0.01 | 82 | 0.94 | 0.01 | 47 | 1.14 | 0.01 | 59 | 1.65 | 0.02 | |
| PFF | 2011 | 5 | 0.65 | 0.03 | | | | 16 | 1.12 | 0.02 | 15 | 1.48 | 0.04 |
| | 2014 | 7 | 0.64 | 0.03 | | | | 21 | 1.01 | 0.03 | 18 | 1.48 | 0.04 |
| | 2017 | 7 | 0.66 | 0.01 | 26 | 0.97 | 0.02 | 10 | 1.13 | 0.03 | 16 | 1.59 | 0.04 |
| MANIG | 2008 | 11 | 0.59 | 0.02 | - | - | - | 46 | 1.07 | 0.01 | | | |
| | 2009 | 12 | 0.66 | 0.03 | - | - | - | 31 | 1.08 | 0.02 | | | |
| | 2010 | 9 | 0.55 | 0.04 | - | - | - | 95 | 1.07 | 0.01 | 11 | 1.47 | 0.02 |
| | 2011 | 16 | 0.62 | 0.02 | - | - | - | 63 | 1.05 | 0.01 | 11 | 1.50 | 0.03 |
| | 2012 | 22 | 0.68 | 0.01 | - | - | - | 79 | 1.10 | 0.01 | 26 | 1.57 | 0.02 |
| | 2013 | 28 | 0.69 | 0.01 | - | - | - | 38 | 1.03 | 0.01 | 21 | 1.53 | 0.01 |
| | 2014 | 28 | 0.69 | 0.01 | - | - | - | 50 | 1.02 | 0.01 | 20 | 1.54 | 0.02 |
| | 2015 | 19 | 0.66 | 0.02 | - | - | - | 58 | 1.07 | 0.01 | 19 | 1.65 | 0.08 |
| | 2016 | 27 | 0.68 | 0.01 | - | - | - | 62 | 1.07 | 0.01 | 31 | 1.57 | 0.03 |
| | 2017 | 25 | 0.70 | 0.02 | - | - | - | 65 | 1.08 | 0.01 | 28 | 1.60 | 0.02 |
| 2018 | 26 | 0.68 | 0.01 | - | - | - | 74 | 1.07 | 0.01 | 51 | 1.57 | 0.02 | |
| 2019 | 12 | 0.72 | 0.02 | - | - | - | 71 | 1.10 | 0.01 | 48 | 1.62 | 0.02 | |
| EAGLE | 2010 | 17 | 0.70 | 0.02 | | | | 64 | 1.10 | 0.01 | 80 | 1.48 | 0.01 |
| | 2013 | 35 | 0.70 | 0.01 | | | | 62 | 1.17 | 0.01 | 51 | 1.62 | 0.02 |
| | 2016 | 29 | 0.69 | 0.02 | 62 | 0.97 | 0.03 | 35 | 1.15 | 0.03 | 71 | 1.72 | 0.02 |
| | 2019 | 33 | 0.68 | 0.01 | 82 | 0.90 | 0.01 | 43 | 1.12 | 0.01 | 72 | 1.63 | 0.01 |

Notes:

1. n_F = number of fish measured for length and weight.
2. SE = standard error.
3. Grey shading indicates a species was not a target species in that year.

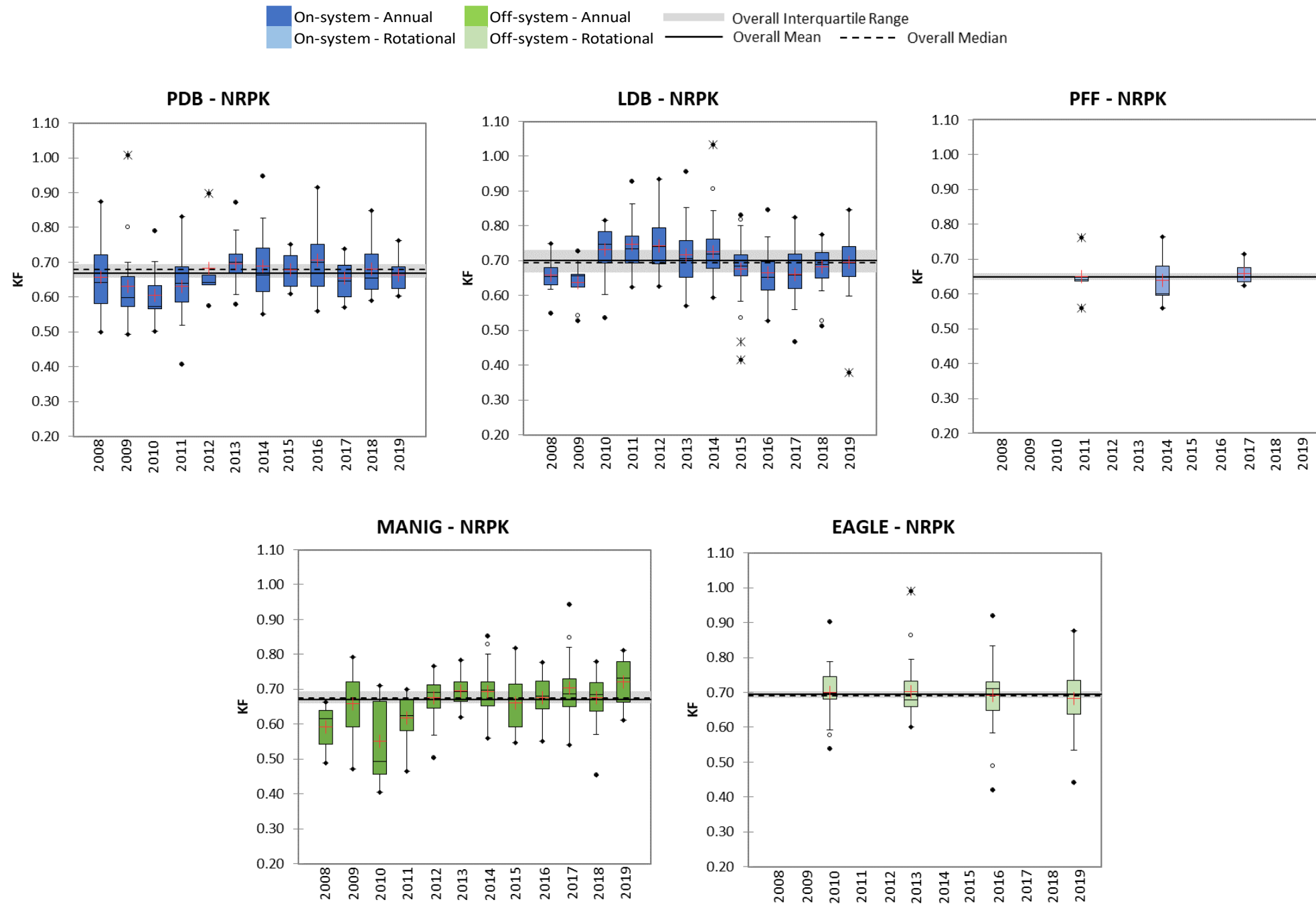


Figure 5.3-1. 2008-2019 Fulton's condition factor (KF) of Northern Pike.

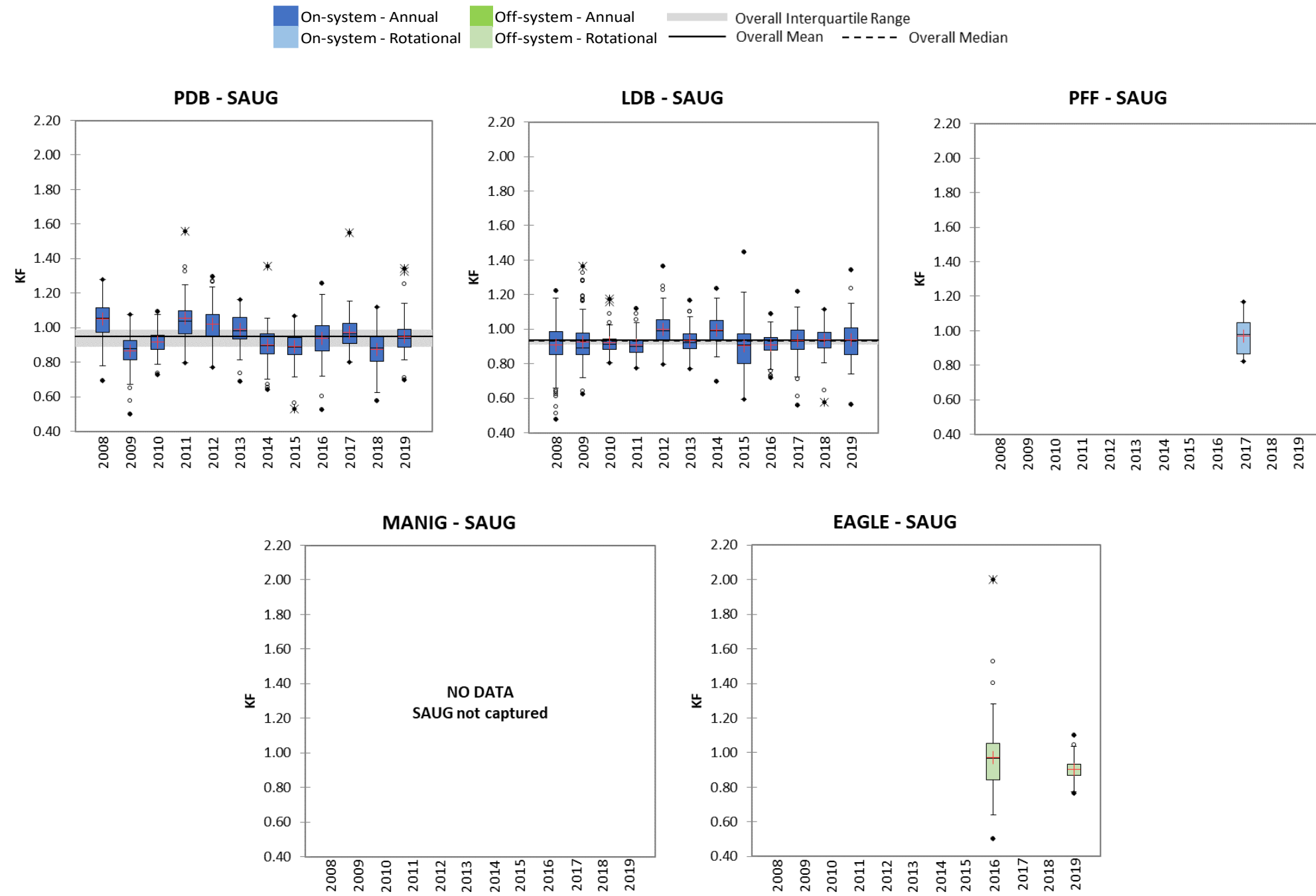


Figure 5.3-2. 2008-2019 Fulton's condition factor (KF) of Sauger.

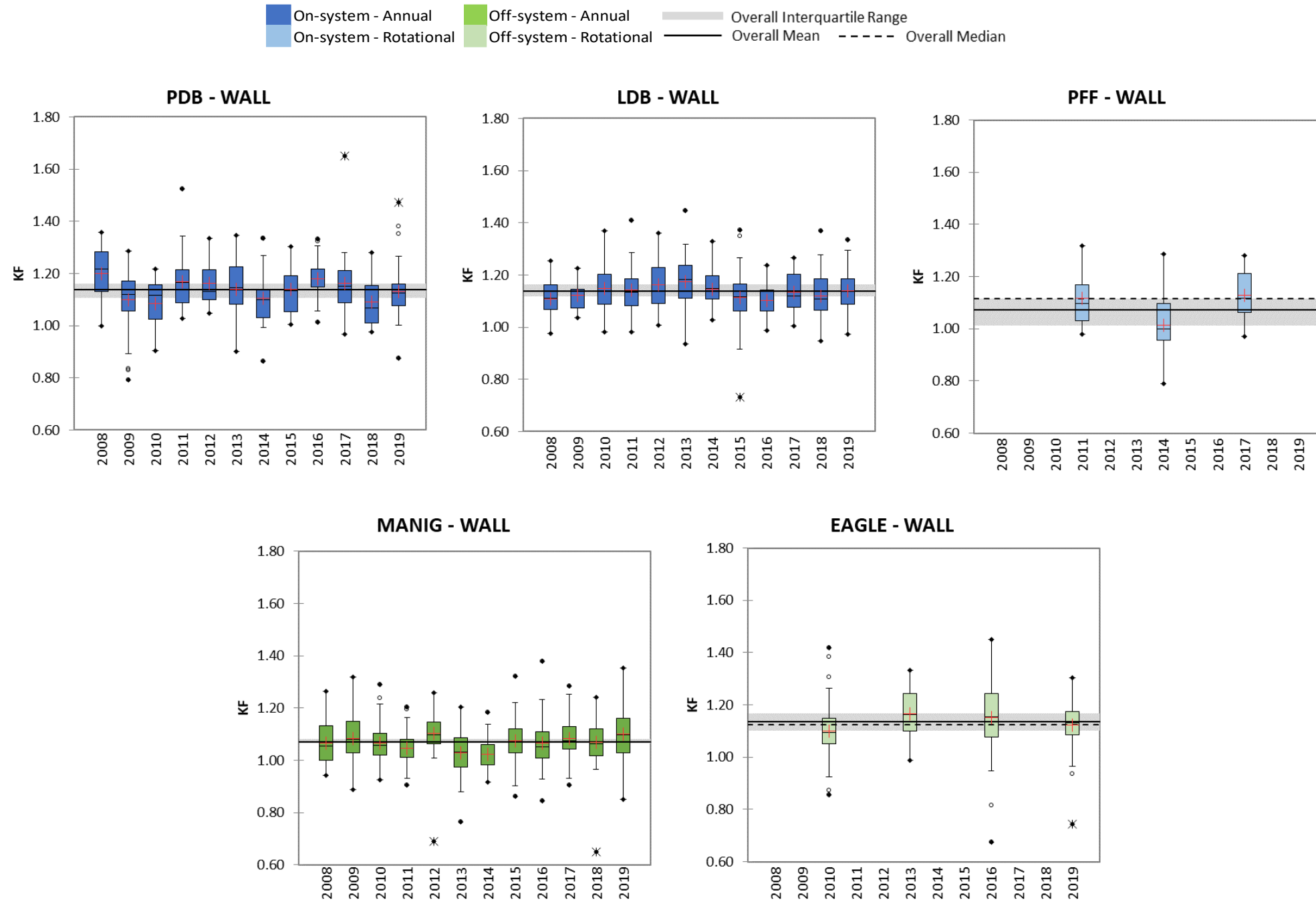


Figure 5.3-3. 2008-2019 Fulton's condition factor (KF) of Walleye.

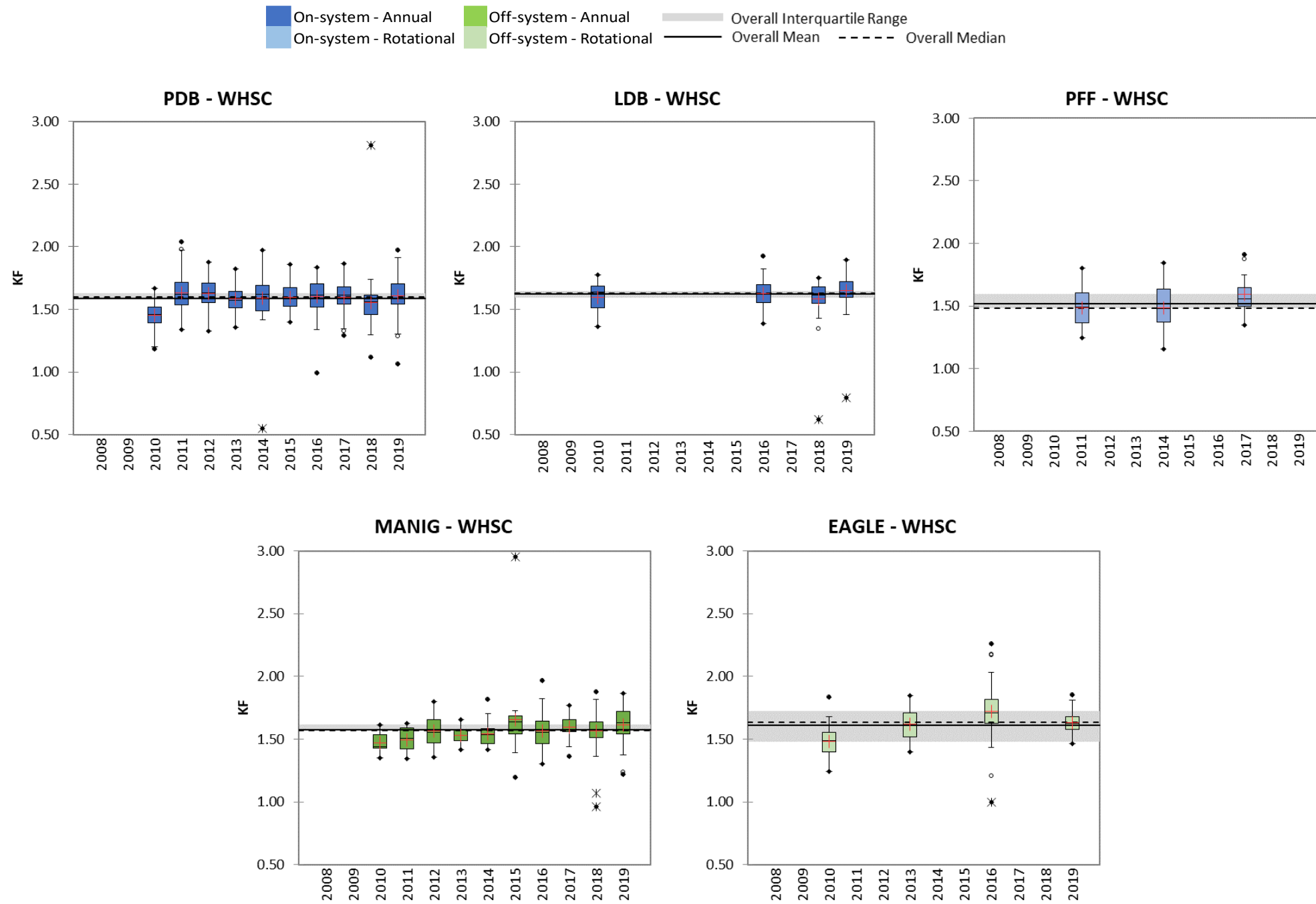


Figure 5.3-4. 2008-2019 Fulton's condition factor (KF) of White Sucker.

5.3.2 RELATIVE WEIGHT

5.3.2.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

The annual mean Wr of Northern Pike greater than 99 mm in total length over the 12 years of monitoring ranged from a low of 77 in 2010 to a high of 89 in 2016 (Table 5.3-2; Figure 5.3-5).

The overall mean Wr was 85, the median was 86, and the IQR was 83-87 (Figure 5.3-5). The annual mean Wr fell within the overall IQR except in 2010 and 2011 when it was below the IQR and in 2016 when it was above the IQR.

Sauger

The annual mean Wr of Sauger greater than 69 mm in total length over the 12 years of monitoring ranged from a low of 81 in 2009 to a high of 104 in 2008 (Table 5.3-2; Figure 5.3-6).

The overall mean Wr was 90, the median was 91, and the IQR was 84-92 (Figure 5.3-6). The annual mean Wr fell within the overall IQR except in 2009 when it was below the IQR and in 2008, 2011, and 2012 when it was above the IQR.

Walleye

The annual mean Wr of Walleye greater than 29 mm in total length over the 12 years of monitoring ranged from a low of 84 in 2010 to a high of 98 in 2008 (Table 5.3-2; Figure 5.3-7).

The overall mean and median Wr were 90 and the IQR was 89-92 (Figure 5.3-7). The annual mean Wr fell within the overall IQR except in 2009, 2010, and 2018 when it was below the IQR and in 2008 and 2016 when it was above the IQR.

White Sucker

The annual mean *Wr* of White Sucker greater than 99 mm in total length over the 10 years of monitoring that it was a target species ranged from a low of 94 in 2010 to a high of 105 in 2011 (Table 5.3-2; Figure 5.3-8).

The overall mean and median *Wr* were 102 and the IQR was 101-104 (Figure 5.3-8). The annual mean *Wr* fell within the overall IQR except in 2010 and 2018 when it was below the IQR and in 2011 when it was above the IQR.

Lac du Bonnet

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

The annual mean *Wr* of Northern Pike greater than 99 mm in total length over the 12 years of monitoring ranged from a low of 80 in 2017 to a high of 91 in 2011 (Table 5.3-2; Figure 5.3-5).

The overall mean *Wr* was 87, the median was 88, and the IQR was 83-89 (Figure 5.3-5). The annual mean *Wr* fell within the overall IQR except in 2009, 2016, and 2017 when it was below the IQR and in 2011 and 2012 when it was above the IQR.

Sauger

The annual mean *Wr* of Sauger greater than 69 mm in total length over the 12 years of monitoring ranged from a low of 86 in 2011 to a high of 95 in 2014 (Table 5.3-2; Figure 5.3-6).

The overall mean *Wr* was 89, the median was 88, and the IQR was 88-89 (Figure 5.3-6). The annual mean *Wr* fell within the overall IQR except in 2008, 2011, and 2015 when it was below the IQR and in 2012, 2014, and 2019 when it was above the IQR.

Walleye

The annual mean *Wr* of Walleye greater than 29 mm in total length over the 12 years of monitoring ranged from a low of 84 in 2018 to a high of 95 in 2019 (Table 5.3-2; Figure 5.3-7).

The overall mean *Wr* was 90, the median was 92, and the IQR was 88-93 (Figure 5.3-7). The annual mean *Wr* fell within the overall IQR except in 2013 and 2018 when it was below the IQR and in 2012 and 2019 when it was above the IQR.

White Sucker

Individual White Sucker were only measured for length and weight sporadically at Lac du Bonnet and Wr is only available for 2010, 2016, 2018, and 2019 (Table 5.3-2). The annual mean Wr of White Sucker greater than 99 mm in total length ranged from a low of 102 in 2018 to a high of 109 in 2016 and 2019 (Table 5.3-2; Figure 5.3-8).

The overall mean Wr was 104, the median was 105, and the IQR was 103-105 (Figure 5.3-8). The annual mean Wr fell within the overall IQR except in 2018 when it was below the IQR.

ROTATIONAL SITES

Pine Falls Forebay

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

The annual mean Wr of Northern Pike greater than 99 mm in total length over the three years of monitoring ranged from a low of 79 in 2014 to a high of 84 in 2017 (Table 5.3-2; Figure 5.3-5).

The overall mean Wr was 82, the median was 80, and the IQR was 79-84 (Figure 5.3-5). The annual mean Wr was equal to or fell within the overall IQR in all three years.

Sauger

Sauger greater than 69 mm in total length had an annual mean Wr of 92 in 2017 (Table 5.3-2; Figure 5.3-6). Sauger was not a target species in the Pine Falls Forebay prior to 2017.

Walleye

The annual mean Wr of Walleye greater than 29 mm in total length over the three years of monitoring ranged from a low of 80 in 2014 to a high of 93 in 2017 (Table 5.3-2; Figure 5.3-7).

The overall mean Wr was 87, the median was 93, and the IQR was 80-93 (Figure 5.3-7). The annual mean Wr was equal to or fell within the overall IQR in all three years.

White Sucker

The annual mean Wr of White Sucker greater than 99 mm in total length over the three years of monitoring ranged from a low of 94 in 2014 to a high of 104 in 2017 (Table 5.3-2; Figure 5.3-8).

The overall mean W_r was 99, the median was 96, and the IQR was 94-104 (Figure 5.3-8). The annual mean W_r was equal to or fell within the overall IQR in all three years.

5.3.2.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

The annual mean W_r of Northern Pike greater than 99 mm in total length over the 12 years of monitoring ranged from a low of 71 in 2010 to a high of 88 in 2019 (Table 5.3-2; Figure 5.3-5).

The overall mean and median W_r were 83 and the IQR was 81-86 (Figure 5.3-5). The annual mean W_r fell within the overall IQR except in 2008, 2010, and 2011 when it was below the IQR and in 2017 and 2019 when it was above the IQR.

Sauger

Sauger were not captured in Manigotagan Lake over the 12 years of monitoring (Table 5.3-2).

Walleye

The annual mean W_r of Walleye greater than 29 mm in total length over the 12 years of monitoring ranged from a low of 80 in 2008 to a high of 87 in 2019 (Table 5.3-2; Figure 5.3-7).

The overall mean W_r was 84, the median was 83, and the IQR was 82-86 (Figure 5.3-7). The annual mean W_r fell within the overall IQR except in 2008 and 2013 when it was below the IQR and in 2019 when it was above the IQR.

White Sucker

The annual mean W_r of White Sucker greater than 99 mm in total length over the 10 years of monitoring that it was a target species ranged from a low of 95 in 2010 to a high of 106 in 2015 (Table 5.3-2; Figure 5.3-8).

The overall mean Wr was 102, the median was 101, and the IQR was 101-105 (Figure 5.3-8). The annual mean Wr fell within the overall IQR except in 2010, 2011, 2013, and 2014 when it was below the IQR and in 2015 when it was above the IQR.

ROTATIONAL SITES

Eaglenest Lake

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

The annual mean Wr of Northern Pike greater than 99 mm in total length over the four years of monitoring ranged from a low of 85 in 2019 to a high of 89 in 2010 (Table 5.3-2; Figure 5.3-5).

The overall mean and median Wr were 86 and the IQR was 85-87 (Figure 5.3-5). The annual mean Wr fell within the overall IQR except in 2010 when it was above the IQR.

Sauger

The annual mean Wr of Sauger greater than 69 mm in total length was 96 in 2016 and 86 in 2019 (Table 5.3-2; Figure 5.3-6). Sauger was not a target species in Eaglenest Lake prior to 2016.

Walleye

The annual mean Wr of Walleye greater than 29 mm in total length over the four years of monitoring ranged from a low of 87 in 2010 to a high of 92 in 2016 (Table 5.3-2; Figure 5.3-7).

The overall mean Wr was 90, the median was 89, and the IQR was 87-90 (Figure 5.3-7). The annual mean Wr was equal to or fell within the overall IQR except in 2016 when it was above the IQR.

White Sucker

The annual mean Wr of White Sucker greater than 99 mm in total length over the four years of monitoring ranged from a low of 96 in 2010 to a high of 110 in 2016 (Table 5.3-2; Figure 5.3-8).

The overall mean Wr was 103, the median was 104, and the IQR was 96-104 (Figure 5.3-8). The annual mean Wr fell within the overall IQR in all years except 2016 when it was above the IQR.

Table 5.3-2. 2008-2019 Relative weight of target species.

| Waterbody | Year | NRPK | | | SAUG | | | WALL | | | WHSC | | |
|-----------|------|-----------------------------|------|-----------------|----------------|------|-----|----------------|------|-----|----------------|------|----|
| | | n _F ¹ | Mean | SE ² | n _F | Mean | SE | n _F | Mean | SE | n _F | Mean | SE |
| PDB | 2008 | 20 | 84 | 3 | 91 | 104 | 2 | 71 | 98 | 2 | | | |
| | 2009 | 30 | 83 | 3 | 120 | 81 | 1 | 80 | 88 | 1 | | | |
| | 2010 | 13 | 77 | 3 | 35 | 85 | 2 | 47 | 84 | 1 | 100 | 94 | 1 |
| | 2011 | 18 | 78 | 3 | 61 | 99 | 2 | 48 | 92 | 2 | 191 | 105 | 1 |
| | 2012 | 12 | 86 | 4 | 150 | 94 | 1 | 61 | 89 | 1 | 108 | 104 | 1 |
| | 2013 | 52 | 87 | 1 | 160 | 91 | 1 | 72 | 89 | 1 | 114 | 101 | 1 |
| | 2014 | 25 | 86 | 3 | 126 | 85 | 1 | 33 | 89 | 2 | 62 | 103 | 1 |
| | 2015 | 17 | 84 | 2 | 146 | 84 | 1 | 56 | 89 | 1 | 98 | 102 | 1 |
| | 2016 | 22 | 89 | 3 | 145 | 90 | 1 | 38 | 93 | 1 | 71 | 104 | 1 |
| | 2017 | 16 | 87 | 3 | 145 | 91 | 1 | 44 | 89 | 1 | 112 | 102 | 1 |
| | 2018 | 19 | 85 | 2 | 127 | 84 | 1 | 38 | 87 | 1 | 99 | 98 | 1 |
| 2019 | 23 | 83 | 1 | 132 | 89 | 1 | 86 | 90 | 1 | 124 | 102 | 1 | |
| LDB | 2008 | 23 | 89 | 4 | 90 | 86 | 1 | 57 | 93 | 3 | | | |
| | 2009 | 12 | 80 | 2 | 167 | 88 | 1 | 66 | 88 | 1 | | | |
| | 2010 | 40 | 89 | 1 | 77 | 88 | 1 | 62 | 88 | 1 | 34 | 103 | 1 |
| | 2011 | 35 | 91 | 2 | 90 | 86 | 1 | 76 | 88 | 2 | | | |
| | 2012 | 56 | 91 | 1 | 92 | 94 | 1 | 123 | 94 | 1 | | | |
| | 2013 | 51 | 88 | 2 | 116 | 88 | 1 | 81 | 87 | 1 | | | |
| | 2014 | 52 | 89 | 2 | 76 | 95 | 1 | 101 | 92 | 1 | | | |
| | 2015 | 39 | 83 | 2 | 77 | 86 | 2 | 76 | 88 | 1 | | | |
| | 2016 | 35 | 82 | 1 | 69 | 88 | 2 | 69 | 92 | 2 | 36 | 105 | 1 |
| | 2017 | 36 | 80 | 2 | 90 | 88 | 1 | 70 | 93 | 2 | | | |
| | 2018 | 36 | 88 | 3 | 129 | 89 | 1 | 88 | 84 | 1 | 45 | 102 | 2 |
| 2019 | 43 | 86 | 1 | 112 | 95 | 2 | 85 | 95 | 2 | 67 | 105 | 1 | |
| PFF | 2011 | 5 | 80 | 3 | | | | 22 | 93 | 2 | 16 | 96 | 2 |
| | 2014 | 7 | 79 | 3 | | | | 50 | 80 | 1 | 24 | 94 | 3 |
| | 2017 | 11 | 84 | 2 | 33 | 92 | 2 | 31 | 93 | 3 | 27 | 104 | 2 |
| MANIG | 2008 | 12 | 74 | 3 | - | - | - | 134 | 80 | 1 | | | |
| | 2009 | 13 | 81 | 3 | - | - | - | 97 | 82 | 1 | | | |
| | 2010 | 11 | 71 | 4 | - | - | - | 155 | 86 | 1 | 11 | 95 | 2 |
| | 2011 | 18 | 77 | 2 | - | - | - | 97 | 82 | 1 | 11 | 97 | 2 |
| | 2012 | 24 | 83 | 2 | - | - | - | 191 | 85 | 1 | 28 | 101 | 1 |
| | 2013 | 28 | 86 | 1 | - | - | - | 100 | 81 | 1 | 22 | 98 | 1 |
| | 2014 | 30 | 86 | 2 | - | - | - | 112 | 82 | 1 | 21 | 100 | 1 |
| | 2015 | 19 | 81 | 2 | - | - | - | 113 | 85 | 1 | 21 | 106 | 5 |
| | 2016 | 31 | 83 | 1 | - | - | - | 109 | 83 | 1 | 32 | 101 | 2 |
| | 2017 | 27 | 87 | 2 | - | - | - | 120 | 86 | 1 | 31 | 103 | 1 |
| | 2018 | 29 | 83 | 2 | - | - | - | 151 | 82 | 1 | 58 | 101 | 1 |
| 2019 | 15 | 88 | 2 | - | - | - | 123 | 87 | 1 | 52 | 105 | 1 | |
| EAGLE | 2010 | 34 | 89 | 2 | | | | 103 | 87 | 1 | 89 | 96 | 1 |
| | 2013 | 54 | 86 | 1 | | | | 108 | 90 | 1 | 66 | 103 | 1 |
| | 2016 | 52 | 87 | 2 | 88 | 96 | 3 | 68 | 92 | 2 | 79 | 110 | 1 |
| | 2019 | 64 | 85 | 1 | 132 | 85 | 1 | 129 | 89 | 2 | 83 | 104 | 1 |

Notes:

1. n_F = number of fish measured for length and weight.
2. SE = standard error.
3. Grey shading indicates a species was not a target species in that year.

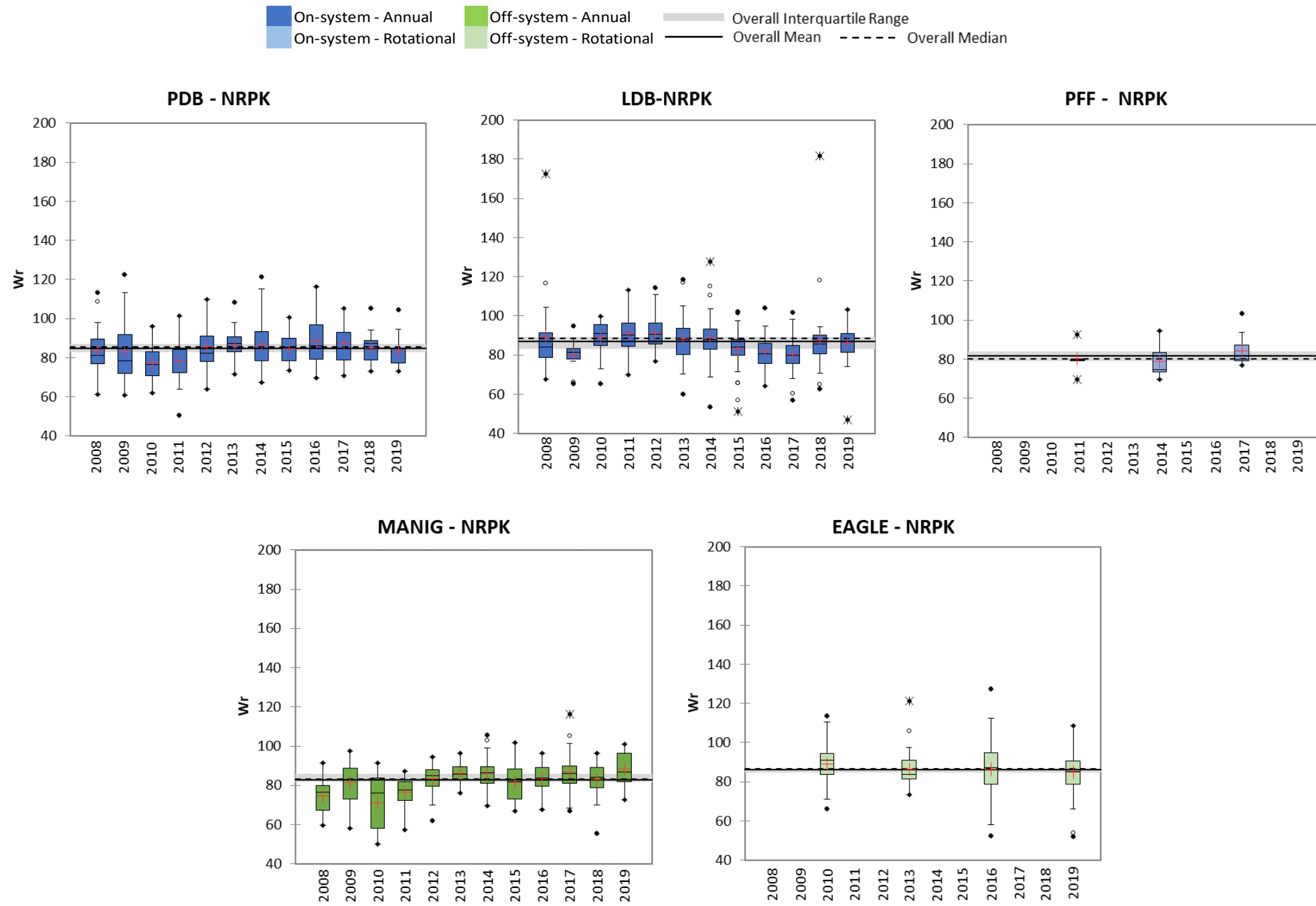


Figure 5.3-5. 2008-2019 Relative weight (Wr) of Northern Pike.

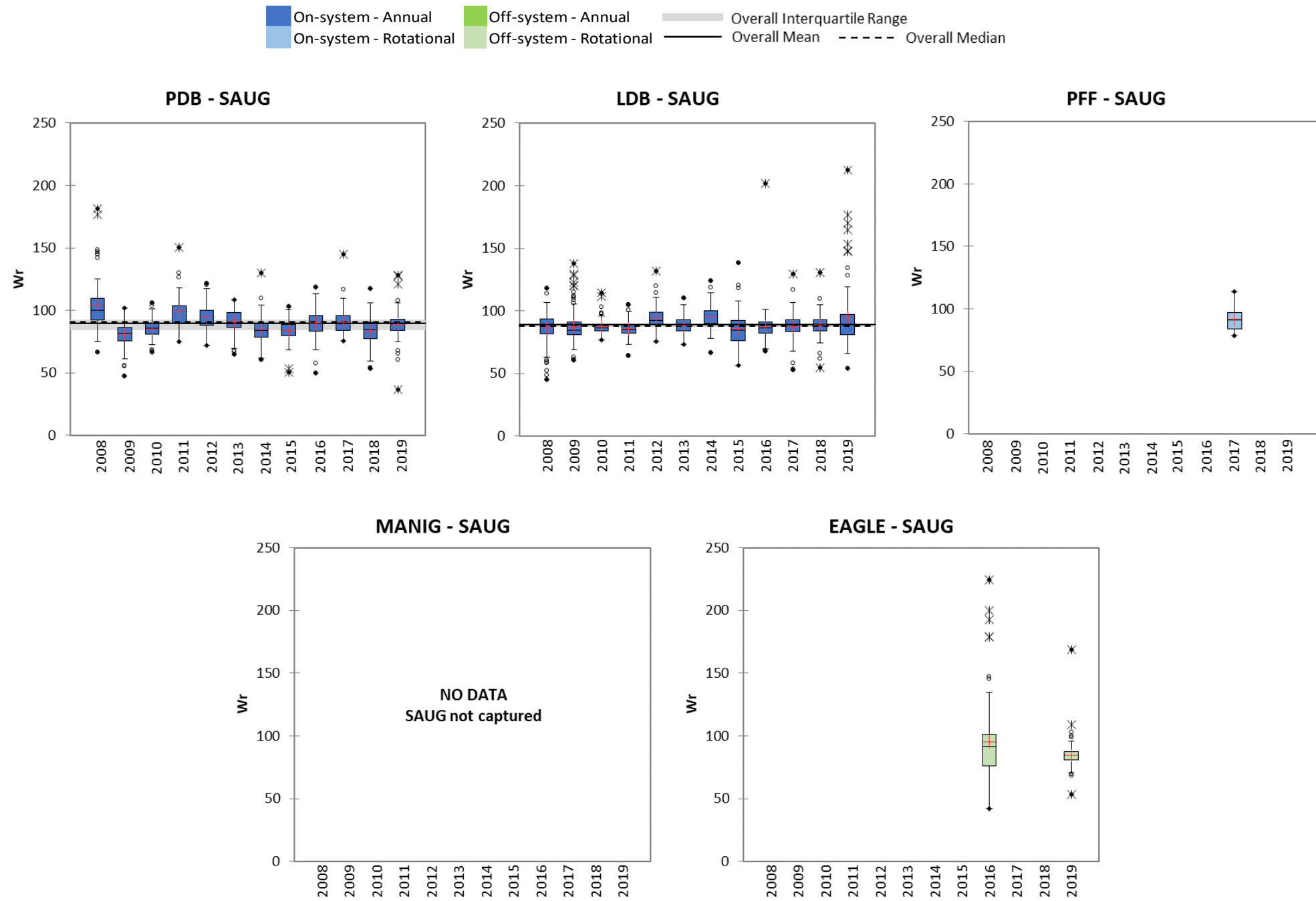


Figure 5.3-6. 2008-2019 Relative weight (Wr) of Sauger.

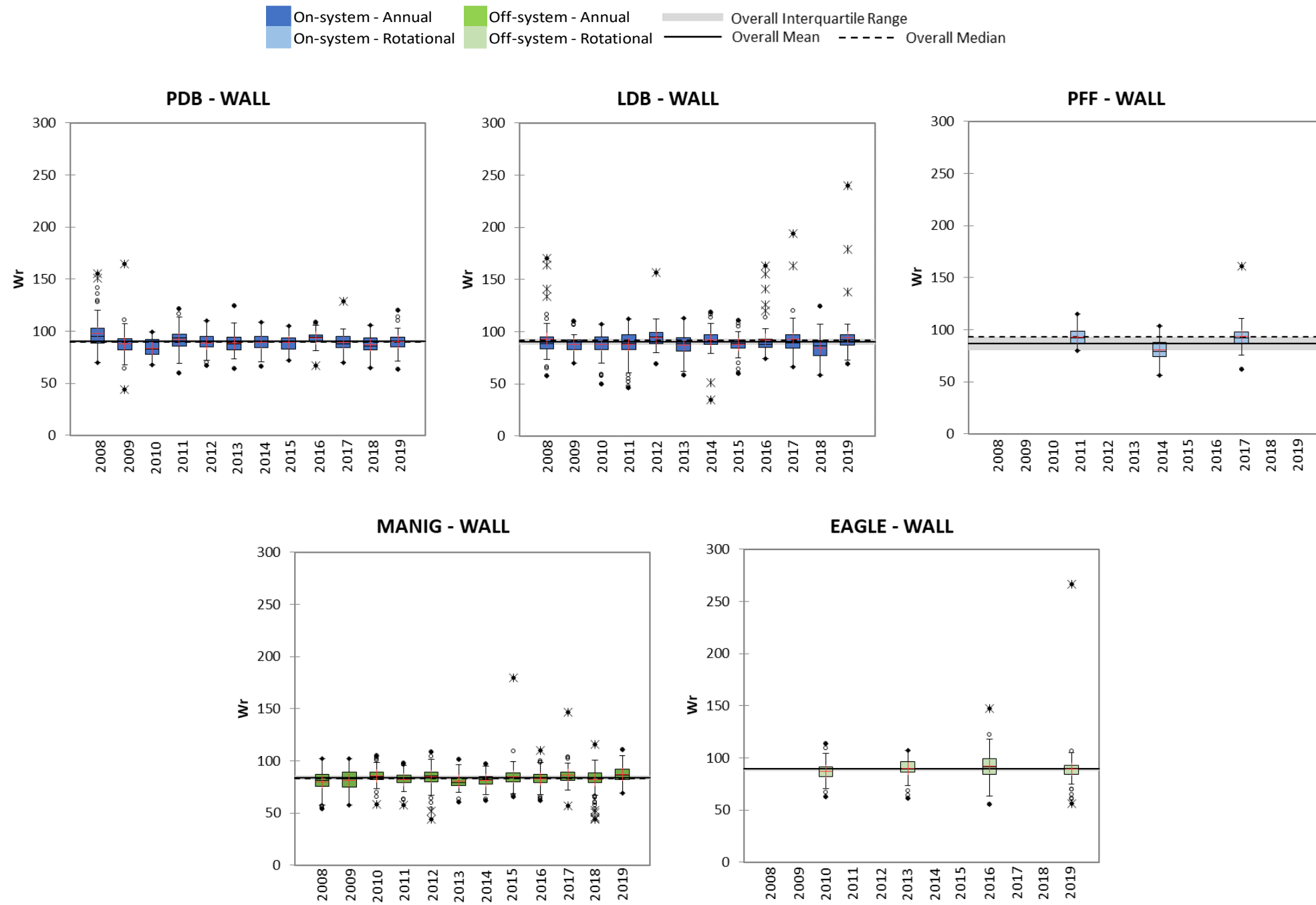


Figure 5.3-7. 2008-2019 Relative weight (Wr) of Walleye.

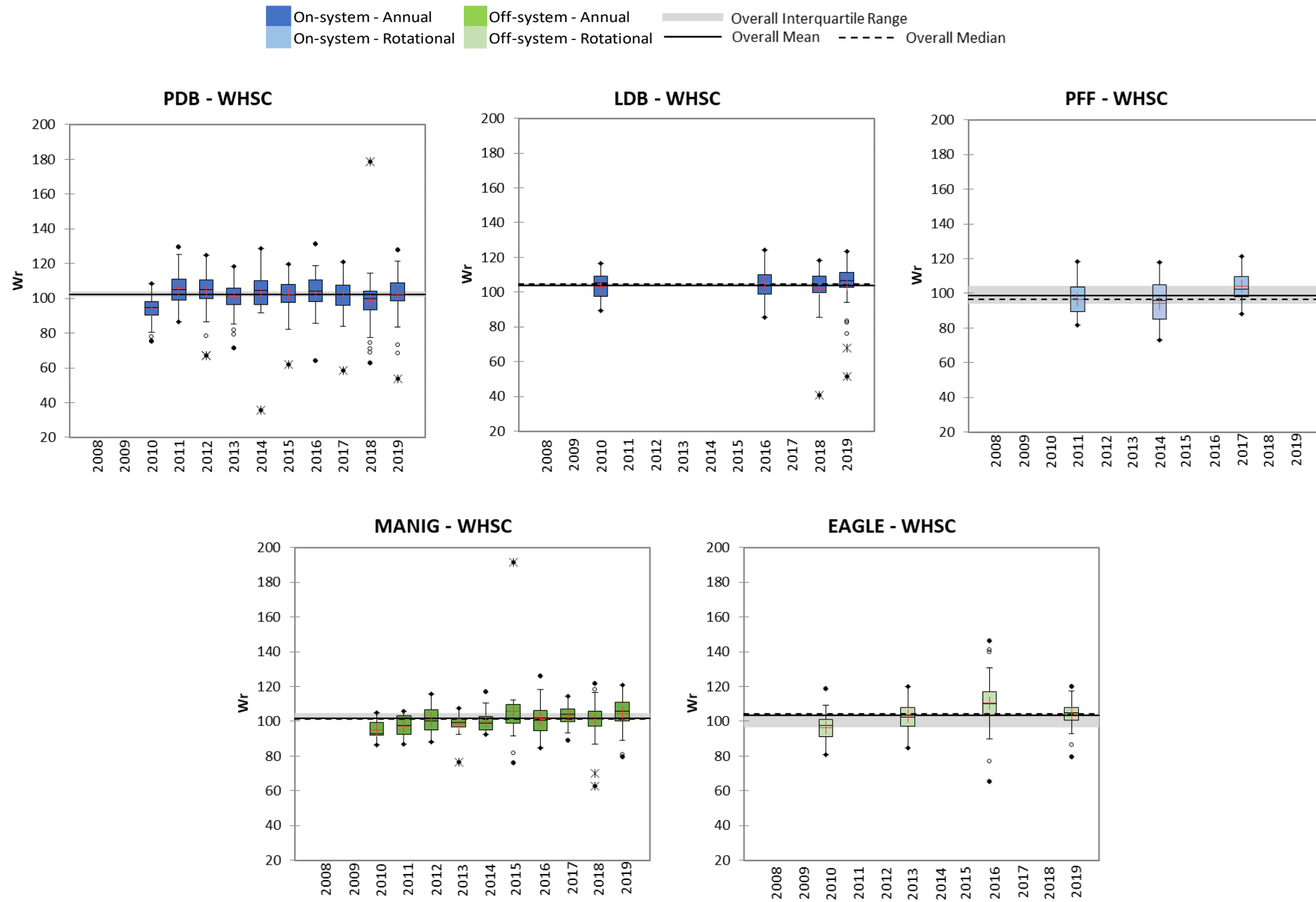


Figure 5.3-8. 2008-2019 Relative weight (Wr) of White Sucker.

5.4 GROWTH

5.4.1 LENGTH-AT-AGE

5.4.1.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

The annual mean FLA of 4-year-old Northern Pike over the 12 years of monitoring ranged from a low of 390 in 2011 to a high of 545 mm in 2014 (Table 5.4-1; Figure 5.4-1).

The overall mean FLA was 452, the median was 435, and the IQR was 421-470 mm (Figure 5.4-1). The annual mean FLA fell within the overall IQR except in 2009 and 2011 when it was below the IQR and in 2008, 2014, and 2016 when it was above the IQR.

Sauger

The annual mean FLA of 3-year-old Sauger over the 12 years of monitoring ranged from a low of 217 in 2008 to a high of 250 mm in 2017 (Table 5.4-1; Figure 5.4-2).

The overall mean and median FLA were 230 and the IQR was 226-230 mm (Figure 5.4-2). The annual mean FLA fell within the overall IQR except in 2008, 2010, and 2012 when it was below the IQR and in 2009, 2017, and 2018 when it was above the IQR.

Walleye

The annual mean FLA of 3-year-old Walleye over the 12 years of monitoring ranged from a low of 234 in 2009 to a high of 263 mm in 2012 (Table 5.4-1; Figure 5.4-3).

The overall mean FLA was 249, the median was 250, and the IQR was 247-259 mm (Figure 5.4-3). The annual mean FLA fell within the overall IQR except in 2008, 2009, and 2011 when it was below the IQR and in 2012, 2015, and 2017 when it was above the IQR.

White Sucker

White Sucker are not aged as part of CAMP.

Lac du Bonnet

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

The annual mean FLA of 4-year-old Northern Pike over the 12 years of monitoring ranged from a low of 509 in 2009 to a high of 707 mm in 2014 (Table 5.4-1; Figure 5.4-1).

The overall mean FLA was 601, the median was 604, and the IQR was 590-622 mm (Figure 5.4-1). The annual mean FLA fell within the overall IQR except in 2009, 2012, and 2015 when it was below the IQR and in 2014 when it was above the IQR.

Sauger

The annual mean FLA of 3-year-old Sauger over the 12 years of monitoring ranged from a low of 222 in 2010 to a high of 261 mm in 2019 (Table 5.4-1; Figure 5.4-2).

The overall mean FLA was 230, the median was 247, and the IQR was 233-253 mm (Figure 5.4-2). The annual mean FLA fell within the overall IQR except in 2010 when it was below the IQR and in 2012, 2017, and 2019 when it was above the IQR.

Walleye

The annual mean FLA of 3-year-old Walleye over the 12 years of monitoring ranged from a low of 282 in 2010 to a high of 347 mm in 2013 (Table 5.4-1; Figure 5.4-3).

The overall mean FLA was 309, the median was 304, and the IQR was 304-320 mm (Figure 5.4-3). The annual mean FLA fell within the overall IQR except in 2009, 2010, and 2011 when it was below the IQR and in 2012, 2013, and 2019 when it was above the IQR.

White Sucker

White Sucker are not aged as part of CAMP.

ROTATIONAL SITES

Pine Falls Forebay

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

The annual mean FLA of 4-year-old Northern Pike over the three years of monitoring ranged from a low of 428 in 2011 to a high of 464 mm in 2014 (Table 5.4-1; Figure 5.4-1).

There were too few 4-year-old Northern Pike captured in the Pine Falls Forebay to calculate the overall metrics.

Sauger

The annual mean FLA of 3-year-old Sauger was 247 mm in 2017 (Table 5.4-1; Figure 5.4-2). Sauger was not a target species in the Pine Falls Forebay prior to 2017.

Walleye

The annual mean FLA of 3-year-old Walleye over the three years of monitoring ranged from a low of 189 in 2011 to a high of 278 mm in 2014 (Table 5.4-1; Figure 5.4-3).

There were too few 3-year-old Walleye captured in the Pine Falls Forebay to calculate the overall metrics.

White Sucker

White Sucker are not aged as part of CAMP.

5.4.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

The annual mean FLA of 4-year-old Northern Pike over the 12 years of monitoring ranged from a low of 380 in 2011 to a high of 539 mm in 2018 (Table 5.4-1; Figure 5.4-1).

The overall mean FLA was 494, the median was 487, and the IQR was 484-521 mm (Figure 5.4-1). The annual mean FLA fell within the overall IQR except in 2011, 2012, and 2019 when it was below the IQR and in 2017 and 2018 when it was above the IQR.

Sauger

Sauger were not captured in Manigotagan Lake over the 12 years of monitoring (Table 5.4-1).

Walleye

The annual mean FLA of 3-year-old Walleye over the 12 years of monitoring ranged from a low of 267 in 2010 to a high of 341 mm in 2008 (Table 5.4-1; Figure 5.4-3).

The overall mean FLA was 297, the median was 295, and the IQR was 288-312 mm (Figure 5.4-3). The annual mean FLA fell within the overall IQR except in 2010, 2013, and 2014 when it was below the IQR and in 2008, 2012, and 2016 when it was above the IQR.

White Sucker

White Sucker are not aged as part of CAMP.

ROTATIONAL SITES

Eaglenest Lake

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

The annual mean FLA of 4-year-old Northern Pike over the four years of monitoring was 373 in 2010 to a high of 549 mm in 2016 (Table 5.4-1; Figure 5.4-1).

The overall mean FLA was 467, the median was 425, and the IQR was 425-510 mm (Figure 5.4-1). The annual mean FLA fell within the overall IQR except in 2010 when it was below the IQR and 2016 was above the IQR.

Sauger

The annual mean FLA of 3-year-old Sauger was 215 in 2016 and 223 mm in 2019 (Table 5.4-1; Figure 5.4-2). Sauger was not a target species in Eaglenest Lake prior to 2016.

Walleye

The annual mean FLA of 3-year-old Walleye over the four years of monitoring ranged from a low of 229 in 2010 to a high of 271 mm in 2013 (Table 5.4-1; Figure 5.4-3).

The overall mean FLA was 259, the median was 263, and the IQR was 263-271 mm (Figure 5.4-3). The annual mean FLA fell within the overall IQR except in 2010 and 2016 when it was below the IQR.

White Sucker

White Sucker are not aged as part of CAMP.

Table 5.4-1. 2008-2019 Fork length-at-age of target species.

| Waterbody | Year | NRPK | | | SAUG | | | WALL | | |
|-----------|------|-----------------------------|------|-----------------|----------------|------|----|----------------|------|----|
| | | n _F ¹ | Mean | SE ² | n _F | Mean | SE | n _F | Mean | SE |
| PDB | 2008 | 2 | 474 | 22 | 5 | 217 | 14 | 4 | 234 | 9 |
| | 2009 | 8 | 415 | 10 | 5 | 242 | 2 | 12 | 234 | 6 |
| | 2010 | 6 | 463 | 10 | 5 | 225 | 6 | - | - | - |
| | 2011 | 1 | 390 | - | - | - | - | 4 | 239 | 8 |
| | 2012 | 1 | 440 | - | 5 | 220 | 11 | 2 | 263 | 10 |
| | 2013 | 17 | 421 | 11 | 27 | 226 | 5 | 16 | 251 | 6 |
| | 2014 | 3 | 545 | 19 | 21 | 230 | 4 | 7 | 247 | 7 |
| | 2015 | 5 | 459 | 17 | 30 | 230 | 4 | 11 | 262 | 10 |
| | 2016 | 8 | 512 | 30 | 14 | 230 | 5 | 5 | 259 | 12 |
| | 2017 | 2 | 435 | 2 | 3 | 250 | 2 | 5 | 260 | 7 |
| | 2018 | 2 | 424 | 25 | 11 | 242 | 4 | 3 | 248 | 9 |
| 2019 | 4 | 470 | 14 | 18 | 229 | 4 | 15 | 250 | 6 | |
| LDB | 2008 | - | - | - | 15 | 243 | 4 | 7 | 309 | 9 |
| | 2009 | 4 | 509 | 19 | 52 | 233 | 2 | 11 | 291 | 9 |
| | 2010 | 7 | 595 | 20 | 5 | 222 | 5 | 8 | 282 | 6 |
| | 2011 | 2 | 604 | 16 | 10 | 242 | 3 | 4 | 301 | 20 |
| | 2012 | 2 | 571 | 39 | 4 | 258 | 11 | 20 | 332 | 6 |
| | 2013 | 7 | 605 | 17 | 16 | 245 | 4 | 2 | 347 | 12 |
| | 2014 | 3 | 707 | 19 | 21 | 247 | 3 | 35 | 304 | 4 |
| | 2015 | 6 | 579 | 16 | 22 | 249 | 4 | 8 | 307 | 9 |
| | 2016 | 6 | 606 | 25 | 8 | 251 | 6 | 6 | 308 | 13 |
| | 2017 | 12 | 590 | 13 | 5 | 258 | 6 | 6 | 304 | 15 |
| | 2018 | 7 | 622 | 33 | 26 | 253 | 4 | 4 | 320 | 8 |
| 2019 | 8 | 622 | 16 | 19 | 261 | 4 | 6 | 328 | 9 | |
| PFF | 2011 | 1 | 428 | - | | | | 1 | 189 | - |
| | 2014 | 1 | 464 | - | | | | 11 | 278 | 10 |
| | 2017 | 4 | 439 | 6 | 5 | 247 | 5 | 4 | 248 | 5 |
| MANIG | 2008 | 3 | 521 | 9 | - | - | - | 11 | 341 | 9 |
| | 2009 | 3 | 506 | 15 | - | - | - | 45 | 295 | 5 |
| | 2010 | - | - | - | - | - | - | 16 | 267 | 9 |
| | 2011 | 1 | 380 | - | - | - | - | 6 | 294 | 9 |
| | 2012 | 3 | 459 | 9 | - | - | - | 7 | 328 | 10 |
| | 2013 | 11 | 484 | 8 | - | - | - | 7 | 280 | 14 |
| | 2014 | 7 | 485 | 11 | - | - | - | 21 | 272 | 9 |
| | 2015 | 9 | 487 | 11 | - | - | - | 22 | 288 | 7 |
| | 2016 | 6 | 506 | 14 | - | - | - | 10 | 319 | 10 |
| | 2017 | 5 | 525 | 11 | - | - | - | 8 | 295 | 3 |
| 2018 | 10 | 539 | 16 | - | - | - | 13 | 323 | 6 | |
| 2019 | 7 | 445 | 8 | - | - | - | 16 | 312 | 7 | |
| EAGLE | 2010 | 3 | 373 | 16 | | | | 8 | 229 | 11 |
| | 2013 | 13 | 425 | 18 | | | | 14 | 271 | 10 |
| | 2016 | 6 | 549 | 34 | 1 | 215 | - | 3 | 244 | 6 |
| | 2019 | 8 | 510 | 22 | 14 | 223 | 7 | 30 | 263 | 3 |

Notes:

1. n_F = number of fish measured for length and weight.
2. SE = standard error.
3. Grey shading indicates that a species was not a target species in that year.

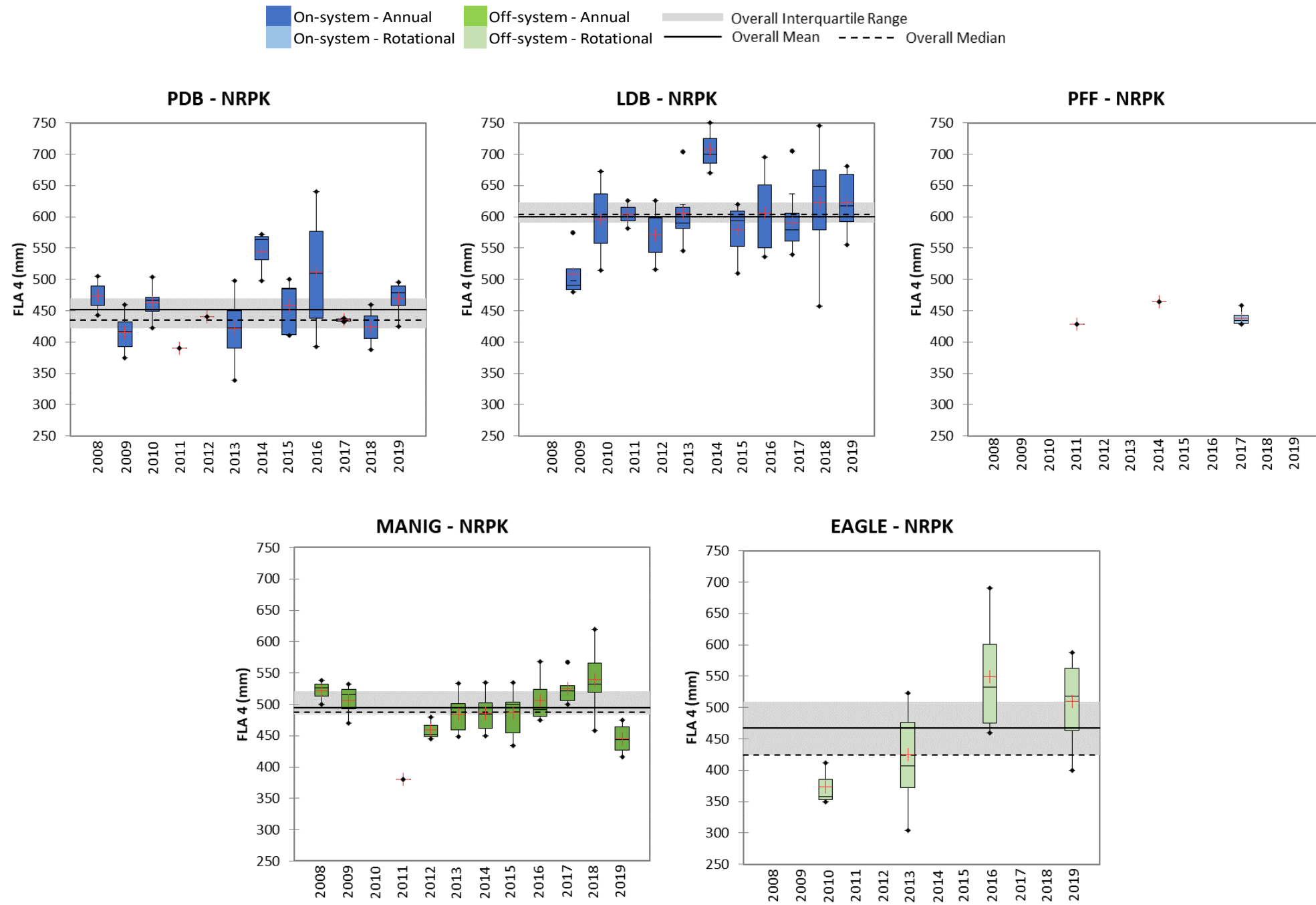


Figure 5.4-1. 2008-2019 Fork length-at-age (FLA) 4 of Northern Pike.

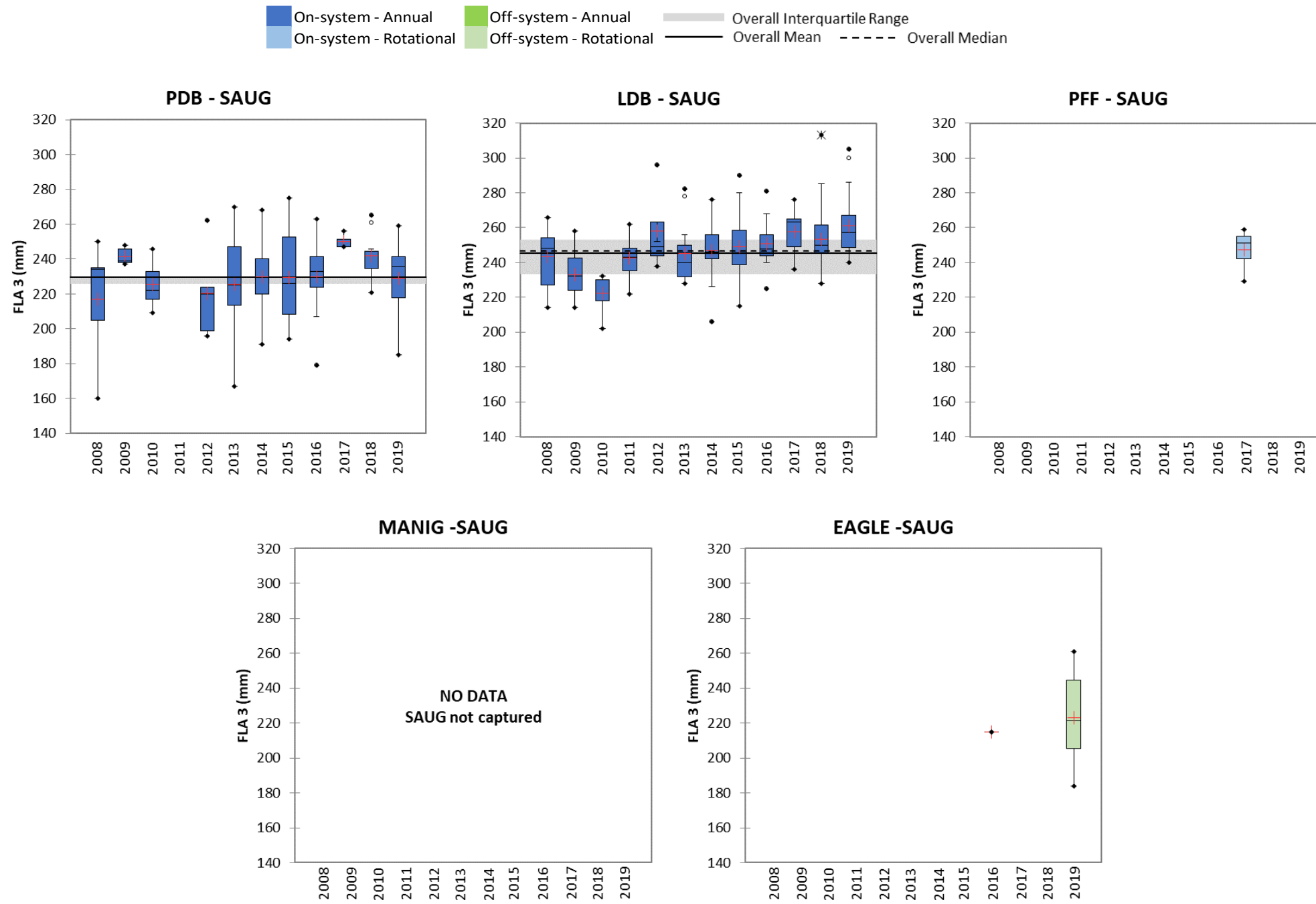


Figure 5.4-2. 2008-2019 Fork length-at-age (FLA) 3 of Sauger.

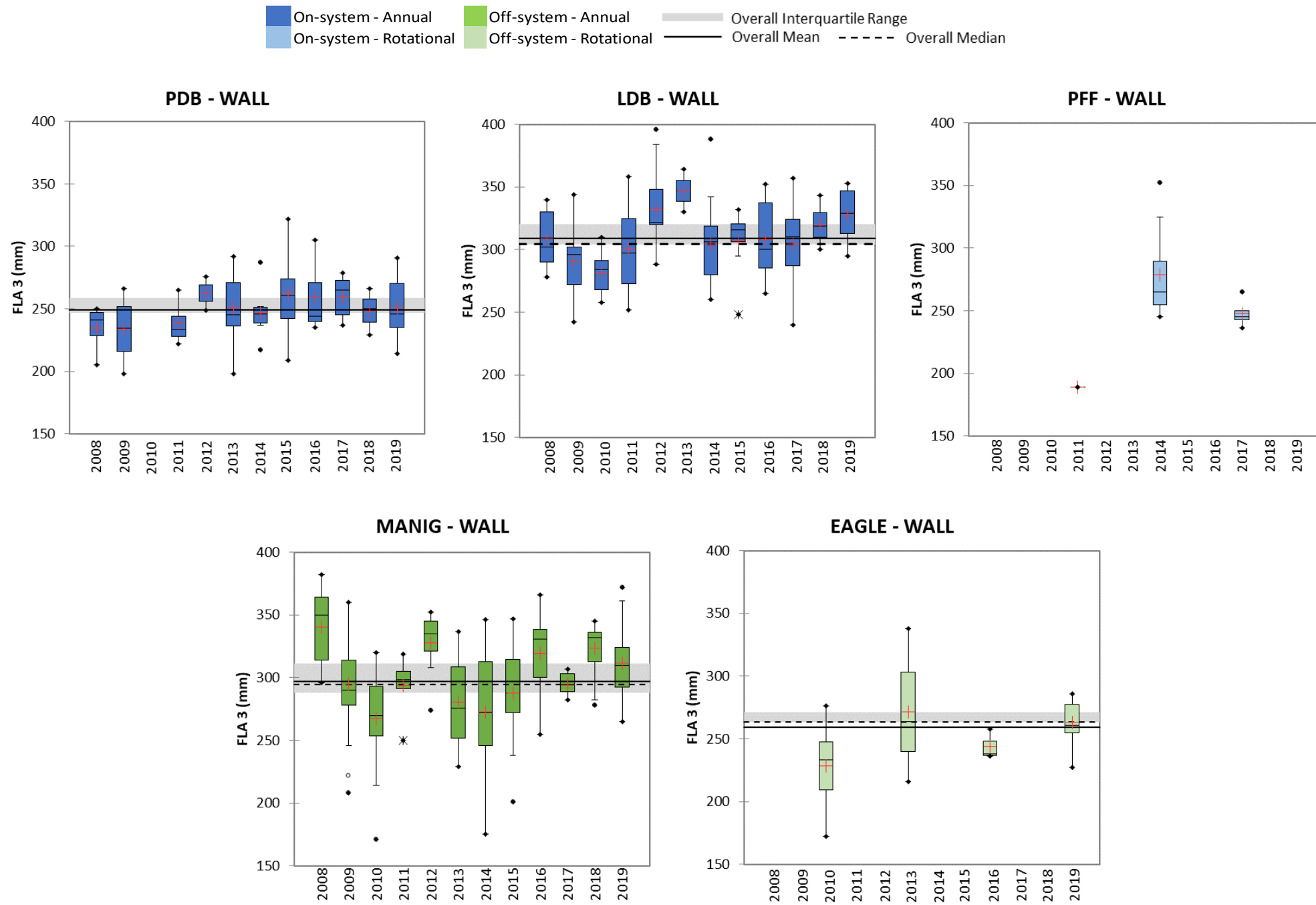


Figure 5.4-3. 2008-2019 Fork length-at-age (FLA) 3 of Walleye.

5.5 RECRUITMENT

5.5.1 RELATIVE YEAR-CLASS STRENGTH

5.5.1.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

The RYCS of Northern Pike over the 12 years of monitoring ranged from a low of 22 for the 2002 cohort to a high of 171 for the 2010 cohort (Figure 5.5-1). There were no missing cohorts from 2002-2014. Particularly weak cohorts (<50) occurred in 2002, 2007, and 2012.

Sauger

The RYCS of Sauger over the 12 years of monitoring ranged from a low of 30 for the 2008 cohort to a high of 233 for the 2006 cohort (Figure 5.5-2). There were no missing cohorts from 2002-2013. A series of weak cohorts (<50) were produced over a three-year period from 2008-2010. Strong cohorts (>100) were produced in 2002, 2003, 2006, 2007, 2011, and 2012.

Walleye

The RYCS of Walleye over the 12 years of monitoring ranged from a low of 46 for the 2008 cohort to a high of 211 for the 2011 cohort (Figure 5.5-3). There were no missing cohorts from 2002-2014. A particularly weak cohort (<50) only occurred in 2008 and strong cohorts (>100) were produced in 2002, 2003, 2006, 2010, and 2011.

White Sucker

White Sucker are not aged as part of CAMP.

Lac du Bonnet

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

The RYCS of Northern Pike over the 12 years of monitoring ranged from a low of 43 for the 2002 cohort to a high of 135 for the 2005 cohort (Figure 5.5-1). There were no missing cohorts from 2002-2014. Particularly strong cohorts (>100) were produced in about half of the years (2003, 2005, 2006, 2011, 2013, and 2014), and only the 2002 cohort was particularly weak (<50).

Sauger

RYCS analysis was not conducted for Sauger from Lac du Bonnet because only a subsample of the catch was aged in some years which is not appropriate for the analysis.

Walleye

The RYCS of Walleye over the 12 years of monitoring ranged from a low of 12 for the 2004 cohort to a high of 281 for the 2011 cohort (Figure 5.5-3). There were no missing cohorts from 2002-2014. Particularly weak cohorts (<50) were produced in 2004, 2010, and 2013. Strong cohorts (>100) were produced in 2003, 2006, 2007, 2011, and 2014.

White Sucker

White Sucker are not aged as part of CAMP.

ROTATIONAL SITES

RYCS analysis requires data be collected in at least three consecutive years and therefore cannot be conducted for rotational waterbodies.

5.5.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

Lake Whitefish

Lake Whitefish was not selected as a target species in any of the Winnipeg River waterbodies even though it is a target species in other CAMP regions.

Northern Pike

The RYCS of Northern Pike over the 12 years of monitoring ranged from a low of 40 for the 2006 cohort to a high of 199 for the 2004 cohort (Figure 5.5-1). There were no missing cohorts from 2002-2014. Strong cohorts (>100) were produced in five years (2003, 2004, 2008, 2011, and 2014) and particularly weak cohorts (<50) produced in two years (2002 and 2006).

Sauger

Sauger were not captured in Manigotagan Lake over the 12 years of monitoring.

Walleye

The RYCS of Walleye over the 12 years of monitoring ranged from a low of 36 for the 2009 cohort to a high of 253 for the 2006 cohort (Figure 5.5-3). There were no missing cohorts from 2002-2014. A series of strong cohorts (>100) were produced from 2005-2007 and again from 2011-2012. A series of weak cohorts (<50) were produced over a three-year period from 2008-2010.

White Sucker

White Sucker are not aged as part of CAMP.

ROTATIONAL SITES

RYCS analysis requires data be collected in at least three consecutive years and therefore cannot be conducted for rotational waterbodies.

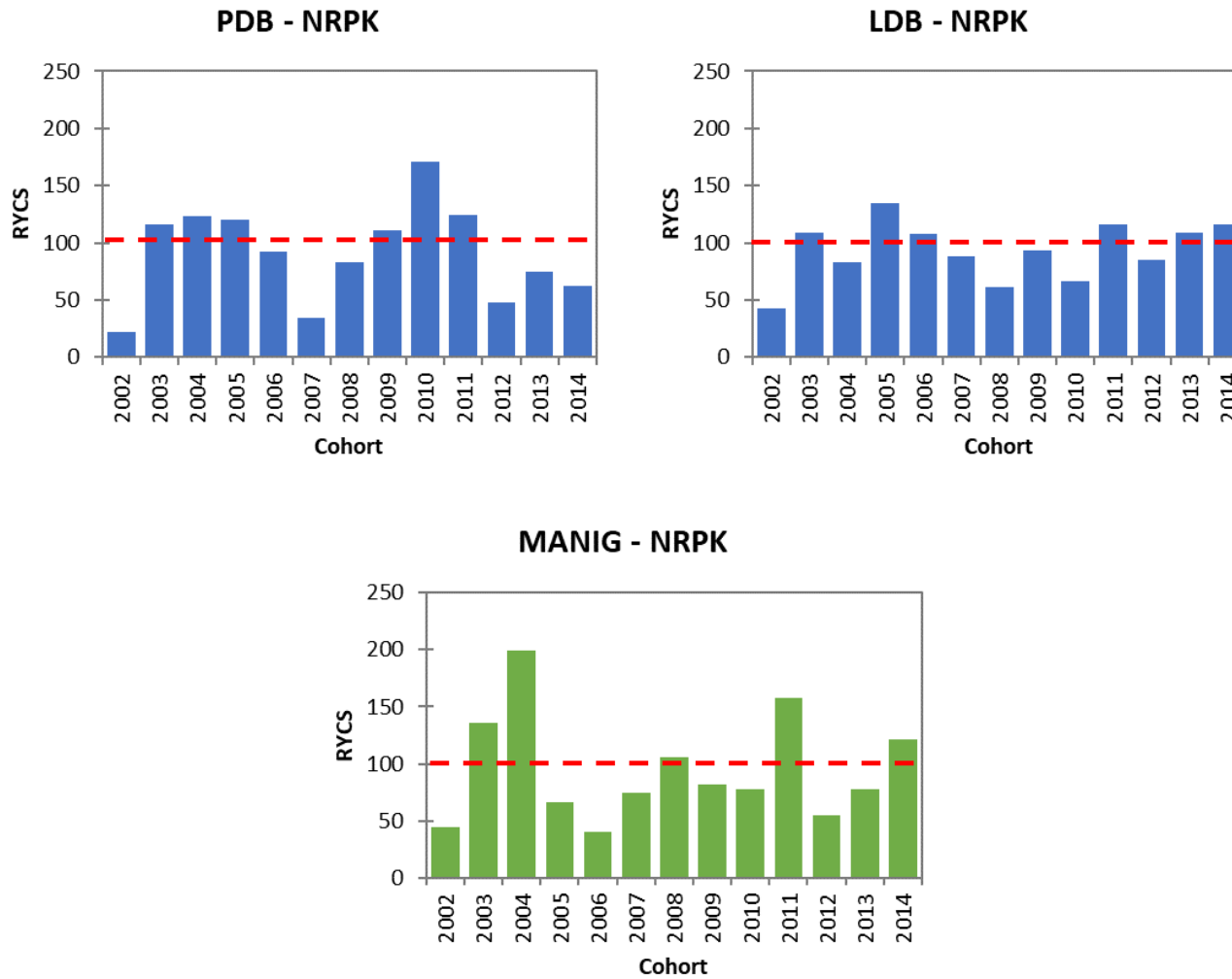
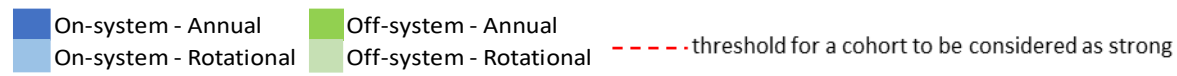


Figure 5.5-1. Relative year-class strength (RYCS) of Northern Pike.

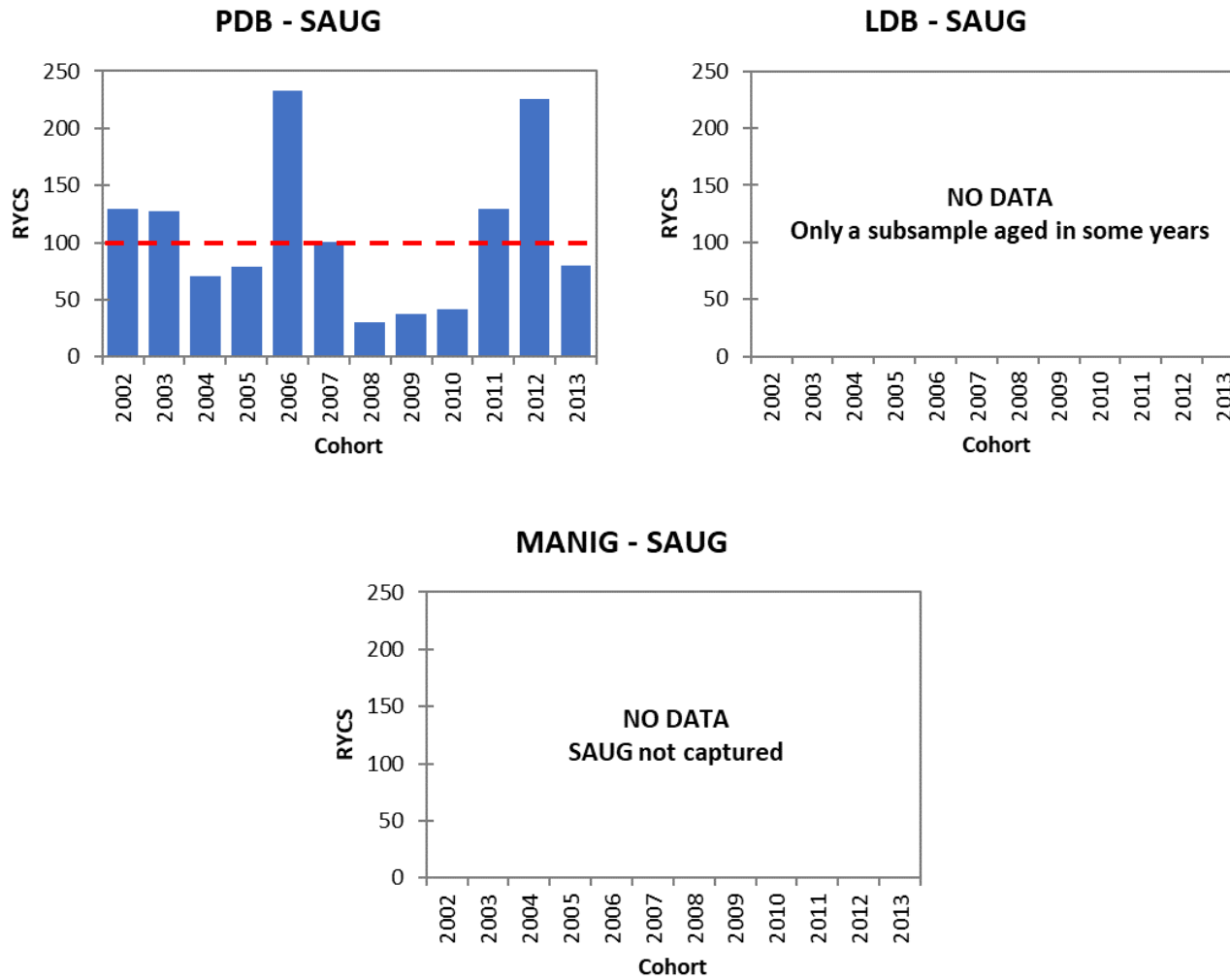
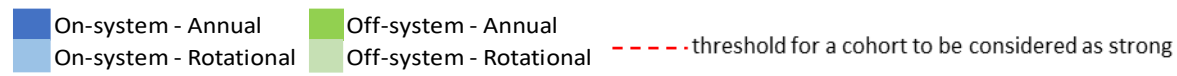


Figure 5.5-2. Relative year-class strength (RYCS) of Sauger.

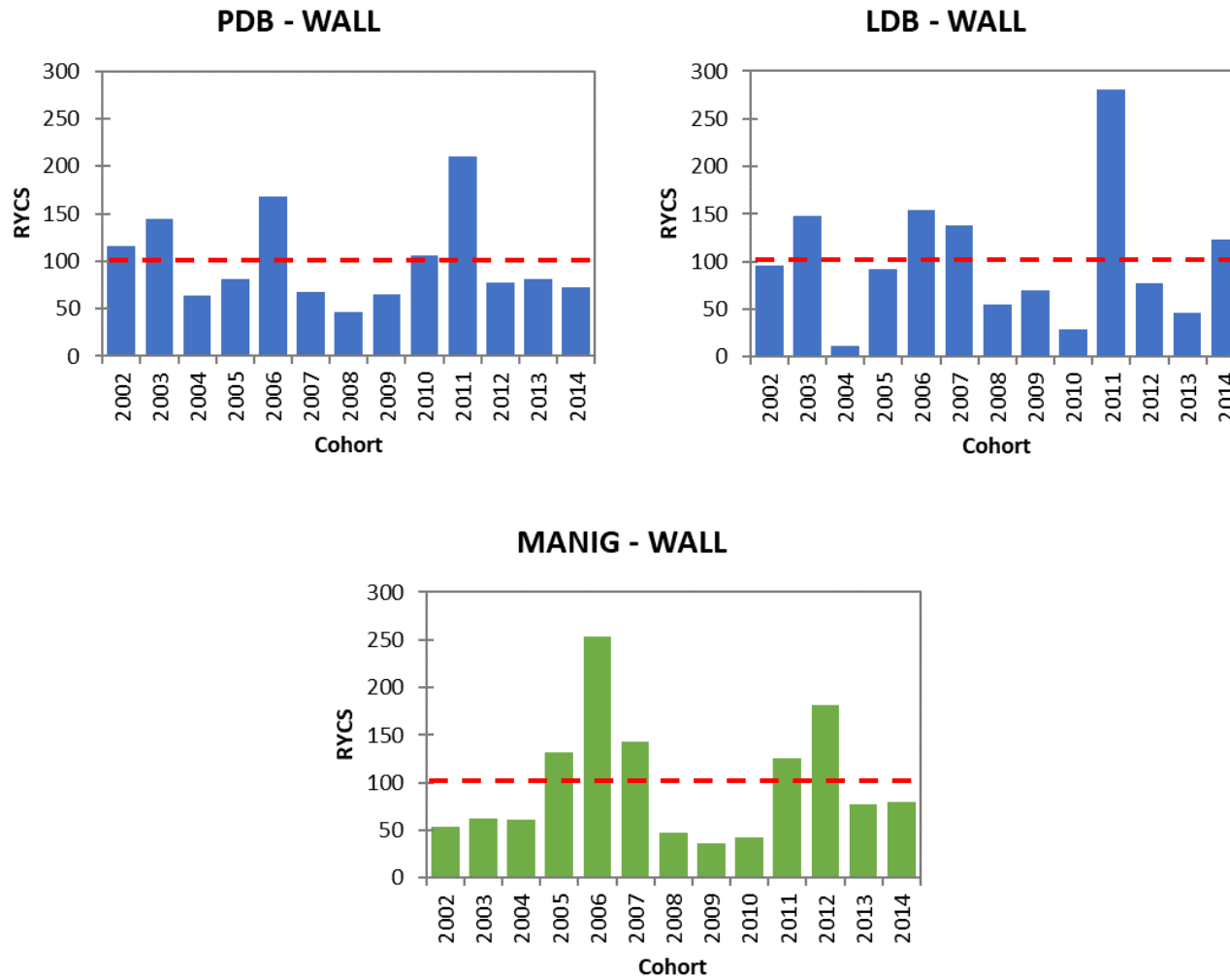
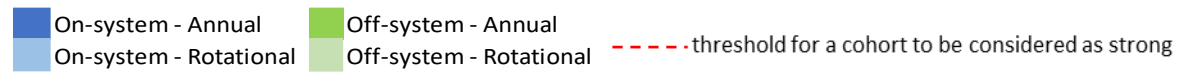


Figure 5.5-3. Relative year-class strength (RYCS) of Walleye.

5.6 DIVERSITY

5.6.1 RELATIVE SPECIES ABUNDANCE

5.6.1.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

A total of 22 fish species were captured in the combined standard and small mesh gangs at the Pointe du Bois Forebay over the 12 years of monitoring (Table 5.6-1) with the number of species caught each year ranging from 16-19 (Tables 5.6-2 and 5.6-3). In one case, a sculpin was not identified to species (unidentified [unid.] sculpin species [*spp.*]).

Standard Gang Index Gill Nets

White Sucker and Yellow Perch (*Perca flavescens*) were the most frequently captured species at the Pointe du Bois Forebay over the 12 years of monitoring, accounting for an average of >25% of the catch (Table 5.6-2). The annual RSA for White Sucker ranged from a low of 17% in 2016 to a high of 39% in 2009. The annual RSA for Yellow Perch ranged from a low of 11% in 2008 to a high of 47% in 2016. Sauger accounted for >25% in two years (2012 and 2014).

Small Mesh Index Gill Nets

The most common species captured in the Pointe du Bois Forebay over the 12 years of monitoring was Spottail Shiner (*Notropis hudsonius*), which accounted for an average of >25% of the catch (Table 5.6-3). The annual RSA for Spottail Shiner ranged from a low of 0% in 2009 to a high of 52% in 2014. Three other species accounted for >25% of the catch in some years: Sauger in 2008, 2009, 2010, and 2016; Yellow Perch in 2011, 2013, 2017, and 2019; and Trout-perch (*Percopsis omiscomaycus*) in 2010 and 2012.

Lac du Bonnet

A total of 23 fish species were captured in the combined standard and small mesh gangs at Lac du Bonnet over the 12 years of monitoring (Table 5.6-1), with the number of species caught each year ranging from 16-18 (Tables 5.6-4 and 5.6-5). In one case, a sculpin was not identified to species. Rainbow Smelt (*Osmerus mordax*) were not captured in Lac du Bonnet.

Standard Gang Index Gill Nets

The catch in standard gangs set in Lac du Bonnet over the 12 years of monitoring were not dominated by any one species, with none of the species accounting for an average of >25% of the catch (Table 5.6-4). Two species accounted for >25% of the catch in some years, Sauger in 2009, 2016, and 2017 and Walleye in 2012.

Small Mesh Index Gill Nets

The most common species captured in Lac du Bonnet over the 12 years of monitoring was Yellow Perch, which accounted for an average of >25% of the catch (Table 5.6-5). The annual RSA for Yellow Perch ranged from a low of 3% in 2016 to a high of 68% in 2015. Four other species accounted for >25% of the catch in some years, Spottail Shiner in 2012, 2013, 2014, 2017, and 2019, Trout-perch in 2016, Black Crappie (*Pomoxis nigromaculatus*) in 2015 and 2017, and Sauger in 2009 and 2016.

ROTATIONAL SITES

Pine Falls Forebay

A total of 17 fish species were captured in the combined standard and small mesh gangs at the Pine Falls Forebay over the three years of monitoring (Table 5.6-1), with the number of species caught each year ranging from 13-14 species (Tables 5.6-6 and 5.6-7).

Standard Gang Index Gill Nets

The catch in standard gangs set in the Pine Falls Forebay over the three years of monitoring was dominated by Channel Catfish, which accounted for an average of \geq 50% of the catch (Table 5.6-6). The annual RSA of Channel Catfish (*Ictalurus punctatus*) ranged from a low of 50% in 2017 to a high of 63% in 2014.

Small Mesh Index Gill Nets

The most common species captured in the Pine Falls Forebay over the three years of monitoring was Trout-perch, which accounted for an average of >25% of the catch (Table 5.6-7). The annual RSA for Trout-perch ranged from a low of 20% in 2011 to a high of 70% in 2014. Two other species accounted for >25% of the catch in some years, Sauger in 2017 and Walleye in 2011.

5.6.1.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

A total of 13 fish species were captured in the combined standard and small mesh gangs at Manigotagan Lake over the 12 years of monitoring (Table 5.6-1), with the number of species caught each year ranging from 7-13 species (Tables 5.6-8 and 5.6-9). Sauger, Rainbow Smelt, and Lake Sturgeon (*Acipenser fulvescens*) were not captured at Manigotagan Lake.

Standard Gang Index Gill Nets

Cisco (*Coregonus artedi*) and Walleye were the most frequently captured species at Manigotagan Lake over the 12 years of monitoring, accounting for an average of >25% of the catch (Table 5.6-2). The annual RSA for Cisco ranged from a low of 27% in 2008 to a high of 50% in 2015. The annual RSA for Walleye ranged from a low of 22% in 2015 to a high of 42% in 2010.

Small Mesh Index Gill Nets

The most common species captured in Manigotagan Lake over the 12 years of monitoring was Walleye, which accounted for an average of >25% of the catch (Table 5.6-3). The annual RSA for Walleye ranged from a low of 12% in 2018 to a high of 100% in 2008. Four other species accounted for >25% of the catch in some years, Emerald Shiner (*Notropis atherinoides*) in 2015, 2016, 2017, and 2018, Cisco in 2011 and 2014, Trout-perch in 2013, and Yellow Perch in 2012.

ROTATIONAL SITES

Eaglenest Lake

A total of 23 fish species were captured in the combined standard and small mesh gangs at Eaglenest Lake over the four years of monitoring (Table 5.6-1), with the number of species caught each year ranging from 16-19 species (Tables 5.6-10 and 5.6-11). Lake Sturgeon were not captured at Eaglenest Lake.

Standard Gang Index Gill Nets

Yellow Perch was the most frequently captured species at Eaglenest Lake over the four years of monitoring, accounting for an average of >25% of the catch (Table 5.6-10). The annual RSA for Yellow Perch ranged from a low of 25% in 2010 to a high of 37% in 2013.

Small Mesh Index Gill Nets

The most common species captured in Eaglenest Lake over the four years of monitoring was Yellow Perch, which accounted for an average of >25% of the catch (Table 5.6-11). The annual RSA for Yellow Perch ranged from a low of 12% in 2010 to a high of 51% in 2019. Three other species accounted for >25% of the catch in some years, Spottail Shiner in 2016, Trout-perch in 2010, and Sauger in 2013.

Table 5.6-1. 2008-2019 Inventory of fish species.

| Family | Species | Abbreviation | Status ¹ | Target | PDB | LDB | PFF | MANIG | EAGLE |
|-----------------|----------------------------|--------------|---------------------|--------|-----|-----|-----|-------|-------|
| Petromyzontidae | Silver Lamprey | SLLM | Native | | • | • | | | • |
| Acipenseridae | Lake Sturgeon ² | LKST | Native | | • | • | • | | |
| Hiodontidae | Mooneye | MOON | Native | | • | • | • | | • |
| | Goldeye | GOLD | Native | | | | • | | • |
| Cyprinidae | Emerald Shiner | EMSH | Native | | • | • | | • | • |
| | Spottail Shiner | SPSH | Native | | • | • | • | • | • |
| Catostomidae | Longnose Sucker | LNSC | Native | | • | • | | | • |
| | White Sucker | WHSC | Native | • | • | • | • | • | • |
| | Shorthead Redhorse | SHRD | Native | | • | • | • | | • |
| | Silver Redhorse | SLRD | Native | | • | • | | | • |
| Ictaluridae | Black Bullhead | BLBL | Native | | | • | | | • |
| | Brown Bullhead | BRBL | Native | | | | | | • |
| | Channel Catfish | CHCT | Native | | | • | • | | |
| Esocidae | Northern Pike | NRPK | Native | • | • | • | • | • | • |
| Osmeridae | Rainbow Smelt | RNSM | Introduced | | • | | • | | • |
| Salmonidae | Cisco | CISC | Native | | • | • | | • | • |
| | Lake Whitefish | LKWH | Native | | • | • | • | • | • |
| Percopsidae | Trout-perch | TRPR | Native | | • | • | • | • | • |
| Gadidae | Burbot | BURB | Native | | • | • | • | • | |
| Cottidae | Mottled Sculpin | MTSC | Native | | | | | • | |
| | Slimy Sculpin | SLSC | Native | | | | | • | |
| | Unid. Sculpin Spp. | - | Native | | • | • | | | |
| Centrarchidae | Rock Bass | RCBS | Native | | • | • | • | | • |
| | Smallmouth Bass | SMBS | Introduced | | • | • | • | • | • |
| | Black Crappie | BLCR | Native | | • | • | | | • |
| Percidae | Yellow Perch | YLPR | Native | | • | • | • | • | • |
| | Logperch | LGPR | Native | | | | | | • |
| | Sauger | SAUG | Native | • | • | • | • | | • |
| | Walleye | WALL | Native | • | • | • | • | • | • |

Notes:

1. Assigned from Stewart and Watkinson (2004).
2. Status under review by Committee on the Status of Endangered Wildlife in Canada (COSEWIC).

Table 5.6-2. 2008-2019 Relative species abundance in standard gang index gill nets in the Pointe du Bois Forebay.

| | | 0% | >0-5% | >5-10% | >10-25% | >25-50% | >50% | | | | | | | |
|-------------|---------|------|-------|--------|---------|---------|------|------|------|------|------|------|------|-------|
| Group | Species | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | Mean |
| Target | LKWH | 0.5% | 1% | 2% | 1% | 0.5% | 0% | 0.3% | 2% | 0.2% | 0.5% | 2% | 1% | 1% |
| | NRPK | 5% | 3% | 5% | 3% | 2% | 7% | 8% | 4% | 5% | 4% | 4% | 3% | 4% |
| | SAUG | 24% | 21% | 13% | 10% | 26% | 18% | 28% | 21% | 16% | 20% | 13% | 17% | 19% |
| | WALL | 15% | 13% | 13% | 8% | 11% | 10% | 10% | 11% | 7% | 10% | 8% | 11% | 11% |
| | WHSC | 38% | 39% | 38% | 32% | 27% | 18% | 22% | 23% | 17% | 27% | 21% | 20% | 27% |
| Lampreys | SLLM | 0.5% | 3% | 5% | 0% | 1% | 0.2% | 2% | 1% | 4% | 2% | 1% | 0.5% | 2% |
| Sturgeon | LKST | 0.2% | 0.2% | 2% | 0.2% | 0% | 0.2% | 0% | 0% | 0% | 0% | 0% | 0% | 0.2% |
| Mooneyes | MOON | 0.5% | 0.7% | 3% | 1% | 11% | 1% | 0.3% | 0% | 0.5% | 2% | 7% | 5% | 3% |
| | GOLD | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Suckers | LNSC | 0.2% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0.02% |
| | SHRD | 4% | 1% | 2% | 0.2% | 0.5% | 0.3% | 0% | 0.2% | 0% | 1% | 1% | 0.5% | 0.9% |
| | SLRD | 0% | 0.2% | 1% | 2% | 0% | 0.5% | 2% | 1% | 0.2% | 1% | 5% | 3% | 1% |
| Catfishes | BLBL | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | BRBL | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | CHCT | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Smelts | RNSM | 0% | 0% | 0% | 0% | 0% | 0.2% | 0% | 0.2% | 0% | 0% | 0% | 0% | 0.0% |
| Coregonids | CISC | 0% | 0.5% | 0.4% | 1% | 2% | 0.2% | 1% | 1% | 0% | 0% | 2% | 1% | 1% |
| Trout-perch | TRPR | 0% | 0% | 0% | 0% | 0% | 0.2% | 0% | 0.2% | 0% | 0% | 0.2% | 0% | 0.1% |
| Codfishes | BURB | 0.5% | 0.5% | 0.4% | 0.3% | 1% | 0.3% | 1% | 1% | 1% | 0% | 1% | 0% | 1% |
| Sunfishes | RCBS | 0% | 2% | 0% | 1% | 0.2% | 1% | 2% | 0.5% | 0.2% | 1% | 0.4% | 2% | 1% |
| | SMBS | 0.2% | 0.2% | 1% | 1% | 1% | 0.6% | 1% | 2% | 1% | 2% | 4% | 2% | 1% |
| | BLCR | 0% | 0% | 0% | 0% | 0% | 0% | 0.7% | 0% | 1% | 1% | 0.4% | 0.5% | 0.3% |
| Perch | YLPR | 11% | 14% | 16% | 40% | 15% | 43% | 22% | 31% | 47% | 27% | 30% | 32% | 27% |

Table 5.6-3. 2008-2019 Relative species abundance in small mesh index gill nets in the Pointe du Bois Forebay.

| | | 0% | >0-5% | >5-10% | >10-25% | >25-50% | >50% | | | | | | | |
|-------------|--------------|------|-------|--------|---------|---------|------|------|------|------|------|------|------|-------|
| Group | Species | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | Mean |
| Sturgeon | LKST | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Mooneyes | MOON | 0% | 0% | 2% | 0% | 0.3% | 0.2% | 0% | 0% | 0% | 0.3% | 2% | 0.2% | 0.4% |
| Minnnows | EMSH | 15% | 0% | 1% | 0% | 0% | 0% | 3% | 0.4% | 1% | 0% | 1% | 0% | 2% |
| | SPSH | 9% | 0% | 14% | 37% | 13% | 26% | 52% | 39% | 16% | 30% | 36% | 46% | 26% |
| Suckers | WHSC | 0% | 0% | 1% | 1% | 1% | 1% | 0% | 1% | 0% | 1% | 1% | 1% | 1% |
| Catfishes | BLBL | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | BRBL | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Pikes | NRPK | 1% | 0% | 0% | 1% | 0.5% | 2% | 1% | 1% | 1% | 0.3% | 0.3% | 0.4% | 0.7% |
| Smelts | RNSM | 0% | 0% | 2% | 0.4% | 0.5% | 0% | 0.4% | 1% | 0% | 0% | 0.3% | 0.2% | 0.4% |
| Coregonids | CISC | 0% | 0% | 0% | 0.2% | 0.2% | 0% | 1% | 2% | 0% | 0.3% | 0% | 0% | 0.3% |
| | LKWH | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Trout-perch | TRPR | 18% | 15% | 28% | 11% | 54% | 13% | 10% | 7% | 17% | 25% | 16% | 13% | 19% |
| Codfishes | BURB | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0.4% | 0% | 0% | 0% | 0% | 0.03% |
| Sculpins | MTSC | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | SLSC | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | Sculpin spp. | 0% | 0% | 0% | 0.2% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0.01% |
| Sunfishes | RCBS | 0% | 0% | 0% | 0% | 0% | 1% | 0.4% | 0% | 1% | 0% | 1% | 0% | 0.2% |
| | SMBS | 0% | 0% | 0% | 0.2% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0.01% |
| | BLCR | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Perch | YLPR | 20% | 0% | 7% | 38% | 14% | 44% | 12% | 19% | 17% | 26% | 23% | 30% | 21% |
| | LGPR | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | SAUG | 26% | 67% | 29% | 5% | 12% | 11% | 19% | 24% | 43% | 18% | 21% | 6% | 23% |
| | WALL | 11% | 18% | 15% | 7% | 4% | 2% | 1% | 4% | 4% | 1% | 1% | 3% | 6% |

Table 5.6-4. 2008-2019 Relative species abundance in standard gang index gill nets in Lac du Bonnet.

| | | 0% | >0-5% | >5-10% | >10-25% | >25-50% | >50% | | | | | | | |
|-------------|---------|------|-------|--------|---------|---------|------|------|------|------|------|------|------|-------|
| Group | Species | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | Mean |
| Target | LKWH | 5% | 4% | 5% | 10% | 9% | 12% | 7% | 10% | 4% | 6% | 3% | 3% | 7% |
| | NRPK | 7% | 3% | 12% | 8% | 12% | 10% | 12% | 10% | 8% | 11% | 8% | 8% | 9% |
| | SAUG | 23% | 33% | 21% | 19% | 18% | 20% | 17% | 20% | 27% | 25% | 22% | 14% | 22% |
| | WALL | 15% | 16% | 17% | 16% | 25% | 12% | 23% | 21% | 14% | 17% | 14% | 15% | 17% |
| | WHSC | 16% | 10% | 11% | 12% | 6% | 11% | 9% | 9% | 9% | 10% | 10% | 13% | 10% |
| Lampreys | SLLM | 0% | 0% | 0% | 0% | 0% | 0.2% | 0% | 0% | 0% | 0% | 0% | 0.0% | 0.02% |
| Sturgeon | LKST | 2% | 1% | 0.3% | 0.5% | 0.4% | 1% | 0% | 3% | 1% | 0% | 3% | 4% | 1% |
| Mooneyes | MOON | 3% | 1% | 9% | 5% | 0% | 2% | 8% | 1% | 1% | 7% | 1% | 1% | 3% |
| | GOLD | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Suckers | LNSC | 0.3% | 0% | 0% | 0% | 6% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 1% |
| | SHRD | 3% | 2% | 7% | 4% | 5% | 7% | 3% | 4% | 3% | 1% | 4% | 4% | 4% |
| | SLRD | 3% | 5% | 6% | 5% | 1% | 0% | 4% | 3% | 2% | 2% | 4% | 3% | 3% |
| Catfishes | BLBL | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | BRBL | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | CHCT | 0.3% | 2% | 0% | 0% | 0% | 0.2% | 0% | 1% | 2% | 0% | 0.2% | 1% | 0.5% |
| Smelts | RNSM | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Coregonids | CISC | 7% | 7% | 3% | 12% | 9% | 8% | 5% | 2% | 7% | 3% | 2% | 5% | 6% |
| Trout-perch | TRPR | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0.3% | 0% | 0% | 0.03% |
| Codfishes | BURB | 0% | 0% | 1% | 0% | 0% | 0.2% | 0% | 0% | 0% | 0.3% | 0% | 0% | 0.1% |
| Sunfishes | RCBS | 2% | 1% | 1% | 1% | 1% | 2% | 2% | 3% | 3% | 1% | 1% | 1% | 2% |
| | SMBS | 1% | 1% | 0.3% | 1% | 0.4% | 1% | 1% | 2% | 0% | 1% | 1% | 1% | 1% |
| | BLCR | 0% | 0% | 0% | 0.2% | 0.2% | 1% | 2% | 0% | 5% | 7% | 6% | 11% | 3% |
| Perch | YLPR | 11% | 15% | 6% | 6% | 7% | 11% | 6% | 11% | 14% | 9% | 20% | 15% | 11% |

Table 5.6-5. 2008-2019 Relative species abundance in small mesh index gill nets in Lac du Bonnet.

| | | 0% | >0-5% | >5-10% | >10-25% | >25-50% | >50% | | | | | | | |
|-------------|--------------|------|-------|--------|---------|---------|------|------|------|------|------|------|------|-------|
| Group | Species | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | Mean |
| Sturgeon | LKST | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Mooneyes | MOON | 1% | 0% | 1% | 0% | 0% | 0% | 1% | 0% | 0% | 0% | 0.2% | 0% | 0.2% |
| Minnnows | EMSH | 7% | 1% | 5% | 19% | 0% | 0% | 14% | 0% | 0% | 9% | 1% | 6% | 5% |
| | SPSH | 22% | 1% | 17% | 16% | 41% | 49% | 52% | 2% | 3% | 29% | 3% | 53% | 24% |
| Suckers | WHSC | 0% | 0% | 0% | 0.4% | 0.2% | 0.1% | 0% | 0.4% | 0% | 0% | 0% | 0% | 0.1% |
| Catfishes | BLBL | 0% | 0% | 0% | 0% | 0.2% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0.02% |
| | BRBL | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Pikes | NRPK | 1% | 1% | 2% | 0.3% | 0% | 0.2% | 1% | 0.3% | 0% | 0% | 0.2% | 0% | 0.4% |
| Smelts | RNSM | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Coregonids | CISC | 0% | 2% | 2% | 0% | 0% | 0% | 2% | 0% | 7% | 0% | 4% | 1% | 1% |
| | LKWH | 0% | 0% | 1% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0.1% |
| Trout-perch | TRPR | 8% | 7% | 19% | 0.3% | 0% | 0.5% | 5% | 1% | 32% | 9% | 12% | 1% | 8% |
| Codfishes | BURB | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Sculpins | MTSC | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | SLSC | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | Sculpin spp. | 0% | 0% | 0% | 0.1% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0.01% |
| Sunfishes | RCBS | 0% | 0% | 0% | 0% | 0% | 0.5% | 0% | 0.3% | 0% | 0% | 0.2% | 0% | 0.1% |
| | SMBS | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0.1% | 0% | 0% | 0.4% | 0% | 0.04% |
| | BLCR | 0% | 0% | 0% | 0% | 0.2% | 0.1% | 4% | 26% | 9% | 37% | 10% | 11% | 8% |
| Perch | YLPR | 37% | 25% | 34% | 61% | 57% | 47% | 17% | 68% | 3% | 4% | 58% | 21% | 36% |
| | LGPR | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | SAUG | 16% | 52% | 11% | 2% | 1% | 2% | 2% | 1% | 28% | 8% | 7% | 8% | 11% |
| | WALL | 10% | 12% | 9% | 1% | 1% | 2% | 3% | 1% | 18% | 5% | 5% | 1% | 6% |

Table 5.6-6. 2008-2019 Relative species abundance in standard gang index gill nets in the Pine Falls Forebay.

| | | 0% | >0-5% | >5-10% | >10-25% | >25-50% | >50% |
|-------------|---------|------|-------|--------|---------|---------|------|
| Group | Species | 2011 | 2014 | 2017 | Mean | | |
| Target | LKWH | 0% | 1% | 1% | 1% | | |
| | NRPK | 3% | 2% | 4% | 3% | | |
| | SAUG | 5% | 7% | 10% | 7% | | |
| | WALL | 11% | 16% | 11% | 13% | | |
| | WHSC | 8% | 8% | 11% | 9% | | |
| Lampreys | SLLM | 0% | 0% | 0% | 0% | | |
| Sturgeon | LKST | 0% | 1% | 0.4% | 0.3% | | |
| Mooneyes | MOON | 6% | 0% | 9% | 5% | | |
| | GOLD | 0% | 1% | 0% | 0.2% | | |
| Suckers | LNSC | 0% | 0% | 0% | 0% | | |
| | SHRD | 1% | 0.3% | 1% | 1% | | |
| | SLRD | 0% | 0% | 0% | 0% | | |
| Catfishes | BLBL | 0% | 0% | 0% | 0% | | |
| | BRBL | 0% | 0% | 0% | 0% | | |
| | CHCT | 60% | 63% | 50% | 57% | | |
| Smelts | RNSM | 0% | 0% | 0% | 0% | | |
| Coregonids | CISC | 0% | 0% | 0% | 0% | | |
| Trout-perch | TRPR | 0% | 0% | 0% | 0% | | |
| Codfishes | BURB | 1% | 0.3% | 0% | 0.3% | | |
| Sunfishes | RCBS | 1% | 0% | 0.4% | 0.5% | | |
| | SMBS | 0% | 1% | 0% | 0.2% | | |
| | BLCR | 0% | 0% | 0% | 0% | | |
| Perch | YLPR | 4% | 2% | 2% | 3% | | |

Table 5.6-7. 2008-2019 Relative species abundance in small mesh index gill nets in the Pine Falls Forebay.

| | | 0% | >0-5% | >5-10% | >10-25% | >25-50% | >50% |
|-------------|--------------|------|-------|--------|---------|---------|------|
| Group | Species | 2011 | 2014 | 2017 | Mean | | |
| Sturgeon | LKST | 10% | 0% | 0% | 3% | | |
| Mooneyes | MOON | 0% | 0% | 7% | 2% | | |
| Minnows | EMSH | 0% | 0% | 0% | 0% | | |
| | SPSH | 20% | 0% | 11% | 10% | | |
| Suckers | WHSC | 0% | 0% | 0% | 0% | | |
| Catfishes | BLBL | 0% | 0% | 0% | 0% | | |
| | BRBL | 0% | 0% | 0% | 0% | | |
| Pikes | NRPK | 0% | 5% | 0% | 2% | | |
| Smelts | RNSM | 0% | 5% | 0% | 2% | | |
| Coregonids | CISC | 0% | 0% | 0% | 0% | | |
| | LKWH | 0% | 0% | 0% | 0% | | |
| Trout-perch | TRPR | 20% | 70% | 44% | 45% | | |
| Codfishes | BURB | 0% | 0% | 0% | 0% | | |
| Sculpins | MTSC | 0% | 0% | 0% | 0% | | |
| | SLSC | 0% | 0% | 0% | 0% | | |
| | Sculpin spp. | 0% | 0% | 0% | 0% | | |
| Sunfishes | RCBS | 0% | 0% | 0% | 0% | | |
| | SMBS | 0% | 0% | 0% | 0% | | |
| | BLCR | 0% | 0% | 0% | 0% | | |
| Perch | YLPR | 10% | 0% | 0% | 3% | | |
| | LGPR | 0% | 0% | 0% | 0% | | |
| | SAUG | 0% | 15% | 26% | 14% | | |
| | WALL | 40% | 5% | 11% | 19% | | |

Table 5.6-8. 2008-2019 Relative species abundance in standard gang index gill nets in Manigotagan Lake.

| | | 0% | >0-5% | >5-10% | >10-25% | >25-50% | >50% | | | | | | | |
|-------------|---------|------|-------|--------|---------|---------|------|------|------|------|------|------|------|-------|
| Group | Species | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | Mean |
| Target | LKWH | 10% | 15% | 2% | 5% | 6% | 2% | 10% | 10% | 7% | 6% | 6% | 4% | 7% |
| | NRPK | 4% | 5% | 4% | 7% | 8% | 8% | 7% | 4% | 7% | 7% | 6% | 5% | 6% |
| | SAUG | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | WALL | 38% | 34% | 42% | 34% | 36% | 26% | 29% | 22% | 24% | 30% | 29% | 38% | 32% |
| | WHSC | 8% | 7% | 4% | 4% | 8% | 7% | 6% | 5% | 8% | 8% | 12% | 17% | 8% |
| Lampreys | SLLM | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Sturgeon | LKST | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Mooneyes | MOON | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | GOLD | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Suckers | LNSC | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | SHRD | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | SLRD | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Catfishes | BLBL | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | BRBL | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | CHCT | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Smelts | RNSM | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Coregonids | CISC | 27% | 30% | 41% | 36% | 39% | 36% | 34% | 50% | 47% | 36% | 33% | 28% | 36% |
| Trout-perch | TRPR | 0% | 0% | 0.4% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0.03% |
| Codfishes | BURB | 8% | 6% | 1% | 12% | 0.3% | 10% | 9% | 4% | 2% | 5% | 4% | 1% | 5% |
| Sunfishes | RCBS | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | SMBS | 1% | 0.4% | 1% | 0% | 0% | 4% | 1% | 1% | 1% | 4% | 4% | 5% | 2% |
| | BLCR | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Perch | YLPR | 5% | 3% | 5% | 4% | 3% | 7% | 4% | 4% | 4% | 3% | 6% | 2% | 4% |

Table 5.6-9. 2008-2019 Relative species abundance in small mesh index gill nets in Manigotagan Lake.

| | | 0% | >0-5% | >5-10% | >10-25% | >25-50% | >50% | | | | | | | |
|-------------|--------------|------|-------|--------|---------|---------|------|------|------|------|------|------|------|------|
| Group | Species | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | Mean |
| Sturgeon | LKST | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Mooneyes | MOON | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Minnnows | EMSH | 0% | 0% | 0% | 0% | 4% | 0% | 0% | 26% | 50% | 50% | 67% | 24% | 18% |
| | SPSH | 0% | 0% | 0% | 0% | 1% | 0% | 4% | 5% | 0% | 0% | 9% | 20% | 3% |
| Suckers | WHSC | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 1% | 0% | 0.1% |
| Catfishes | BLBL | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | BRBL | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Pikes | NRPK | 0% | 0% | 0% | 0% | 1% | 9% | 12% | 3% | 6% | 0% | 0% | 0% | 3% |
| Smelts | RNSM | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Coregonids | CISC | 0% | 20% | 15% | 60% | 8% | 11% | 36% | 3% | 0% | 10% | 1% | 0% | 14% |
| | LKWH | 0% | 0% | 3% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0.2% |
| Trout-perch | TRPR | 0% | 0% | 8% | 0% | 8% | 36% | 16% | 13% | 13% | 0% | 7% | 16% | 10% |
| Codfishes | BURB | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Sculpins | MTSC | 0% | 0% | 0% | 0% | 1% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0.1% |
| | SLSC | 0% | 0% | 0% | 0% | 1% | 0% | 0% | 3% | 0% | 0% | 0% | 0% | 0.3% |
| | Sculpin spp. | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 10% | 0% | 0% | 1% |
| Sunfishes | RCBS | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | SMBS | 0% | 20% | 0% | 0% | 3% | 0% | 0% | 0% | 0% | 0% | 2% | 12% | 3% |
| | BLCR | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| Perch | YLPR | 0% | 0% | 0% | 0% | 31% | 4% | 0% | 0% | 0% | 0% | 2% | 8% | 4% |
| | LGPR | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | SAUG | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% | 0% |
| | WALL | 100% | 60% | 75% | 40% | 43% | 41% | 32% | 49% | 31% | 30% | 12% | 20% | 44% |

Table 5.6-10. 2008-2019 Relative species abundance in standard gang index gill nets in Eaglenest Lake.

| | | 0% | >0-5% | >5-10% | >10-25% | >25-50% | >50% |
|-------------|---------|------|-------|--------|---------|---------|------|
| Group | Species | 2010 | 2013 | 2016 | 2019 | Mean | |
| Target | LKWH | 2% | 2% | 5% | 1% | 2% | |
| | NRPK | 7% | 9% | 9% | 10% | 9% | |
| | SAUG | 13% | 10% | 10% | 10% | 11% | |
| | WALL | 21% | 16% | 11% | 18% | 17% | |
| | WHSC | 17% | 11% | 13% | 13% | 14% | |
| Lampreys | SLLM | 3% | 0% | 0% | 0.3% | 0.8% | |
| Sturgeon | LKST | 0% | 0% | 0% | 0% | 0% | |
| Mooneyes | MOON | 7% | 6% | 2% | 9% | 6% | |
| | GOLD | 0% | 0% | 1% | 0% | 0.2% | |
| Suckers | LNSC | 0% | 0% | 3% | 0% | 1% | |
| | SHRD | 0.4% | 0.2% | 0% | 0.2% | 0.2% | |
| | SLRD | 1% | 1% | 1% | 2% | 1% | |
| Catfishes | BLBL | 0% | 2% | 0% | 0% | 1% | |
| | BRBL | 0% | 0% | 17% | 3% | 5% | |
| | CHCT | 0% | 0% | 0% | 0% | 0% | |
| Smelts | RNSM | 0.2% | 0% | 0% | 0% | 0.1% | |
| Coregonids | CISC | 0.2% | 3% | 1% | 0.5% | 1.1% | |
| Trout-perch | TRPR | 0% | 0% | 0% | 0% | 0% | |
| Codfishes | BURB | 0% | 0% | 0% | 0% | 0% | |
| Sunfishes | RCBS | 1% | 1% | 1% | 1% | 1% | |
| | SMBS | 1% | 2% | 1% | 3% | 2% | |
| | BLCR | 0.2% | 0.2% | 0% | 0.2% | 0.1% | |
| Perch | YLPR | 25% | 37% | 26% | 29% | 29% | |

Table 5.6-11. 2008-2019 Relative species abundance in small mesh index gill nets in Eaglenest Lake.

| | | 0% | >0-5% | >5-10% | >10-25% | >25-50% | >50% |
|-------------|--------------|------|-------|--------|---------|---------|------|
| Group | Species | 2010 | 2013 | 2016 | 2019 | Mean | |
| Sturgeon | LKST | 0% | 0% | 0% | 0% | 0% | |
| Mooneyes | MOON | 1% | 0.3% | 1% | 0.2% | 0.4% | |
| Minnows | EMSH | 0% | 0% | 0% | 0.2% | 0.1% | |
| | SPSH | 7% | 8% | 26% | 24% | 16% | |
| Suckers | WHSC | 1% | 0.3% | 0% | 0.2% | 0.3% | |
| Catfishes | BLBL | 0% | 0% | 0% | 0% | 0% | |
| | BRBL | 0% | 0% | 1% | 0% | 0.3% | |
| Pikes | NRPK | 3% | 1% | 1% | 1% | 1% | |
| Smelts | RNSM | 0% | 0% | 0% | 0% | 0% | |
| Coregonids | CISC | 0% | 0% | 1% | 0% | 0.2% | |
| | LKWH | 0% | 0% | 0% | 0% | 0% | |
| Trout-perch | TRPR | 46% | 18% | 18% | 5% | 22% | |
| Codfishes | BURB | 0% | 0% | 0% | 0% | 0% | |
| Sculpins | MTSC | 0% | 0% | 0% | 0% | 0% | |
| | SLSC | 0% | 0% | 0% | 0% | 0% | |
| | Sculpin spp. | 0% | 0% | 0% | 0% | 0% | |
| Sunfishes | RCBS | 0% | 0% | 0% | 0.2% | 0.1% | |
| | SMBS | 0% | 0% | 0% | 0% | 0% | |
| | BLCR | 0% | 0% | 0% | 0% | 0% | |
| Perch | YLPR | 12% | 44% | 31% | 51% | 34% | |
| | LGPR | 0% | 0% | 0% | 0.2% | 0.1% | |
| | SAUG | 18% | 25% | 19% | 15% | 19% | |
| | WALL | 12% | 4% | 3% | 3% | 6% | |

5.6.2 HILL'S EFFECTIVE RICHNESS

5.6.2.1 ON-SYSTEM SITES

ANNUAL SITES

Pointe du Bois Forebay

The Hill's effective species richness over the 12 years of monitoring ranged from a low of 5.8 in 2013 to a high of 9.9 species in 2017 (Table 5.6-12; Figure 5.6-1).

The overall mean Hill's index value was 7.4, the median was 7.2, and the IQR was 6.3-8.1 species (Figure 5.6-1). The annual mean Hill's index value fell within the overall IQR except in 2009, 2011, and 2013 when it was below the IQR and in 2010, 2017, and 2018 when it was above the IQR.

Lac du Bonnet

The Hill's effective species richness over the 12 years of monitoring ranged from a low of 5.5 in 2013 to a high of 11.8 species in 2014 (Table 5.6-12; Figure 5.6-1).

The overall mean Hill's index value was 8.7, the median was 8.8, and the IQR was 7.1-10.3 species (Figure 5.6-1). The annual mean Hill's index value fell within the overall IQR except in 2011, 2013, and 2015 when it was below the IQR and in 2008, 2010, and 2014 when it was above the IQR.

ROTATIONAL SITES

Pine Falls Forebay

The Hill's effective species richness over the three years of monitoring ranged from a low of 4.3 in 2014 to a high of 6.2 species in 2017 (Table 5.6-12; Figure 5.6-1).

The overall mean Hill's index value was 5.1, the median was 4.8, and the IQR was 4.5-5.5 species (Figure 5.6-1). The annual mean Hill's index value was below the IQR in 2014 and was above the IQR in 2017.

5.6.2.2 OFF-SYSTEM SITES

ANNUAL SITES

Manigotagan Lake

The Hill's effective species richness over the 12 years of monitoring ranged from a low of 3.7 in 2010 to a high of 7.3 species in 2018 (Table 5.6-12; Figure 5.6-1).

The overall mean Hill's index value was 5.4, the median was 5.3, and the IQR was 5.1-5.7 species (Figure 5.6-1). The annual mean Hill's index value fell within the overall IQR except in 2010, 2011, and 2016 when it was below the IQR and in 2013 and 2018 when it was above the IQR.

ROTATIONAL SITES

Eaglenest Lake

The Hill's effective species richness over the four years of monitoring ranged from a low of 7.2 in 2013 to a high of 9.7 species in 2016 (Table 5.6-12; Figure 5.6-1).

The overall mean Hill's index value was 8.3, the median was 8.0, and the IQR was 7.4-8.9 species (Figure 5.6-1). The annual mean Hill's index value fell within the overall IQR except in 2013 when it was below the IQR and in 2016 when it was above the IQR.

Table 5.6-12. 2008-2019 Hill's effective species richness.

| Waterbody | Year | n_F^1 | n_{spp}^2 | Value |
|-----------|------|---------|-------------|-------|
| PDB | 2008 | 537 | 16 | 6.9 |
| | 2009 | 444 | 16 | 5.9 |
| | 2010 | 364 | 18 | 8.6 |
| | 2011 | 1181 | 18 | 6.2 |
| | 2012 | 1018 | 16 | 7.0 |
| | 2013 | 1204 | 17 | 5.8 |
| | 2014 | 522 | 18 | 7.9 |
| | 2015 | 656 | 17 | 7.4 |
| | 2016 | 598 | 16 | 6.3 |
| | 2017 | 765 | 16 | 9.9 |
| | 2018 | 775 | 19 | 9.0 |
| LDB | 2008 | 438 | 18 | 10.7 |
| | 2009 | 451 | 17 | 7.4 |
| | 2010 | 421 | 17 | 11.4 |
| | 2011 | 1543 | 18 | 6.4 |
| | 2012 | 1041 | 16 | 7.3 |
| | 2013 | 1712 | 18 | 5.5 |
| | 2014 | 605 | 16 | 11.8 |
| | 2015 | 1105 | 17 | 5.9 |
| | 2016 | 481 | 16 | 10.0 |
| | 2017 | 631 | 17 | 10.2 |
| | 2018 | 940 | 18 | 8.0 |
| PFF | 2011 | 204 | 13 | 4.8 |
| | 2014 | 332 | 14 | 4.3 |
| | 2017 | 283 | 13 | 6.2 |

Table 5.6-12. continued.

| Waterbody | Year | n _F ¹ | n _{spp} ² | Value |
|-----------|------|-----------------------------|-------------------------------|-------|
| MANIG | 2008 | 356 | 8 | 5.1 |
| | 2009 | 286 | 8 | 5.2 |
| | 2010 | 311 | 9 | 3.7 |
| | 2011 | 294 | 7 | 4.7 |
| | 2012 | 432 | 13 | 5.4 |
| | 2013 | 350 | 9 | 6.1 |
| | 2014 | 391 | 10 | 5.7 |
| | 2015 | 476 | 12 | 5.1 |
| | 2016 | 453 | 10 | 5.0 |
| | 2017 | 397 | 10 | 5.7 |
| | 2018 | 581 | 11 | 7.3 |
| 2019 | 333 | 11 | 5.7 | |
| EAGLE | 2010 | 668 | 17 | 8.6 |
| | 2013 | 910 | 16 | 7.2 |
| | 2016 | 749 | 16 | 9.7 |
| | 2019 | 1090 | 19 | 7.5 |

Notes:

1. n_F = number of fish caught in standard and small mesh gill nets.
2. n_{spp} = number of species caught in standard and small mesh gill nets.

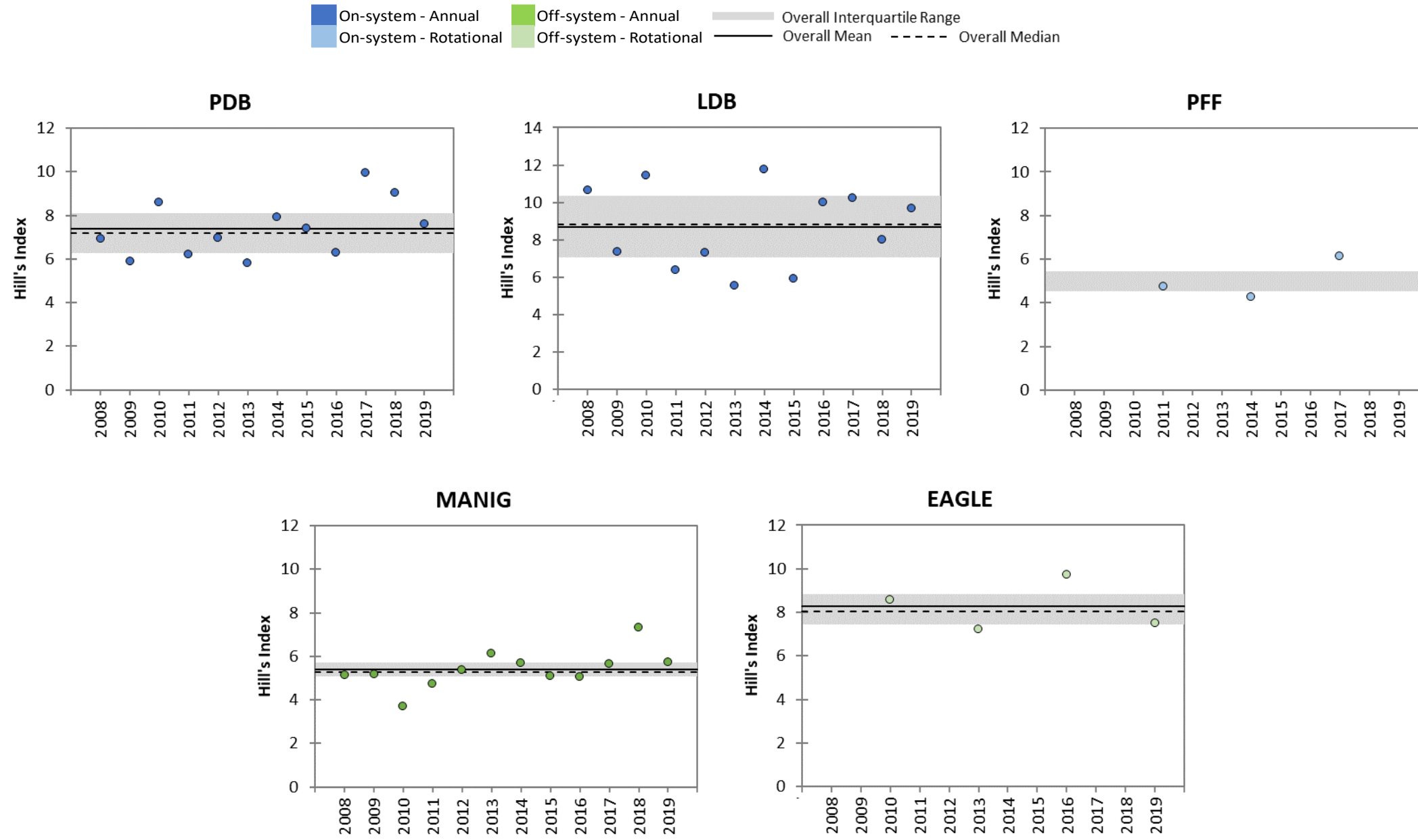


Figure 5.6-1. 2008-2019 Hill's effective species richness.

APPENDIX 5-1. GILLNETTING SITE INFORMATION AND LOCATIONS

The following is a summary of sampling locations over the 12 years of monitoring in the Winnipeg River Region:

Pointe du Bois Forebay

- Gill nets were set at the target locations in all 12 years at all but three locations:
 - GN-04 was not set in 2009 and GN/SN-12 was not set in 2014 due to a combination of high water velocities and inconsistent depths at these sites.
 - GN/SN-16 was added as a target location in 2009.

Lac du Bonnet

- Gill nets were set at the target locations in all 12 years except in 2017 when fishing could not be completed at GN-09 and GN-10 because of a boat malfunction.

Pine Falls Forebay

- Gill nets were set at the target locations in all four years.

Manigotagan Lake

- Gill nets were set at the target locations in all 12 years.

Eaglenest Lake

- Gill nets were set at the target locations in all four years at all but two locations:
 - Site GN-07 was not set in 2013 as gill nets were set at an alternative location to try and increase the amount of deep-water habitat covered.
 - GN/SN-14 was added as a target location after the Pilot Program starting in 2013.

Table A5-1-1. 2008-2019 Set information for gillnetting sites.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|-----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| PDB | GN-01 | 15 | 326434 | 5579359 | 14-Jul-08 | 22.0 | 7.3 | 7.9 | 18.0 |
| | GN-03 | 15 | 324982 | 5580785 | 14-Jul-08 | 26.7 | 11.5 | 6.2 | 19.0 |
| | GN-04 | 15 | 323895 | 5579878 | 14-Jul-08 | 27.5 | 18.0 | 6.9 | 18.0 |
| | GN-05 | 15 | 323540 | 5577759 | 15-Jul-08 | 24.7 | 14.9 | 14.7 | 20.0 |
| | GN-07 | 15 | 322553 | 5578660 | 16-Jul-08 | 23.8 | 5.5 | 6.1 | 18.0 |
| | GN-10 | 15 | 322201 | 5577408 | 17-Jul-08 | 22.4 | 12.8 | 13.5 | 18.0 |
| | GN-11 | 15 | 320098 | 5578585 | 17-Jul-08 | 26.2 | 4.5 | 5.6 | 20.0 |
| | GN-12 | 15 | 319766 | 5577488 | 17-Jul-08 | 20.4 | 11.4 | 16.6 | 19.0 |
| | GN-15 | 15 | 319474 | 5575530 | 19-Jul-08 | 23.3 | 10.1 | 7.5 | 19.0 |
| | SN-01 | 15 | 326409 | 5579338 | 14-Jul-08 | 20.9 | 5.8 | 7.3 | 18.0 |
| | SN-05 | 15 | 323569 | 5577793 | 15-Jul-08 | 25.6 | 10.8 | 14.9 | 20.0 |
| | SN-12 | 15 | 319788 | 5577462 | 17-Jul-08 | 21.2 | - | - | 19.0 |
| | GN-01 | 15 | 326366 | 5579325 | 14-Jul-09 | 46.3 | 7.2 | 7.9 | 17.0 |
| | GN-03 | 15 | 324972 | 5580891 | 17-Jul-09 | 25.6 | 6.0 | 5.0 | 16.0 |
| | GN-05 | 15 | 323487 | 5577708 | 17-Jul-09 | 24.8 | 15.4 | 15.2 | 16.0 |
| | GN-07 | 15 | 322561 | 5578643 | 13-Jul-09 | 22.3 | 6.6 | 5.1 | 20.5 |
| | GN-10 | 15 | 322191 | 5577415 | 18-Jul-09 | 25.6 | 12.5 | 12.6 | 16.5 |
| | GN-11 | 15 | 320112 | 5578578 | 15-Jul-09 | 48.8 | 5.1 | 4.5 | 17.0 |
| | GN-12 | 15 | 319738 | 5577533 | 18-Jul-09 | 24.7 | 11.0 | 12.5 | 16.0 |
| | GN-15 | 15 | 319493 | 5575532 | 19-Jul-09 | 22.8 | 6.9 | 7.4 | 16.5 |
| | GN-16 | 15 | 323786 | 5581047 | 20-Jul-09 | 26.9 | 4.8 | 4.6 | 16.5 |
| | SN-01 | 15 | 326386 | 5579343 | 14-Jul-09 | 46.3 | 7.1 | 7.2 | 17.0 |
| | SN-05 | 15 | 323529 | 5577737 | 17-Jul-09 | 24.8 | 14.2 | 15.4 | 16.0 |
| | SN-12 | 15 | 319755 | 5577493 | 18-Jul-09 | 24.7 | 11.4 | 11.0 | 16.0 |
| SN-16 | 15 | 323839 | 5581029 | 20-Jul-09 | 26.9 | 4.5 | 4.8 | 16.5 | |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|-----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| PDB | GN-01 | 15 | 326330 | 5579312 | 13-Jul-10 | 25.1 | 7.6 | 6.5 | 19.0 |
| | GN-03 | 15 | 324805 | 5580920 | 15-Jul-10 | 27.3 | 5.6 | 6.9 | 19.0 |
| | GN-04 | 15 | 324032 | 5579877 | 13-Jul-10 | 24.4 | 19.6 | 15.8 | 19.0 |
| | GN-05 | 15 | 323449 | 5577738 | 14-Jul-10 | 29.8 | 15.8 | 16.0 | 19.0 |
| | GN-07 | 15 | 322557 | 5578674 | 16-Jul-10 | 26.2 | 6.5 | 6.2 | 19.0 |
| | GN-10 | 15 | 322233 | 5577384 | 16-Jul-10 | 26.6 | 12.3 | 13.6 | 19.0 |
| | GN-11 | 15 | 320126 | 5578551 | 17-Jul-10 | 24.0 | 4.7 | 6.0 | 19.0 |
| | GN-12 | 15 | 319643 | 5577488 | 17-Jul-10 | 24.1 | 13.0 | 9.7 | 19.0 |
| | GN-15 | 15 | 319538 | 5575533 | 17-Jul-10 | 24.0 | 6.9 | 5.9 | 19.0 |
| | GN-16 | 15 | 324027 | 5581109 | 15-Jul-10 | 28.6 | 5.4 | 4.9 | 19.0 |
| | SN-01 | 15 | 326300 | 5579283 | 13-Jul-10 | 25.1 | 7.9 | 7.6 | 19.0 |
| | SN-05 | 15 | 323495 | 5577776 | 14-Jul-10 | 29.8 | 12.8 | 15.8 | 19.0 |
| | SN-12 | 15 | 319610 | 5577472 | 17-Jul-10 | 24.1 | 12.1 | 13.0 | 19.0 |
| | SN-16 | 15 | 324067 | 5581119 | 15-Jul-10 | 28.6 | 5.2 | 5.4 | 19.0 |
| | GN-01 | 15 | 326500 | 5579322 | 8-Jul-11 | 21.6 | 6.4 | 7.4 | 22.0 |
| | GN-03 | 15 | 324935 | 5580917 | 8-Jul-11 | 23.3 | 6.2 | 4.3 | 22.5 |
| | GN-04 | 15 | 323902 | 5579858 | 9-Jul-11 | 26.3 | 22.5 | 10.2 | 20.0 |
| | GN-05 | 15 | 323450 | 5577817 | 12-Jul-11 | 28.8 | 8.8 | 21.5 | 21.0 |
| | GN-07 | 15 | 322595 | 5578555 | 9-Jul-11 | 29.0 | 6.6 | 6.4 | 20.0 |
| | GN-10 | 15 | 322220 | 5577494 | 13-Jul-11 | 26.8 | 11.1 | 10.8 | 21.0 |
| | GN-11 | 15 | 320114 | 5578570 | 12-Jul-11 | 26.7 | 3.8 | 5.5 | 21.0 |
| | GN-12 | 15 | 319733 | 5577542 | 13-Jul-11 | 26.7 | 9.1 | 10.3 | 21.0 |
| | GN-15 | 15 | 319513 | 5575527 | 13-Jul-11 | 25.0 | 6.5 | 3.7 | 21.0 |
| | GN-16 | 15 | 323983 | 5581088 | 10-Jul-11 | 26.5 | 5.0 | 5.4 | 20.5 |
| SN-01 | 15 | 326333 | 5579342 | 8-Jul-11 | 21.6 | 4.9 | 6.4 | 22.0 | |
| SN-05 | 15 | 323486 | 5577827 | 12-Jul-11 | 28.8 | 9.1 | 8.8 | 21.0 | |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|-----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| PDB | SN-12 | 15 | 319766 | 5577565 | 13-Jul-11 | 26.7 | 8.8 | 9.1 | 21.0 |
| | SN-16 | 15 | 323955 | 5581096 | 10-Jul-11 | 26.5 | 4.9 | 5.0 | 20.5 |
| | GN-01 | 15 | 326386 | 5579353 | 15-Jul-12 | 25.2 | 6.9 | 8.0 | 22.0 |
| | GN-03 | 15 | 324862 | 5580941 | 17-Jul-12 | 25.0 | 6.2 | 4.5 | 21.0 |
| | GN-04 | 15 | 323975 | 5579923 | 21-Jul-12 | 24.8 | 19.6 | 22.2 | 22.5 |
| | GN-05 | 15 | 323378 | 5577702 | 22-Jul-12 | 24.5 | 17.4 | 14.1 | 22.5 |
| | GN-07 | 15 | 322530 | 5578696 | 17-Jul-12 | 24.1 | 4.7 | 5.3 | 22.0 |
| | GN-10 | 15 | 322142 | 5577503 | 21-Jul-12 | 25.9 | 13.1 | 12.5 | 23.0 |
| | GN-11 | 15 | 320170 | 5578508 | 15-Jul-12 | 24.3 | 5.0 | 5.6 | 21.5 |
| | GN-12 | 15 | 319627 | 5577501 | 20-Jul-12 | 26.3 | 14.2 | 10.1 | 22.0 |
| | GN-15 | 15 | 319609 | 5575621 | 20-Jul-12 | 27.8 | 6.8 | 6.8 | 22.0 |
| | GN-16 | 15 | 324065 | 5581123 | 16-Jul-12 | 24.3 | 5.2 | 5.0 | 21.0 |
| | SN-01 | 15 | 326418 | 5579376 | 15-Jul-12 | 26.3 | 5.3 | 6.9 | 22.0 |
| | SN-05 | 15 | 323354 | 5577686 | 22-Jul-12 | 24.0 | 16.4 | 17.4 | 22.5 |
| | SN-12 | 15 | 319627 | 5577501 | 20-Jul-12 | 26.6 | 13.1 | 14.2 | 22.0 |
| | SN-16 | 15 | 324104 | 5581139 | 16-Jul-12 | 24.8 | 5.1 | 5.2 | 21.0 |
| | GN-01 | 15 | 326306 | 5579301 | 19-Jul-13 | 24.8 | 7.5 | 6.8 | 21.5 |
| | GN-03 | 15 | 324850 | 5580959 | 16-Jul-13 | 27.8 | 5.5 | 7.5 | 22.5 |
| | GN-04 | 15 | 324005 | 5579797 | 20-Jul-13 | 24.3 | 21.0 | 21.8 | 21.5 |
| | GN-05 | 15 | 323480 | 5577771 | 17-Jul-13 | 29.3 | 14.1 | 16.7 | 21.5 |
| GN-07 | 15 | 322629 | 5578571 | 15-Jul-13 | 24.7 | 6.9 | 5.2 | 23.0 | |
| GN-10 | 15 | 322220 | 5577404 | 19-Jul-13 | 25.8 | 12.1 | 13.5 | 22.0 | |
| GN-11 | 15 | 320115 | 5578559 | 18-Jul-13 | 27.9 | 4.1 | 5.6 | 21.0 | |
| GN-12 | 15 | 319624 | 5577502 | 18-Jul-13 | 28.9 | 14.2 | 9.0 | 21.5 | |
| GN-15 | 15 | 319669 | 5575457 | 20-Jul-13 | 24.0 | 4.9 | 6.1 | 21.0 | |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|-----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| PDB | GN-16 | 15 | 324012 | 5581115 | 16-Jul-13 | 30.1 | 5.0 | 4.6 | 21.5 |
| | SN-01 | 15 | 326290 | 5579280 | 19-Jul-13 | 24.8 | 7.9 | 7.5 | 21.5 |
| | SN-05 | 15 | 323503 | 5577789 | 17-Jul-13 | 29.3 | 10.3 | 14.1 | 21.5 |
| | SN-12 | 15 | 319598 | 5577482 | 18-Jul-13 | 28.9 | 12.5 | 14.2 | 21.5 |
| | SN-16 | 15 | 324045 | 5581119 | 16-Jul-13 | 30.1 | 5.0 | 5.0 | 21.5 |
| | GN-01 | 15 | 326413 | 5579364 | 24-Jul-14 | 23.2 | 8.8 | 6.9 | 19.5 |
| | GN-03 | 15 | 324981 | 5580853 | 21-Jul-14 | 24.3 | 5.8 | 5.9 | 20.0 |
| | GN-04 | 15 | 323956 | 5579893 | 20-Jul-14 | 24.0 | 23.2 | 20.2 | 18.5 |
| | GN-05 | 15 | 323328 | 5577731 | 22-Jul-14 | 24.0 | 15.2 | 17.4 | 19.0 |
| | GN-07 | 15 | 322624 | 5578570 | 17-Jul-14 | 21.2 | 5.6 | 7.3 | 18.0 |
| | GN-10 | 15 | 322131 | 5577475 | 15-Jul-14 | 20.8 | 12.1 | 13.2 | 18.5 |
| | GN-11 | 15 | 320207 | 5578469 | 18-Jul-14 | 23.6 | 6.0 | 5.4 | 18.5 |
| | GN-15 | 15 | 319651 | 5575446 | 25-Jul-14 | 23.9 | 6.6 | 4.9 | 21.0 |
| | GN-16 | 15 | 323880 | 5581093 | 19-Jul-14 | 26.1 | 5.3 | 4.8 | 18.0 |
| | SN-01 | 15 | 326258 | 5579292 | 24-Jul-14 | 23.2 | 9.8 | 8.8 | 19.5 |
| | SN-05 | 15 | 323458 | 5577783 | 22-Jul-14 | 24.0 | 12.1 | 15.2 | 19.0 |
| | SN-16 | 15 | 324024 | 5581115 | 19-Jul-14 | 26.1 | 5.2 | 5.3 | 18.0 |
| | GN-01 | 15 | 326320 | 5579283 | 15-Jul-15 | 24.7 | 7.6 | 6.5 | 19.5 |
| | GN-03 | 15 | 324861 | 5580951 | 15-Jul-15 | 25.4 | 5.8 | 5.5 | 20.0 |
| | GN-04 | 15 | 323904 | 5579813 | 14-Jul-15 | 23.8 | 22.8 | 18.6 | 19.5 |
| | GN-05 | 15 | 323465 | 5577769 | 14-Jul-15 | 26.4 | 15.1 | 18.6 | 20.0 |
| | GN-07 | 15 | 322555 | 5578678 | 17-Jul-15 | 27.3 | 6.3 | 6.0 | 20.0 |
| | GN-10 | 15 | 322215 | 5577402 | 18-Jul-15 | 24.0 | 12.6 | 13.8 | 20.0 |
| | GN-11 | 15 | 320200 | 5578474 | 13-Jul-15 | 22.5 | 5.8 | 4.6 | 22.5 |
| GN-12 | 15 | 319641 | 5577498 | 16-Jul-15 | 26.2 | 13.5 | 9.8 | 20.0 | |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|-----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| PDB | GN-15 | 15 | 319663 | 5575572 | 13-Jul-15 | 23.5 | 5.8 | 4.4 | 22.5 |
| | GN-16 | 15 | 324024 | 5581097 | 17-Jul-15 | 25.8 | 5.1 | 5.0 | 19.5 |
| | SN-01 | 15 | 326297 | 5579266 | 15-Jul-15 | 24.7 | 7.9 | 7.6 | 19.5 |
| | SN-05 | 15 | 323505 | 5577770 | 14-Jul-15 | 26.4 | 13.1 | 15.1 | 20.0 |
| | SN-12 | 15 | 319615 | 5577478 | 16-Jul-15 | 26.2 | 12.1 | 13.5 | 20.0 |
| | SN-16 | 15 | 324056 | 5581108 | 17-Jul-15 | 25.8 | 5.4 | 5.1 | 19.5 |
| | GN-01 | 15 | 326304 | 5579288 | 12-Jul-16 | 25.8 | 7.5 | 6.1 | 20.0 |
| | GN-03 | 15 | 324846 | 5580961 | 11-Jul-16 | 23.8 | 3.1 | 5.5 | 21.0 |
| | GN-04 | 15 | 323839 | 5579798 | 12-Jul-16 | 24.5 | 20.0 | 22.1 | 20.0 |
| | GN-05 | 15 | 323464 | 5577764 | 14-Jul-16 | 27.8 | 14.9 | 16.0 | 21.0 |
| | GN-07 | 15 | 322538 | 5578694 | 14-Jul-16 | 26.4 | 4.6 | 6.2 | 20.5 |
| | GN-10 | 15 | 322197 | 5577382 | 15-Jul-16 | 29.9 | 12.4 | 12.1 | 21.5 |
| | GN-11 | 15 | 320130 | 5578531 | 11-Jul-16 | 23.7 | 4.9 | 6.1 | 21.0 |
| | GN-12 | 15 | 319638 | 5577503 | 15-Jul-16 | 25.8 | 14.1 | 11.4 | 21.0 |
| | GN-15 | 15 | 319604 | 5575451 | 15-Jul-16 | 24.4 | 4.2 | 6.0 | 21.0 |
| | GN-16 | 15 | 324003 | 5581091 | 13-Jul-16 | 27.0 | 4.8 | 4.7 | 19.5 |
| | SN-01 | 15 | 326275 | 5579276 | 12-Jul-16 | 25.8 | 8.3 | 7.5 | 20.0 |
| | SN-05 | 15 | 323499 | 5577775 | 14-Jul-16 | 28.2 | 12.8 | 14.9 | 21.0 |
| | SN-12 | 15 | 319613 | 5577473 | 15-Jul-16 | 26.2 | 11.7 | 14.1 | 21.0 |
| | SN-16 | 15 | 324031 | 5581108 | 13-Jul-16 | 27.3 | 5.3 | 4.8 | 19.5 |
| | GN-01 | 15 | 326403 | 5579378 | 19-Jul-17 | 26.3 | 7.5 | 6.2 | 20.5 |
| | GN-03 | 15 | 324860 | 5580946 | 19-Jul-17 | 26.7 | 5.7 | 5.6 | 21.0 |
| | GN-04 | 15 | 323857 | 5579821 | 18-Jul-17 | 27.6 | 19.8 | 21.6 | 20.5 |
| GN-05 | 15 | 323342 | 5577714 | 21-Jul-17 | 27.9 | 12.7 | 16.8 | 21.5 | |
| GN-07 | 15 | 322547 | 5578690 | 21-Jul-17 | 25.3 | 4.8 | 6.0 | 21.5 | |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|-----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| PDB | GN-10 | 15 | 322218 | 5577383 | 17-Jul-17 | 24.3 | 12.4 | 12.7 | 21.0 |
| | GN-11 | 15 | 320103 | 5578574 | 17-Jul-17 | 23.9 | 5.6 | 5.9 | 21.5 |
| | GN-12 | 15 | 319718 | 5577578 | 20-Jul-17 | 25.0 | 12.4 | 10.8 | 20.5 |
| | GN-15 | 15 | 319666 | 5575569 | 21-Jul-17 | 24.9 | 6.0 | 4.0 | 20.5 |
| | GN-16 | 15 | 323878 | 5581063 | 18-Jul-17 | 26.8 | 4.8 | 4.6 | 20.5 |
| | SN-01 | 15 | 326269 | 5579274 | 19-Jul-17 | 26.3 | 7.5 | 8.5 | 20.5 |
| | SN-05 | 15 | 323511 | 5577760 | 21-Jul-17 | 27.9 | 12.7 | 12.8 | 21.5 |
| | SN-12 | 15 | 319601 | 5577467 | 20-Jul-17 | 25.0 | 12.4 | 11.3 | 20.5 |
| | SN-16 | 15 | 324037 | 5581111 | 18-Jul-17 | 26.8 | 4.8 | 5.2 | 20.5 |
| | GN-01 | 15 | 326407 | 5579377 | 17-Jul-18 | 28.3 | 6.3 | 7.5 | 21.0 |
| | GN-03 | 15 | 324847 | 5580943 | 18-Jul-18 | 27.4 | 6.4 | 3.8 | 21.5 |
| | GN-04 | 15 | 323890 | 5579782 | 16-Jul-18 | 25.8 | 20.4 | 20.6 | 21.0 |
| | GN-05 | 15 | 323453 | 5577770 | 18-Jul-18 | 26.3 | 15.4 | 17.9 | 22.0 |
| | GN-07 | 15 | 322553 | 5578681 | 18-Jul-18 | 25.5 | 6.2 | 8.7 | 21.5 |
| | GN-10 | 15 | 322146 | 5577500 | 19-Jul-18 | 23.1 | 12.1 | 13.2 | 22.0 |
| | GN-11 | 15 | 320124 | 5578562 | 19-Jul-18 | 23.8 | 3.3 | 6.3 | 23.0 |
| | GN-12 | 15 | 319721 | 5577562 | 16-Jul-18 | 26.8 | 9.4 | 11.0 | 21.0 |
| | GN-15 | 15 | 319673 | 5575618 | 15-Jul-18 | 25.3 | 5.8 | 3.8 | 23.0 |
| | GN-16 | 15 | 323948 | 5581073 | 15-Jul-18 | 24.3 | 4.7 | 5.7 | 22.5 |
| | SN-01 | 15 | 326301 | 5579289 | 17-Jul-18 | 28.3 | 7.5 | 8.7 | 21.0 |
| | SN-05 | 15 | 323490 | 5577776 | 18-Jul-18 | 26.3 | 12.9 | 15.4 | 22.0 |
| | SN-12 | 15 | 319623 | 5577474 | 16-Jul-18 | 26.8 | 11.0 | 9.3 | 21.0 |
| | SN-16 | 15 | 324092 | 5581132 | 15-Jul-18 | 24.3 | 5.1 | 5.7 | 22.5 |
| GN-01 | 15 | 326327 | 5579332 | 17-Jul-19 | 27.0 | 7.7 | 3.9 | 21.5 | |
| GN-03 | 15 | 324825 | 5580957 | 17-Jul-19 | 26.3 | 5.5 | 5.5 | 21.5 | |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|-----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| PDB | GN-04 | 15 | 323880 | 5579820 | 20-Jul-19 | 24.8 | 20.5 | 21.2 | 21.0 |
| | GN-05 | 15 | 323494 | 5577771 | 19-Jul-19 | 25.4 | 12.8 | 17.2 | 21.5 |
| | GN-07 | 15 | 322556 | 5578667 | 16-Jul-19 | 22.4 | 6.5 | 6.7 | 22.0 |
| | GN-10 | 15 | 322138 | 5577498 | 16-Jul-19 | 21.8 | 12.0 | 13.0 | 22.0 |
| | GN-11 | 15 | 320111 | 5578568 | 19-Jul-19 | 26.8 | 3.8 | 6.3 | 21.5 |
| | GN-12 | 15 | 319614 | 5577488 | 20-Jul-19 | 24.2 | 12.6 | 9.4 | 21.0 |
| | GN-15 | 15 | 319633 | 5575472 | 20-Jul-19 | 23.8 | 5.3 | 6.8 | 21.0 |
| | GN-16 | 15 | 324039 | 5581128 | 18-Jul-19 | 25.6 | 5.4 | 4.9 | 21.5 |
| | SN-01 | 15 | 326298 | 5579318 | 17-Jul-19 | 27.9 | 7.7 | 7.7 | 21.5 |
| | SN-05 | 15 | 323534 | 5577779 | 19-Jul-19 | 25.4 | 11.0 | 12.8 | 21.5 |
| | SN-12 | 15 | 319591 | 5577471 | 20-Jul-19 | 23.6 | 11.8 | 12.6 | 21.0 |
| SN-16 | 15 | 324068 | 5581146 | 18-Jul-19 | 25.0 | 5.3 | 5.4 | 21.5 | |
| LDB | GN-01 | 15 | 300757 | 5586695 | 21-Sep-08 | 18.4 | 13.2 | 13.3 | 17.1 |
| | GN-02 | 15 | 301067 | 5588844 | 21-Sep-08 | 19.9 | 3.9 | 7.9 | 17.3 |
| | GN-03 | 15 | 297646 | 5586239 | 21-Sep-08 | 21.0 | 8.0 | 7.9 | 16.6 |
| | GN-04 | 15 | 298543 | 5583161 | 22-Sep-08 | 20.8 | 5.0 | 6.1 | 16.4 |
| | GN-05 | 15 | 294154 | 5584383 | 22-Sep-08 | 20.0 | 7.5 | 7.7 | 16.1 |
| | GN-06 | 15 | 293387 | 5587195 | 22-Sep-08 | 22.0 | 5.2 | 5.2 | 16.1 |
| | GN-07 | 15 | 291311 | 5583254 | 23-Sep-08 | 19.0 | 4.5 | 5.9 | 16.3 |
| | GN-08 | 15 | 288429 | 5586292 | 23-Sep-08 | 19.0 | 2.0 | 11.4 | 17.0 |
| | GN-09 | 15 | 288004 | 5582048 | 24-Sep-08 | 18.5 | 8.0 | 3.0 | 17.0 |
| | GN-10 | 15 | 286599 | 5579621 | 24-Sep-08 | 21.0 | 5.0 | 3.9 | 17.0 |
| | SN-01 | 15 | 300612 | 5586709 | 21-Sep-08 | 18.4 | 13.2 | 13.2 | 17.1 |
| | SN-04 | 15 | 298654 | 5583047 | 22-Sep-08 | 20.8 | 5.0 | 5.0 | 16.4 |
| | SN-08 | 15 | 288592 | 5586328 | 23-Sep-08 | 19.0 | 2.0 | 2.0 | 17.0 |
| | GN-01 | 15 | 300801 | 5586766 | 20-Sep-09 | 17.8 | 17.0 | 17.0 | 20.4 |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|-----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| LDB | GN-02 | 15 | 301078 | 5588621 | 20-Sep-09 | 18.1 | 6.5 | 3.5 | 21.3 |
| | GN-03 | 15 | 297768 | 5586211 | 21-Sep-09 | 20.5 | 7.9 | 7.9 | 20.3 |
| | GN-04 | 15 | 298678 | 5583062 | 21-Sep-09 | 20.1 | 1.7 | 5.8 | 20.0 |
| | GN-05 | 15 | 294285 | 5584308 | 22-Sep-09 | 18.4 | 7.8 | 7.6 | 19.9 |
| | GN-06 | 15 | 293285 | 5587369 | 22-Sep-09 | 18.6 | 4.5 | - | 19.9 |
| | GN-07 | 15 | 291360 | 5583363 | 23-Sep-09 | 19.9 | 4.4 | 5.1 | 19.6 |
| | GN-08 | 15 | 288602 | 5586327 | 23-Sep-09 | 20.5 | 1.6 | 11.5 | 19.3 |
| | GN-09 | 15 | 288150 | 5582048 | 24-Sep-09 | 20.2 | 3.5 | 8.0 | 20.3 |
| | GN-10 | 15 | 286655 | 5579739 | 24-Sep-09 | 21.1 | 5.3 | 5.0 | 19.9 |
| | SN-01 | 15 | 300801 | 5586766 | 20-Sep-09 | 17.8 | 17.0 | 17.0 | 20.4 |
| | SN-04 | 15 | 298678 | 5583062 | 21-Sep-09 | 20.1 | 1.7 | 1.7 | 20.0 |
| | SN-08 | 15 | 288602 | 5586327 | 23-Sep-09 | 20.5 | 4.4 | 4.4 | 19.3 |
| | GN-01 | 15 | 300801 | 5586766 | 21-Sep-10 | 21.8 | 15.6 | 15.1 | 13.8 |
| | GN-02 | 15 | 301078 | 5588621 | 21-Sep-10 | 20.5 | 6.5 | 7.8 | 13.5 |
| | GN-03 | 15 | 297768 | 5586211 | 21-Sep-10 | 22.0 | 9.2 | 8.8 | 13.3 |
| | GN-04 | 15 | 298678 | 5583062 | 22-Sep-10 | 21.8 | 7.3 | 3.7 | 12.4 |
| | GN-05 | 15 | 294285 | 5584308 | 22-Sep-10 | 21.3 | 9.0 | 8.7 | 12.9 |
| | GN-06 | 15 | 293285 | 5587369 | 22-Sep-10 | 24.3 | 6.5 | 6.0 | 12.9 |
| | GN-07 | 15 | 291360 | 5583363 | 23-Sep-10 | 17.9 | 6.1 | 7.3 | 12.8 |
| | GN-08 | 15 | 288602 | 5586327 | 23-Sep-10 | 19.3 | 2.5 | 7.8 | 13.7 |
| GN-09 | 15 | 288150 | 5582048 | 20-Sep-10 | 19.8 | 1.5 | 9.1 | 14.4 | |
| GN-10 | 15 | 286655 | 5579739 | 20-Sep-10 | 16.3 | 5.5 | 5.8 | 14.5 | |
| SN-01 | 15 | 300801 | 5586766 | 21-Sep-10 | 21.8 | 15.6 | 15.1 | 13.8 | |
| SN-04 | 15 | 298678 | 5583062 | 22-Sep-10 | 21.8 | 7.3 | 7.3 | 12.4 | |
| SN-08 | 15 | 288602 | 5586327 | 23-Sep-10 | 19.3 | 2.5 | 2.5 | 13.7 | |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|-----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| LDB | GN-01 | 15 | 300801 | 5586766 | 26-Sep-11 | 20.0 | 13.1 | 13.1 | 15.9 |
| | GN-02 | 15 | 301078 | 5588621 | 26-Sep-11 | 19.0 | 2.7 | 6.7 | 16.4 |
| | GN-03 | 15 | 297768 | 5586211 | 26-Sep-11 | 23.6 | 7.6 | 6.4 | 15.3 |
| | GN-04 | 15 | 298678 | 5583062 | 26-Sep-11 | 22.0 | 1.0 | 5.9 | 14.8 |
| | GN-05 | 15 | 294285 | 5584308 | 27-Sep-11 | 18.2 | 7.6 | 7.6 | 16.5 |
| | GN-06 | 15 | 293285 | 5587369 | 27-Sep-11 | 17.5 | 5.2 | 5.5 | 15.3 |
| | GN-07 | 15 | 291360 | 5583363 | 27-Sep-11 | 20.3 | 6.1 | 5.8 | 16.4 |
| | GN-08 | 15 | 288602 | 5586327 | 28-Sep-11 | 24.7 | 1.8 | 13.1 | 15.8 |
| | GN-09 | 15 | 288150 | 5582048 | 28-Sep-11 | 20.8 | 7.6 | 1.8 | 15.9 |
| | GN-10 | 15 | 286655 | 5579739 | 28-Sep-11 | 18.1 | 5.2 | 4.3 | 15.8 |
| | SN-01 | 15 | 300801 | 5586766 | 26-Sep-11 | 20.0 | 13.1 | 13.1 | 15.9 |
| | SN-04 | 15 | 298678 | 5583062 | 26-Sep-11 | 22.0 | 1.0 | 1.0 | 14.8 |
| | SN-08 | 15 | 288602 | 5586327 | 28-Sep-11 | 24.7 | 1.8 | 1.8 | 15.8 |
| | GN-01 | 15 | 300801 | 5586766 | 23-Sep-12 | 22.2 | 13.1 | 12.8 | 14.4 |
| | GN-02 | 15 | 301078 | 5588621 | 23-Sep-12 | 20.0 | 8.2 | 5.6 | 14.3 |
| | GN-03 | 15 | 297768 | 5586211 | 24-Sep-12 | 18.3 | 8.4 | 7.6 | 12.9 |
| | GN-04 | 15 | 298678 | 5583062 | 23-Sep-12 | 22.8 | 1.8 | 5.5 | 12.8 |
| | GN-05 | 15 | 294285 | 5584308 | 24-Sep-12 | 19.5 | 7.6 | 7.6 | 12.9 |
| | GN-06 | 15 | 293285 | 5587369 | 24-Sep-12 | 20.3 | 5.5 | 5.2 | 12.9 |
| | GN-07 | 15 | 291360 | 5583363 | 24-Sep-12 | 21.3 | 3.1 | 5.6 | 13.3 |
| | GN-08 | 15 | 288602 | 5586327 | 25-Sep-12 | 23.8 | 1.9 | 12.8 | 13.4 |
| | GN-09 | 15 | 288150 | 5582048 | 25-Sep-12 | 20.5 | 7.9 | 2.7 | 13.8 |
| | GN-10 | 15 | 286655 | 5579739 | 25-Sep-12 | 18.2 | 5.5 | 4.1 | 14.0 |
| | SN-01 | 15 | 300801 | 5586766 | 23-Sep-12 | 22.2 | 13.1 | 13.1 | 14.4 |
| SN-04 | 15 | 298678 | 5583062 | 23-Sep-12 | 22.8 | 1.8 | 1.8 | 12.8 | |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|-----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| LDB | SN-08 | 15 | 288602 | 5586327 | 25-Sep-12 | 23.8 | 1.9 | 1.9 | 13.4 |
| | GN-01 | 15 | 300612 | 5586709 | 23-Sep-13 | 21.5 | 12.8 | 12.8 | 17.2 |
| | GN-02 | 15 | 301067 | 5588844 | 23-Sep-13 | 20.3 | 5.5 | 6.4 | 17.2 |
| | GN-03 | 15 | 297646 | 5586239 | 23-Sep-13 | 25.5 | 7.7 | 7.6 | 17.1 |
| | GN-04 | 15 | 298654 | 5583047 | 23-Sep-13 | 23.3 | 1.7 | 5.9 | 16.3 |
| | GN-05 | 15 | 294154 | 5584383 | 24-Sep-13 | 17.6 | 7.6 | 7.6 | 16.9 |
| | GN-06 | 15 | 293387 | 5587195 | 24-Sep-13 | 16.1 | 5.5 | 5.3 | 16.9 |
| | GN-07 | 15 | 291311 | 5583254 | 24-Sep-13 | 18.3 | 4.4 | 5.5 | 17.1 |
| | GN-08 | 15 | 288592 | 5586328 | 25-Sep-13 | 24.0 | 1.8 | 13.7 | 17.4 |
| | GN-09 | 15 | 288131 | 5582036 | 25-Sep-13 | 21.5 | 7.9 | 1.5 | 17.7 |
| | GN-10 | 15 | 286604 | 5579676 | 25-Sep-13 | 19.8 | 6.1 | 4.9 | 17.7 |
| | SN-01 | 15 | 300612 | 5586709 | 23-Sep-13 | 21.5 | 12.8 | 12.9 | 17.2 |
| | SN-04 | 15 | 298654 | 5583047 | 23-Sep-13 | 23.3 | 1.7 | 1.7 | 16.3 |
| | SN-08 | 15 | 288592 | 5586328 | 25-Sep-13 | 24.0 | 1.8 | 1.8 | 17.4 |
| | GN-01 | 15 | 300612 | 5586709 | 22-Sep-14 | 21.8 | 12.8 | 12.8 | 13.0 |
| | GN-02 | 15 | 301067 | 5588844 | 21-Sep-14 | 21.0 | 5.5 | 6.4 | 13.5 |
| | GN-03 | 15 | 297646 | 5586239 | 21-Sep-14 | 19.5 | 7.7 | 7.6 | 12.8 |
| | GN-04 | 15 | 298654 | 5583047 | 21-Sep-14 | 23.4 | 1.7 | 5.9 | 12.8 |
| | GN-05 | 15 | 294154 | 5584383 | 22-Sep-14 | 20.2 | 7.6 | 7.6 | 13.5 |
| | GN-06 | 15 | 293387 | 5587195 | 22-Sep-14 | 19.5 | 5.5 | 5.3 | 13.4 |
| | GN-07 | 15 | 291311 | 5583254 | 22-Sep-14 | 21.1 | 4.4 | 5.5 | 13.8 |
| | GN-08 | 15 | 288592 | 5586328 | 23-Sep-14 | 21.0 | 1.8 | 13.7 | 13.8 |
| | GN-09 | 15 | 288131 | 5582036 | 23-Sep-14 | 18.6 | 7.9 | 1.5 | 13.4 |
| | GN-10 | 15 | 286604 | 5579676 | 23-Sep-14 | 17.2 | 6.1 | 4.9 | 14.0 |
| SN-01 | 15 | 300612 | 5586709 | 22-Sep-14 | 21.8 | 12.8 | 12.9 | 13.0 | |
| SN-04 | 15 | 298654 | 5583047 | 21-Sep-14 | 23.4 | 1.7 | 1.7 | 12.8 | |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|-----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| LDB | SN-08 | 15 | 288592 | 5586328 | 23-Sep-14 | 21.0 | 1.8 | 1.8 | 13.8 |
| | GN-01 | 15 | 300612 | 5586709 | 20-Sep-15 | 20.5 | 13.0 | 12.5 | 17.4 |
| | GN-02 | 15 | 301067 | 5588844 | 20-Sep-15 | 19.3 | 6.3 | 4.7 | 17.2 |
| | GN-03 | 15 | 297646 | 5586239 | 20-Sep-15 | 22.5 | 8.0 | 7.3 | 17.2 |
| | GN-04 | 15 | 298654 | 5583047 | 21-Sep-15 | 20.5 | 1.8 | 5.6 | 17.1 |
| | GN-05 | 15 | 294154 | 5584383 | 21-Sep-15 | 19.7 | 7.5 | 7.3 | 16.9 |
| | GN-06 | 15 | 293387 | 5587195 | 21-Sep-15 | 20.8 | 4.3 | 5.2 | 17.1 |
| | GN-07 | 15 | 291311 | 5583254 | 21-Sep-15 | 21.5 | 3.9 | 5.0 | 17.5 |
| | GN-08 | 15 | 288592 | 5586328 | 22-Sep-15 | 23.0 | 10.0 | 3.0 | 17.0 |
| | GN-09 | 15 | 288131 | 5582036 | 22-Sep-15 | 20.5 | 8.9 | 4.0 | 17.0 |
| | GN-10 | 15 | 286604 | 5579676 | 22-Sep-15 | 18.5 | 4.6 | 3.9 | 17.1 |
| | SN-01 | 15 | 300612 | 5586709 | 20-Sep-15 | 20.5 | 1.8 | 1.8 | 17.4 |
| | SN-04 | 15 | 298654 | 5583047 | 21-Sep-15 | 20.5 | 13.0 | 13.0 | 17.1 |
| | SN-08 | 15 | 288592 | 5586328 | 22-Sep-15 | 23.0 | 10.0 | 10.0 | 17.0 |
| | GN-01 | 15 | 300612 | 5586709 | 20-Sep-16 | 23.5 | 13.4 | 13.4 | 16.9 |
| | GN-02 | 15 | 301067 | 5588844 | 19-Sep-16 | 23.3 | 6.3 | 5.3 | 17.2 |
| | GN-03 | 15 | 297646 | 5586239 | 19-Sep-16 | 23.5 | 7.7 | 7.7 | 17.0 |
| | GN-04 | 15 | 298654 | 5583047 | 19-Sep-16 | 21.5 | 5.5 | 6.0 | 16.4 |
| | GN-05 | 15 | 294154 | 5584383 | 20-Sep-16 | 19.3 | 7.5 | 7.6 | 16.4 |
| | GN-06 | 15 | 293387 | 5587195 | 20-Sep-16 | 19.5 | 5.3 | 5.5 | 16.4 |
| | GN-07 | 15 | 291311 | 5583254 | 21-Sep-16 | 19.6 | 6.0 | 5.7 | 17.0 |
| | GN-08 | 15 | 288592 | 5586328 | 21-Sep-16 | 20.0 | 10.2 | 1.8 | 16.6 |
| | GN-09 | 15 | 288131 | 5582036 | 21-Sep-16 | 18.3 | 4.0 | 8.0 | 17.2 |
| GN-10 | 15 | 286604 | 5579676 | 22-Sep-16 | 19.6 | 5.8 | 5.5 | 16.9 | |
| SN-01 | 15 | 300612 | 5586709 | 20-Sep-16 | 23.5 | 13.4 | 13.4 | 16.9 | |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|-----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| LDB | SN-04 | 15 | 298654 | 5583047 | 19-Sep-16 | 21.5 | 5.5 | 5.5 | 16.4 |
| | SN-08 | 15 | 288592 | 5586328 | 21-Sep-16 | 20.0 | 10.2 | 10.2 | 16.6 |
| | GN-01 | 15 | 300801 | 5586766 | 18-Sep-17 | 22.6 | - | 13.2 | 16.2 |
| | GN-02 | 15 | 301078 | 5588621 | 18-Sep-17 | 23.0 | 9.7 | 9.2 | 16.2 |
| | GN-03 | 15 | 297768 | 5586211 | 19-Sep-17 | 22.0 | 7.9 | 7.8 | 15.7 |
| | GN-04 | 15 | 298678 | 5583062 | 18-Sep-17 | 22.5 | - | 5.6 | 14.8 |
| | GN-05 | 15 | 294285 | 5584308 | 19-Sep-17 | 21.8 | 7.3 | 7.5 | 14.9 |
| | GN-06 | 15 | 293285 | 5587369 | 19-Sep-17 | 22.6 | 4.4 | 4.5 | 15.8 |
| | GN-07 | 15 | 291360 | 5583363 | 20-Sep-17 | 22.2 | - | - | 15.2 |
| | GN-08 | 15 | 288602 | 5586327 | 20-Sep-17 | 21.8 | - | - | 15.8 |
| | SN-01 | 15 | 300801 | 5586766 | 18-Sep-17 | 22.6 | 13.0 | - | 16.2 |
| | SN-04 | 15 | 298678 | 5583062 | 18-Sep-17 | 22.5 | 4.7 | - | 14.8 |
| | SN-08 | 15 | 288602 | 5586327 | 20-Sep-17 | 21.8 | - | - | 15.8 |
| | GN-01 | 15 | 300718 | 5586678 | 17-Sep-18 | 24.2 | 13.2 | 13.1 | 18.0 |
| | GN-02 | 15 | 301127 | 5588669 | 18-Sep-18 | 18.3 | 9.8 | 9.8 | 18.5 |
| | GN-03 | 15 | 297889 | 5586283 | 18-Sep-18 | 18.7 | 8.0 | 7.9 | 18.5 |
| | GN-04 | 15 | 298573 | 5583134 | 18-Sep-18 | 18.8 | 5.9 | 5.2 | 18.5 |
| | GN-05 | 15 | 294161 | 5584269 | 17-Sep-18 | 22.4 | 7.6 | 7.6 | 17.4 |
| | GN-06 | 15 | 293462 | 5587348 | 17-Sep-18 | 22.6 | 5.6 | 4.9 | 17.2 |
| | GN-07 | 15 | 291425 | 5583354 | 19-Sep-18 | 22.7 | 3.4 | 4.7 | 17.0 |
| GN-08 | 15 | 288535 | 5586306 | 19-Sep-18 | 23.3 | 2.9 | 7.6 | 17.2 | |
| GN-09 | 15 | 288137 | 5582074 | 20-Sep-18 | 20.8 | 5.8 | 8.2 | 17.3 | |
| GN-10 | 15 | 286633 | 5579760 | 20-Sep-18 | 21.0 | 5.3 | 5.3 | 17.3 | |
| SN-01 | 15 | 300694 | 5586658 | 17-Sep-18 | 24.2 | 13.1 | | 18.0 | |
| SN-04 | 15 | 298555 | 5583136 | 18-Sep-18 | 18.8 | 5.9 | | 18.5 | |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| LDB | SN-08 | 15 | 288535 | 5586306 | 19-Sep-18 | 23.3 | 2.6 | | 17.2 |
| | GN-01 | 15 | 300716 | 5586910 | 16-Sep-19 | 22.9 | - | - | 17.2 |
| | GN-02 | 15 | 301074 | 5588706 | 16-Sep-19 | 23.6 | - | - | 17.5 |
| | GN-03 | 15 | 297692 | 5586320 | 17-Sep-19 | 21.6 | - | - | 17.7 |
| | GN-04 | 15 | 298617 | 5583087 | 16-Sep-19 | 21.9 | - | - | 17.4 |
| | GN-05 | 15 | 294178 | 5584309 | 17-Sep-19 | 20.8 | - | - | 18.0 |
| | GN-06 | 15 | 293306 | 5587454 | 17-Sep-19 | 21.8 | - | - | 18.8 |
| | GN-07 | 15 | 291409 | 5583417 | 18-Sep-19 | 18.9 | - | - | 18.2 |
| | GN-08 | 15 | 288559 | 5586301 | 19-Sep-19 | 28.8 | - | - | 18.1 |
| | GN-09 | 15 | 288099 | 5582006 | 18-Sep-19 | 19.0 | - | - | 18.2 |
| | GN-10 | 15 | 286482 | 5579738 | 19-Sep-19 | 21.7 | 5.3 | 5.3 | 18.2 |
| | SN-01 | 15 | 300584 | 5586893 | 16-Sep-19 | 22.9 | - | - | 17.2 |
| | SN-04 | 15 | 298665 | 5583077 | 16-Sep-19 | 21.9 | - | - | 17.4 |
| | SN-08 | 15 | 288590 | 5586317 | 19-Sep-19 | 28.8 | - | - | 18.1 |
| PFF | GN-01 | 14 | 701309 | 5604645 | 15-Jul-11 | 22.1 | 7.7 | 5.0 | 21.5 |
| | GN-02 | 14 | 701396 | 5604838 | 15-Jul-11 | 22.8 | 15.0 | 15.5 | 21.5 |
| | GN-03 | 14 | 702466 | 5603192 | 16-Jul-11 | 24.8 | 11.4 | 8.6 | 21.5 |
| | GN-05 | 14 | 702878 | 5601645 | 17-Jul-11 | 23.8 | 7.1 | 7.2 | 22.0 |
| | GN-07 | 14 | 706060 | 5600416 | 17-Jul-11 | 26.9 | 14.8 | 10.5 | 23.0 |
| | GN-09 | 14 | 701645 | 5605432 | 18-Jul-11 | 26.3 | 15.2 | 15.2 | 23.0 |
| | GN-10 | 14 | 706297 | 5600715 | 19-Jul-11 | 23.6 | 7.1 | 2.4 | 23.0 |
| | GN-11 | 14 | 701806 | 5604275 | 19-Jul-11 | 24.0 | 3.8 | 2.6 | 24.0 |
| | GN-13 | 14 | 704600 | 5600217 | 20-Jul-11 | 27.2 | 7.9 | 6.9 | 23.0 |
| | SN-01 | 14 | 701341 | 5604635 | 15-Jul-11 | 22.1 | 5.6 | 7.7 | 21.5 |
| | SN-03 | 14 | 702484 | 5603222 | 16-Jul-11 | 24.3 | 11.7 | 11.4 | 21.5 |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|-----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| PFF | SN-07 | 14 | 706079 | 5600308 | 17-Jul-11 | 25.3 | 15.5 | 14.8 | 22.0 |
| | GN-01 | 14 | 701293 | 5604672 | 16-Jul-14 | 18.6 | - | - | 18.0 |
| | GN-02 | 14 | 701556 | 5604852 | 16-Jul-14 | 19.4 | - | - | 18.0 |
| | GN-03 | 14 | 702670 | 5603400 | 15-Jul-14 | 18.8 | 17.0 | 8.5 | 17.0 |
| | GN-05 | 14 | 702871 | 5601580 | 15-Jul-14 | 17.3 | 10.0 | 7.7 | 17.0 |
| | GN-07 | 14 | 706167 | 5600389 | 17-Jul-14 | 21.3 | 4.1 | 10.5 | 18.0 |
| | GN-09 | 14 | 701658 | 5605403 | 16-Jul-14 | 18.0 | - | - | 18.0 |
| | GN-10 | 14 | 706332 | 5600706 | 17-Jul-14 | 22.7 | 5.4 | 7.3 | 18.0 |
| | GN-11 | 14 | 701911 | 5604235 | 15-Jul-14 | 20.2 | 5.5 | 8.5 | 17.0 |
| | GN-13 | 14 | 704588 | 5600180 | 14-Jul-14 | 22.7 | 5.5 | 8.5 | 16.0 |
| | SN-01 | 14 | 701316 | 5604659 | 16-Jul-14 | 18.6 | - | - | 18.0 |
| | SN-03 | 14 | 702675 | 5603425 | 15-Jul-14 | 18.8 | 17.0 | 17.0 | 17.0 |
| | SN-07 | 14 | 706170 | 5600351 | 17-Jul-14 | 21.3 | 4.1 | 4.1 | 18.0 |
| | GN-01 | 14 | 701176 | 5604727 | 24-Jul-17 | 23.8 | 9.2 | 8.8 | 22.0 |
| | GN-02 | 14 | 701565 | 5604854 | 24-Jul-17 | 24.8 | 13.4 | 22.4 | 22.0 |
| | GN-03 | 14 | 702675 | 5603433 | 25-Jul-17 | 23.7 | 15.7 | 7.7 | 22.0 |
| | GN-05 | 14 | 702868 | 5601580 | 25-Jul-17 | 24.9 | 10.2 | 7.3 | 22.0 |
| | GN-07 | 14 | 706145 | 5600404 | 26-Jul-17 | 25.0 | 11.1 | 10.1 | 22.0 |
| | GN-09 | 14 | 701654 | 5605413 | 23-Jul-17 | 24.8 | 14.6 | 8.6 | 21.5 |
| | GN-10 | 14 | 706271 | 5600728 | 25-Jul-17 | 22.4 | 7.3 | 3.8 | 23.0 |
| | GN-11 | 14 | 702025 | 5604270 | 23-Jul-17 | 24.4 | 6.6 | 4.8 | 23.0 |
| | GN-13 | 14 | 704603 | 5600212 | 26-Jul-17 | 24.0 | 7.7 | 6.8 | 22.0 |
| | SN-01 | 14 | 701312 | 5604671 | 24-Jul-17 | 23.8 | 9.2 | 8.7 | 22.0 |
| SN-03 | 14 | 702675 | 5603460 | 25-Jul-17 | 23.7 | 15.7 | 14.3 | 22.0 | |
| SN-07 | 14 | 706160 | 5600378 | 26-Jul-17 | 25.0 | 11.1 | 8.2 | 22.0 | |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|----------|----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| MANIG | GN-01 | 15 | 318961 | 5640604 | 7-Sep-08 | 15.8 | 18.2 | 17.5 | 17.0 |
| | GN-02 | 15 | 320044 | 5639418 | 7-Sep-08 | 16.8 | 6.7 | 3.7 | 17.0 |
| | GN-03 | 15 | 317163 | 5639414 | 8-Sep-08 | 18.6 | 5.7 | 13.0 | 17.0 |
| | GN-04 | 15 | 317424 | 5636260 | 8-Sep-08 | 21.7 | 3.5 | 1.4 | 16.0 |
| | GN-05 | 15 | 316202 | 5637849 | 9-Sep-08 | 15.5 | 20.5 | 20.3 | 16.5 |
| | GN-06 | 15 | 314841 | 5640180 | 9-Sep-08 | 16.2 | 15.0 | 14.0 | 16.5 |
| | SN-01 | 15 | 318961 | 5640604 | 7-Sep-08 | 15.8 | 18.2 | 18.2 | 17.0 |
| | SN-03 | 15 | 317163 | 5639414 | 8-Sep-08 | 18.6 | 3.3 | 5.7 | 17.0 |
| | GN-01 | 15 | 318955 | 5640676 | 1-Sep-09 | 17.4 | 19.0 | 19.0 | 18.2 |
| | GN-02 | 15 | 320137 | 5639470 | 1-Sep-09 | 18.7 | 7.6 | 4.7 | 18.0 |
| | GN-03 | 15 | 317135 | 5639444 | 2-Sep-09 | 20.9 | 1.0 | 13.0 | 19.7 |
| | GN-04 | 15 | 317454 | 5636318 | 2-Sep-09 | 23.4 | 4.5 | 1.5 | 18.5 |
| | GN-05 | 15 | 316224 | 5637833 | 3-Sep-09 | 17.6 | 21.7 | 21.5 | 22.7 |
| | GN-06 | 15 | 314846 | 5640208 | 3-Sep-09 | 18.4 | 16.0 | 14.8 | 22.0 |
| | SN-01 | 15 | 318955 | 5640676 | 1-Sep-09 | 17.4 | 19.0 | 19.0 | 18.2 |
| | SN-03 | 15 | 317135 | 5639444 | 2-Sep-09 | 20.9 | 1.0 | 1.0 | 19.7 |
| | GN-01 | 15 | 318918 | 5640658 | 7-Sep-10 | 16.8 | 19.0 | 19.0 | 16.0 |
| | GN-02 | 15 | 320077 | 5639435 | 7-Sep-10 | 18.0 | 7.3 | 5.2 | 16.0 |
| | GN-03 | 15 | 317176 | 5639330 | 8-Sep-10 | 20.2 | 11.0 | 3.9 | 16.0 |
| | GN-04 | 15 | 317422 | 5636276 | 8-Sep-10 | 23.5 | 4.3 | 1.9 | 16.0 |
| | GN-05 | 15 | 315996 | 5637930 | 9-Sep-10 | 41.5 | 22.0 | 21.0 | 16.0 |
| | GN-06 | 15 | 314778 | 5640225 | 9-Sep-10 | 39.4 | 16.0 | 13.0 | 16.0 |
| | SN-01 | 15 | 318950 | 5640638 | 7-Sep-10 | 16.8 | 19.0 | 19.0 | 16.0 |
| | SN-03 | 15 | 317138 | 5639436 | 8-Sep-10 | 20.2 | 11.0 | 3.9 | 16.0 |
| GN-01 | 15 | 319087 | 5640536 | 8-Sep-11 | 22.4 | 16.0 | 17.0 | 18.5 | |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| MANIG | GN-02 | 15 | 320163 | 5639343 | 8-Sep-11 | 22.4 | 3.9 | 4.3 | 18.5 |
| | GN-03 | 15 | 317108 | 5639251 | 7-Sep-11 | 23.8 | 13.0 | 8.4 | 19.0 |
| | GN-04 | 15 | 317408 | 5636217 | 6-Sep-11 | 19.9 | 9.3 | 2.0 | 21.0 |
| | GN-05 | 15 | 316021 | 5637932 | 6-Sep-11 | 18.3 | 20.0 | 20.0 | 21.0 |
| | GN-06 | 15 | 314711 | 5640265 | 7-Sep-11 | 25.4 | 14.0 | 13.0 | 19.0 |
| | SN-01 | 15 | 319094 | 5640464 | 8-Sep-11 | 22.4 | 16.0 | 16.0 | 18.5 |
| | SN-03 | 15 | 317250 | 5639192 | 7-Sep-11 | 23.8 | 4.1 | 8.4 | 19.0 |
| | GN-01 | 15 | 318877 | 5640698 | 10-Sep-12 | 22.1 | 18.7 | 18.7 | 17.0 |
| | GN-02 | 15 | 320078 | 5639433 | 10-Sep-12 | 21.4 | 6.9 | 6.8 | 17.0 |
| | GN-03 | 15 | 317194 | 5639432 | 11-Sep-12 | 23.9 | 6.2 | 12.3 | 17.0 |
| | GN-04 | 15 | 317418 | 5636273 | 12-Sep-12 | 24.9 | 3.5 | 4.1 | 16.0 |
| | GN-05 | 15 | 316047 | 5637936 | 12-Sep-12 | 25.4 | 20.5 | 20.4 | 16.0 |
| | GN-06 | 15 | 314777 | 5640223 | 11-Sep-12 | 46.2 | 14.9 | 12.8 | 17.0 |
| | SN-01 | 15 | 318901 | 5640679 | 10-Sep-12 | 22.0 | 18.7 | 18.3 | 17.0 |
| | SN-03 | 15 | 317161 | 5639448 | 11-Sep-12 | 23.9 | 3.1 | 6.2 | 16.0 |
| | GN-01 | 15 | 318933 | 5640672 | 9-Sep-13 | 22.5 | 17.6 | 17.4 | 19.0 |
| | GN-02 | 15 | 320083 | 5639437 | 9-Sep-13 | 23.7 | 6.0 | 6.3 | 19.0 |
| | GN-03 | 15 | 317166 | 5639338 | 10-Sep-13 | 24.5 | 10.0 | 6.6 | 19.5 |
| | GN-04 | 15 | 317423 | 5636278 | 11-Sep-13 | 25.8 | 2.9 | 3.5 | 19.0 |
| | GN-05 | 15 | 316015 | 5637938 | 10-Sep-13 | 26.1 | 20.1 | 19.9 | 19.5 |
| | GN-06 | 15 | 314779 | 5640223 | 11-Sep-13 | 23.7 | 13.1 | 12.1 | 19.0 |
| | SN-01 | 15 | 318957 | 5640646 | 9-Sep-13 | 22.5 | 17.7 | 17.6 | 19.0 |
| | SN-03 | 15 | 317201 | 5639292 | 10-Sep-13 | 24.5 | 12.1 | 10.0 | 19.5 |
| GN-01 | 15 | 318933 | 5640697 | 8-Sep-14 | 23.2 | 18.0 | 17.8 | 17.0 | |
| GN-02 | 15 | 320046 | 5639431 | 8-Sep-14 | 23.6 | 7.0 | 4.0 | 17.0 | |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|-----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| MANIG | GN-03 | 15 | 317198 | 5639412 | 9-Sep-14 | 24.2 | 8.0 | 15.0 | 16.0 |
| | GN-04 | 15 | 317431 | 5636234 | 10-Sep-14 | 24.5 | 3.5 | 2.6 | 16.0 |
| | GN-05 | 15 | 315982 | 5637938 | 9-Sep-14 | 23.8 | 20.7 | 20.8 | 16.0 |
| | GN-06 | 15 | 314836 | 5640208 | 10-Sep-14 | 23.6 | 15.6 | 12.3 | 16.0 |
| | SN-01 | 15 | 318980 | 5640642 | 8-Sep-14 | 23.2 | 18.0 | 18.0 | 17.0 |
| | SN-03 | 15 | 317157 | 5639443 | 9-Sep-14 | 24.2 | 6.0 | 8.0 | 16.0 |
| | GN-01 | 15 | 318849 | 5640694 | 14-Sep-15 | 22.6 | 18.4 | 37.9 | 17.0 |
| | GN-02 | 15 | 320088 | 5639461 | 14-Sep-15 | 23.3 | 7.1 | 5.8 | 16.0 |
| | GN-03 | 15 | 317156 | 5639399 | 15-Sep-15 | 24.3 | 6.8 | 13.2 | 16.0 |
| | GN-04 | 15 | 317440 | 5636255 | 16-Sep-15 | 24.6 | 4.0 | 3.6 | 15.0 |
| | GN-05 | 15 | 315992 | 5637967 | 15-Sep-15 | 22.8 | 25.9 | 34.0 | 15.0 |
| | GN-06 | 15 | 314862 | 5640180 | 16-Sep-15 | 24.3 | 17.9 | 17.6 | 16.0 |
| | SN-01 | 15 | 318939 | 5640650 | 14-Sep-15 | 22.6 | 18.5 | 18.4 | 17.0 |
| | SN-03 | 15 | 317151 | 5639442 | 15-Sep-15 | 24.3 | 3.8 | 6.8 | 16.0 |
| | GN-01 | 15 | 318917 | 5640683 | 12-Sep-16 | 24.7 | 38.6 | 29.3 | 18.0 |
| | GN-02 | 15 | 320074 | 5639445 | 12-Sep-16 | 24.3 | 6.6 | 4.2 | 18.0 |
| | GN-03 | 15 | 317174 | 5639419 | 13-Sep-16 | 23.6 | 6.0 | 12.4 | 16.0 |
| | GN-04 | 15 | 317423 | 5636258 | 14-Sep-16 | 22.8 | 3.4 | 2.3 | 16.0 |
| | GN-05 | 15 | 315989 | 5637931 | 13-Sep-16 | 24.0 | 20.7 | 20.2 | 16.5 |
| | GN-06 | 15 | 314698 | 5640135 | 14-Sep-16 | 24.0 | 14.9 | 14.9 | 16.0 |
| | SN-01 | 15 | 318939 | 5640652 | 12-Sep-16 | 24.7 | 18.1 | 38.6 | 18.0 |
| | SN-03 | 15 | 317150 | 5639443 | 13-Sep-16 | 23.6 | 5.1 | 6.0 | 16.0 |
| | GN-01 | 15 | 318907 | 5640656 | 13-Sep-17 | 24.8 | 17.4 | 17.4 | 17.5 |
| | GN-02 | 15 | 320055 | 5639466 | 13-Sep-17 | 24.4 | 6.0 | 4.9 | 17.5 |
| GN-03 | 15 | 317179 | 5639422 | 12-Sep-17 | 26.1 | 5.5 | 12.0 | 17.0 | |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|-----------|-----------|----------------------------------|-----------------|------|-------------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| MANIG | GN-04 | 15 | 317456 | 5636271 | 11-Sep-17 | 24.5 | 2.9 | 1.9 | 18.0 |
| | GN-05 | 15 | 315983 | 5637959 | 12-Sep-17 | 27.1 | 18.5 | 20.1 | 16.5 |
| | GN-06 | 15 | 314869 | 5640137 | 11-Sep-17 | 23.8 | 14.9 | 14.6 | 17.0 |
| | SN-01 | 15 | 318934 | 5640632 | 13-Sep-17 | 24.8 | 17.4 | 17.4 | 17.5 |
| | SN-03 | 15 | 317158 | 5639443 | 12-Sep-17 | 26.1 | 5.5 | 4.1 | 17.0 |
| | GN-01 | 15 | 318913 | 5640663 | 10-Sep-18 | 23.6 | 17.1 | 19.7 | 18.0 |
| | GN-02 | 15 | 320065 | 5639413 | 10-Sep-18 | 21.8 | 5.6 | 5.1 | 17.5 |
| | GN-03 | 15 | 317172 | 5639398 | 11-Sep-18 | 26.9 | 5.6 | 12.3 | 17.0 |
| | GN-04 | 15 | 317431 | 5636267 | 12-Sep-18 | 25.5 | 2.4 | 2.9 | 15.5 |
| | GN-05 | 15 | 316010 | 5637920 | 12-Sep-18 | 24.3 | 19.7 | 19.8 | 16.5 |
| | GN-06 | 15 | 314767 | 5640228 | 11-Sep-18 | 26.0 | 13.9 | 10.5 | 17.0 |
| | SN-01 | 15 | 318945 | 5640644 | 10-Sep-18 | 23.6 | 17.0 | 17.1 | 18.0 |
| | SN-03 | 15 | 317155 | 5639429 | 11-Sep-18 | 26.9 | 4.2 | 5.6 | 17.0 |
| | SN-04 | 15 | 317402 | 5636261 | 12-Sep-18 | 25.5 | 2.3 | 2.4 | 15.5 |
| | GN-01 | 15 | 318927 | 5640658 | 9-Sep-19 | 19.1 | 16.4 | 17.5 | 16.5 |
| | GN-02 | 15 | 320113 | 5639339 | 9-Sep-19 | 23.9 | 6.4 | 2.0 | 15.0 |
| | GN-03 | 15 | 317164 | 5639420 | 12-Sep-19 | 24.1 | 5.7 | 11.6 | 16.0 |
| | GN-04 | 15 | 317331 | 5636331 | 11-Sep-19 | 24.8 | 3.3 | 1.9 | 15.0 |
| | GN-05 | 15 | 316028 | 5637939 | 11-Sep-19 | 21.3 | 20.3 | 19.8 | 15.5 |
| | GN-06 | 15 | 314783 | 5640218 | 12-Sep-19 | 24.3 | 15.2 | 11.8 | 15.0 |
| SN-01 | 15 | 318946 | 5640643 | 9-Sep-19 | 18.9 | 16.2 | 16.4 | 16.5 | |
| SN-03 | 15 | 317144 | 5639437 | 12-Sep-19 | 23.5 | 3.2 | 5.7 | 16.0 | |
| SN-04 | 15 | 317438 | 5636296 | 11-Sep-19 | 24.6 | 3.4 | 3.3 | 15.0 | |
| EAGLE | GN-02 | 15 | 342972 | 5578980 | 20-Jul-10 | 25.2 | 3.2 | 3.3 | 19.0 |
| | GN-03 | 15 | 341990 | 5576703 | 19-Jul-10 | 24.4 | 7.8 | 8.1 | 19.5 |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| EAGLE | GN-04 | 15 | 343679 | 5577547 | 20-Jul-10 | 28.3 | 12.9 | 14.2 | 20.0 |
| | GN-05 | 15 | 341358 | 5573610 | 21-Jul-10 | 28.6 | 7.8 | 10.9 | 20.0 |
| | GN-06 | 15 | 341263 | 5576055 | 21-Jul-10 | 26.9 | 2.4 | 5.5 | 20.0 |
| | GN-07 | 15 | 345398 | 5569004 | 22-Jul-10 | 27.8 | 4.2 | 3.9 | 19.5 |
| | GN-08 | 15 | 346310 | 5569611 | 22-Jul-10 | 25.0 | 12.3 | 16.2 | 19.5 |
| | GN-09 | 15 | 342276 | 5571804 | 23-Jul-10 | 26.3 | 4.8 | 3.8 | 19.5 |
| | GN-10 | 15 | 342879 | 5571475 | 23-Jul-10 | 24.5 | 7.4 | 14.5 | 19.5 |
| | GN-11 | 15 | 340695 | 5570949 | 24-Jul-10 | 27.3 | 3.0 | 3.1 | 21.0 |
| | GN-12 | 15 | 340769 | 5571569 | 24-Jul-10 | 24.5 | 8.9 | 10.3 | 21.0 |
| | SN-04 | 15 | 343718 | 5577472 | 20-Jul-10 | 28.3 | 12.6 | 12.9 | 20.0 |
| | SN-08 | 15 | 346239 | 5569587 | 22-Jul-10 | 25.0 | 11.7 | 12.3 | 19.5 |
| | SN-11 | 15 | 340623 | 5570926 | 24-Jul-10 | 27.3 | 2.9 | 3.0 | 21.0 |
| | GN-02 | 15 | 342035 | 5578932 | 22-Jul-13 | 23.6 | 3.9 | 3.7 | 20.5 |
| | GN-03 | 15 | 341951 | 5576658 | 26-Jul-13 | 26.0 | 8.1 | 8.4 | 20.0 |
| | GN-04 | 15 | 343691 | 5577527 | 26-Jul-13 | 23.6 | 12.8 | 13.9 | 20.0 |
| | GN-05 | 15 | 341335 | 5573649 | 26-Jul-13 | 24.0 | 7.6 | 11.5 | 20.0 |
| | GN-06 | 15 | 341284 | 5576116 | 22-Jul-13 | 24.8 | 5.3 | 4.9 | 21.0 |
| | GN-08 | 15 | 346274 | 5569642 | 23-Jul-13 | 24.6 | 13.5 | 11.8 | 21.0 |
| | GN-09 | 15 | 342220 | 5571764 | 23-Jul-13 | 26.3 | 5.3 | 4.2 | 21.0 |
| | GN-10 | 15 | 342873 | 5571537 | 24-Jul-13 | 25.7 | 7.7 | 13.2 | 20.5 |
| | GN-11 | 15 | 340667 | 5570941 | 24-Jul-13 | 24.5 | 3.1 | 3.3 | 21.0 |
| | GN-12 | 15 | 340725 | 5571532 | 25-Jul-13 | 25.5 | 8.3 | 9.4 | 20.5 |
| | GN-14 | 15 | 341362 | 5574269 | 25-Jul-13 | 24.8 | 5.1 | 4.3 | 20.5 |
| | SN-04 | 15 | 343613 | 5577611 | 26-Jul-13 | 23.6 | 12.8 | 12.8 | 20.0 |
| | SN-08 | 15 | 346245 | 5569659 | 23-Jul-13 | 24.6 | 12.9 | 13.5 | 21.0 |
| | SN-11 | 15 | 340639 | 5570924 | 24-Jul-13 | 24.5 | 3.0 | 3.1 | 21.0 |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|-----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| EAGLE | SN-14 | 15 | 341395 | 5574276 | 25-Jul-13 | 24.8 | 4.8 | 5.1 | 20.5 |
| | GN-02 | 15 | 342007 | 5578923 | 19-Jul-16 | 24.4 | 3.3 | 3.4 | 20.0 |
| | GN-03 | 15 | 341963 | 5576649 | 23-Jul-16 | 24.6 | 8.1 | 8.3 | 20.5 |
| | GN-04 | 15 | 343670 | 5577534 | 19-Jul-16 | 22.5 | 12.6 | 13.2 | 21.0 |
| | GN-05 | 15 | 341427 | 5573682 | 22-Jul-16 | 23.4 | 7.8 | 11.6 | 20.5 |
| | GN-06 | 15 | 341301 | 5576111 | 23-Jul-16 | 22.7 | 5.6 | 5.4 | 20.5 |
| | GN-07 | 15 | 345368 | 5569037 | 20-Jul-16 | 26.6 | 4.3 | 2.8 | 20.0 |
| | GN-08 | 15 | 346381 | 5569686 | 20-Jul-16 | 25.5 | 14.2 | 15.1 | 20.0 |
| | GN-09 | 15 | 342316 | 5571775 | 21-Jul-16 | 22.4 | 3.3 | 6.6 | 20.5 |
| | GN-10 | 15 | 342855 | 5571526 | 20-Jul-16 | 22.8 | 10.6 | 13.7 | 20.0 |
| | GN-11 | 15 | 340707 | 5570941 | 21-Jul-16 | 25.2 | 2.8 | 3.1 | 19.0 |
| | GN-12 | 15 | 340688 | 5571517 | 21-Jul-16 | 25.9 | 7.9 | 8.9 | 19.0 |
| | GN-14 | 15 | 341338 | 5574271 | 22-Jul-16 | 21.8 | 5.1 | 4.3 | 21.0 |
| | SN-04 | 15 | 343705 | 5577529 | 19-Jul-16 | 22.5 | 12.5 | 12.6 | 21.0 |
| | SN-08 | 15 | 346376 | 5569719 | 20-Jul-16 | 25.5 | 13.7 | 14.2 | 20.0 |
| | SN-11 | 15 | 340667 | 5570933 | 21-Jul-16 | 25.2 | 2.3 | 2.8 | 19.0 |
| | SN-14 | 15 | 341347 | 5574282 | 22-Jul-16 | 21.8 | 5.0 | 5.1 | 21.0 |
| | GN-02 | 15 | 341999 | 5578938 | 27-Jul-19 | 23.5 | 2.7 | 2.6 | 20.5 |
| | GN-03 | 15 | 341948 | 5576656 | 27-Jul-19 | 23.9 | 7.5 | 7.4 | 20.5 |
| | GN-04 | 15 | 343705 | 5577495 | 22-Jul-19 | 24.3 | 11.4 | 13.0 | 22.0 |
| | GN-05 | 15 | 341348 | 5573666 | 26-Jul-19 | 25.0 | 7.1 | 9.9 | 20.5 |
| | GN-06 | 15 | 341234 | 5576100 | 22-Jul-19 | 24.8 | 2.1 | 4.9 | 21.5 |
| | GN-07 | 15 | 345436 | 5568966 | 23-Jul-19 | 26.5 | 3.4 | 3.7 | 21.0 |
| GN-08 | 15 | 346280 | 5569632 | 23-Jul-19 | 25.9 | 11.9 | 9.7 | 21.0 | |
| GN-09 | 15 | 342328 | 5571778 | 24-Jul-19 | 25.3 | 5.9 | 3.4 | 21.0 | |

Table A5-1-1. continued.

| Location | Site | UTM Coordinates | | | Set Date | Set Duration (h) ¹ | Water Depth (m) | | Set Water Temperature (°C) |
|----------|-------|-----------------|---------|----------|-----------|-------------------------------|-----------------|------|----------------------------|
| | | Zone | Easting | Northing | | | Start | End | |
| EAGLE | GN-10 | 15 | 342869 | 5571520 | 25-Jul-19 | 24.3 | 9.3 | 13.0 | 22.0 |
| | GN-11 | 15 | 340672 | 5571047 | 25-Jul-19 | 24.5 | 2.9 | 2.7 | 22.0 |
| | GN-12 | 15 | 340724 | 5571587 | 24-Jul-19 | 24.9 | 7.8 | 8.6 | 22.0 |
| | GN-14 | 15 | 341255 | 5574291 | 26-Jul-19 | 24.8 | 4.5 | 3.5 | 20.5 |
| | SN-04 | 15 | 343727 | 5577470 | 22-Jul-19 | 23.8 | 11.3 | 11.4 | 22.0 |
| | SN-08 | 15 | 346256 | 5569643 | 23-Jul-19 | 25.2 | 11.1 | 11.9 | 21.0 |
| | SN-11 | 15 | 340644 | 5571050 | 25-Jul-19 | 25.0 | 3.0 | 2.9 | 22.0 |
| | SN-14 | 15 | 341366 | 5574279 | 26-Jul-19 | 24.6 | 4.0 | 4.5 | 20.5 |

Notes:

1. Gill nets that were set for >36 h (red font) were excluded from the data analysis for abundance and diversity metrics.

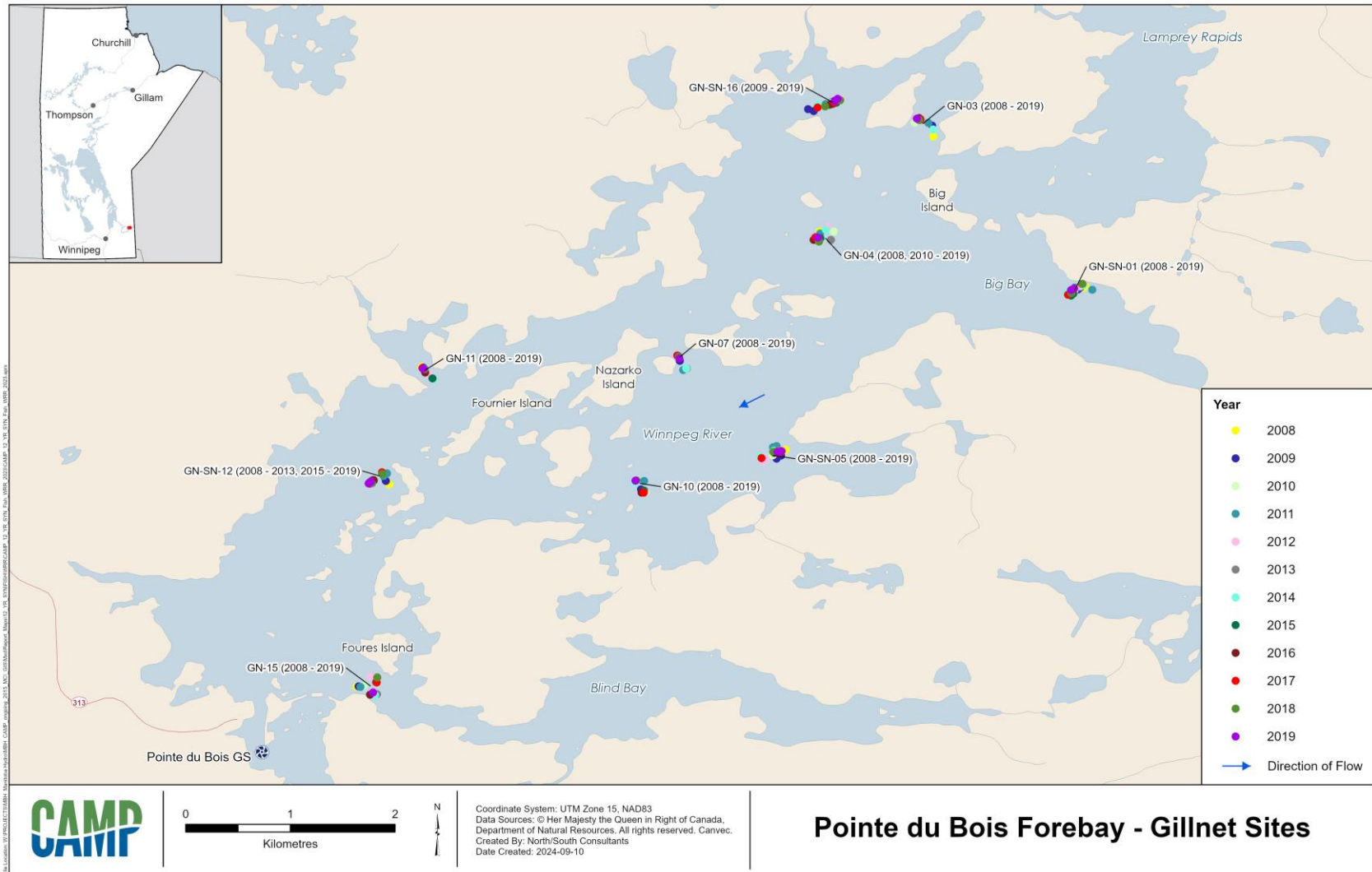


Figure A5-1-1. 2008-2019 Gillnetting sites in the Pointe du Bois Forebay.

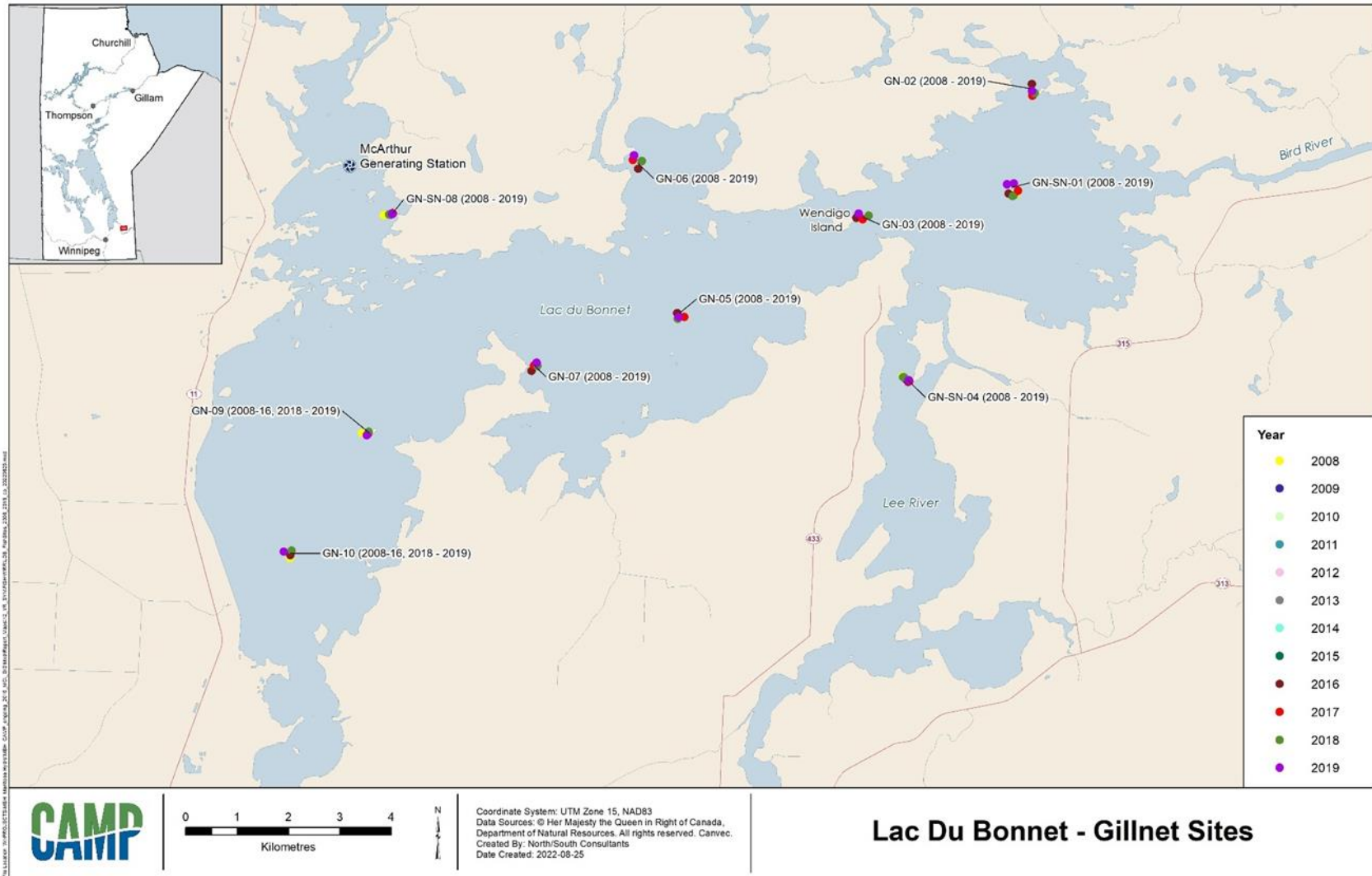


Figure A5-1-2. 2008-2019 Gillnetting sites in Lac du Bonnet.

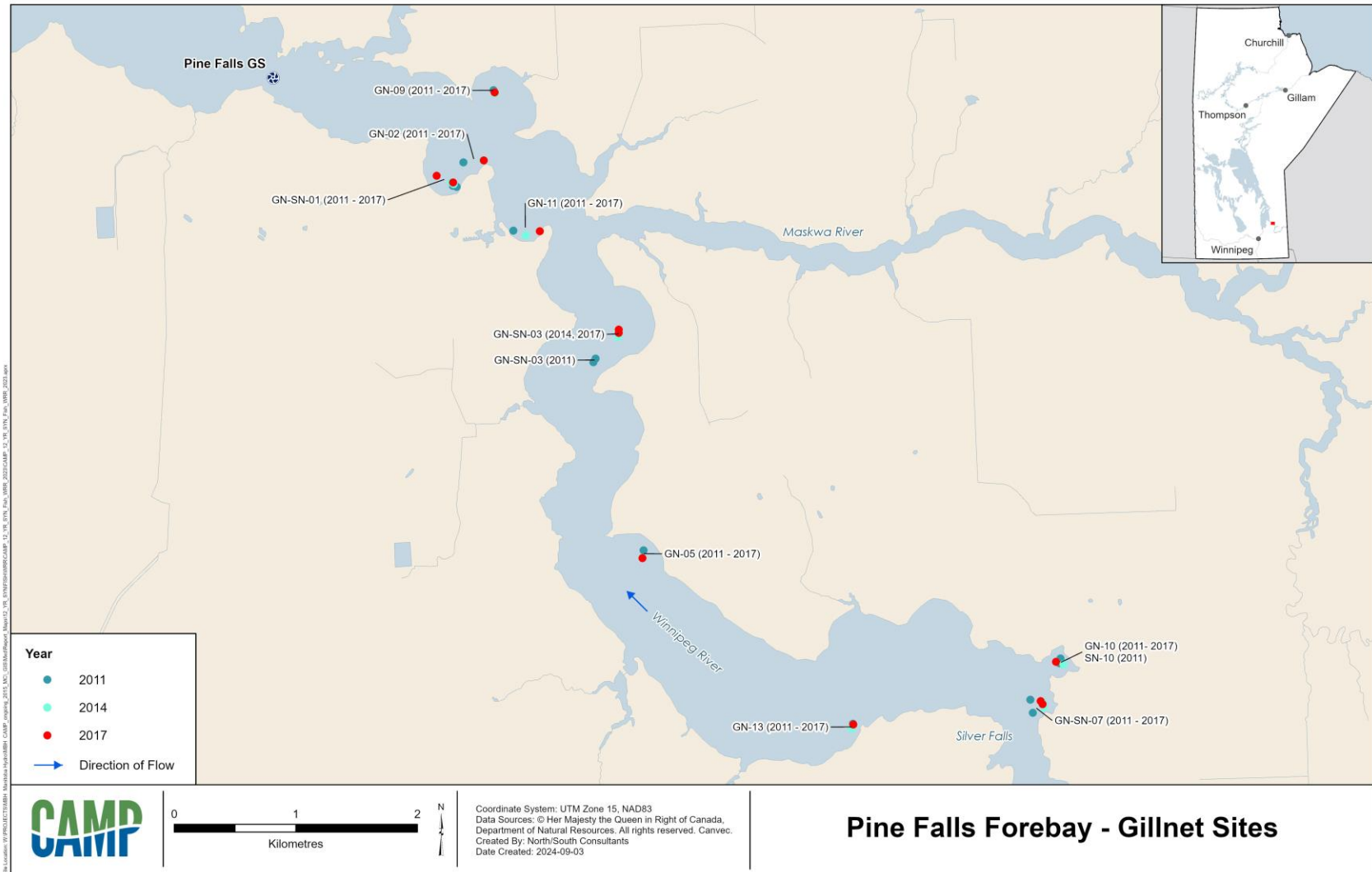


Figure A5-1-3. 2011-2017 Gillnetting sites in the Pine Falls Forebay.

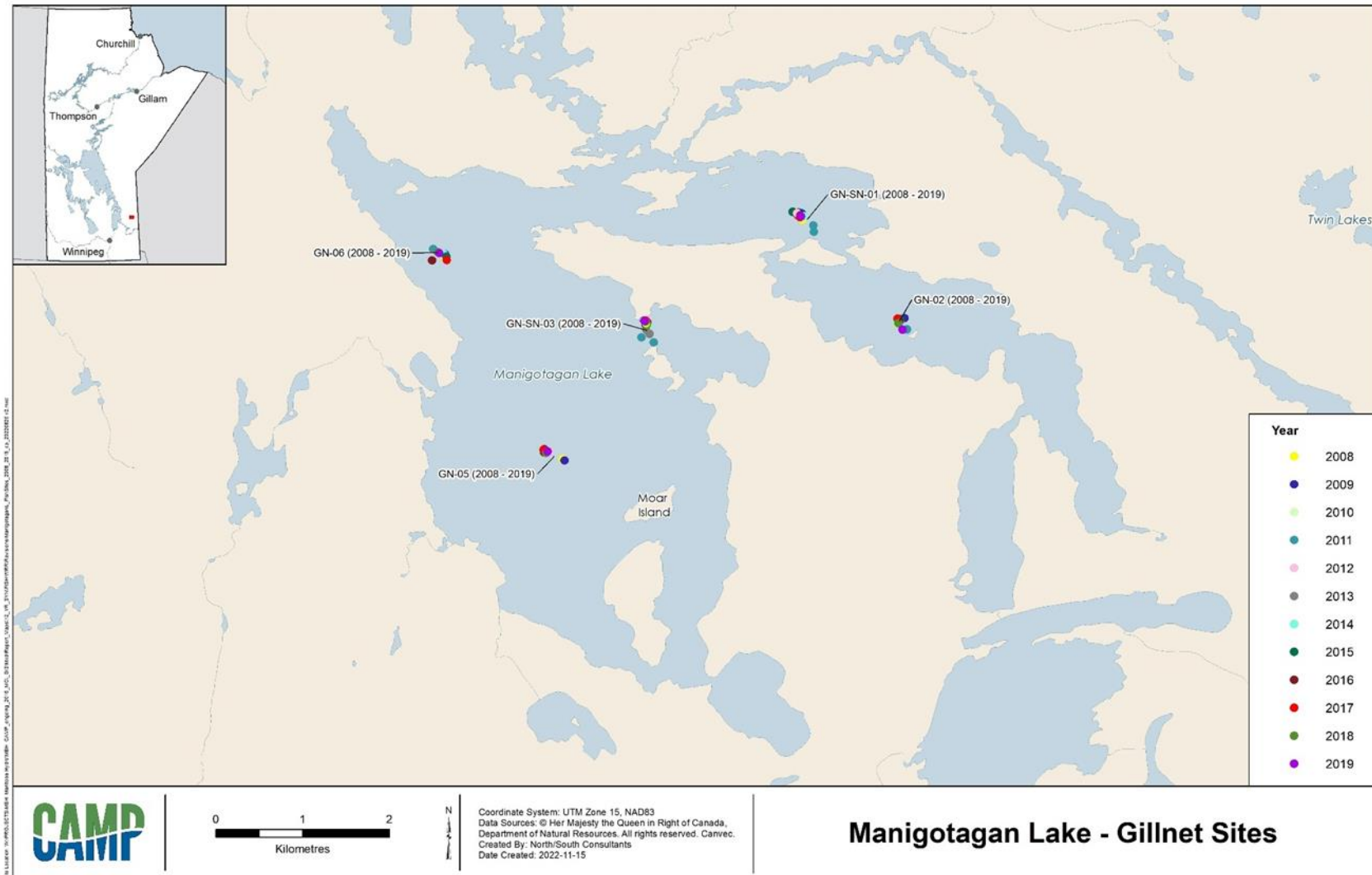


Figure A5-1-4. 2008-2019 Gillnetting sites in Manigotagan Lake.

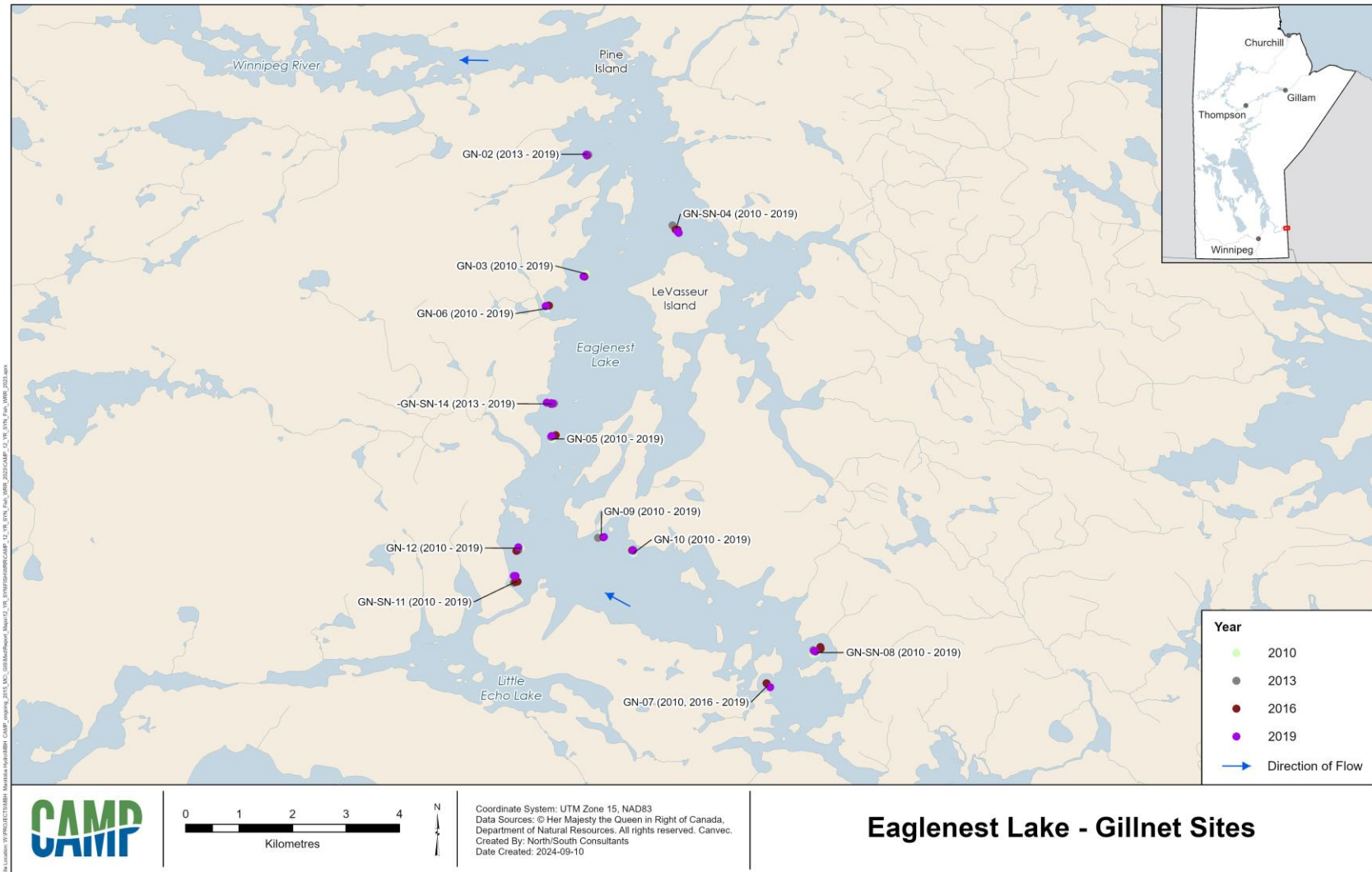


Figure A5-1-5. 2010-2019 Gillnetting sites in Eglene Lake.

6.0 MERCURY IN FISH

6.1 INTRODUCTION

The following presents the results of fish mercury monitoring conducted from 2008-2019 in the Winnipeg River Region. Fish mercury sampling was conducted on a three-year rotation beginning in 2010 at the on-system Pointe du Bois Forebay and the off-system Manigotagan Lake (Table 6.1-1; Figure 6.1-1).

Mercury concentrations are measured in muscle tissue of commercially important fish species – Northern Pike, Walleye, and Lake Whitefish. Monitoring of mercury in 1-year-old Yellow Perch is also conducted as a potential early indicator of changes in mercury in the food web. Samples of fish muscle are collected during the conduct of fish community monitoring. Mercury is analysed in the trunk muscle of Northern Pike, Lake Whitefish, and Walleye selected over a range of fork lengths. Yearling Yellow Perch are analyzed for mercury as carcass with the head, pelvic and pectoral girdles, caudal fin, and digestive tract removed.

There were two departures from the planned field sampling schedule during the 12-year period:

- Lake Whitefish were sampled for mercury analysis from Manigotagan Lake in 2011 because of low catches in 2010; and
- Lake Whitefish were not sampled for mercury analysis from the Pointe du Bois Forebay as scheduled in 2019.

Two metrics were selected for detailed reporting: arithmetic mean mercury concentrations; and, length-standardized mean mercury concentrations (also referred to as “standard mean(s)”; Table 6.1-2). Standard lengths varied by species as follows: Lake Whitefish (350 mm); Northern Pike (550 mm); and Walleye (400 mm). As CAMP targets a specific age class of Yellow Perch, fish captured for this component are inherently of a limited size range; therefore, length-standardization for this species was not undertaken.

A detailed description of the program design and sampling methods is provided in Technical Document 1, Section 2.6.

Table 6.1-1. 2008-2019 Inventory of fish mercury sampling.

| Waterbody/Area | Sampling Year | | | | | | | | | | | |
|----------------|---------------|------|------|----------------|------|------|------|------|------|------|------|----------------|
| | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
| PDB | | | ● | | | ● | | | ● | | | ● ² |
| MANIG | | | ● | ● ¹ | | ● | | | ● | | | ● |

Notes:

1. Lake Whitefish only; samples collected in 2011 due to low catches in 2010.
2. Lake Whitefish were not collected for mercury analysis.

Table 6.1-2. Mercury in fish indicators and metrics.

| Indicator | Metric | Units |
|-----------------|--|-------------------------|
| Mercury in Fish | • Arithmetic mean mercury concentration | Parts per million (ppm) |
| | • Length-standardized mean mercury concentration of large-bodied species | ppm |

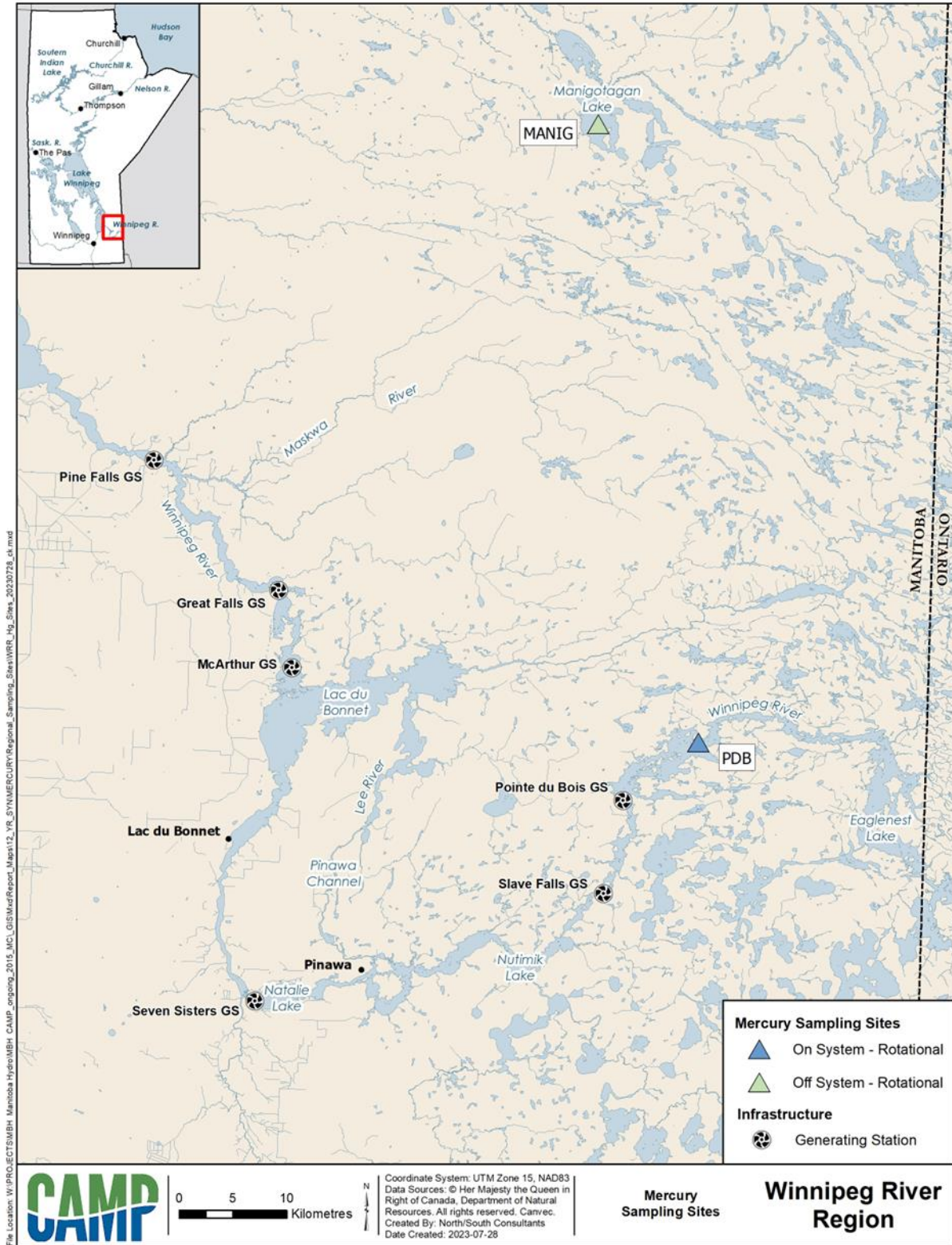


Figure 6.1-1. 2008-2019 Fish mercury sampling sites.

6.2 MERCURY IN FISH

6.2.1 MERCURY CONCENTRATIONS IN FISH

6.2.1.1 ON-SYSTEM SITES

ANNUAL SITES

There are no waterbodies in the Winnipeg River Region that are monitored for fish mercury annually.

ROTATIONAL SITES

Pointe du Bois Forebay

Lake Whitefish

The arithmetic mean mercury concentration of Lake Whitefish over the four years of monitoring ranged from a low of 0.045 parts per million (ppm) in 2013 to a high of 0.096 ppm in 2010 (Table 6.2-1). Lake Whitefish were not sampled for mercury in 2019 (Figure 6.2-1).

Northern Pike

The arithmetic mean mercury concentration of Northern Pike over the four years of monitoring ranged from a low of 0.313 ppm in 2013 to a high of 0.508 ppm in 2010 (Table 6.2-1). The mercury concentration typically increased with fork length, although there was variation in the mercury concentration of Northern Pike of the same length (Figure 6.2-2).

Walleye

The arithmetic mean mercury concentration of Walleye over the four years of monitoring ranged from a low of 0.283 ppm in 2019 to a high of 0.652 ppm in 2010 (Table 6.2-1). The mercury concentration typically increased with fork length, although there was variation in the mercury concentration of Walleye of the same length (Figure 6.2-3).

Yellow Perch

The arithmetic mean mercury concentration of 1-year-old Yellow Perch over the three years of monitoring was 0.017 ppm in 2013, 0.030 ppm in 2016, and 0.023 ppm in 2019 (Figure 6.2-4).

6.2.1.2 OFF-SYSTEM SITES

ANNUAL SITES

There are no waterbodies in the Winnipeg River Region that are monitored for fish mercury annually.

ROTATIONAL SITES

Manigotagan Lake

Lake Whitefish

The arithmetic mean mercury concentration of Lake Whitefish over the four years of monitoring ranged from a low of 0.044 ppm in 2019 to a high of 0.126 ppm in 2013 (Table 6.2-1). There was variation in the mercury concentration of Lake Whitefish of the same length (Figure 6.2-1).

Northern Pike

The arithmetic mean mercury concentration in Northern Pike over the four years of monitoring ranged from a low of 0.419 ppm in 2013 to a high of 1.18 ppm in 2010 (Table 6.2-1). The mercury concentration typically increased with fork length, although there was variation in the mercury concentration of Northern Pike of the same length (Figure 6.2-2).

Walleye

The arithmetic mean mercury concentration in Walleye over the four years of monitoring ranged from a low of 0.291 ppm in 2016 to a high of 0.530 ppm in 2013 (Table 6.2-1). The mercury concentration typically increased with fork length, although there was variation in the mercury concentration of Walleye of the same length (Figure 6.2-3).

Yellow Perch

The arithmetic mean mercury concentration of 1-year-old Yellow Perch over the three years of monitoring was 0.059 ppm in 2013 and 0.025 ppm in 2016 (Figure 6.2-4). None of the Yellow Perch collected for mercury analysis in 2019 were one-year-old.

Table 6.2-1. 2010-2019 Fork length, age, and mercury concentrations of Lake Whitefish, Northern Pike, and Walleye.

| Species | Waterbody | Year | Fork Length (mm) | | | | | Age (years) | | | | | Mercury (ppm) | | | | | | |
|---------|-----------|------|------------------|------|------------------|------------------|-----------------|-------------|------|-----|-----|----|---------------|-------|-------|-------|-------|----------------------------|---------------------|
| | | | n ¹ | Mean | Min ² | Max ² | SE ³ | n | Mean | Min | Max | SE | n | Mean | Min | Max | SE | Standard Mean ⁴ | 95% CL ⁵ |
| LKWH | PDB | 2010 | 4 | 403 | 320 | 560 | 54 | 4 | 7 | 3 | 15 | 3 | 4 | 0.096 | 0.046 | 0.230 | 0.045 | 0.053 | 0.028-0.101 |
| | | 2013 | 3 | 363 | 297 | 431 | 39 | 0 | - | - | - | - | 3 | 0.045 | 0.037 | 0.055 | 0.005 | not significant | |
| | | 2016 | 4 | 430 | 334 | 470 | 32 | 4 | 6 | 4 | 6 | 1 | 4 | 0.062 | 0.055 | 0.071 | 0.003 | not significant | |
| | | 2019 | 0 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| | MANIG | 2011 | 13 | 371 | 254 | 508 | 18 | 12 | 13 | 5 | 30 | 2 | 13 | 0.096 | 0.073 | 0.135 | 0.006 | not significant | |
| | | 2013 | 5 | 383 | 216 | 450 | 43 | 5 | 13 | 2 | 20 | 3 | 5 | 0.126 | 0.060 | 0.210 | 0.026 | not significant | |
| | | 2016 | 32 | 406 | 248 | 482 | 12 | 31 | 12 | 1 | 27 | 1 | 32 | 0.100 | 0.040 | 0.325 | 0.009 | 0.076 | 0.065-0.089 |
| | | 2019 | 12 | 354 | 238 | 482 | 22 | 11 | 9 | 4 | 26 | 2 | 12 | 0.044 | 0.026 | 0.091 | 0.006 | 0.041 | 0.034-0.050 |
| NRPK | PDB | 2010 | 17 | 490 | 235 | 897 | 34 | 17 | 5 | 2 | 10 | 0 | 17 | 0.508 | 0.110 | 1.23 | 0.078 | 0.558 | 0.435-0.716 |
| | | 2013 | 35 | 495 | 339 | 782 | 21 | 34 | 5 | 3 | 9 | 0 | 35 | 0.313 | 0.086 | 1.05 | 0.042 | 0.341 | 0.298-0.390 |
| | | 2016 | 32 | 550 | 231 | 883 | 25 | 32 | 4 | 1 | 7 | 0 | 32 | 0.461 | 0.114 | 0.980 | 0.037 | 0.440 | 0.409-0.473 |
| | | 2019 | 28 | 474 | 286 | 927 | 28 | 28 | 4 | 2 | 8 | 0 | 28 | 0.330 | 0.079 | 1.22 | 0.053 | 0.389 | 0.335-0.451 |
| | MANIG | 2010 | 11 | 583 | 520 | 770 | 24 | 11 | 6 | 5 | 10 | 1 | 11 | 1.18 | 0.584 | 1.69 | 0.121 | not significant | |
| | | 2013 | 28 | 509 | 411 | 640 | 10 | 28 | 5 | 3 | 6 | 0 | 28 | 0.419 | 0.144 | 0.840 | 0.036 | 0.494 | 0.420-0.581 |
| | | 2016 | 31 | 554 | 365 | 725 | 16 | 31 | 5 | 2 | 12 | 0 | 31 | 0.437 | 0.101 | 1.13 | 0.044 | 0.376 | 0.327-0.431 |
| | | 2019 | 14 | 537 | 416 | 781 | 35 | 14 | 5 | 3 | 8 | 0 | 14 | 0.422 | 0.158 | 1.21 | 0.086 | 0.385 | 0.317-0.466 |
| WALL | PDB | 2010 | 36 | 375 | 128 | 710 | 27 | 28 | 11 | 2 | 27 | 1 | 36 | 0.652 | 0.098 | 1.92 | 0.076 | 0.649 | 0.585-0.719 |
| | | 2013 | 37 | 405 | 156 | 662 | 22 | 37 | 8 | 2 | 18 | 1 | 37 | 0.419 | 0.124 | 1.12 | 0.048 | 0.365 | 0.323-0.414 |
| | | 2016 | 36 | 393 | 183 | 574 | 20 | 36 | 7 | 2 | 17 | 1 | 36 | 0.467 | 0.145 | 1.66 | 0.055 | 0.425 | 0.369-0.488 |
| | | 2019 | 36 | 336 | 123 | 551 | 21 | 36 | 6 | 1 | 16 | 1 | 36 | 0.283 | 0.052 | 0.792 | 0.033 | 0.324 | 0.280-0.376 |
| | MANIG | 2010 | 53 | 351 | 116 | 596 | 15 | 53 | 5 | 0 | 14 | 0 | 53 | 0.396 | 0.095 | 1.36 | 0.038 | 0.429 | 0.386-0.477 |
| | | 2013 | 36 | 368 | 131 | 657 | 26 | 36 | 7 | 1 | 18 | 1 | 36 | 0.530 | 0.107 | 1.89 | 0.070 | 0.522 | 0.465-0.585 |
| | | 2016 | 38 | 372 | 121 | 570 | 18 | 38 | 5 | 1 | 14 | 0 | 38 | 0.291 | 0.063 | 0.954 | 0.034 | 0.283 | 0.246-0.326 |
| | | 2019 | 36 | 410 | 219 | 665 | 21 | 34 | 7 | 2 | 17 | 1 | 36 | 0.325 | 0.068 | 1.22 | 0.053 | 0.226 | 0.192-0.266 |

Notes:

1. n = sample size.
2. Min = minimum; Max = maximum.
3. SE = standard error.
4. For standard lengths of 350 mm for LKWH, 550 mm for NRPK, and 400 mm for WALL.
5. CL = confidence limits.

Table 6.2-2. 2013-2019 Fork length and mercury concentrations of 1-year-old Yellow Perch.

| Species | Waterbody | Year | n ¹ | Fork Length (mm) | | | | Mercury (ppm) | | | |
|---------|-----------|------|----------------|------------------|------------------|------------------|-----------------|---------------|--------|--------|-------|
| | | | | Mean | Min ² | Max ² | SE ³ | Mean | Min | Max | SE |
| YLPR | PDB | 2013 | 19 | 78 | 70 | 83 | 1 | 0.017 | <0.010 | 0.029 | 0.002 |
| | | 2016 | 8 | 81 | 70 | 90 | 1 | 0.030 | 0.018 | 0.044 | 0.003 |
| | | 2019 | 9 | 82 | 78 | 88 | 1 | 0.023 | 0.0195 | 0.0289 | 0.001 |
| | MANIG | 2013 | 2 | 90 | 88 | 92 | 0 | 0.059 | 0.055 | 0.062 | 0.002 |
| | | 2016 | 9 | 65 | 60 | 68 | 0 | 0.025 | 0.017 | 0.031 | 0.002 |
| | | 2019 | 0 | - | - | - | - | - | - | - | - |

Notes:

1. n = sample size.
2. Min = minimum; Max = maximum.
3. SE = standard error.

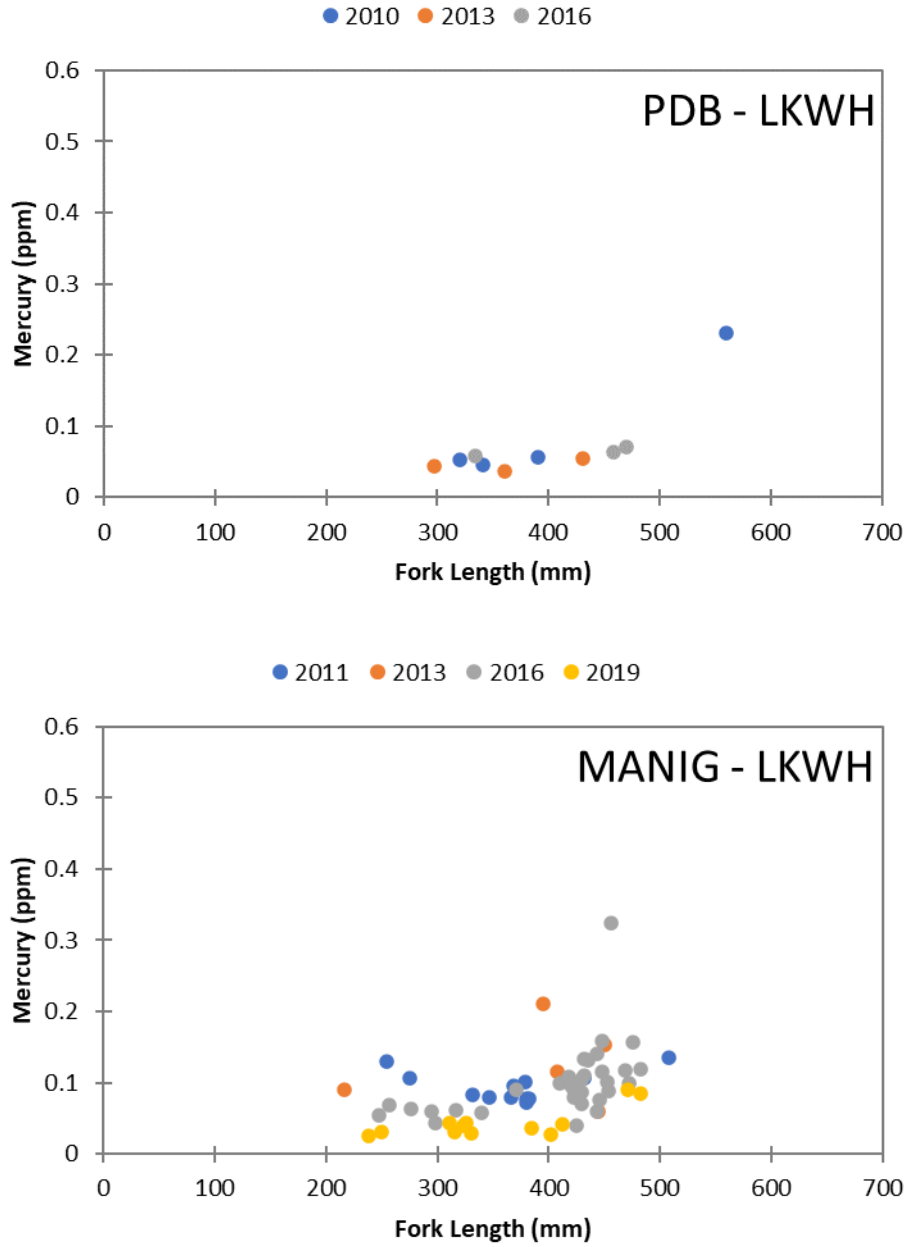


Figure 6.2-1. 2010-2019 Mercury concentration versus fork length of Lake Whitefish.

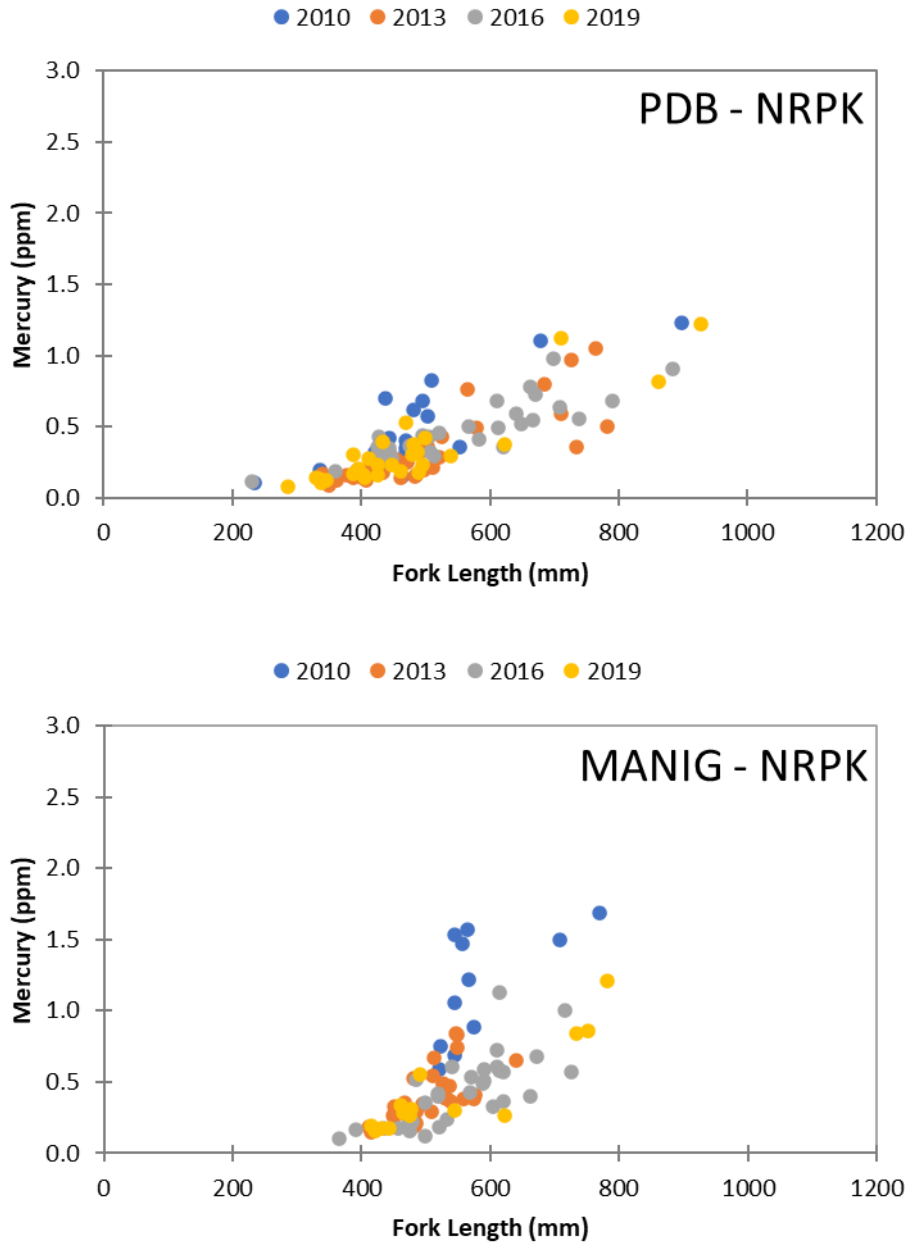


Figure 6.2-2. 2010-2019 Mercury concentration versus fork length of Northern Pike.

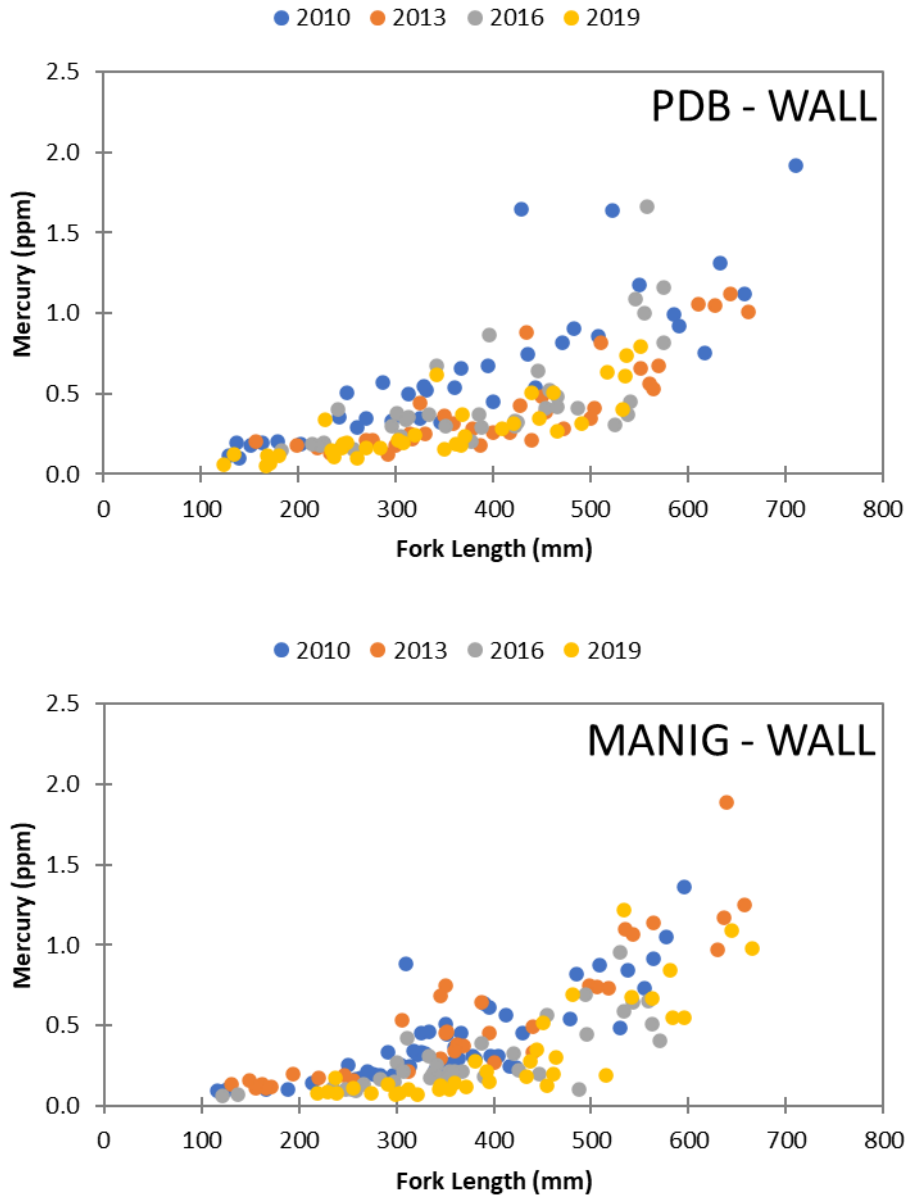


Figure 6.2-3. 2010-2019 Mercury concentration versus fork length of Walleye.

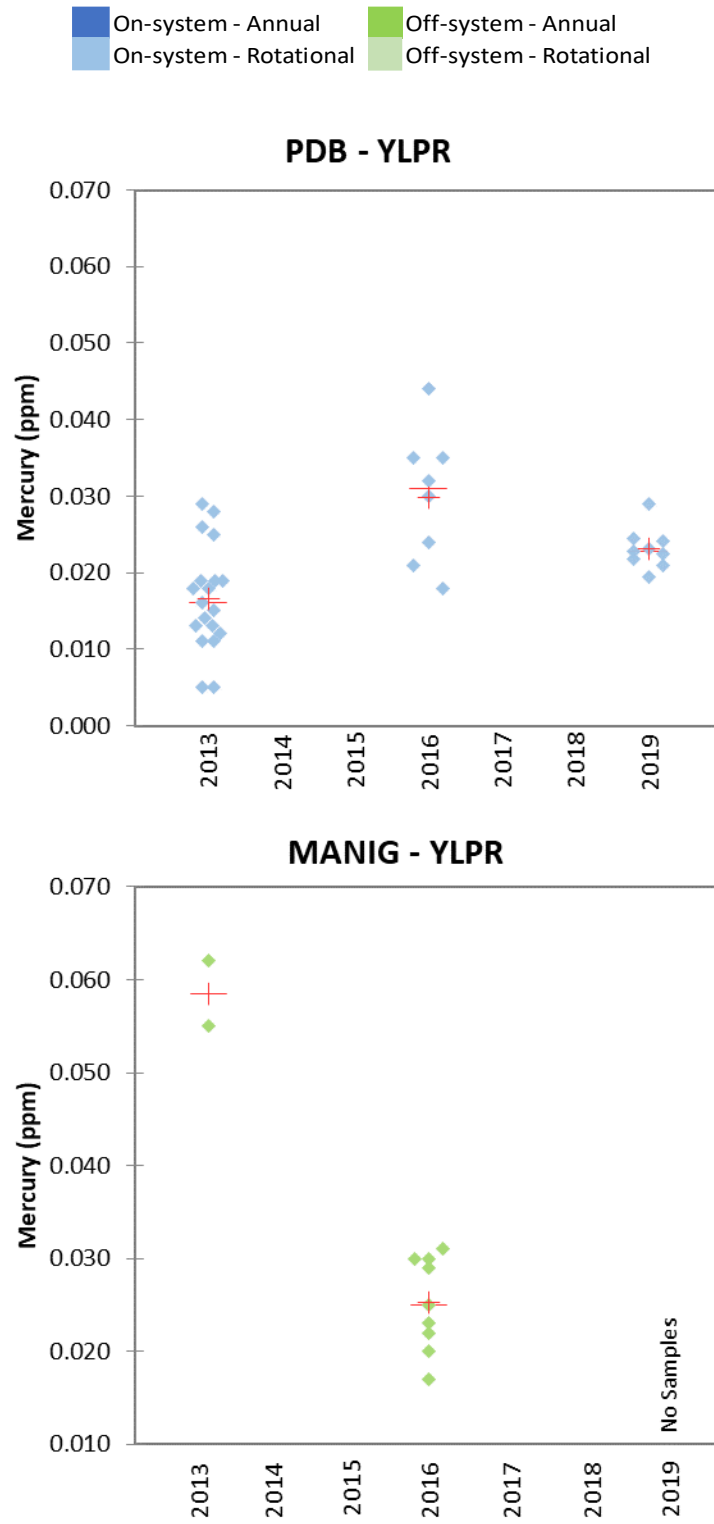


Figure 6.2-4. 2013-2019 Mercury concentrations of one-year-old Yellow Perch.

6.2.2 LENGTH-STANDARDIZED MEAN CONCENTRATION

6.2.2.1 ON-SYSTEM SITES

ANNUAL SITES

There are no waterbodies in the Winnipeg River Region that are monitored for fish mercury annually.

ROTATIONAL SITES

Pointe du Bois Forebay

Lake Whitefish

The length-standardized mean mercury concentration of a 350 mm Lake Whitefish was 0.053 ppm in 2010 (Figure 6.2-5). A standard mean could not be calculated for the other years because there was not a significant relationship between mercury concentration and fork length for the few Lake Whitefish analyzed for mercury in 2013 and 2016, and no Lake Whitefish were analyzed for mercury in 2019 (Table 6.2-1).

Northern Pike

The length-standardized mean mercury concentration of a 550 mm Northern Pike over the four years of monitoring ranged from a low of 0.341 ppm in 2013 to a high of 0.558 ppm in 2010 (Figure 6.2-6).

The overall mean concentration was 0.432 ppm, the median concentration was 0.414 ppm, and the interquartile range (IQR) was 0.377–0.470 ppm (Figure 6.2-6). The annual mean mercury concentration fell within the IQR except in 2013 when it was below and in 2010 when it was above.

Walleye

The length-standardized mean mercury concentration of a 400 mm Walleye over the four years of monitoring ranged from a low of 0.324 ppm in 2019 to a high of 0.649 ppm in 2010 (Figure 6.2-7).

The overall mean concentration was 0.441 ppm, the median concentration was 0.395 ppm, and the IQR was 0.355–0.481 ppm (Figure 6.2-7). The annual mean mercury concentration fell within the IQR except in 2019 when it was below and in 2010 when it was above.

6.2.2.2 OFF-SYSTEM SITES

ANNUAL SITES

There are no waterbodies in the Winnipeg River Region that are monitored for fish mercury annually.

ROTATIONAL SITES

Manigotagan Lake

Lake Whitefish

The length-standardized mean mercury concentration of a 350 mm Lake Whitefish was 0.041 ppm in 2019 and 0.076 ppm in 2016 (Figure 6.2-5). A standard mean could not be calculated for the other years because there was not a significant relationship between mercury concentration and fork length for the few Lake Whitefish analyzed for mercury (Table 6.2-1).

Northern Pike

The length-standardized mean mercury concentration of a 550 mm Northern Pike over the four years of monitoring ranged from a low of 0.376 ppm in 2016 to a high of 0.494 ppm in 2013 (Figure 6.2-6). A standard mean could not be calculated for 2010 because there was not a significant relationship between mercury concentration and fork length for the few Northern Pike analyzed for mercury (Table 6.2-1).

The overall mean concentration was 0.418 ppm, the median concentration was 0.385 ppm, and the IQR was 0.380–0.439 ppm (Figure 6.2-6). The annual mean mercury concentration fell within the IQR except in 2016 when it was below and in 2013 when it was above.

Walleye

The length-standardized mean mercury concentration of a 400 mm Walleye over the four years of monitoring ranged from a low of 0.226 ppm in 2019 to a high of 0.522 ppm in 2013 (Figure 6.2-7).

The overall mean concentration was 0.365 ppm, the median concentration was 0.356 ppm, and the IQR was 0.269–0.452 ppm (Figure 6.2-7). The annual mean mercury concentration fell within the IQR except in 2019 when it was below and in 2013 when it was above.

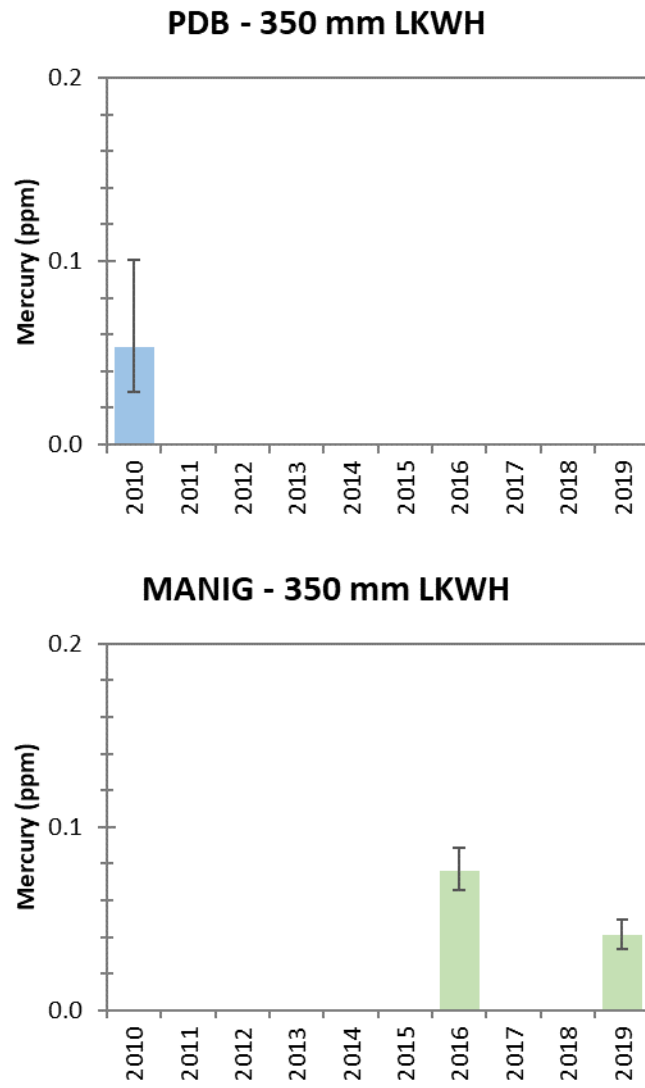
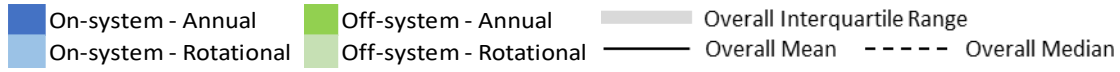


Figure 6.2-5. 2010-2019 Length-standardized mean mercury concentrations ($\pm 95\%$ confidence interval) of Lake Whitefish.

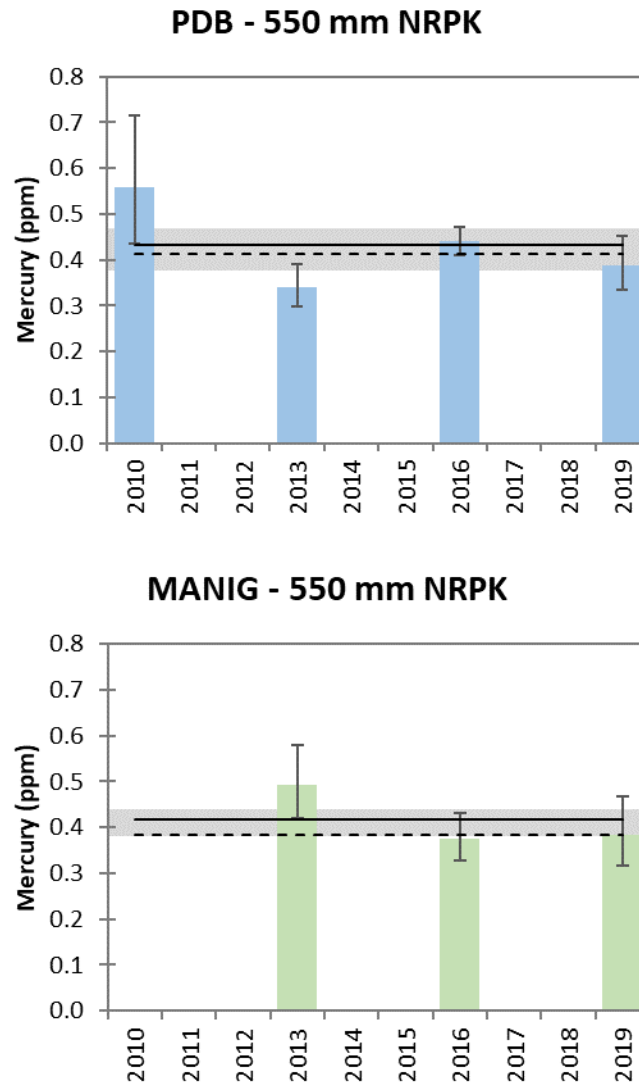
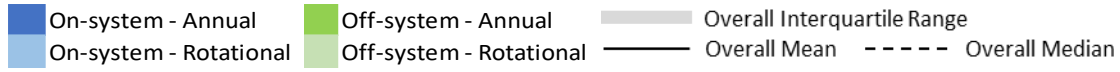


Figure 6.2-6. 2010-2019 Length-standardized mean mercury concentrations ($\pm 95\%$ confidence interval) of Northern Pike.

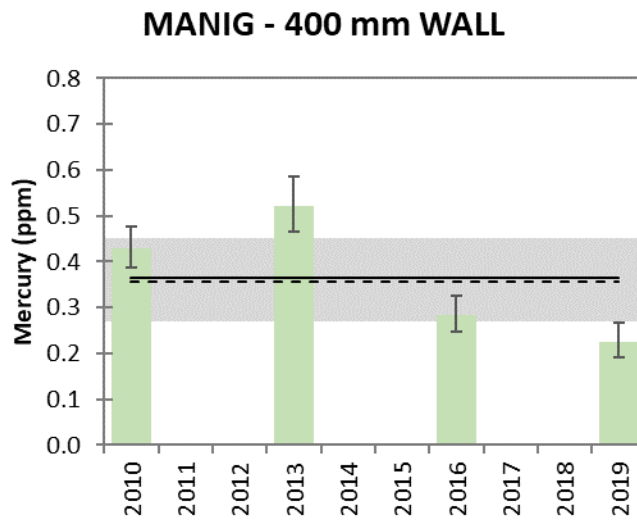
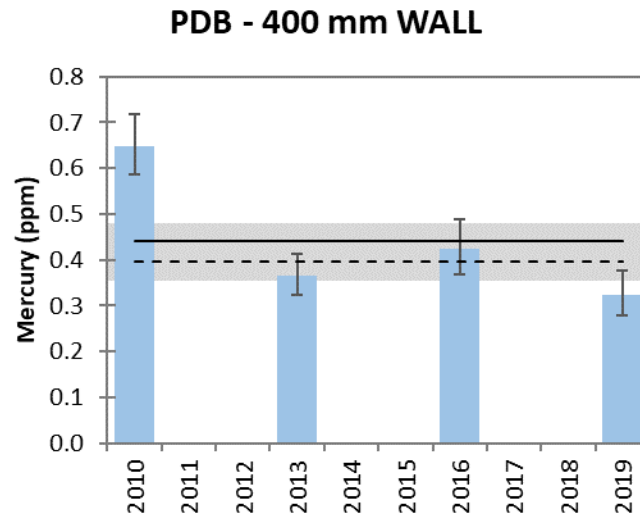
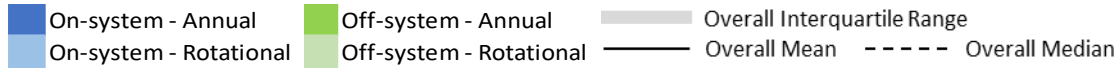


Figure 6.2-7. 2010-2019 Length-standardized mean mercury concentrations ($\pm 95\%$ confidence interval) of Walleye.

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